



# Consequences of oxidative damage on the fatty acid profile in muscle of *Cichlasoma amazonarum* acutely exposed to copper

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**Abstract** Rapid industrialization results in the production of large quantities of waste that are commonly discharged into water bodies, leading to the damage of the aquatic ecosystem and freshwater organisms. Copper (Cu) can induce oxidative damage in fish muscle, the main fish portion that is consumed by humans. However, the responses of the Amazon fish *Cichlasoma amazonarum* and its capacity to withstand acute Cu concentrations found in Amazon water around mines remain unknown. Thus, the aim of this study was to evaluate whether exposure to Cu causes muscle oxidative stress and/or oxidative damage and impairs the fillet fatty acid profile of *C. amazonarum* acutely exposed to Cu found in Amazon waters around mines. Muscle reactive oxygen species and protein

carbonylation levels were significantly higher in fish exposed to 1500 µg/L Cu compared with the control group, while muscle lipid peroxidation levels were significantly higher in fish exposed to 500, 750, and 1500 µg/L Cu compared with control group. Muscle antioxidant capacity against peroxy radical's levels and glutathione peroxidase activity were significantly lower in fish exposed to 1500 µg/L Cu compared with the control group, while muscle superoxide dismutase activity was significantly lower in fish exposed to 750 and 1500 µg/L Cu compared with control group. The total content of saturated fatty acids was significantly higher in fish exposed to 1500 µg/L Cu compared with the control group, while the total content of monounsaturated fatty acids and sum of n3 fatty acids were significantly lower in fish exposed to 1500 µg/L Cu compared with control group. No significant difference was observed regarding muscle catalase, glutathione S-transferase, and glutathione reductase activities. Based on these lines of evidence, the results of this comprehensive study agree with the initial hypothesis that the exposure to Cu found in Amazon water around mines induces oxidative damage and inhibits enzymatic and non-enzymatic antioxidant response in the muscle of *C. amazonarum* exposed to high Cu levels. Moreover, the impairment of the fillet fatty acid profile appears to be mediated by oxidative damage, representing a negative impact on fish health.

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## Introduction

Metal pollution is largely recognized as a serious threat to the aquatic ecosystem due to the complex sources, persistent and toxic behavior, and bio-accumulation properties (He et al. 2019). Large quantities of metals have been released into aquatic environments worldwide primarily by anthropogenic sources (such as industrial drainages, mining, vehicle emissions, agricultural runoff, sewage disposal, and dam construction) and by natural sources (erosion and weathering of rocks) (Jafari et al. 2018; Hader et al. 2020; Fry et al. 2020), making metals a group of contaminants with high ecological importance due to their serious risk for animal and human health through direct ingestion, dermal contact, and food intake (Zhang et al. 2018).

Copper (Cu) is an abundant metal that occurs as a natural element due to weathering of rocks and volcano eruptions. It is used across the world in many industrial applications (ore refining, electroplating process, applications of fertilizers and pesticides) and consumer products, and consequently, it has been extracted in very high quantities (Siqueira-Gay et al. 2020). This extraction may cause elevation of waterborne Cu concentration and contamination of the aquatic ecosystem (Fry et al. 2020), which generates concerns due to its deleterious effects on the aquatic ecosystem when present in higher levels. Although physiologically essential, waterborne Cu at elevated concentrations can be toxic to fish, causing serious cellular injuries mediated principally by oxidative stress and oxidative damage (Abou Anni et al. 2019; Tesser et al. 2020). Oxidative stress is a reversible alteration of redox status in cell compartments. Oxidative stress can damage cellular molecules (via lipid peroxidation and protein carbonylation), and in so doing, can serve as a cellular signal (Viña et al. 2018). When the redox condition of the cell modulates to a more reduced state, there is an irreversible (i.e., non-physiological) process known as oxidative damage. In other words, oxidative stress precedes oxidative damage that occurs when reducing systems can no longer accommodate the rapid rate of cell component oxidation (Viña et al. 2018). To reduce or avoid the oxidative damage caused by excessive reactive oxygen species (ROS) generation, the antioxidant defense system (enzymatic or non-enzymatic) can be activated as a survival strategy to eliminate ROS overproduced during exposure to waterborne Cu (Boareto et al. 2018; Shekh et al. 2020). In this sense, the first line of antioxidant

defense includes the enzymes superoxide dismutase (SOD) and catalase (CAT), while the second line of the antioxidant system includes glutathione peroxidase (GPx), glutathione reductase (GR), and glutathione S-transferase (GST), as well as non-enzymatic mechanisms that maintain the balance of oxidation-reduction reactions and consequently protection against oxidative damage (Sies et al. 2017). On the other hand, the antioxidant defense system can also be inhibited by Cu levels, which contribute to the intensification of Cu toxicity and fish mortality, as observed by Tesser et al. (2020) for curimatá (*Prochilodus lineatus*) and Zebral et al. (2019) for killifish (*Poecilia vivipara*). To the best of our knowledge, the responses of *Cichlasoma amazonarum* and its capacity to withstand acute Cu exposure in terms of oxidative stress are unknown. Our hypothesis is that the muscle antioxidant system of *C. amazonarum* would be unable to cope with acute Cu exposure levels found in Amazon waters around mines.

Fish is considered a healthy and important dietary resource for many populations worldwide. It is a rich source of health-friendly fatty acids, especially omega-3 (n3) and omega-6 (n6) polyunsaturated fatty acids (PUFA), as well as docosahexaenoic (DHA) and eicosapentaenoic (EPA) acids (Mohanty et al. 2019). Beyond the importance for human nutrition, fatty acids are the fundamental structural components of almost all forms of lipids, acting as precursors of bioactive molecules, and have structural and functional role lined to fish reproduction, osmoregulation, growth, and promoting fat-soluble vitamin absorption and transportation in the body (Sissener et al. 2020). However, due to the capability of fatty acids to accumulate in edible tissues, they might have undesirable effects on quality and freshness of fish flesh due to increase on total content of fats that augments the risk of rancidness, and be responsible for metal toxicity-related health risks in humans during consumption of fish (Das et al. 2018; Afridi et al. 2019). Despite muscle not being an active tissue in accumulating metals compared with other tissues, it is commonly analyzed because it is the main fish part consumed by humans, and although *C. amazonarum* is not commonly consumed by the human population, this species can provide a perspective and projection about meat quality of Amazon fishes linked to fatty acid profile during waterborne Cu exposure. Also, *C. amazonarum* is a fish species that develops a role in the food chain in the aquatic ecosystem

that it is consumed by other big fishes, as pirarucu (*Arapaima gigas*), one of the most consumed fishes by Amazonian population (Hrbek et al. 2005); i.e., the Cu levels present in *C. amazonarum* can be deposited in pirarucu and consequently affecting humans. A recent study conducted by Das et al. (2018) revealed that long-term metal exposure (lead, Cu, cadmium, nickel, and zinc) alters the fatty acid composition of Indian major carps (*Labeo rohita*, *Catla catla*, and *Cirrhinus cirrhosus*) compromising fillet quality and revealing a direct relationship between oxidative damage/impairment on antioxidant defense system and changes of muscle fatty acid profile. Thus, our hypothesis is that acute exposure to Cu can affect the muscle fatty acid profile of *C. amazonarum*, and that oxidative damage can be involved in this process.

Based on these lines of evidence, the aim of this study was to evaluate whether exposure to Cu causes muscle oxidative stress and/or oxidative damage and impairs the fillet fatty acid profile of *C. amazonarum* acutely exposed to Cu found in Amazon waters around mines.

## Material and methods

### Chemical

Copper sulfate ( $\text{CuSO}_4$ ; 99% chemical purity; molecular weight 159.609 g/mol) was purchased from Sigma-Aldrich (St. Louis, MO, USA). A Cu stock solution (dissolved in water) was prepared from  $\text{CuSO}_4$  and spiked into the exposure waters 24 h before test, to allow equilibration.

### Fish maintenance and experimental design

*Cichlasoma amazonarum* adults were bought from a fish farm (Amazonas, Brazil) and transferred to the Laboratory of Ecophysiology and Molecular Evolution (Amazonas, Brazil). A period of 30 days of acclimation was used to recover from stress and to acclimatize to laboratory conditions in continuously aerated 3000-L tanks with running water, being fed daily until satiation with commercial pelleted food (36% protein). Feeding was suspended 24 h prior to experiments.

A total of 108 juvenile *C. amazonarum* (mean weight,  $20.94 \pm 4.01$  g; mean length,  $8.14 \pm 1.09$  cm) were randomly transferred into eighteen 30-L glass

aquaria and divided into six groups ( $n = 6$  in each group, in triplicate), with increasing nominal Cu levels, as follows: 0.0 (control), 25, 250, 500, 750, and 1500  $\mu\text{g/L}$ , for 96 h. Fish were fasted over the course of the test, mortality was recorded every day, and dead fish were removed. The Cu nominal levels used in this study were based on the environmental concentrations of Cu found in the Amazon water around mines (Santana and Barroncas, 2007), as well as in a curve of Cu concentrations used for other Amazon fishes, as cardinal tetra (*Paracheirodon axelrodi*) (Crémazy et al. 2016) and *C. amazonarum* (Baldissera et al., 2020). The water quality variables were evaluated daily and remained within acceptable limits throughout the experimental period, as follows: temperature ( $28.7 \pm 0.4$  °C), dissolved oxygen ( $6.3 \pm 0.3$  mg/L), and pH ( $5.7 \pm 0.3$ ). The Ethics Committee of INPA authorized all experimental procedures under the number 004/2018.

Water was collected in order to quantify the total Cu content on 0 h (beginning of the experiment), being determined by atomic absorption spectrometry using graphite furnace technique (Perkin-Elmer AAnalyst 800 AA spectrophotometer, Norwalk, CT, USA), as described in detail by Crémazy et al. (2016). The Cu concentrations in water of each group in the analyzed moments are represented in Table 1. Waterborne Cu levels of the control group was explained by Cu content on laboratory water, that is,  $17.2 \pm 1.6$   $\mu\text{g/L}$ .

### Muscle collection

After 96 h Cu exposure, the muscle of two juveniles from each tank ( $n = 6$  per treatment) was collected under anesthesia with tricaine methanesulfonate (MS-222) at 1.5 g/L followed by spinal cord sectioning.

### Muscle oxidative stress-related parameters

A glass Potter homogenizer was used to homogenize muscle tissue with 10 mM Tris-HCl buffer pH 7.4. Thereafter, the homogenate was centrifuged at 2000g for 10 min, and the supernatants were collected and stored at  $-20$  °C until analyses. The techniques proposed by LeBel et al. (1992), Monserrat et al. (2003), and Reznick and Packer (1994) were used to determine muscle ROS, lipid peroxidation (LPO), and carbonyl protein levels, respectively. The results were expressed

**Table 1** Concentrations of copper (Cu) in water tanks of *Cichlasoma amazonarum* at 0 (beginning of experiment) and 96 h (end of experiment)

Nominal waterborne Cu levels ( $\mu\text{g/L}$ )	Real waterborne Cu levels ( $\mu\text{g/L}$ )	
	0 h	96 h
0.0 (control)	19.00 $\pm$ 0.41	17.25 $\pm$ 00.48
25	29.45 $\pm$ 0.21	29.47 $\pm$ 0.21
250	278.875 $\pm$ 2.06	258.75 $\pm$ 1.56
500	474.00 $\pm$ 0.14	367.00 $\pm$ 11.68
750	783.25 $\pm$ 7.08	434.00 $\pm$ 4.22
1500	1679.01 $\pm$ 22.01	1058.50 $\pm$ 23.79

as U DCF/mg of protein,  $\mu\text{mol}$  CHP/g of tissue, and nmol of carbonyls formed/mg of protein, respectively.

The methods proposed by Amado et al. (2009), Marklund and Marklund (1974), Aebi (1984), Paglia and Valentine (1967), Habig et al. (1974), and Carlberg and Mannervik (1985) were used to measure muscle antioxidant capacity against peroxy radicals (ACAP) levels, as well as SOD, CAT, GPx, GR, and GST activities, respectively. The results were presented as follows: fluorescence units/mg of protein, as U SOD/mg of protein,  $\mu\text{mol}$  CAT/mg of protein, nmol NADPH/h/mg of protein,  $\mu\text{mol}$  GS-DNB/min/mg of protein, and U GR/mg of protein, respectively.

Protein content was evaluated by the Coomassie blue G dye method using bovine serum albumin as the standard (Read and Northcote 1981).

### Muscle fatty acid profile

The total lipids were extracted from muscle samples according to Bligh and Dyer (1959) method, and the lipids were transesterified/esterified to fatty acid methyl esters (FAME) as proposed by Hartman and Lago (1973), as described in detail by Baldissera et al. (2020). Results were expressed as a percentage of the total area of the chromatograms.

### Muscle Cu levels

Muscle samples were decomposed with 2.0 mL 65% nitric acid ( $\text{HNO}_3$ ), as previously reported by Baldissera et al. (2020). The Cu levels were determined using an inductively coupled plasma mass spectrometer (ICP-MS) (Elan DCR II, Perkin-Elmer, Canada) equipped with a concentric nebulizer (Meinhard, USA), a cyclonic spray chamber (Glass Expansion, Inc., Australia), and a quartz torch with a quartz injector tube (2 mm

diameter). The operational parameters to determine Cu levels via ICP-MS were as follows: radiofrequency power and plasma (1300 W), main flow rate (15 L/min), auxiliary flow rate (1.2 L/min), nebulizer flow rate (1.04 L/min), sampling cone and skimmer (Pt), and element monitored ( $^{65}\text{Cu}$ ).

### Statistical analyses

Normality and homogeneity of variances for all data were tested by the Shapiro–Wilk’s and Levene’s tests, respectively. Data obtained were compared using a one-way analysis of variance (ANOVA) followed by the Tukey post hoc analysis. Significance was set at  $p < 0.05$ , and data were presented as mean  $\pm$  standard error.

## Results

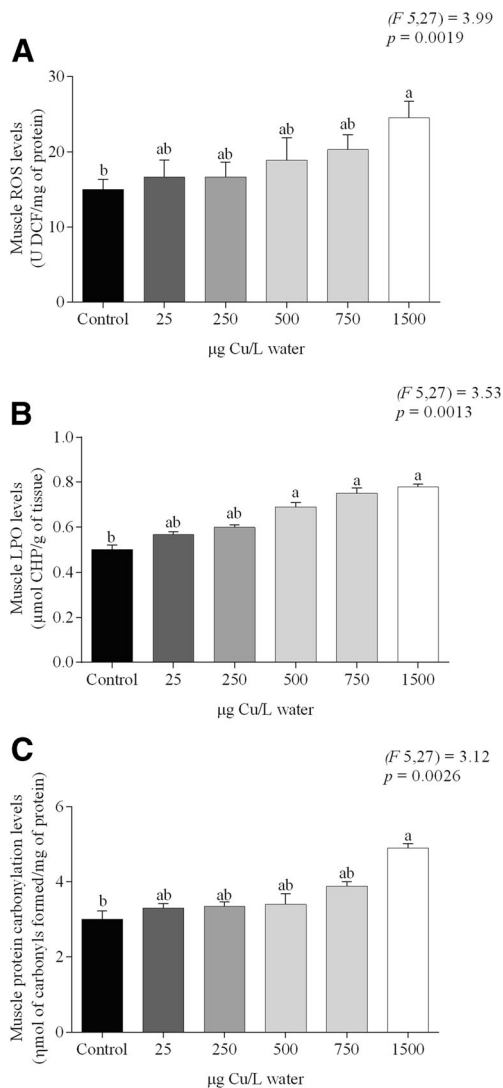
### Mortality

No mortality was observed after 24, 48, and 72 h of exposure to different Cu levels. On the other hand, 1 fish exposed to 750  $\mu\text{g}$  Cu/L and 2 fish exposed to 1500  $\mu\text{g}$  Cu/L water died after 96 h exposure.

### Muscle oxidative stress-related parameters

Muscle ROS and protein carbonylation levels were significantly higher in fish exposed to 1500  $\mu\text{g/L}$  Cu compared with the control group, while muscle LPO levels were significantly higher in fish exposed to 500, 750, and 1500  $\mu\text{g/L}$  Cu compared with control group (Fig. 1).

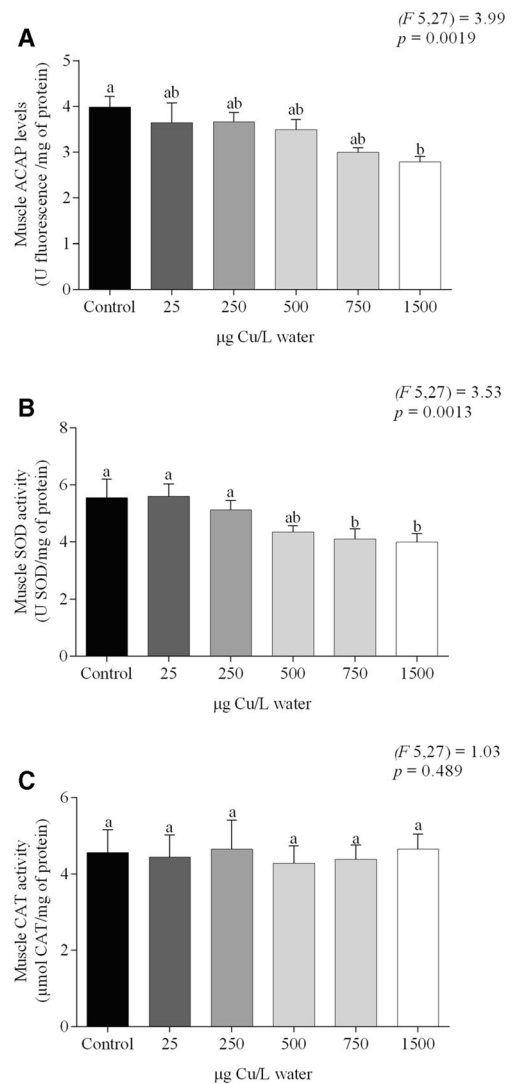
Muscle ACAP levels were significantly lower in fish exposed to 1500  $\mu\text{g/L}$  Cu compared with the control



**Fig. 1** Muscle reactive oxygen species (ROS) (a), lipid peroxidation (LPO) (b), and protein carbonylation (c) levels of *Cichlasoma amazonarum* adults maintained under control condition (no copper addition in the water) or exposed to waterborne copper for 96 h. Bars that do not share a common superscript letter differ significantly using a one-way analysis of variance (ANOVA) for independent samples followed by Tukey post hoc test ( $p < 0.05$ ;  $n = 6$  per group)

group, while muscle SOD activity was significantly lower in fish exposed to 750 and 1500 µg/L Cu compared with control group and groups exposed to 25 and 250 µg/L Cu. No significant difference was observed between groups regarding muscle CAT activity (Fig. 2).

Muscle GPx activity was significantly lower in fish exposed to 1500 µg/L Cu compared with the control group and groups exposed to 25 and 250 µg/L Cu. No

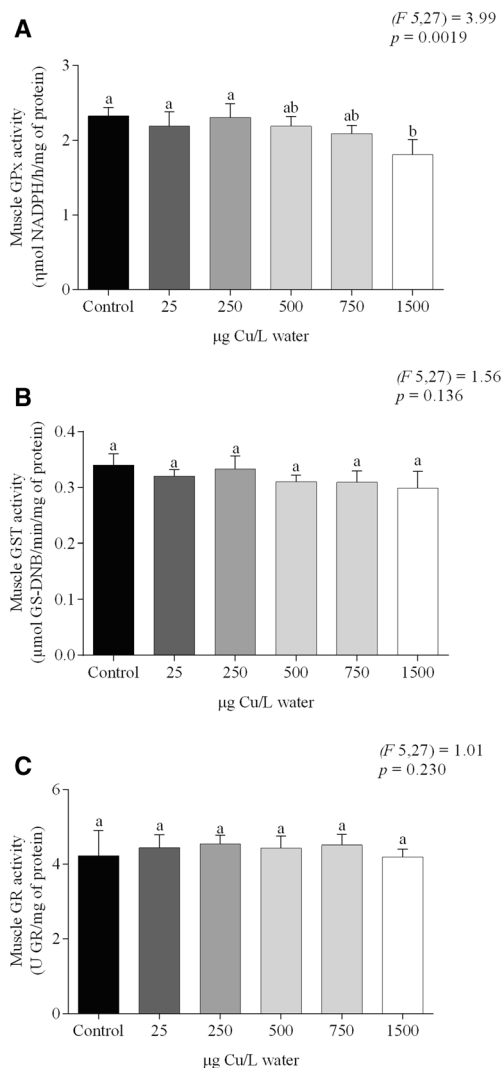


**Fig. 2** Muscle antioxidant capacity against peroxy radicals (ACAP) (a) levels, superoxide dismutase (SOD) (b), and catalase (CAT) (c) activities of *Cichlasoma amazonarum* adults maintained under control condition (no copper addition in the water) or exposed to waterborne copper for 96 h. Bars that do not share a common superscript letter differ significantly using a one-way analysis of variance (ANOVA) for independent samples followed by Tukey post hoc test ( $p < 0.05$ ;  $n = 6$  per group)

significant difference was observed between groups regarding muscle GST and GR activities (Fig. 3).

#### Muscle fatty acid profile

Total SFA was significantly highest, while the total of MUFA and omega 6 (n6) were significantly lowest in the muscle of fish exposed to 1500 µg/L Cu (Table 2).



**Fig. 3** Muscle glutathione peroxidase (GPx) (a), glutathione S-transferase (GST) (b), and glutathione reductase (GR) (c) activities of *Cichlasoma amazonarum* adults maintained under control condition (no copper addition in the water) or exposed to waterborne copper for 96 h. Bars that do not share a common superscript letter differ significantly using a one-way analysis of variance (ANOVA) for independent samples followed by Tukey post hoc test ( $p < 0.05$ ;  $n = 6$  per group)

Muscle levels of C10:0 (capric acid), C13:0 (tridecanoic acid), and C18:0 (stearic acid) fatty acids were significantly highest in fish exposed to 1500 µg/L Cu. Muscle levels of C16:1 (palmitoleic acid) and C18:1n7c (vaccenic acid) fatty acids were significantly lowest in fish exposed to 1500 µg/L Cu, while the levels of C18:1n9c (oleic acid) was significantly lower in fish exposed to 1500 µg/L Cu compared with control group and fish exposed to 500 µg/L Cu. Muscle levels of

C18:2n6 (linoleic acid) fatty acid were significantly lowest in fish exposed to 1500 µg/L Cu, while the levels of C22:5n3 (docosapentaenoic acid) and C22:6n3 (docosahexaenoic acid) fatty acids were significantly lower in fish exposed to 500, 750, and 1500 µg/L Cu compared with control group (Table 3).

#### Muscle Cu levels

Muscle Cu levels were significantly highest in fish exposed to 1500 µg/L Cu (Table 4).

#### Discussion

The current study clearly demonstrates that short-term exposure to different Cu concentrations induced muscle oxidative damage of *C. amazonarum* via alteration in oxidative homeostasis and impairment of enzymatic and non-enzymatic antioxidant defense system. Moreover, our data revealed that some fatty acids in muscle fish were affected by exposure to Cu concentrations as a consequence of excessive muscle lipid peroxidation and impairment of antioxidant system, which can affect the fatty acid quality and freshness value of fish fillet.

Oxidative stress is considered an important mode of toxic action for Cu in many fish species, being considered as a molecular initiation event or an early key event in adverse outcomes of exposure to Cu. However, some evidence has suggested that differences in several aspects of Cu toxicity, including oxidative stress, may be related to species-specific physiological abilities to regulate the redox system (Shekh et al. 2020). In this sense, oxidative stress-related parameters were evaluated in *C. amazonarum* in order to investigate the sensitivity of this fish species to oxidative stress when exposed to waterborne Cu levels found in Amazon water around mines. In the present study, muscle ROS levels were significantly higher in fish exposed to the highest waterborne Cu levels tested (1500 µg/L), which suggests an increase in free radical formation during acute exposure to high Cu levels. In agreement with our observations, a study conducted by Jiang et al. (2015) revealed a significant increase on muscle ROS levels of common carp (*Cyprinus carpio*) exposed for 96 h to waterborne Cu level (560 µg/L), concluding that Cu exposure depleted the ROS scavenging ability of fish and oxidized the lipids and proteins in the muscle. As a consequence of muscle ROS overproduction, we observed a

**Table 2** Total saturated fatty acids (SFA), monounsaturated fatty acids (MUFA), polyunsaturated fatty acids (PUFA), omega 3 (n3), omega 6 (n6), and omega 9 (n9) in the meat of *Cichlasoma**amazonarum* adults maintained under control condition (no copper addition in the water), or exposed to waterborne copper for 96 h

Fatty acids/groups	0.0 µg/L (control)	25 µg/L	250 µg/L	500 µg/L	750 µg/L	1500 µg/L	<i>F</i> and <i>p</i> values
∑ SFA	59.20 ± 1.28 <sup>b</sup>	58.96 ± 1.69 <sup>b</sup>	61.07 ± 1.31 <sup>b</sup>	57.64 ± 1.56 <sup>b</sup>	57.16 ± 2.98 <sup>b</sup>	64.95 ± 1.44 <sup>a</sup>	<i>2.541; 0.048</i>
∑ MUFA	34.01 ± 1.09 <sup>a</sup>	34.39 ± 1.05 <sup>a</sup>	32.99 ± 1.26 <sup>a</sup>	34.97 ± 1.52 <sup>a</sup>	34.16 ± 1.66 <sup>a</sup>	27.55 ± 8.88 <sup>b</sup>	<i>4.369; 0.006</i>
∑ PUFA	7.51 ± 0.71 <sup>a</sup>	8.01 ± 0.34 <sup>a</sup>	6.84 ± 0.71 <sup>a</sup>	7.39 ± 0.52 <sup>a</sup>	8.55 ± 0.12 <sup>a</sup>	6.55 ± 0.12 <sup>a</sup>	2.11; 0.078
∑ n3	1.49 ± 0.18 <sup>a</sup>	1.00 ± 0.15 <sup>a</sup>	1.05 ± 0.25 <sup>a</sup>	1.01 ± 0.12 <sup>a</sup>	1.04 ± 0.29 <sup>a</sup>	1.38 ± 0.07 <sup>a</sup>	1.154; 0.364
∑ n6	6.70 ± 0.21 <sup>a</sup>	6.40 ± 0.25 <sup>a</sup>	6.23 ± 0.09 <sup>a</sup>	6.57 ± 0.48 <sup>a</sup>	6.14 ± 0.27 <sup>a</sup>	5.26 ± 0.02 <sup>b</sup>	<i>2.987; 0.047</i>
∑ n9	29.39 ± 0.53 <sup>a</sup>	28.96 ± 1.19 <sup>a</sup>	27.75 ± 0.71 <sup>a</sup>	30.54 ± 1.05 <sup>a</sup>	30.35 ± 1.56 <sup>a</sup>	26.08 ± 1.10 <sup>a</sup>	2.366; 0.073

Values accompanied by different letters in the same line are statistically different, considering  $p < 0.05$ . *F* and *p* values highlighted in Italic values indicate significant effects of Cu

significant increase on muscle lipid peroxidation (500, 750, and 1500 µg/L Cu) and protein carbonylation (1500 µg/L Cu) compared with control group, which reveals that excessive muscle ROS formation elicits the oxidation of lipid and proteins, contributing to Cu toxic effects. Tesser et al. (2020) also reported muscle lipid damage in curimatá (*C. lineatus*) exposed 96 h to 24 µg/L Cu, concluding that lipid oxidation contributes to impairment of membrane fluidity and permeability and consequently, increased fish mortality. At the best of our knowledge, there are no reports about the effects of Cu on muscle protein damage. However, Ransberry et al. (2016) reported that killifish (*Fundulus heteroclitus*) exposed for 96 h to a Cu concentration (25 µg/L) presented gill protein damage via augmentation on protein carbonylation levels, being a consequence of excessive free radicals production. Under the tested condition (exposure time, Cu level, and analyzed tissue), it is possible to suggest that *C. amazonarum* is little affected in terms of oxidative damage, being more sensitive to lipid peroxidation than protein damage when exposed to different Cu levels.

In response to the increase of free radical formation, including ROS, the enzymatic or non-enzymatic antioxidant system can be upregulated or downregulated, and this response exerts a significant role in Cu toxicity. Thus, to evaluate whether Cu-induced ROS overproduction is related to inhibited scavenging capability, we determined the ROS scavenge ability in fish muscle via measurement of antioxidant enzymes and non-enzymes. In consequence of exacerbated muscle ROS production, fish muscle ACAP levels (1500 µg/L) and muscle SOD (750 and 1500 µg/L) activity were significantly inhibited by Cu exposure, indicating a minor

capacity to scavenge the Cu-induced ROS overproduction. As our observation, Jiang et al. (2015) showed that muscle SOD activity is inhibited in common carp (*C. carpio*) exposed for 96 h to 560 µg/L, which contributed to Cu-induced lipid and protein damage. The antioxidant enzyme SOD acts as the first line of defense against oxygen toxicity through the conversion of superoxide anion to hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) and water (H<sub>2</sub>O) (Feng et al. 2015), and its inhibition indicates a minor capacity to eliminate an important ROS molecule, which can also explain its muscle increase in fish exposed to high Cu levels. In our study, no significant difference was observed regarding muscle CAT activity, which reveals that this enzyme is not involved on ROS elimination in the tested conditions, since it plays an important role in ROS detoxification during waterborne Cu exposure via degradation of H<sub>2</sub>O<sub>2</sub> into H<sub>2</sub>O and O<sub>2</sub> (Diaz-Alba et al. 2017). Also, it is important to emphasize that the absence of significant difference on muscle CAT activity may be a consequence of no significant increase on H<sub>2</sub>O<sub>2</sub>, since this enzyme changes in response to alteration on H<sub>2</sub>O<sub>2</sub> levels. However, it is important to emphasize that both conditions have been found in the literature; i.e., CAT activity can be inhibited or stimulated, and these conditions are dependent on the length of time of exposure, Cu concentration, tissue analyzed, and fish species (Kim et al. 2014). Moreover, a significant decrease in muscle GPx activity was found in *C. amazonarum* exposed to high waterborne Cu concentration (1500 µg/L), as observed by Jiang et al. (2015) in the muscle of common carp exposed for 96 h to 560 µg/L. This enzyme presents widespread cellular distribution and catalyzes the reduction of both H<sub>2</sub>O<sub>2</sub> and organic peroxides, indicating that GPx is one of the

**Table 3** The fillet fatty acid profile (g/kg) of *Cichlasoma amazonarum* adults maintained under control condition (no copper addition in the water) or exposed to waterborne copper for 96 h

Fatty acids/groups	0.0 µg/L (control)	25 µg/L	250 µg/L	500 µg/L	750 µg/L	1500 µg/L	<i>F</i> and <i>p</i> values
C10:0	0.09 ± 0.007 <sup>b</sup>	0.11 ± 0.019 <sup>b</sup>	0.14 ± 0.007 <sup>b</sup>	0.08 ± 0.014 <sup>b</sup>	0.08 ± 0.015 <sup>b</sup>	0.19 ± 0.02 <sup>a</sup>	<i>9.80; 0.0001</i>
C13:0	0.0 ± 0.0 <sup>b</sup>	0.0 ± 0.0 <sup>b</sup>	0.0 ± 0.0 <sup>b</sup>	0.0 ± 0.0 <sup>b</sup>	0.0 ± 0.0 <sup>b</sup>	0.14 ± 0.08 <sup>a</sup>	<i>5.11; 0.0007</i>
C14:0	1.21 ± 0.08 <sup>a</sup>	1.28 ± 0.12 <sup>a</sup>	1.12 ± 0.15 <sup>a</sup>	1.26 ± 0.10 <sup>a</sup>	1.12 ± 0.15 <sup>a</sup>	0.92 ± 0.28 <sup>a</sup>	1.45; 0.241
C15:0	0.10 ± 0.02 <sup>a</sup>	0.16 ± 0.008 <sup>a</sup>	0.15 ± 0.01 <sup>a</sup>	0.15 ± 0.02 <sup>a</sup>	0.12 ± 0.02 <sup>a</sup>	0.15 ± 0.01 <sup>a</sup>	1.69; 0.194
C16:0	37.59 ± 0.74 <sup>a</sup>	37.33 ± 1.09 <sup>a</sup>	39.28 ± 0.96 <sup>a</sup>	36.62 ± 1.16 <sup>a</sup>	34.18 ± 1.19 <sup>a</sup>	39.29 ± 0.96 <sup>a</sup>	1.21; 0.199
C17:0	0.13 ± 0.024 <sup>a</sup>	0.17 ± 0.03 <sup>a</sup>	0.12 ± 0.01 <sup>a</sup>	0.11 ± 0.02 <sup>a</sup>	0.13 ± 0.02 <sup>a</sup>	0.12 ± 0.01 <sup>a</sup>	0.800; 0.560
C18:0	20.50 ± 1.29 <sup>b</sup>	19.52 ± 1.16 <sup>b</sup>	19.98 ± 0.60 <sup>b</sup>	19.17 ± 1.23 <sup>b</sup>	18.72 ± 0.34 <sup>b</sup>	25.58 ± 2.27 <sup>a</sup>	<i>2.971; 0.032</i>
C20:0	0.11 ± 0.06 <sup>a</sup>	0.12 ± 0.02 <sup>a</sup>	0.10 ± 0.02 <sup>a</sup>	0.08 ± 0.06 <sup>a</sup>	0.10 ± 0.08 <sup>a</sup>	0.09 ± 0.05 <sup>a</sup>	1.031; 0.094
C16:1	2.73 ± 0.44 <sup>a</sup>	2.68 ± 0.41 <sup>a</sup>	2.21 ± 0.24 <sup>a</sup>	2.74 ± 0.45 <sup>a</sup>	2.16 ± 0.04 <sup>a</sup>	0.86 ± 0.14 <sup>b</sup>	<i>3.041; 0.032</i>
C17:1	0.05 ± 0.03 <sup>a</sup>	0.04 ± 0.01 <sup>a</sup>	0.06 ± 0.02 <sup>a</sup>	0.04 ± 0.02 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.03 ± 0.02 <sup>a</sup>	0.321; 0.234
C18:1t	0.01 ± 0.01 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.02 ± 0.01 <sup>a</sup>	0.874; 0.111
C18:1n7c	24.23 ± 1.37 <sup>a</sup>	24.83 ± 0.52 <sup>a</sup>	22.91 ± 1.56 <sup>a</sup>	25.92 ± 1.34 <sup>a</sup>	24.95 ± 1.81 <sup>a</sup>	19.16 ± 1.41 <sup>b</sup>	<i>2.871; 0.043</i>
C18:1n9c	1.83 ± 0.19 <sup>a</sup>	1.60 ± 0.11 <sup>ab</sup>	1.57 ± 0.07 <sup>ab</sup>	1.89 ± 0.20 <sup>a</sup>	1.61 ± 0.07 <sup>ab</sup>	1.03 ± 0.07 <sup>b</sup>	<i>3.696; 0.016</i>
C20:1n9	0.05 ± 0.011 <sup>a</sup>	0.03 ± 0.011 <sup>a</sup>	0.06 ± 0.013 <sup>a</sup>	0.06 ± 0.013 <sup>a</sup>	0.05 ± 0.011 <sup>a</sup>	0.05 ± 0.021 <sup>a</sup>	0.428; 0.822
C22:1n9	4.35 ± 0.54 <sup>a</sup>	4.49 ± 0.28 <sup>a</sup>	4.40 ± 0.51 <sup>a</sup>	4.29 ± 0.39 <sup>a</sup>	4.97 ± 0.73 <sup>a</sup>	5.15 ± 1.39 <sup>a</sup>	0.268; 0.926
C24:1n9	0.35 ± 0.06 <sup>a</sup>	0.35 ± 0.05 <sup>a</sup>	0.35 ± 0.08 <sup>a</sup>	0.32 ± 0.06 <sup>a</sup>	0.34 ± 0.07 <sup>a</sup>	0.35 ± 0.14 <sup>a</sup>	0.250; 0.983
C18:2n6	6.32 ± 0.18 <sup>a</sup>	6.44 ± 0.39 <sup>a</sup>	5.94 ± 0.08 <sup>a</sup>	6.34 ± 0.51 <sup>a</sup>	6.79 ± 0.72 <sup>a</sup>	4.82 ± 0.11 <sup>b</sup>	<i>2.983; 0.043</i>
C18:3n3	0.08 ± 0.02 <sup>a</sup>	0.09 ± 0.03 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.08 ± 0.03 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.234; 0.212
C18:3n6	0.04 ± 0.01 <sup>a</sup>	0.05 ± 0.01 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.06 ± 0.03 <sup>a</sup>	0.334; 0.183
C20:2n6	0.06 ± 0.04 <sup>a</sup>	0.09 ± 0.03 <sup>a</sup>	0.07 ± 0.02 <sup>a</sup>	0.05 ± 0.02 <sup>a</sup>	0.07 ± 0.03 <sup>a</sup>	0.06 ± 0.01 <sup>a</sup>	0.455; 0.212
C20:3n6	0.24 ± 0.04 <sup>a</sup>	0.24 ± 0.06 <sup>a</sup>	0.24 ± 0.03 <sup>a</sup>	0.22 ± 0.03 <sup>a</sup>	0.23 ± 0.06 <sup>a</sup>	0.26 ± 0.06 <sup>a</sup>	0.395; 0.846
C20:5n3	0.01 ± 0.002 <sup>a</sup>	0.02 ± 0.002 <sup>a</sup>	0.02 ± 0.002 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	0.0 ± 0.0 <sup>a</sup>	1.332; 0.098
C22:5n3	0.16 ± 0.02 <sup>a</sup>	0.14 ± 0.02 <sup>a</sup>	0.19 ± 0.03 <sup>ab</sup>	0.18 ± 0.05 <sup>b</sup>	0.10 ± 0.05 <sup>b</sup>	0.08 ± 0.01 <sup>b</sup>	<i>3.168; 0.034</i>
C22:6n3	1.34 ± 0.18 <sup>a</sup>	1.17 ± 0.03 <sup>a</sup>	0.86 ± 0.20 <sup>ab</sup>	0.74 ± 0.07 <sup>b</sup>	0.74 ± 0.17 <sup>b</sup>	0.65 ± 0.05 <sup>b</sup>	<i>3.426; 0.027</i>

Capric acid (C10:0), tridecanoic acid (C13:0), myristic acid (C14:0), pentadecanoic acid (C15:0), palmitic acid (C16:0), heptadecanoic acid (C17:0), stearic acid (C18:0), arachidic acid (C20:0), palmitoleic acid (C16:1), heptadecenoic acid (C17:1), oleic acid trans isomer (C18:1t), vaccenic acid (C18:1n7c), cis-9-octadecenoic acid (oleic acid) (C18:1n9c), cis-11-eicosenoic acid (C20:1n9), erucic acid (C22:1n9), nervonic acid (C24:1n9), linoleic acid (C18:2n6), alpha-linolenic acid (C18:3n3), gamma-linolenic acid (C18:3n6), cis-11,14-eicosadienoic acid (C20:2n6), cis-8,11,14-eicosatrienoic acid (C20:3n6), eicosapentaenoic acid (C20:5n3), docosapentaenoic acid (C22:5n3), and docosahexaenoic acid (C22:6n3). Values accompanied by different letters in the same line are statistically different considering  $p < 0.05$ . *F* and *p* values highlighted in italics indicate a signification effect of Cu

most important enzymes of the antioxidant defense system against Cu-induced oxidative damage (Díaz-Alba et al. 2017). Moreover, GPx plays a key role in

protecting cell against lipid damage, and its depression can also explain the lipid damage observed in *C. amazonarum* exposed to Cu levels. Finally, no

**Table 4** Concentrations of copper (Cu) in the muscle of *Cichlasoma amazonarum* adults after 96 h of exposure

Group/tissue	Waterborne Cu levels (µg/L)						<i>F</i> and <i>p</i> values
	0.0 (control)	25	250	500	750	1500	
Cu in muscle (µg/g)	0.12 ± 0.011 <sup>b</sup>	0.15 ± 0.011 <sup>b</sup>	0.15 ± 0.012 <sup>b</sup>	0.15 ± 0.014 <sup>b</sup>	0.15 ± 0.02 <sup>b</sup>	0.21 ± 0.001 <sup>a</sup>	<i>6.741; 0.0005</i>

Values accompanied by different letters in the same line are statistically different, considering  $p < 0.05$ . *F* and *p* values highlighted in italics indicate a signification effect of Cu



significant difference was observed regarding muscle GST and GR activities, as reported by Jiang et al. (2015) that verified that both enzymatic activities were not significantly affected in common carp (*C. carpio*) exposed for 96 h to 600 µg/L Cu. Also, it is important to highlight that reduced activities of antioxidant enzymes would result in the accumulation of ROS, confirmed by a sharp increase in LPO and protein carbonyl content in *C. amazonarum*. In the process, the weakened antioxidant defense system could not eliminate or neutralize the excessive ROS, which may further, in turn, inhibit enzyme activities or even degrade the enzymes through ROS-induced enzyme oxidation and protein damage (Yuan et al. 2017). Moreover, it is important to emphasize that Cu was able to accumulate only in the muscle of fish exposed to 1500 µg/L Cu, i.e., at the same level that caused an increase in ROS levels and impairment on the muscle antioxidant system. This allows us to hypothesize a direct effect of Cu accumulation on Cu-induced oxidative damage.

Recently, a study conducted by Das et al. (2018) revealed that alteration in oxidative balance and diminished activity of antioxidant enzymes are involved in change in the muscle fatty acid composition of Indian carps exposed to different metals. In this sense, we decided to evaluate whether acute exposure to Cu concentrations is capable of altering the muscle fatty acid composition of *C. amazonarum*. In the present study, exposure to a high Cu concentration (1500 µg/L) caused a significant increase on SFA level, while the total contents of MUFA and sum of n6 were significantly lower than the other groups of fatty acids, i.e., alterations on muscle fatty acid composition occurred at the concentration that Cu-induced oxidative damage and alterations on the muscle antioxidant system, which suggests a relationship between oxidative damage/alteration on the antioxidant system and alterations on the muscle fatty acid profile of *C. amazonarum*. Partially in agreement with our observations, Afridi et al. (2019) reported that Mrigal (*Cirrhinus mrigala*) exposed for 96 h to 1340 µg/L Cu presented a significant increase on the total content of muscle SFA, while no significant difference was observed regarding the total of MUFA and PUFA levels, as well as the sum of n3 and n6 fatty acids. Besides the effect of oxidative damage, these results can also be explained and can occur by two reasons: the first is linked to lipid mobilization to meet the energy demand through the oxidation process to help the fish to cope with challenge imposed by metals,

as reviewed by Javed and Usmani (2019) for fish exposure to other metals. The second can be linked to a self-regulating and compensative mechanism of the fish muscle membrane to maintain the structural and functional stability when exposed to metals, as observed by Xiang et al. (2020) in gills of common carp exposed to silver. It is very important to emphasize that exposure to 500, 750, and 1500 µg/L Cu elicited a significant reduction on EPA and DHA fatty acids, two important fatty acids for fish health linked to the improvement of fish growth, enzyme activity, and cell signaling process (Calder and Yaqoob 2009). Thus, these alterations on the muscle fatty acid profile can negatively impair fish health, and probably affect human health.

Although *C. amazonarum* is not a fish used for human food, these data are an alarm about the possible quality of fillet associated with the composition of fatty acid in other fish in the Amazon. The fatty acid composition of the fish fillet is an important trait for consumers because some MUFA and PUFA fatty acids promote health (DiNicolantonio 2017), while some SFA are considered unhealthy (Mensink 2016). An increase of total content of SFA in the muscle of fish exposed to 1500 µg/Cu can be considered a negative impact for human nutrition due to their effects on cholesterol metabolism and higher LDL cholesterol levels, both of which are risk factors for coronary heart disease (CHD) (Mensink 2016). Similarly, a reduction in the total content of MUFAs can also be negative for human health since their consumption has been related to a reduction in cardiovascular diseases because they decrease inflammation, reduce triglycerides, and total cholesterol levels (Michielsen et al. 2019). Regarding specific fatty acids, only 9 from a total of 24 were significantly altered in the muscle of *C. amazonarum* exposed to Cu concentrations. In disagreement with our observations, Afridi et al. (2019) did not observe a significant difference in any of these fatty acids in the muscle of mrigal exposed for 96 h to 1340 µg/L Cu, but only a significant increase on C14:0 and a significant reduction on C22:0 (docosanoic acid) fatty acids compared with fish unexposed to Cu. Although without significant effects on the total content of PUFA, exposure to 500, 750, and 1500 µg/L Cu elicited a significant reduction on EPA and DHA fatty acids, which represents a concern for consumers. A meta-analysis by Kotwal et al. (2012) revealed that the consumption of EPA and DHA protects against cardiovascular death. Moreover, a meta-

analysis of 31 clinical trials conducted by Delgado-Lista et al. (2012) revealed that consumption of marine-derived EPA and DHA for a period of 6 months reduced any cardiovascular event by 10%, and lowered risk of cardiac death and fatal or non-fatal coronary events by 9% and 18%, respectively, evidencing the importance of these fatty acids for human nutrition. In summary, exposure to Cu concentrations impairs the fish muscle fatty acid profile, which can cause negative impacts for fish consumers.

Based on this evidence, the results of this comprehensive study agree with the initial hypothesis that the exposure to waterborne Cu induced oxidative damage and inhibited enzymatic and non-enzymatic antioxidant response in the muscle of *C. amazonarum* exposed to high Cu levels. Moreover, the impairment of the fillet fatty acid profile appears to be mediated by oxidative damage, representing a negative impact on fish health and is a concern for fish consumers.

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#### Compliance with ethical standards

**Competing interests** The authors declare that they have no competing interests.

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