

Photosynthetic characteristics and physiological plasticity of an *Aphanizomenon flos-aquae* (Cyanobacteria, Nostocaceae) winter bloom in a deep oligo-mesotrophic lake (Lake Stechlin, Germany)

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Abstract In winter of 2009/2010, *Aphanizomenon flos-aquae* bloomed in the ice and snow covered oligo-mesotrophic Lake Stechlin, Germany. The photosynthesis of the natural population was measured at eight temperatures in the range of 2–35°C, at nine different irradiance levels in the range of 0–1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$ PAR at each applied temperature. The photoadaptation parameter (I_k) and the maximum photosynthetic rate (P_{max}) correlated positively with the temperature between 2 and 30°C, and there was a remarkable drop in both parameters at 35°C. The low I_k at low temperatures enabled the active photosynthesis of overwintering populations at low irradiance levels under ice and snow cover. The optimum of the photosynthesis was above 20°C at irradiances above 150 $\mu\text{mol m}^{-2} \text{s}^{-1}$. At lower irradiance levels (7.5–30 $\mu\text{mol m}^{-2} \text{s}^{-1}$), the photosynthesis was the

most intensive in the temperature range of 2–5°C. The interaction between light and temperature allowed the proliferation of *A. flos-aquae* in Lake Stechlin resulting in winter water bloom in this oligo-mesotrophic lake. The applied 2°C is the lowest experimental temperature ever in the photosynthesis/growth studies of *A. flos-aquae*, and the results of the P–I and P–T measurements provide novel information about the tolerance and physiological plasticity of this species.

Keywords *Aphanizomenon flos-aquae* · P–I and P–T characteristics · Oligo-mesotrophic Lake Stechlin · Cyanobacterial bloom · Winter · Ice cover · Physiological plasticity

Introduction

The dynamics of phytoplankton populations are controlled by multiple factors including physiological and evolutionary adaptations, environmental and biological processes. The interactions between the different factors are important in understanding the response of organisms to these variables. The effects of temperature on phytoplankton cellular processes and growth are well known (e.g. Reynolds, 1984), being temperature an important, but not the only factor in determining the occurrence of a particular algal species. The physiological adaptations may contribute to the success of certain phytoplankton taxa to become efficient competitors, as demonstrated on several

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species of cyanobacteria (e.g. Padisák, 1997; Price et al., 1998). Cyanobacterial dominance commonly occurs in eutrophic water bodies (e.g. Yamamoto & Nakahara, 2009a) at water temperatures above 20°C due to the high temperature optima of their growth (e.g. Reynolds, 1984; Wilhelm & Adrian, 2008), but the success of cyanobacteria is a result of their response to interactions between different environmental constraints (e.g. Dokulil & Teubner, 2000).

Aphanizomenon flos-aquae (L.) Ralfs is a filamentous and heterocyclic cyanobacterium, which is capable of N₂ fixation (e.g. De Nobel et al., 1998; Reynolds et al., 2002). It produces akinetes that enable survival in unfavourable growth conditions (Yamamoto & Nakahara, 2007). The temperature optimum of its growth is above 20°C (e.g. Uehlinger, 1981; Dokulil & Teubner, 2000), but it depends on the light intensities at which the population was grown (Konopka & Brock, 1978). The species prefers shallow eutrophic freshwaters (e.g. Yamamoto & Nakahara, 2009a, b), and since external load reduction resulted in nitrogen scarcity in many lakes, relative biomass share of N-fixing genera, like *Aphanizomenon*, increased in many lakes during the last decades (Reynolds et al., 2002; Wagner & Adrian, 2009). *Aphanizomenon flos-aquae* cells contain gas vesicles, which have a significant role in regulating buoyancy (e.g. Walsby, 1994). It can accumulate in the illuminated layer near the water surface, developing a surface bloom (e.g. Dokulil & Teubner, 2000; Preußel et al., 2009). Temperature and light intensity are key factors in the regulation the growth of Nostocales (Mehnert et al., 2010). The competition between *A. flos-aquae* and other cyanobacterial species for light at different temperatures was investigated in several studies (e.g. De Nobel et al., 1998; Huisman et al., 1999; Yamamoto & Nakahara, 2006; Wagner & Adrian, 2009). The current knowledge about the temperature- and light-dependent growth of *A. flos-aquae* in temperature regions >15°C is well documented, but information is scarce about its winter occurrence. Wildman et al. (1975) found all developmental stages of *A. flos-aquae* in samples from an ice-covered lake in November. In another study, some filaments of the species were present in the phytoplankton throughout the winter months until the onset of exponential growth in mid-May in Kinnego Bay, Ireland (Jones, 1979). The possibility that these ‘overwintering’ filaments may have been the result of continuous

germination of akinetes through the winter was discounted, since the observed filaments in winter and early spring were mostly long filaments consisted of many cells (Jones, 1979).

Aphanizomenon flos-aquae appeared in the oligomesotrophic, deep Lake Stechlin (Germany) first in 2001. Until 2006, the species was seasonal and sporadic between early summer and the autumnal overturn. Since 2006, it provided biomass peaks in late summers. In 2009, the biomass of the late summer peak was 400–500 µg wet weight L⁻¹. In autumn a drop of the biomass was observed, but after the autumnal overturn the population of *A. flos-aquae* began to grow, and provided another biomass peak in winter with 915–920 µg wet weight L⁻¹. In December, a macroscopically visible surface bloom of *A. flos-aquae* developed along the shorelines of the lake. Between December 2009 and January 2010, *A. flos-aquae* contributed 87–90% to total biomass therefore developed a sufficiently long-lasting winter equilibrium phase (Padisák et al., 2010) which continued in February 2010 when 17 cm thick ice developed on the lake and was covered by 20 cm snow. According to our knowledge about the species, appearance of *A. flos-aquae* in Lake Stechlin, and its winter bloom under ice cover were unexpected phenomena. They triggered the questions that the filaments of *A. flos-aquae* were physiologically active or not and that the bloom under winter conditions was a result of its low-temperature tolerance or low-temperature preference.

This study focused on the question of how temperature and irradiance level influence the photosynthetic activity of *A. flos-aquae*. We assumed that the interactions between temperature and the light availability would

- (i) cause corresponding changes in the P–I (photosynthesis–irradiance) parameters of *A. flos-aquae*,
- (ii) determine the occurrence of the species under extreme conditions, and
- (iii) due to its physiologically distinctive features it could reach high biomass with active photosynthesis in an ice-covered lake.

To test these hypotheses, the photosynthetic activity of a natural population of *A. flos-aquae* collected on 26th February 2010 in Lake Stechlin was investigated at eight different temperatures (2–35°C) and at nine

different irradiance levels ($0\text{--}1,320 \mu\text{mol m}^{-2} \text{s}^{-1}$) in laboratory.

Materials and methods

Site description, sampling

Lake Stechlin is a medium-sized (4.2 km^2), deep (z_{mean} : 23.3 m; z_{max} : 69.5 m) lake without surface inflows at $53^\circ 10' \text{N}$ latitude, $13^\circ 02' \text{E}$ longitude and 84.5 m a.s.l. in Brandenburg, Germany. The lake has glacial origin, was originally oligotrophic, but during the last decade it turned oligo-mesotrophic. In February 2010, the following concentrations of nutrient were measured: SRP: $8.5 \mu\text{g L}^{-1}$; TP: $27.7 \mu\text{g L}^{-1}$; $\text{NO}_3^- \text{-N}$: $58.7 \mu\text{g L}^{-1}$ (APHA, 1998).

Phytoplankton samples were taken on 26th February 2010 at a sampling site in front of the Leibniz-Institute of Freshwater Ecology and Inland Fisheries (Fig. 1), where the water depth was 26 m. At the sampling occasion, temperature and photosynthetically active radiation (PAR) (Li 192A Underwater Quantum Sensor, LI-COR Biosciences, Lincoln, Nebraska, USA) profiles were taken in the water column of the lake: the sensors were sank into different depth through a leak, while leak was recovered with ice and snow to get the real in situ temperature and light values during the measurements. The in situ temperature ranged between 0.3 and 2°C in the upper 10 m layer. The euphotic zone expanded to depth of 10–12 m. Samples were taken with a Van

Dorn sampler in 2 m increments from the euphotic depths then were mixed (integrated sample). The sampling location and light field (under 17 cm thick ice and 20 cm snow cover) are shown on Fig. 1. Vertical light attenuation coefficient, K_d (m^{-1}) of different layers of the water column was calculated with the Lambert–Beer function (Kirk, 1994) from simultaneous measurements of irradiances per metre (0–20 m) in the field at time of the sampling. The K_d was in the upper 1 m layer the highest, 1.33 m^{-1} . The average K_d of the layers between 1 and 20 m was $0.42 \pm 0.15 \text{ m}^{-1}$; the euphotic layer was lying at the depth of 11 m (Fig. 1).

The samples were kept in the dark and cool for 24–36 h, while they were transported from Germany to the location of the laboratory measurements at the University of Pannonia, Veszprém, Hungary. Subsamples of 300 ml were preserved in Lugol solution for microscopic counting (Utermöhl, 1958). The chlorophyll *a* concentrations of the samples were measured according to Wetzel & Likens (2000).

P–I and P–T measurements

The measurements were carried out in a laboratory incubation system (Üveges et al., 2011). The photosynthetic activity of the natural population was measured by the ^{14}C method (Steemann-Nielsen, 1952). The rate of photosynthesis was determined by adding $\text{NaH}^{14}\text{CO}_3$ with known activity ($0.099181\text{--}0.114185 \text{ MBq}$) to 15 ml of samples in 20 ml scintillation vials (Econo Glass Vial, PerkinElmer, Waltham, MA, USA). The

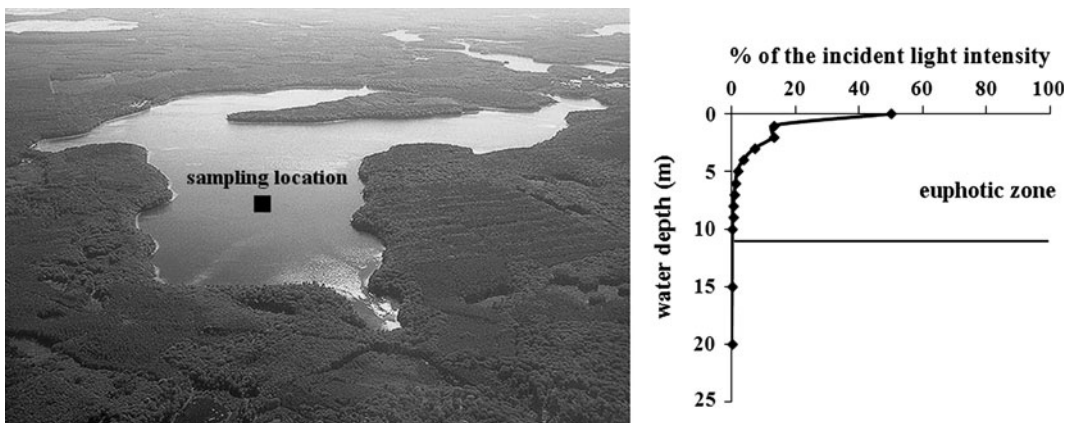


Fig. 1 Lake Stechlin and the sampling location in the lake. On the chart right the light field of the water column, covered by 17 cm thick ice and 20 cm snow, was shown at the sampling

occasion. On 26th February 2010, the irradiance level was $134 \mu\text{mol m}^{-2} \text{s}^{-1}$ at 0 m (immediately below the ice), which corresponded to the 50% of the incident irradiance level

vials were pre-incubated at appropriate temperatures for 1 h before the addition of radioisotope to equilibrate the contents to the experimental condition. Three vials were incubated as replicates at each irradiance level. The photosynthesis of the natural population was measured for 2 h after the addition of radioisotope at nine irradiance levels (0; 7.5; 30; 75; 150; 260; 500; 920; 1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$ PAR) at each applied temperature (2; 5; 10; 15; 20; 25; 30; 35°C). In all experiments, three vials were wrapped in aluminium foil to serve as dark control; they were incubated at 0 $\mu\text{mol m}^{-2} \text{s}^{-1}$ irradiance. After the 2 h of incubation and removing of the externally attached radioisotope, the incorporation of ^{14}C into algal protein was measured in each vial with liquid scintillation analyzer (Packard Tri-Carb 3180TR/SL, GMI Inc., Ramsey, MN, USA). Non-photosynthetic C uptake, which was determined in control vials kept in the dark, was subtracted from photosynthetic C uptake.

To fit curves for photosynthesis vs. irradiance data and to determine the P–I parameters (Table 1), the GraFit software was applied (GraFit by R. Leatherbarrow, 1989–1992 Erithacus Software Ltd.) with the equation of Platt et al. (1980). The Platt et al. (1980) equation fits most photoinhibition data fairly well:

$$P = P_s \left(1 - \exp^{-(\alpha I/P_s)} \right) \left(\exp^{-(\beta I/P_s)} \right),$$

where P is the photosynthetic rate at irradiance I , P_s is the maximum photosynthesis obtained in the absence of photoinhibition, and α and β are parameters describing the initial slope and the photoinhibited section of the P–I curve, respectively. Photosynthetic parameters like maximal photosynthetic rate (P_{\max})

and photoadaptation parameter ($I_k = P_{\max}/\alpha$) were derived from the previous parameters (Table 1).

Q_{10}

The Q_{10} model (Ahlgren, 1987) was employed to describe the relationship between photosynthesis and temperature:

$$Q_{10} = \left(P_{I(T_2)} / P_{I(T_1)} \right)^{(10/(T_2-T_1))},$$

where $P_{I(T_2)}$ and $P_{I(T_1)}$ are photosynthetic activity at given irradiance level (I) at two temperatures, T_2 and T_1 , respectively. In fact, algal growth rates increase up to the optimal temperature beyond which they decrease due to stressful conditions, so Q_{10} was calculated from the linear section of the photosynthesis–temperature (P–T) curves.

Results

Phytoplankton sample

On 26th February 2010, the chlorophyll a content of the integrated sample from 0 to 10 m upper layer of Lake Stechlin was $4.39 \pm 0.40 \mu\text{g chl } a \text{ L}^{-1}$. According to the microscopic analysis, *A. flos-aquae* dominated in the samples; the aggregated filaments of *A. flos-aquae* provided more than 99% of the total biomass. The rest 1% was provided by the diatom *Stephanodiscus neoastraea*. Therefore, from ecophysiological point of

Table 1 Parameter values obtained by nonlinear regression of the photosynthesis–irradiance data for the natural population of winter blooming *Aphanizomenon flos-aquae* in Lake Stechlin (Germany)

T	α	P_s	β	P_{\max}	I_k
2	0.195 (0.153)	4.31 (0.15)	0.0032 (0.0004)	3.96	20
5	0.143 (0.008)	6.15 (0.17)	0.0041 (0.0004)	5.39	38
10	0.160 (0.011)	12.98 (0.69)	0.0076 (0.0014)	10.70	67
15	0.132 (0.007)	13.96 (0.61)	0.0056 (0.0009)	11.69	89
20	0.138 (0.014)	15.34 (1.21)	0.0024 (0.0015)	14.04	102
25	0.121 (0.008)	20.98 (1.78)	0.0041 (0.0020)	18.07	149
30	0.113 (0.005)	22.46 (1.46)	0.0037 (0.0015)	19.42	172
35	0.109 (0.018)	10.45 (1.15)	0.0000 (0.0012)	10.57	97

SE is given in brackets for fitted parameters; SE was not calculated for the derived parameters, P_{\max} and I_k . T , temperature (°C); α , initial slope ($\mu\text{g C L}^{-1} \text{h}^{-1} (\mu\text{mol m}^{-2} \text{s}^{-1})^{-1}$); P_s , maximum photosynthetic rate obtained at the lack of photoinhibition ($\mu\text{g C L}^{-1} \text{h}^{-1}$); β , photoinhibition parameter ($\mu\text{g C L}^{-1} \text{h}^{-1} (\mu\text{mol m}^{-2} \text{s}^{-1})^{-1}$); I_k , photoadaptation parameter ($\mu\text{mol m}^{-2} \text{s}^{-1}$)

view the natural samples can be considered as a monoculture of *A. flos-aquae*.

P–I characteristics

The maximum photosynthetic rate (P_{\max}) varied between 3.96 and 19.42 $\mu\text{g C L}^{-1} \text{h}^{-1}$ (Fig. 2; Table 1). The population reached the highest P_{\max} at 30°C, and the lowest at 2°C. At 2 and 5°C the fixed carbon was 20 and 27% of that fixed at 30°C. A substantial decrease in the P_{\max} was measured between 10 and 5°C: *A. flos-aquae* photosynthesised at 50% of the maximum rate at 10°C. At 35°C, the P_{\max} declined to 54% of that measured at 30°C. In the temperature range of 2–30°C, the P_{\max} correlated positively with the temperature ($r = 0.975$; $P < 0.001$; $n = 7$).

The initial slope of the P–I curve (Fig. 2), α correlated negatively ($r = -0.876$; $P < 0.001$; $n = 7$) with the temperature; varied between 0.109 and 0.195 $\mu\text{g C L}^{-1} \text{h}^{-1} (\mu\text{mol m}^{-2} \text{s}^{-1})^{-1}$. Photoinhibition was observed at different irradiance levels in the experiments carried out at different temperatures (Fig. 2). At 2°C, the community was inhibited at the

irradiance of 500 $\mu\text{mol m}^{-2} \text{s}^{-1}$ by 20% and at 1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$ by 55% relative to photosynthesis at 75 $\mu\text{mol m}^{-2} \text{s}^{-1}$. At 25°C, the photosynthesis was inhibited only at the highest irradiance levels, at 920 and 1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$. The strongest photoinhibition was observed in the experiment carried out at 10°C, the community was inhibited at 1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$ by 54% relative to photosynthesis at irradiance of 260 $\mu\text{mol m}^{-2} \text{s}^{-1}$, in the region of P_{\max} .

The photoadaptation parameter, I_k showed a strong positive correlation with the temperature in the range of 2–30°C ($r = 0.985$; $P < 0.001$; $n = 7$).

P–T characteristics

In the irradiance range of 260–1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$, the carbon uptake increased with the increasing temperature until the temperature reached the 30°C (Fig. 3A), strong positive correlation was observed in all cases ($P < 0.001$; $n = 21$): $r = 0.989$ at 1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$; $r = 0.986$ at 920 $\mu\text{mol m}^{-2} \text{s}^{-1}$; $r = 0.983$ at 500 $\mu\text{mol m}^{-2} \text{s}^{-1}$; and $r = 0.954$ at 260 $\mu\text{mol m}^{-2} \text{s}^{-1}$. The photosynthesis at these irradiance levels was inhibited by 41–48% at

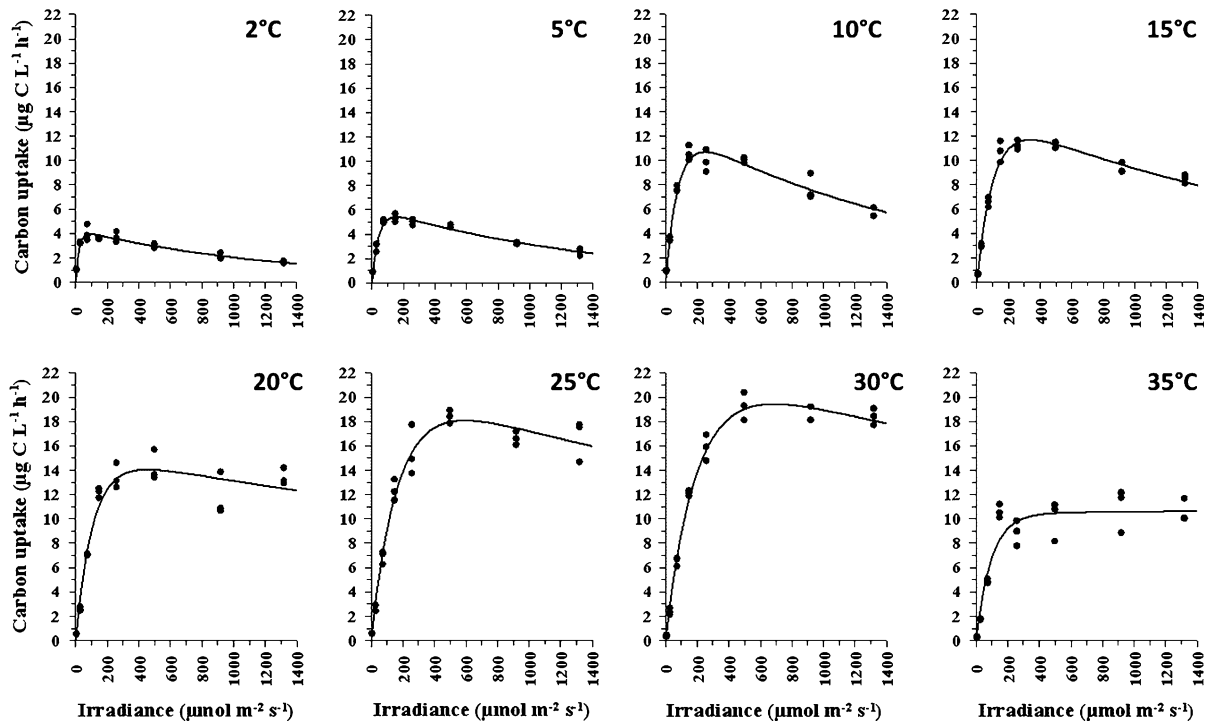


Fig. 2 Photosynthesis–irradiance curves of natural population of *Aphanizomenon flos-aquae* collected in Lake Stechlin, on 26th February 2010 at different experimental temperatures

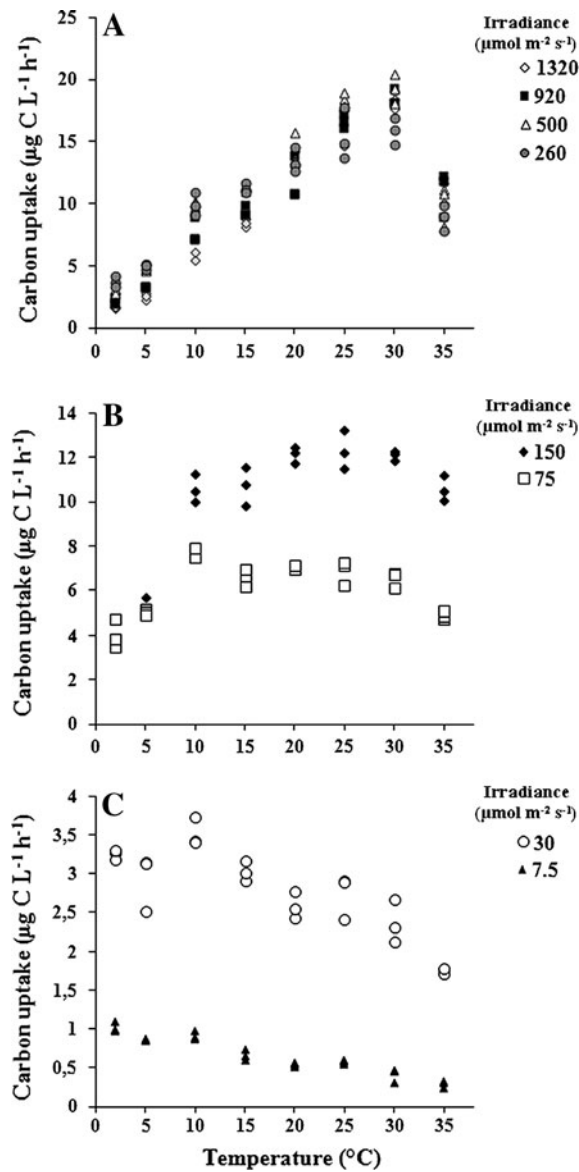


Fig. 3 Photosynthetic activity of natural winter population of *Aphanizomenon flos-aquae* as a function of temperature at different light intensities

35°C relative to that at 30°C. Strong positive correlation between carbon uptake and temperature was found only at the upper temperature region (2–10°C) at the irradiances of 150 and 75 $\mu\text{mol m}^{-2} \text{s}^{-1}$ ($r = 0.984$ and $r = 0.971$; $P < 0.001$; $n = 9$) (Fig. 3B). At low irradiance levels, the photosynthetic activity correlated negatively with the temperature (Fig. 3C). At irradiance level of 30 $\mu\text{mol m}^{-2} \text{s}^{-1}$, the carbon uptake did not change remarkably with the increasing temperature, but above

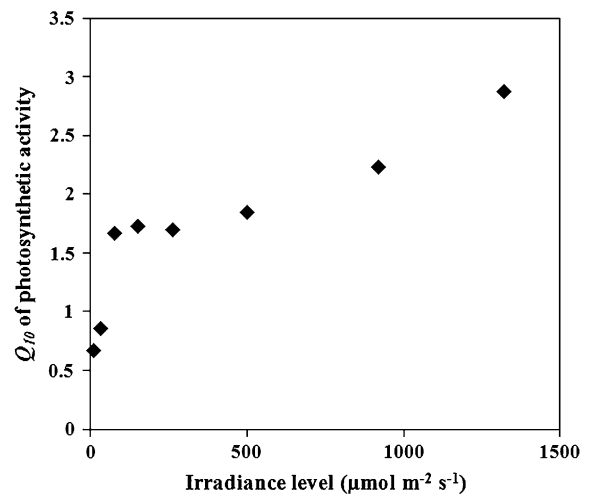


Fig. 4 Q_{10} of the photosynthetic activity of natural winter population of *Aphanizomenon flos-aquae* at different irradiance levels

15°C a negative correlation was observed ($r = -0.851$; $P < 0.001$; $n = 15$). At the lowest irradiance level (7.5 $\mu\text{mol m}^{-2} \text{s}^{-1}$), the photosynthesis of *A. flos-aquae* showed a strong negative correlation with the temperature in the whole temperature range ($r = -0.955$; $P < 0.001$; $n = 24$) (Fig. 3C).

Q_{10} of the photosynthetic activity

Q_{10} values computed for the linear section of the P–T curve (Fig. 3) varied from 0.67 to 2.88 (Fig. 4). At the lowest applied irradiances the photosynthesis did not show positive correlation to the temperature, the Q_{10} was less than 1. At the mid-irradiance levels (75–260 $\mu\text{mol m}^{-2} \text{s}^{-1}$) the average Q_{10} was 1.70 ± 0.03 , which suggests lower variation of photosynthesis with temperature, than at the highest irradiance levels (500–1,320 $\mu\text{mol m}^{-2} \text{s}^{-1}$), where Q_{10} varied from 1.85 to 2.88.

Discussion

After the first appearance of *A. flos-aquae* in Lake Stechlin (Padisák et al., 2010), the life cycle of the species corresponded to the usual patterns described by Yamamoto & Nakahara (2009b) until 2009, when it overwintered with filaments near to the water surface. Overwintering filaments of the species were observed

in previous studies (e.g. Simona, 2003; Yamamoto, 2009), however, physiological activity of the observed winter populations has not been studied. Yamamoto & Nakahara (2007) explained the winter appearance with the high biomass in summer which did not disappear completely in winter. It was certainly not the case for the Lake Stechlin *A. flos-aquae* population in winter of 2009, since morphology of summer and autumn filaments were clearly different: in contrast to the summer filaments, those found in winter did not include heretocytes. In a study from 2009, Yamamoto & Nakahara (2009a) concluded that the increase of population density of *A. flos-aquae* in winter might be a result of the adaptation of the species to low water temperatures.

The influence of temperature on the cellular metabolism of the species was addressed in several studies. Konopka & Brock (1978) isolated *Aphanizomenon* from Lake Mendota, and studied the influence of temperature on rate of carbon uptake under low-light intensities. They found, that the optimal temperature of *A. flos-aquae* for photosynthesis was 20°C, but at 10°C *Aphanizomenon* was still photosynthesising at 8% of the maximum rate. In other studies, the optimum temperature of *Aphanizomenon*'s growth varied between 20 and 28°C similar to our results at saturating and sub-saturating irradiances, but in these studies the lower limit was found between 4 and 10°C (Uehlinger, 1981; Tsujimura et al., 2001). We assume that the pre-experiment growth temperature could affect the temperature responses of *A. flos-aquae*, as it has been shown in case of *Anabaena* (Scherer et al., 1981). This was confirmed by the results of Yamamoto & Nakahara (2005, 2006). They found, that the growth of *A. flos-aquae* maintained at 20°C, ceased at water temperature 11–12°C, and the lowest temperature at which *A. flos-aquae* could grow was 14°C. Konopka & Brock (1978) observed significant photosynthesis of natural samples at 4°C, but those samples were collected or maintained at higher temperatures than the natural samples in this study. At the highest temperature (35°C) applied in our experiment, there was a drop in the photosynthesis of the natural population of *A. flos-aquae*; the species showed moderate photosynthetic activity, similar to the results of Butterwick et al. (2005). Previously the photosynthetic activity of overwintering natural populations of *A. flos-aquae* was not studied at such a low temperature (2°C) and irradiance ($7.5 \mu\text{mol m}^{-2} \text{s}^{-1}$),

therefore our findings in this temperature and irradiance levels cannot be compared to results of previous studies.

According to the P–T curves, the photosynthetic activity correlated negatively with the temperature under light-limited conditions. New and interesting result is that at the lowest irradiance level ($7.5 \mu\text{mol m}^{-2} \text{s}^{-1}$) the temperature optimum was at 2°C, and it was between 2 and 10°C at the irradiance of $30 \mu\text{mol m}^{-2} \text{s}^{-1}$. For biological functions, the Q_{10} value is ranges generally between 1 and 3 (e.g. Bissinger et al., 2008). According to our results, the Q_{10} values at the applied lowest irradiances were <1 (0.67–0.87), and at irradiances $>150 \mu\text{mol m}^{-2} \text{s}^{-1}$ it was between 1.67 and 2.88, similarly to other studies. These values suggest responsiveness to temperature changes only at mid and higher irradiance levels. These results can change our previous knowledge about the temperature dependence of the photosynthetic activity, and suggest that the responsiveness to temperature changes can be affected by other environmental parameters, like irradiance level. The interaction between temperature and the light availability affected the photosynthetic characteristics of *A. flos-aquae*, but the physiological distinctive features and survival strategies of the species contributed to the success at such extreme conditions.

The ability of *A. flos-aquae* to adapt to light-limited conditions was found higher than other cyanobacterial species (De Nobel et al., 1998). According to our study, the photoadaptation parameter was at each temperature low, the values were similar to those measured by De Nobel et al. (1998). The natural population of *A. flos-aquae* was more susceptible to photoinhibition due to the low-light conditions they grow in Lake Stechlin. At 2°C, the photoadaptation parameter was very low ($20 \mu\text{mol m}^{-2} \text{s}^{-1}$), but the photosynthetic efficiency was at this temperature the highest in the light-limited region, which suggests a competitive advantage of *A. flos-aquae* in case of light-limited conditions. The increase in the photosynthetic efficiency could have been caused by the changes in the ratio of phycobilin/chlorophyll *a* (Vincent, 2007). Photoinhibition was observed at each temperature (except 35°C), the poor ability to utilize high irradiances indicates the adaptation to low irradiance levels. Photoinhibition by PAR usually becomes increasingly severe at low temperatures (e.g. Krause, 1994), but in this study we found low

photoinhibition in the low-temperature range. The low photoinhibition at low temperatures can be attributed to the changed pigment composition and function of the photosystems (Lazarova et al., 2009); similar responses were observed in polar cyanobacteria to withstand the extremes of their environment (e.g. Vincent, 2007).

The deficiency of this study is that the pigment composition of the community was not examined; only the chlorophyll *a* was measured. According to previous studies cyanobacteria which contain phycoerythrin and phycoerythrocyanin, like *A. flos-aquae*, have competitive advantage in light-limited environments (Huisman et al., 1999). Phycoerythrin and phycoerythrocyanin synthesis did appear to be affected by light intensity (Gervais et al., 1997) and composition (Huisman et al., 1999). The light availability under the ice and snow cover was very low in Lake Stechlin in winter 2009/2010, the irradiance varied between 15 and 130 $\mu\text{mol m}^{-2} \text{s}^{-1}$ in the upper 3 m layer during the day. Assuming that the spectral composition of light was similar in Lake Stechlin to that in Lake Pääjärvi (Lei et al., 2011), the green light dominated in the euphotic layer. At low-light conditions, *A. flos-aquae* could absorb green light (525–600 nm range) more efficiently because of its pigment composition (De Nobel et al., 1998; Huisman et al., 1999), therefore the spectral composition of the light under the ice and snow cover was favourable to its development. Despite our results show that the physiological plasticity of the species could enable them to develop deep chlorophyll maxima (DCM), in previous studies the appearance of the species in deeper water layers was not described nor in winter either in summer periods.

According to our results, the studied *A. flos-aquae* population had very similar ecophysiological characteristics to high-latitude cyanobacteria, which are dominant in cold ecosystems (Vincent, 2000): (i) *A. flos-aquae* could grow over a wide temperature range; (ii) at low temperatures it grows at slow rates; (iii) the low photoadaptation parameter enables the growth in dim light environment. Most of the previous studies about cyanobacteria supported the tenet that the optimum of photosynthesis and growth are above 20°C (e.g. Paerl & Huisman, 2008). This was also observed for cyanobacteria isolated from Antarctica: the species could grow over a wide temperature range (5–30°), but most of them were unable to grow at

temperatures <5°C. Therefore, the polar cyanobacteria species are thought to be originated from warmer temperate regions (Seaburg et al., 1981). During long period of their evolution, they developed various adaptive mechanisms (e.g. Vincent, 2000), which contributed to their success and dominance in cold ecosystems. Microbial communities of cold environments are often unusual and intrinsically interesting because they have been subject to long periods of isolation with relatively low levels of disturbance (Vyverman et al., 2010). Not only the climate change (e.g. Vincent et al., 2009, 2011), but the species with high ecophysiological plasticity and success may have profound effects on the structure and efficiency of the food webs. *A. flos-aquae*, because of its ecophysiological plasticity and two temperature optima of its photosynthesis, can spread in habitats where its appearance was inconceivable according to our previous knowledge. The spread of different cyanobacterial species in temperate and polar regions at higher latitudes is often explained with the climate change, but an ongoing cyanobacterial adaptation cannot be ignored. The ecophysiological characteristics of species, like *A. flos-aquae*, may help us to understand the previous adaptation of cyanobacteria to cold environments, and their success in invasion and in extreme environments.

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