

Hydro-chemical characterization of groundwater and evaluation of health risk assessment for fluoride contamination areas in the eastern blocks of Purulia district, India

Puja Chowdhury¹ 💿 · Bhabani Prasad Mukhopadhyay¹ · Siperna Nayak¹ · Amit Bera¹

Received: 13 July 2020 / Accepted: 8 October 2021 / Published online: 31 October 2021 © The Author(s), under exclusive licence to Springer Nature B.V. 2021

Abstract

Purulia is a hard rock terrain where water scarcity as well as water quality degradation has been a major threat for the past few decades. The prolong use of fluoride contaminated groundwater causes serious health issues in human. The study was performed during the post-monsoon period for better understanding of the general geochemistry along with the fluoride contamination of groundwater and analysis of the health risk factor for the local dwellers. The physico-chemical parameters were analysed during the field work and rest in the laboratory using the standard procedures. The statistical mean values of the cations Ca+2; Na+; Mg+2; K+; and Fe+2 are 91.53, 42.3, 31.76, 3.58 and 0.93 mg/l and for anions HCO₃⁻, Cl⁻, SO₄⁻, NO₃⁻ and F⁻ are 231.67, 106.81, 82.83, 31.52 and 4.06 mg/l, respectively. Fluoride is one of the important trace elements in groundwater, and the value ranges from 1.30 to 7 mg/l with an average of 4.06 mg/l in the study region. According to the piper plot, the water type is 80% Ca-HCO₃, 19% mixed CaMgCl and 1% CaCl type. Gibbs diagram indicates that the rock-water interaction is the most dominating mechanism prevailing in this region. The health risk assessment is revealed based upon the values of hazard quotient for ingestion (HQin) and dermal pathway (HQde) for the different age groups. The results from the research show that 6-12 months babies are more exposed to health risk through direct consumption of contaminated water in the study region.

Keywords Groundwater chemistry \cdot Fluoride contamination \cdot Gibbs diagram \cdot Chloroalkaline indices \cdot Health risk assessment

Puja Chowdhury puja.chowdhury91@gmail.com

¹ Department of Earth Sciences, Indian Institute of Engineering Science and Technology, Shibpur, Howrah 711103, West Bengal, India

1 Introduction

Groundwater occurs beneath the earth surface almost everywhere either connected to one single or many other local aquifer systems possessing similar characteristics (Vasanthavigar et al., 2010). Extraction of large quantity of groundwater has led to severe water crisis in different parts of the world (Asoka et al., 2017; Das, 2019; Rodell et al., 2009; Zolekar et al., 2021). Degradation of groundwater level has its negative impact on groundwater quality also (Bera et al., 2020, 2021; Biswas et al., 2020). Groundwater quality has been investigated by many researchers (Abdul-Wahab et al., 2020; Chakraborty et al., 2021; Chitsazan et al., 2019; Gaikwad et al., 2020; Islam et al., 2018; Jalees et al., 2020; Kalawapudi et al., 2019; Singaraja et al., 2016; Thivya et al., 2013). The contamination of groundwater is a major issue nowadays occurring both in rural and urban areas with large number of potential sources (Jayapraskh et al. 2008; Vasanthavigar et al., 2010). Fluoride ion occurs in all waters as a trace element mostly (Gaciri et al. 1993; Farooq 2018). It is an incompatible lithophile element and electronegative in nature (Faure, 1991). The drinking water which contains high concentrations of fluoride shows adverse effect on human health that has been reported from many different terrains of the world (Adimalla & Qian, 2019; Emenika et al., 2017; Narsimha & Sudarshan, 2018a, 2018b; Shaji et al., 2007). The impact of fluoride toxicity in human health depends mainly on its concentration in drinking water and the period for which the water has been exposed to the body surface in the different age groups independently (Dregne, 1967; Viswanthan et al., 2009). Frencken (1992) reported that the fluoride concentration ranges from 100 to 1000 mg/l of water in different rocks types—igneous, metamorphic and sedimentary rocks. International regions of Africa, Syria, Jordan, Sudan, Kenya, Uganda, Mexico, Turkey, China, Korea, India are highly been affected by fluorosis (Grech, 1966; Tekle-Haimanot et al., 1987; Gizaw, 1996; Carrilo-Rivera et al., 2002; Oruc, 2003; Guo et al., 2007; Chen et al., 2017; Adimalla et al. 2018b; Adimalla, 2019; Adimalla & Li, 2019).

From the view of geochemistry, it has been observed that certain minerals contain fluoride in their structures which are apatite $[Ca_5(PO_4)_3F]$, fluorite $[CaF_2]$, apophyllite, topaz [Al₂F₂(SiO₄)], sphene, cryolite [Na₃AlF₆], villiaumite [NaF], some silicates such as amphiboles like hornblende $[(Ca,Na)_{2,3}(Mg,Fe,Al)_{5}(Al,Si)_{8}O_{2,2}(OH,F)_{2}]$ and clay minerals like muscovite, biotite, kaolinite [Al₂Si₂O₅(OH)₄], vermiculite [(Mg,Fe,Al)₃(Al,Si)₄O₁₀(OH)], montmorillonite (Subba Adimalla & Qian, 2019; Chae et al., 2006; Rao & Devadas, 2003; Sivasankar et al., 2016). The occurrence of fluoride in groundwater is mostly due to geological sources with lower concentrations of anthropogenic sources either from agricultural fertilizers or industries sewages (Das et al., 2016; Kumar et al., 2016, 2020; Singh et al., 2020). Fluoride concentration depends mainly upon factors such as temperature, pH, cationic-anionic exchange with the aquifer materials, characteristics of the geological formations, solubility of minerals possessing fluoride in their structure and dissolution of those minerals through some geochemical processes (Adimalla & Qian, 2019; Raju et al., 2009). Naseen et al. (2010) reported that granite acts as a good source of fluoride in many Archean hard rock terrains, where during weathering processes granite leaches to form kaolin and other clay minerals which enriches the fluoride concentration in the groundwater.

In India during early 1930s, only 4 states were identified with higher fluoride concentration in the waters leading to endemic fluorosis, but now at present nearly 21 states are highly been affected (Adimalla et al., 2018a, 2018b, 2018c; Narsimha & Rajitha, 2018). An estimation of about 60 million people which includes children of about 6 million suffers from dental and skeletal fluorosis due to the consumption of fluoride rich water (Raju et al., 2009). The problem of fluorosis is increasing day by day throughout India within a very short span of time, and it is leading to an alarming health risk factor in the country (Adimalla & Venkatayogi, 2017; Adimalla et al., 2018a, 2018b, 2018c; Ayoob & Gupta, 2006). In West Bengal, 43 blocks in 7 districts are highly affected by higher fluoride concentration (>1.5 mg/l) (Mandal & Sanyal, 2019). According to WBPHED 2006, the rural population of the state are at high risk consisting nearly 7.41 million which is about 12% of the total rural populations. Chakrabarti and Ray (2013) reported that over 65 blocks from 8 districts are at high risk due to higher fluoride concentrations. Purulia is a drought-prone arid to semi-arid region of Archean granitic terrain which is affected by the endemic fluorosis. According to Chakraborti and Ray (2013), the fluoride value in the groundwater varied from 0.126 to 8.16 ppm over the blocks of Hura, Purulia 1 and Purulia 2. The exact number of people and the age groups mostly facing the risk of fluoride is unknown due to lack of systematic survey and understanding of the spatial distribution over Purulia (Jha et al., 2013).

Researches related to the fluoride contaminated water and the health risk on humans in Purulia have been done previously by some researchers. Mondal et al. (2013) studied the different water quality parameters and stated the fluoride contaminated zones in Purulia district. Samal et al. (2015) stated about the fluoride intake through the food chain in adults and infants and also the effect of elevated fluoride in the soil structure. Bhattacharya (2016) discussed about the concentration of fluoride in water and soil in the six different blocks (Hura, Puncha, Kashipur, Santuri, Raghunathpur-I and Manbazar-I) of Purulia. Farooq et al. (2018) discussed about the spatial distribution of fluoride concentrations in the groundwater of Purulia-I and Purulia-II blocks. Mandal and Sanyal (2019) with the help of GIS techniques studied the spatial distribution of fluoride concentrations in the waters of Purulia. Researches related to the use of the health risk model for fluoride contaminated water and its effects of on the different age groups of humans in detailed manner have not been performed earlier in Purulia.

The main objectives of this research are to understand the general groundwater quality based upon the different physical, chemical parameters and identification of the potential sources that leads to the higher proportions of different cations–anions in water. Secondly, identification of the fluoride vulnerable areas and assessment of the health risk factor for the different age groups- babies, young children, teenagers, adults and aged people both by consumption and dermal contact pathways. The results obtained from this study would further help the policy makers and the planners for preparing groundwater management programs and scientific techniques to improve the health of the local dwellers in the coming future.

2 Study area

The study area is situated in the eastern and the north-eastern bocks of Purulia district which has an area of about 1545 km². The area lies within the latitudes of 23°31'44'' N and 23°39'47'' N and the longitudes between 86°32'57''E and 86°51'50''E. Purulia is a hard rock terrain over which the tropic of cancer passes. It shows semi-arid tropical monsoonal climate. The area is a drought-prone region of West Bengal which is a semi-arid in nature (Archarya & Nag, 2013). The selected study area consists of Para, Kashipur, Raghunathpur-I, Raghunathpur II, Santuri and Neturia blocks of Purulia (Fig. 1). The climate in Purulia varies from hot, dry summer to cold winters with scanty rainfall. The average



Fig. 1 Location map of the study area

temperature of the area ranged from 7 °C during winter to 38 °C in summers (Fig. 2). The annual rainfall varies from 1100 to 1500 mm. The area mainly consists of the Pre-Cambrian highly metamorphosed rocks which act as the basement rocks and a small portion of Gondwana sedimentary rocks in the north-eastern parts of the district. The metamorphic rocks mainly include the granite gneiss, biotite granite gneiss, calc-granulite, ultrabasics, meta-basics, meta-sedimentary rocks, pegmatites and quartz veins (Baidya, 1992). Generally, the E-W trending strike of the formations is predominant with moderate to steep northerly and southerly dipping beds (Acharya et al., 2014). Aquifers are recharged during the monsoon through precipitation. According to CGWB, the modes of occurrence of groundwater in Purulia are: (i) weathered mantle and (ii) fractured zone of the hard rock. Near the lithological boundaries of biotite gneiss, phyllites, micaceous bodies and phyllites



Fig. 2 Temporal distribution of rainfall and temperature of the study area

exist higher permeability zones as small pockets (Saha, 1997). According to the total number of population for the 6 blocks of Purulia under study is 812196 and the population density is 525/km². The population density in the Para block is comparatively high (642/km²), and Santuri block shows comparatively low population density (437/km²). Purulia is one of the economically backward districts of West Bengal. Agriculture is the main occupation for the majority of the population. The rivers in this region are mainly rainfed. Therefore, during the monsoon seasons the water levels are quite high in the rivers, whereas during the dry seasons the rivers mostly dry out and the farmers have to dependent on the reservoirs and the groundwater irrigation for farming. Water scarcity is still a major threat in the land of Purulia during the summer seasons. Even after 73 years of independence, the people of Purulia still have to struggle for the basic amenities of life which are food and water.

3 Materials and methods

Groundwater level determines the water status in an area. The water level fluctuates with season as it depends mainly on the availability of rainfall. The investigation was carried out in the month of December 2019 for the post-monsoon period. Depth to water level was measured for 60 cylindrical fitted tube wells around different villages in the blocks of Para, Kashipur, Santuri, Neturia, Raghunathpur-I and Raghunathpur-II (Fig. 3). The physical parameters like the electrical conductivity (EC), pH, temperature, total dissolved solids (TDS) are measured for each sample using Hanna Multi-parameter waterproof meter (HI98194) in the field itself. The total alkalinity for each sample was measured in the field using AQUASOL Alkalinity kit (AE-214). The samples of water were collected from the 60 tube wells in 500 ml bottles for testing in the laboratory. The water was collected in fresh bottles after pumping it for 7–8 min, so that the water stored in the casing is lost and we get the fresh sample from the aquifer itself. The samples were analysed by standardized methods for different cationic and anionic concentrations which are Mg⁺², Na⁺, K^+ , Ca^{+2} , $HCO_3 + CO_3^-$, Cl^- , SO_4^{-2} , NO_3^- , $Fe^{+2/+3}$, F^- , pH, EC, TDS and temperature using different instruments. Sodium and potassium are measured by digital flame photometer (Systronics 128), chloride by argentometric method, calcium and magnesium by titrimetric method. Iron is estimated by using Move-100 with Spectroquant reagents. Fluoride (F^{-}) concentration in groundwater is detected using the ion-selective electrode method (APHA, 1995). The different cationic and anionic values and the physical parameter values are given in Table 1 for the study area. The analysed geochemical data are plotted in the Hill Piper diagram (Piper, 1944) which helps in understanding the groundwater facies. The analysed values of the above-mentioned parameters were compared with standardized values recommended by WHO (2011) guidelines. Gibbs diagram (Schoeller, 1967) for the analysed samples was plotted to determine the mechanism or the processes which controls the geochemistry of the water.

For validating the water quality, charge balance error (CBE) is calculated for the samples using the following equation

$$%CBE = \frac{\Sigma Cations - \Sigma Anions}{\Sigma Cations - \Sigma Anions} \times 100$$
(1)

Here, all the cationic and anionic values are expressed in meq/l. The CBE for all the collected samples is found to be within the acceptable limit, i.e. within $\pm 10\%$.



Fig. 3 Spatial distribution of depth to water level in the study area

3.1 Health risk assessment

Health risk assessment model is utilized for understanding the impact of higher or lower concentration of different ions on the human health (Adimalla & Qian, 2019; Adimalla et al., 2018a, 2018b, 2018c). The United States Environmental Agency (US EPA, 1989 and Moya et.al, 2011) has introduced the health risk model for evaluating the potential impact of harmful chemicals and its risk on different age groups of mankind through different pathways either directly or indirectly. The calculation of the hazard quotient (HQ) for the 60 different locations on the different age groups: 6-12 months, 1-5 years, 5-10 years, 10-15 years, 15-18 years, 18-21 years, 21-65 years, >65 years, was performed to get a detailed evaluation. In this model, the assessment is computed based on the results associated with the hazard quotient (HQ) for the different age groups both by direct consumption, i.e. ingestion and dermal contact pathway using Eqs. (2)–(5):

$$CDD_{in} = \frac{C_{fw} \times IR_W \times EF_r \times ED}{BW \times AT}$$
(2)

$$CDD_{de} = \frac{C_{fw} \times EF_r \times K \times ESA \times ED \times CF}{BW \times AT}$$
(3)

Table 1 Physico-	chemical	parameter	rs of groundv	water samples	in the study area (.	EC in μS/cn	a, remaining	g mg/l, exc	ept pH)					
Location	Hq	EC	TDS	ΗT	$HCO_3 + CO_3^-$	CI-	$SO4^{2-}$	NO3 ⁻	Ч	Ca ⁺²	Na^+	Mg^{+2}	\mathbf{K}^+	Fe^{+2}
Mekhyada	6.3	512	327.68	272.84	300	85.08	30	50	5	42	22.62	40.82	1.94	0.5
Mirgipahari	6.9	1000	640	431.62	300	170.16	150	75	5	119.2	46.15	32.56	3.26	0.4
Kashipur	6.6	705	451.2	245.81	400	99.26	70	0.5	4.5	65.6	54.01	19.93	2.69	7
Kapista	7	292	186.88	173.91	200	42.54	50	0.5	ю	30	24.7	24.06	1.11	1
Gaganabad	7.6	669	447.36	333.75	300	56.72	40	1	4	83.2	31.3	30.62	2.84	0.3
Kirtankhola	7.1	406	259.84	737.76	200	28.36	20	1.5	4.5	67.2	15.91	138.51	0.75	0.3
Indrabil	6.9	1200	768	458.63	200	198.52	120	70	4	122.8	75.93	36.94	2.64	0.3
Sihika	7.1	1177	753.28	506.56	400	148.89	120	1.5	3.2	140	75.73	38.15	3.24	2
Rangunigota	7.3	759	485.76	383.74	400	42.54	20	2.5	0	83.2	44.5	42.77	1.99	0.4
Kalapathar	7.2	658	421.12	236.78	200	92.17	90	1	5	68	52.84	16.28	2.2	1.5
Lara more	Г	1940	1241.6	877.32	400	340.32	130	50	5.2	226.4	76.9	75.82	3.33	0.3
Kustar	6.9	1401	896.64	681.52	300	205.61	100	60	5	158.8	40.23	69.26	7.89	0.5
Beldi	6.9	414	264.96	245.80	300	35.45	20	2.5	9	60.8	20.75	22.84	1.55	0.3
Nutundih	7.02	364	232.96	224.90	200	35.45	30	2.2	1.4	33.2	17.36	34.51	8.26	1.5
Ramkanali	6.3	1680	1075.2	423.65	200	177.25	130	50	5.7	108	67.06	37.42	6.24	0.4
Garh panchkot	6.6	557	356.48	307.71	200	35.45	20	1	2.5	88	9.3	21.38	3.08	7
Panchyet	7.07	550	352	309.74	200	28.36	40	2.5	1.5	88.4	17.64	21.63	1.9	1.5
Kisalajod	7.38	829	530.56	325.75	200	56.72	90	2.5	1.4	76.8	67.52	32.56	1.72	0.05
Neturia	7.4	1000	640	269.80	200	48.63	120	7.5	9	68.4	27.06	24.06	3.61	0.05
Madhukunda	6.86	779	625.28	443.61	300	163.07	180	1	5.5	128.4	103.68	29.89	3.03	9
Murulia	7.15	1473	942.72	362.68	300	70.9	600	6.5	5	106.8	30.1	23.33	1.18	0.4
Madhubanpur	6.83	1500	096	759.31	200	248.15	140	80	7	231.2	87.2	44.23	17.06	0.3
Murardi	6.85	1406	899.84	301.77	300	92.17	110	70	2.9	72.8	45.17	29.16	2.49	1
Jiyathol	6.8	490	313.6	140.93	100	92.17	80	2.5	1.5	18.8	26.06	22.84	0.63	0
Santuri	7.38	864	552.96	238.85	200	42.54	40	1	1.4	49.2	17.23	28.19	1.98	3
Saloni	6.95	789	504.96	682.35	400	361.59	140	50	3.5	219.2	107.39	32.81	7.3	0.3
Kesardih	7.3	595	380.8	144.93	200	35.45	40	40	б	28	15.74	18.23	2.27	0.1

(continued)
.
Ð
P
Тa

Location	Hq	EC	TDS	ΗT	$HCO_3 + CO_3^{-}$	CI-	$SO4^{2-}$	NO3-	Ŀ	Ca^{+2}	Na^+	Mg^{+2}	\mathbf{K}^+	Fe^{+2}
Sadhusaltora	7.2	944	604.16	228.83	180	56.72	60	50	6.8	50	23.14	25.27	1.35	0.4
Kalukar	7.3	241	154.24	114.93	200	28.36	30	50	4	26.4	18.67	11.91	1.6	0.7
Nodiha	7.1	704	450.56	354.70	200	85.08	50	09	1.5	98	32.35	26.73	3.25	-
Malanjdi	6.5	941	602.24	447.58	200	177.25	70	09	9	132	36.21	28.67	2.79	-
Aamchalor	٢	370	236.8	191.87	200	85.08	20	50	5	45.2	16.12	19.2	1.16	0.3
Lipannia	7.18	1491	954.24	339.76	200	106.35	110	50	5.4	80.4	50.13	33.78	2.17	1
Rongomatiya	6.9	457	292.48	215.83	200	163.07	30	50	4	51.6	34	21.14	1.83	0.5
Para	6.7	816	522.24	294.95	200	113.44	50	50	4.7	96.8	61.22	12.94	1.53	0
Hariharpur	6.8	371	237.44	187.83	100	56.72	40	09	4	54	25.87	12.88	1.08	1
Dubra	7.1	1636	1047.04	1273.76	300	432.49	120	75	5	406	82.36	63.18	10.37	0
Paharigora	7.1	750	480	374.74	300	<i>96.11</i>	70	50	4	88	45.82	37.67	3.14	0.5
Shyampur	6.4	478	305.92	222.83	100	<i>77.99</i>	80	2.5	2.5	53.2	39.97	21.87	1.12	1.5
Santaldih	7.3	861	551.04	343.73	200	92.17	80	7.5	4	87.2	69.17	30.62	9.72	0.4
Kamargora	6.9	1577	1009.28	764.23	400	212.7	160	75	4.5	251.2	49.8	33.29	3.49	-
Cheliyama	Ζ	1247	798.08	503.60	300	148.89	130	65	9	131.2	77.36	42.77	39.74	0.3
Dandua	7.1	486	311.04	205.83	100	49.63	80	50	3.8	47.2	36.57	21.38	1.7	0.3
Gobranda	6.7	538	344.32	164.91	100	113.44	80	40	9	28.8	17.73	22.6	3.55	0.7
Raghunathpur	7.09	153	97.92	256.79	20	70.9	110	40	٢	68.4	37.7	20.9	1.3	1.5
Bundla	7.04	1023	654.72	278.74	200	56.72	40	50	4.8	82	15.61	17.98	3.9	0.3
Durmut	6.34	342	218.88	195.88	200	42.54	40	10	1.4	35.6	13	26	2.39	1.5
Naragoria	6.95	179	114.56	85.93	100	21.27	30	2	3.5	16.8	14.9	10.69	1.03	0.5
Achkoda	6.7	216	138.24	131.91	100	28.36	20	30	5.5	24.4	15.57	17.25	1.48	0.4
Sidpur	6.8	475	304	210.84	200	49.63	50	20	4.8	52.8	31.25	19.2	2.73	1
Bhurkundabari	7.13	623	398.72	281.80	400	28.36	20	5	5.8	66.4	50.77	28.19	4.09	0.3
Dhakibana	6.3	238	152.32	87.96	200	14.18	40	50	4	12.8	8.02	13.61	1.51	1

(continued)
5
ľ
Ta

 $\underline{\textcircled{O}}$ Springer

lable I (continut	(DS													
Location	Ηd	EC	TDS	HT	$HCO_3 + CO_3^{-}$	CI-	$SO4^{2-}$	NO3 ⁻	! Ľ	Ca ⁺²	Na^+	Mg^{+2}	\mathbf{K}^+	Fe^{+2}
Vet bhiltora	6.47	1406	899.84	806.39	300	283.6	110	50	6.7	191.6	60.34	79.7	1.56	1
Adra	7.04	1351	864.64	541.51	300	163.07	160	50	2.8	153.6	52.53	38.39	1.28	0.3
Joychandi	7.2	1000	640	256.84	200	49.63	40	70	1.4	45.2	31.62	34.99	1.48	0.5
Sakhabartal	7.5	604	386.56	190.88	100	28.36	50	0.5	4.5	43.6	14.36	19.93	0.68	2.5
Mosairdih	6.94	1306	835.84	295.74	200	70.9	50	30	1.3	85.2	46.67	20.17	1.39	1
Jiyara	7.15	874	559.36	215.82	200	35.45	40	5	1.4	61.6	27.05	15.07	1.09	1
Jhapra	6.95	1663	1064.32	806.40	400	311.96	150	0.5	4	198	86.09	75.82	2.27	0.7
Gopinathpur	7.3	945	604.8	160.88	200	42.54	40	0.5	2.5	43.6	29.36	12.64	1.02	2.5

$$HQ_{in} = \frac{CDD_{in}}{RfD}$$
(4)

$$HQ_{de} = \frac{CDD_{de}}{RfD}$$
(5)

where CDD_{in} —chronic dose via ingestion pathway in daily basis (mg/kg/day); CDD_{de} —chronic daily dose via dermal exposure pathway (mg/kg/day); C_{fw} —concentration of fluoride in drinking water(mg/l); EF_r —frequency of exposure (days/years); ED—exposure duration (years); BW—body weight (kg); AT—resident time (days/years); ESA—exposed skin area (cm²); *K*—skin adherence factor; CF—conversion factor (l/cm³); and RfD—reference dose of fluoride (0.06 mg/kg/day) taken from the International Risk Information System (IRIS) and (US EPA 1989). The detailed value of the parameters used for the calculation is given in Table 2 (US EPA 1989; Chen et al., 2017; Adimalla Li & Qian 2018; Emenike et al., 2018; Adimalla & Qian, 2019).

4 Results and discussion

4.1 Quality based on the general parameters

The basic physio-chemical parameters for each location are given in Table 1 which are analysed individually and systematically for understanding the general water quality. The analysed data are compared with the standard ranges of WHO (2011) to assess the water quality and its suitability for drinking purpose.

pH acts as an indicator of water strength. It also determines the capacity of water in reacting with the acidic and alkaline materials present within. The pH value ranges from 6.3 to 7.5 with an average value of 6.96. Most of the water samples are neutral to weakly alkaline in nature. According to WHO (2011), the acceptable range of pH varied between 6.5 and 7.5. All the samples show that the values are within the acceptable limits for drinking purpose.

TDS ranges from 97.92 to 1241.60 mg/l with a mean of 539.12 mg/l in the study area. According WHO (2011), the highest permissible limit is 500 mg/l, but 1500 mg/l is accepted as the maximum extended desirable limits for drinking purpose. 51.67% of the samples are within 500 mg/l, and 48.33% are within the maximum desirable limits. Todd (1980) classified the groundwater as freshwater when the TDS value is less than 1000 mg/l; brackish water when TDS value ranged from 1000 to 10,000 mg/l; saline water when TDS ranged from 10,000 mg/l to 1,000,000 mg/l and lastly as brine water when TDS ranged greater than 1,000,000 mg/l. In the study area, 80% of the samples are fresh and 20% are brackish in nature.

Electrical conductivity (EC) is the electric current conveyed by the water. EC values ranged from 153 to 1940 μ s/cm with an average value of 834.38 μ s/cm. EC can be classed into three types—type I (low) when the EC is less than 1500 μ s/cm; type II (moderate) where EC ranges from 1500 to 3000 μ s/cm; and lastly as type III (high) where EC value is greater than 3000 μ s/cm (Sarath Logeshkumaran et al., 2015; Prasanth et al., 2012). All the samples are within the WHO desirable limits of 1500 μ s/cm. Higher EC values in the groundwater are an indication of rock–water interaction mechanism dominance.

arameters	Units	6-12 months	1-5 years	5-10 years	10-15 years	15-18 years	18-21 years	21-65 years	>65 years
ngestion pathway									
ngestion rate (IRw)	L/day	1	1.16	1.32	1.82	1.78	2.34	2.94	2.73
Exposure frequency (Efr)	Days/year	365	365	365	365	365	365	365	365
Exposure duration (ED)	Year	9	9	9	9	9	30	30	30
3 ody weight (BW)	kg	9.1	16.6	29.3	54.2	67.6	67.6	78.8	80
Average time (Ati)	Days	2190	2190	2190	2190	2190	10,950	10,950	10,950
Concentration of element (Cfw)	mg /l	Present study							
Dermal pathway									
Skin surface area (SA)	cm^2	4500	6800	10,500	15,700	18,000	19,550	19,800	19,400
Exposure time (ET)	h/event	0.54	0.54	0.54	0.54	0.54	0.71	0.71	0.71
Exposure frequency (Efr)	Days/year	350	350	350	350	350	350	350	350
Exposure duration	Year	9	9	9	9	9	30	30	30
Conversion factor (CF)	L/cm^3	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001
3 ody weight (BW)	kg	9.1	16.6	29.3	54.2	67.6	67.6	78.8	80
Average time (Ati)	Days	2190	2190	2190	2190	2190	10,950	10,950	10,950
Skin adherence factor (Kp)	cm/h	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001
Concentration of element (Cfw)	mg /l	Present study							

 Table 2
 Different values of the parameters used for the health risk assessment

Total hardness (TH) ranged from 87.96 to 1273.76 mg/l with a mean value of 359.27 mg/l. The most desirable limit is 100 mg/l, and the maximum allowable limit is 500 mg/l as given by WHO (2011). 80% of the samples are within the maximum allowable limits, while rest 20% show very high hardness value which is within 1500 mg/l. The location Dubra shows the highest value of total hardness where the Ca and Mg concentrations are very high. The higher concentration of total hardness can cause scaling in the pots and boilers used in different industries; clogging in the irrigation pipes and sometimes causes health problems like kidney failure (Vasanthavigar et al., 2010). The total hardness is determined using the following equation (Todd, 1980);

$$TH(CaCO_3)mg/l = (2.497)Ca + (4.115)Mg$$
(6)

According to the mean value of the cationic and anionic concentrations in the collected samples from the study area are arranged in the following decreasing order of occurrences- for cations $Ca^{+2}>Na^+>Mg^{+2}>K^+>Fe^{+2}$ and anions as $HCO_3^->Cl^->SO_4^->NO_3^->F^-$.

The Ca⁺² value ranges from 16.8 to 251.2 mg/l with a mean value of 91.53 mg/l in the study area. 10% of the samples show very high concentrations of Ca⁺² exceeding the limit of 200 mg/l given by WHO (2011). The rest remaining 90% samples are within maximum allowable limits. According to WHO, the most desirable limit for calcium is within 75 mg/l and 58% of the samples are within the desirable ranges; rest falls within the maximum allowable limits. The higher concentration of Ca⁺² can be predicted due to the lithology which is mostly granitic gneiss and amphibolite schist.

The Na⁺ concentration ranges from 9.3 to 107.39 mg/l with an average of 42.3 mg/l which are within the maximum allowable limit of 200 mg/l for the drinking purposes as given by WHO (2011). The chief sources of sodium are the Archean gneisses and the weathering of the crystalline rocks.

The value of Mg^{+2} ranged from 10.62 mg/l to 138.51 mg/l with a mean of 31.76 mg/l. The maximum allowable limit is 150 mg/l, and all the samples show values which are within the limit. The major Mg^{+2} sources in the water are the magnesium-bearing minerals in the country rock, domestic and industrial wastes (Adimalla & Qian, 2019; Adimalla & Venkatayogi, 2017; Marghade et al., 2011). Ca⁺² and Mg⁺² are very much essential for human body, but when taken in excess, it can cause adverse effect on human health (Adimalla & Qian, 2019).

Potassium (K⁺) ranges from 0.75 to 39.75 mg/l with an average of 3.58 mg/l in the study area. The location Cheliyama shows the highest value of 39.75 mg/l. 98% of the samples are within the desirable limits of 12 mg/l. According to He and Macgregor (2008), the K⁺ ion in the groundwater is present as a very necessary trace element to maintain stability in the human body.

Iron is present generally in the groundwater as a trace element. According to WHO (2011), the acceptable limit is 0.3 mg/l, but in the study area most of the samples show higher concentrations of iron. The value ranges from 0.2 to 2.5 mg/l of a mean value of 0.93 mg/l. Higher concentrations of iron lead to reddishness in the water colour and kidney problems and deteriorate the quality of utensils leaving a stain behind. The probable sources of iron are the iron–magnesium-bearing amphiboles, pyroxene gneisses and biotite schist. Rock–water interaction and weathering of these iron-rich minerals and rocks lead to higher concentration iron.

Bicarbonate (HCO_3^-) value ranges from 100 to 500 mg/l with a mean value of 231.67 mg/l. No direct negative effect on human health has been reported yet due to HCO_3^- .

Cl⁻ concentration ranged from 14.18 to 432.49 mg/l with an average value of 106.81 mg/l. Excess of Cl⁻ causes deleterious effect on human body and also causes pollution (Adimalla & Venkatayogi, 2017; Adimalla et al., 2018a, 2018b, 2018c; Marghade et al., 2011). Here, it is the second most dominating anion, where 90% of the samples are within the desirable limit of 200 mg/l, but most of the samples are within the maximum desirable limit of 600 mg/l. Tiwari et al., 2014; Adimalla and Li (2019) stated that the higher chloride (Cl⁻) concentration is mostly due to the effluents from domestic wastes, wastes from leaked septic tanks and breakdown of the chloride bearing minerals.

Sulphate (SO₄) in the study region ranges from 20 to 600 mg/l, with an average of 82.83 mg/l. The maximum allowable limit is 250 mg/l (WHO). The highest value is observed in Muradi which is 600 mg/l.

The NO_3^- concentration in the samples ranged from 0.5 to 80 mg/l with a mean value of 31.52 mg/l. According to WHO, the desirable limit is 45 mg/l and only 40% of the samples exceed the desirable limit. In several studies, it has been observed that there are some correlations between nitrate concentration in the water and practice of agricultural pattern (Adimalla & Li, 2019; Debernardi et al., 2008). Zang et al. (2018) stated that NO_3^- pollution in groundwater has been spread all over the world. Diseases like methemoglobinemia, gastric cancer, birth malfunction, hypertension are mostly caused due to higher concentrations of nitrate (Bao et al., 2017; Fan, 2011; Spalding & Exner, 1993). The source of nitrate has been identified by many workers to be the chemicals used as fertilizers, leakage of the septic tanks and sewage systems (Datta & Tayagi 1996; Subba Rao, 2003). Mostly during post-monsoon period, agricultural activities are favourably done in the study region; therefore, use of nitrogenous fertilizers is high which causes higher concentration of NO_3^- at some places.

4.2 Fluoride contamination in groundwater and its spatial distribution

Fluoride is an incompatible lithophile, electronegative element which forms complexes with polyvalent metal ions present in the water (Faure, 1991). Fluoride is one of the important elements found in the groundwater as trace. The concentration of fluoride in the study area ranged from 1.30 to 7 mg/l with an average value of 4.06 mg/l/. The concentration is mostly high in almost all locations of the study area which should be within 1.5 mg/l for drinking water stated by WHO. Edmunds and Smedley (2013) stated the best range suited for fluoride in the drinking water is 1 mg/l. Fluoride concentration in water can be classified into the following classes: <1 mg/l as low fluoride; 1–1.5 mg/l acceptable limit of fluoride; and > 1.5 mg/as high fluoride. It is observed that 16.67% of the collected samples are within the acceptable limits, whereas the rest 83% shows higher values of fluoride up to a range of 7 mg/l.

Jiang et al. (2014) stated that the geochemistry of water possessing higher elevated concentrations of fluoride is mostly associated with weakly alkaline pH, moderate TDS and higher concentrations of HCO₃ and Na⁺. Alkaline condition promotes dissolution of minerals that contains fluoride in their structure like CaF₂ in groundwater. Chen et al. (2016) stated the following equations in an alkaline to weakly alkaline groundwater, and increased OH⁻ concentration causes the reaction (7), then to the corresponding dissolution of CaF₂ in reaction (8)

$$Ca(OH)_2 \leftrightarrow Ca^{+2} + 2OH^-$$
 (7)

$$\operatorname{CaF}_2 \leftrightarrow \operatorname{Ca}^{+2} + 2\mathrm{F}^-$$
 (8)



Fig. 4 a-d The graphs depicting the correlation between fluoride and the other physico-chemical parameters like pH, sodium, bicarbonate and nitrate

Figure 4a–d depicts the correlation between the different physio-chemical parameters in the study area. Fluoride shows a positive correlation with HCO_3^- ($r^2=0.0045$); Na⁺ ($r^2=0.052$) and NO₃⁻ ($r^2=0.112$). The sources of fluoride are considered to be muscovite, biotite and granite gneisses which are present through the area. Researchers have predicted that the probable reason for elevated fluoride concentrations is mainly due to the dissolution reaction and the exchange reactions (Jack et al., 2000; Liu et al. 2015).

The spatial distribution of fluoride concentration in the study area is prepared using the IDW Interpolation technique, with the help of spatial analysis tool of ArcGIS 10.0, and Jenks's Natural Break method has been used to classify them. Using Jenk's data segmentation method, best arrangement value of different classes can be determined (Zhang et al., 2012). This method reduces intra-class variation and increases inter-class variation (Jenks, 1967).

The area is distributed into 6 zones—less than 1.5; 1.5-2; 2-3; 3-4; 4-5; and 5-7 mg/l (Fig. 5). The ranges vary from 1 to 7 mg/l, and most of the villages show ranges > 3 mg/l. Only in few villages 10 stations show concentration within the safe permissible limit. This distribution is made to study the region more minutely and accurately.

4.3 Piper plot

Piper plot helps in understanding the chemistry of groundwater and its evolution processes. It determines the dominant water type (Piper, 1944). The diagram consists of one diamond-shaped field at the centre and two triangular fields at the two sides. The cations and anions are expressed as percentage of the total cations and anions present in mg/l. The left triangle



Fig. 5 Spatial distribution map of fluoride concentrations in the study area

represents the cationic plot, whereas the right triangle represents the anionic plot. The anions are Cl⁻, SO₄⁻² and HCO₃⁻⁺+CO₃⁻², and the cations are Ca⁺², Mg⁺² and Na⁺ + K⁺ which are projected into the central diamond-shaped field that determines the water type or the water facies. From Fig. 6, it is observed that 95% of the sample shows alkaline earth metals (Ca⁺²+Mg⁺²) exceed the alkali cations (Na⁺ + K⁺) in zone 3 and 85% of the samples are the weak acids (HCO₃⁻⁺+CO₃⁻²) represented as zone 2 over the strong acids (Cl⁻+SO₄⁻²). Therefore, based upon the cation and anion distribution the dominant water types are 80% Ca-HCO₃; 19% mixed CaMgCl; and 1% CaCl type. Ca⁺² and Mg⁺² are dominant as the result of weathering of silicate rocks like granite gneiss (Adimalla et al., 2018a, 2018b, 2018c). These samples also show high EC with high TDS which determines the major role of rock–water interaction along with longer residential time (Dehbandi et al., 2018; Edmunds & Smedley, 2013).

4.4 Gibbs plot demonstrating the chemical weathering

Gibbs diagram (Gibbs, 1970) helps in understanding the relationship possesses between the aquifer lithology and the water composition. It consists of three different distinct fields that are evaporation dominance, rock-water or rock dominance and the precipitation dominance. For cations, TDS and $(Na^+ + K^+)/(Na^+ + K^+ + Ca)$ are plotted, whereas for anions TDS and $(Cl^-/Cl^- + HCO_3)$ are plotted to determine the



The characterisation of the different zones in piper plot-

- 1- Strong acids exceed weak acids
- 2- Weak acids exceed strong acids
- 3- Alkaline earths exceed alkalies
- 4- Alkalies exceed alkaline earths
- 5- Ca-HCO₃ type water
- 6- Na-Cl type water
- 7- Mixed Ca-Mg-Cl type of water

Fig. 6 Piper plot for the collected water samples

prevailing mechanisms leading to the variable concentrations of the ion in the groundwater and its impact on the surrounding country rock. From Fig. 7, it is observed that the samples fall mostly in the rock-water interaction dominance field. Hard rock regions possessing low rainfall and high temperature mostly show rock-water interaction dominance (Adimalla & Li 2018; Subba Rao et al., 2017; Adimalla et al., 2018a, 2018b, 2018c). Therefore, it can be inferred that higher rate of residence time or slow water percolation intensely enhances the ionic concentration in the groundwater (Chen et al., 2017; Subba Rao et al., 2017; Adimalla 2018).

Furthermore, a plot has been made for better understanding of the weathering pattern. Ca + Mg and $HCO_3 + SO_4$ values have been plotted (Fig. 8) which indicates that 88% of the samples fall in the silicate weathering region and the rest in carbonate weathering region. The process that is dominant in this region is the rock–water interaction along with silicate weathering (Fig. 9).

8. Mixed Ca-Na-HCO₃ type of water 9. Na-HCO₃ type of water 10. Ca-Cl type of water



Fig. 7 Gibbs plot for the collected water samples

4.5 Ion exchange by hydrogeochemical analysis

Schoeller (1967) proposed a mathematical relation for better understanding of the ionic exchange taking place between the groundwater and the aquifer host rock material is called the chloro-alkaline indices (CAI). This method provides a quantitative analysis about the direct and reverse ion exchange processes which tells us about the groundwater chemistry. CAI (I) and CAI (II) are calculated using the following equations which are expressed in meg/l (Adimalla et al., 2018a, 2018b, 2018c).

$$CAI(I) = [CI^{-} - (Na^{+} + K^{+})/CI^{-}]$$
(9)

$$CAI(II) = \left[CI^{-} - \left(Na^{+} + K^{+} \right) / \left(SO_{4}^{-2} + HCO_{3}^{-} + CO_{3}^{-2} + NO_{3}^{-} \right) \right]$$
(10)

 $\underline{\textcircled{O}}$ Springer



Fig. 8 The scatter plot between $(HCO_3 + SO_4)$ and (Ca + Mg)



Fig. 9 The CAI (I) and CAI (II) values for the different locations

When CAI (I) and CAI (II) indices are positive, then the ionic exchange is between $(Na^+ + K^+)$ of the groundwater and $(Ca^{+2} + Mg^{+2})$ of the aquifer material or the host rock, and then, it is called reverse exchange method. On the other hand, in a direct ion exchange

method both the CAI (I) and CAI (II) indices are negative when $(Na^+ + K^+)$ cations of the aquifer material/host rock exchange with $(Ca^{+2} + Mg^{+2})$ of the groundwater (Adimalla et al., 2018a, 2018b, 2018c; Schoeller, 1967).

The CAI (I) and CAI (II) indices for the study area range from (-0.934 to 0.796) and (-0.062 to 0.686), respectively. The values are plotted in bar graphs in Fig. 7. The plots reveal that 5% of the sample shows reverse ionic exchange, whereas 95% of the samples show direct ionic exchange process. Therefore, the dominant ionic exchange that governs the chemistry of the groundwater is reverse exchange or chloro-alkaline equilibrium.

4.6 Fluoride health risk factor assessment

The health risk assessment is done based upon the model given in USEPA (Moya et al., 2011; USEPA, 1989). This model is used to calculate the non-carcinogenic (Adimalla & Li, 2019; Chen et al., 2017) effects of fluoride for the different age groups which include the babies, children, teenagers, adolescence, grown-up adults and the aged people, i.e. 6-12 months; 1-5 years; 5-10 years; 10-15 years; 15-18 years; 18-21 years; 21-65 years; and > 65 years for more detailed investigations. Fluoride is very much essential for the protection of teeth from demineralization which are caused by the acids produced by the different bacteria and sugar present in the mouth. Children require fluoride to protect the newly formed teeth, and adults need for lowering the decay of teeth. If fluoride is taken within 1.5 mg/l, it helps in demineralization and enamel strengthening (Ahada & Suthar, 2019). Many hard rock regions of India are affected by the excess of fluoride concentrations in drinking water which causes some of the epidemic diseases like dental and skeletal fluorosis (Brindha et al., 2011; Adimalla & Qian, 2019). Apart from dental and skeletal fluorosis, the intake of excess of fluoride causes other acute disorders such as respiratory failure, general paralysis, lowering of blood pressure, weight loss, anaemia and adverse effects on reproduction (Dissanayake 1991; Freni 1994; Chen et al., 2017; Adimalla & Qian, 2019). Guo et al. (2007) discussed on the effects of fluoride in the male fertility level and also stated that children consuming > 2 mg/l of fluoride showed poor intelligence test than the ones consuming within the standard limit of fluoride. Similarly, many towns and municipalities of Purulia with high population are highly been effected by the elevated concentrations of fluoride in groundwater (Mandal & Sanyal, 2019). This study considers the concentration of fluoride consumed on daily basis through drinking and dermal contact as bathwater source or other domestic purpose. The health risk is computed for the different age groups based on the HQ values through ingestion and dermal pathway. The health risk is computed for the different age groups based on the HQ values through ingestion and dermal pathway. The calculated HQ values for each age group are plotted using simple Kriging interpolation method, and the indexes are prepared ranging between the highest and lowest HQ values, to understand the distribution of health risk in the study area (Figs. 10 and 11).

4.7 Ingestion pathway

USEPA Exposure Factor Handbook (Moya et.al, 2011; USEPA, 1989) provides the values of the different key parameters given in Table 2, which are used to compute the health risk with analysed fluoride concentrations for the different villages in Purulia. The estimated HQ_{in} values are calculated using the formulas (2–5) as described under material and methods for the different age groups—6–12 months; 1–5 years; 5–10 years; 10–15 years;



Fig. 10 a-h HQ (in) values for the different age groups through ingestion pathway in the study area

15–18 years; 18–21 years; 21–65 years; and > 65 years, which are given in Table S1, see supplementary file, and shown in Fig. 10a–h. The HQ_{in} value ranges from 2.38 to 12.82 with an average of 7.4 for 6–12 months babies/infants. The HQ_{in} value for the age group 1–5 years young children varies from 7.02 to 1.30 with an average value of 4.1. The HQ_{in} value for the age group 5–10 years children varies from 0.97 to 5.25 with a mean value of 3.0. The HQ_{in} value ranges from 0.72 to 3.91 with an average of 2.3 for the age group 10–15 years teenagers. The HQ_{in} value for the age group 15–18 years young people varies from 0.57 to 3.07 with a mean value of 1.7. The HQ_{in} value ranges from 0.75 to 4.03



Fig. 10 (continued)

with an average of 2.3 for the age group 18–21 years young adults. The HQ_{in} value for the age group 21–65 years adults ranges from 0.80 to 4.35 with a mean value of 2.5. The HQ_{in} value ranges from 0.73 to 3.98 with an average of 2.3 for the age group greater than 65 years aged people.

The acceptable limit or the safe limit for HQ_{in} is below 1, but most of the locations show much higher value some up to 12.8. Based upon the mean HQ_{in} values of the non-carcinogenic fluoride health risk on the different age groups can be arranged in the following decreasing order: 6–12 months > 1–5 years > 5–10 years > 21–65 years > 18–21 years > abo



Fig. 11 a-h HQ(de) values for the different age groups through ingestion pathway in the study area



Fig. 11 (continued)

ve 65 years > 10–15 years > 15–18 years. From the observation, it can be inferred that the age groups 6–12 months babies and the 1–5 years children are most likely to suffer from the health complications and diseases due to direct consumption of groundwater with high concentration of fluoride in the long run. The health risk factor of 1–5 years is 2.4, 3.2, 4.2, 3.2, 2.9 and 3.2 times greater than the age groups 5–11 years; 11–16 years; 16–21 years; 21–65 years; and > 65 years, respectively, due to consumption of the elevated fluoride contaminated water. Nearly 85% of health problems arises due to consumption of uncleaned and contaminated water; therefore, ingestion pathway would be responsible for 70% of global fluorosis (Dissanayake 1991; Freni 1994). The age group 6–12 months babies is at high health risk than the other age groups in the study area, so proper planning is necessary for water management. Many workers from India (Adimalla & Qian, 2019; Singh et al., 2018) and globally (Chen et al., 2017) reported similar results in which children are more prone to health risk. The probable reason could be due to the lower body weight of the children and constant exposure to fluoride contaminated water through ingestion. So, proper planning is necessary for water management for the proper and healthy future.

4.8 Dermal pathway

The estimated HQ_{de} values are calculated using the formulas (2–5) as described under material and methods for the different age groups: 6–12 months; 1–5 years; 5–10 years; 10–15 years; 15–18 years; 18–21 years; 21–65 years; and>65 years, which are given in Table S2, see supplementary file, and the HQ(de) values are shown in Fig. 11a–h. The HQ_{de} value ranges from 0.005 to 0.029 with an average value of 0.017 for 6–12 months babies. The HQ_{de} value varies from 0.004 to 0.024; 0.004 to 0.021; 0.003 to 0.017; 0.002 to 0.016; 0.004 to 0.022; 0.003 to 0.019; and 0.003 to 0.019 with the mean values of 0.014; 0.012; 0.010; 0.009; 0.013; 0.012; and 0.011, respectively, for the age groups: 1–5 years; 5–10 years; 10–15 years; 15–18 years; 18–21 years; 21–65 years; and>65 years, respectively. All the HQ_{de} values for the age groups are within the safe limit which is within 1. The dermal contact pathway effect is quite lower than the direct consumption ingestion pathway. Therefore, it can be concluded that the principal pathway for non-carcinogenic fluoride in the human body is through direct ingestion or consumption of the groundwater with elevated fluoride concentrations. The calculated HQ_{in} and HQ_{de} values can act as baseline information for the different authorities for making the policies related to water sources and its management.

5 Conclusion

Water is the basic need in life, so the hydro-chemical characterization of groundwater is very much essential for understanding its suitability for drinking and other purposes. Groundwater samples were collected from the eastern portion of Purulia district for estimating the health risk in the different age groups due to high fluoride concentrations. It is observed that the general trend of the ionic concentrations based on their mean values is arranged as $Ca^{+2} > Na^{+} > Mg^{+2} > K^{+} > Fe^{+2}$ for cations and $HCO_3^- > Cl^- > SO_4^- > NO_3^- > F$ -for anions. The water is weakly alkaline in nature with moderate TH and TDS value in the overall area. The concentration of iron and fluoride varies from 0.2 to 2.5 ppm and 1.30 to 7 mg/l which deteriorates the drinking water quality. The mechanism which mostly controls the groundwater geochemistry is the rock-water interaction and the weathering of the silicate minerals. 95% of the samples show ionic exchange through direct process and the rest 5% by reverse ionic exchange. The health risk assessment for non-carcinogenic fluoride contaminated water reveals that the pathway through which people are been affected is direct consumption or ingestion pathway. It is observed that the HQ_{in} values have exceeded the safe limit (less than 1) for all the age groups. 6-12 months babies are more prone to health complications related to elevated concentrations fluoride. Therefore, proper and necessary steps must be taken to reduce the health risk and reduce the contamination effect on human health. However, this study would help the governmental and the non-governmental associations for making rules and regulations for groundwater managements on the scientific background. This would in the long run improve the health of the people in this region and would provide safe drinking water.

Supplementary Information The online version contains supplementary material available at https://doi.org/10.1007/s10668-021-01911-1.

Acknowledgements The authors are grateful to the State Water Investigation Directorate (SWID), Kolkata, for helping in analysis of the samples and Agro-irrigation officers, Belguma, for providing information's related to Purulia district. Special thanks to Mr. Amit Bera for his unconditional help throughout the manuscript proceedings. The authors also extend their thanks to anonymous reviewers for the valuable constructive comments and suggestion.

References

- Abdul-Wahab, D., Adomako, D., Abass, G., Adotey, D. K., Anornu, G., & Ganyaglo, S. (2020). Hydrogeochemical and isotopic assessment for characterizing groundwater quality and recharge processes in the Lower Anayari catchment of the Upper East Region, Ghana. *Environment, Development and Sustainability*, 23(4), 5297–5315.
- Acharya, T., & Nag, S. K. (2013). Study of groundwater prospects of the crystalline rocks in Purulia District, West Bengal, India using remote sensing data. *Earth Resources*, 1(2), 54–59.
- Acharya, T., Prasad, R., & Chakrabarti, S. (2014). Evaluation of regional fracture properties for groundwater development using hydrolithostructural domain approach in variably fractured hard rocks of Purulia district, West Bengal India. *Journal of Earth System Science*, 123(3), 517–529.

- Adimalla, N. (2019). Groundwater quality for drinking and irrigation purposes and potential health risks assessment: a case study from semi-arid region of South India. *Exposure and Health*, 11(2), 109–123.
- Adimalla, N., & Li, P. (2019). Occurrence, health risks, and geochemical mechanisms of fluoride and nitrate in groundwater of the rock-dominant semi-arid region, Telangana State, India. *Human and Ecological Risk Assessment: An International Journal*, 25(1–2), 81–103.
- Adimalla, N., Li, P., & Qian, H. (2018). Evaluation of groundwater contamination for fluoride and nitrate in semi-arid region of Nirmal Province, South India: a special emphasis on human health risk assessment (HHRA). *Human and Ecological Risk Assessment: An International Journal.*
- Adimalla, N., Li, P., & Venkatayogi, S. (2018b). Hydrogeochemical evaluation of groundwater quality for drinking and irrigation purposes and integrated interpretation with water quality index studies. *Environmental Processes*, 5(2), 363–383.
- Adimalla, N., & Qian, H. (2019). Hydrogeochemistry and fluoride contamination in the hard rock terrain of central Telangana, India: analyses of its spatial distribution and health risk. SN Applied Sciences, 1(3), 202.
- Adimalla, N., & Venkatayogi, S. (2017). Mechanism of fluoride enrichment in groundwater of hard rock aquifers in Medak, Telangana State South India. *Environmental Earth Sciences*, 76(1), 45.
- Adimalla, N., Vasa, S. K., & Li, P. (2018c). Evaluation of groundwater quality, Peddavagu in Central Telangana (PCT), South India: an insight of controlling factors of fluoride enrichment. *Modeling Earth Systems and Environment*, 4(2), 841–852.
- Apha, A. (1995). WEF, 1998. Standard methods for the examination of water and wastewater, 20.
- Ahada, C. P., & Suthar, S. (2019). Assessment of human health risk associated with high groundwater fluoride intake in southern districts of Punjab India. *Exposure and Health*, 11(4), 267–275.
- Asoka, A., Gleeson, T., Wada, Y., & Mishra, V. (2017). Relative contribution of monsoon precipitation and pumping to changes in groundwater storage in India. *Nature Geoscience.*, 10(2), 109–117.
- Ayoob, S., & Gupta, A. K. (2006). Fluoride in drinking water: a review on the status and stress effects. Critical Reviews in Environmental Science and Technology, 36(6), 433–487.
- Baidya, T. K. (1992). Apatite-magnetite deposit in the Chhotanagpur gneissic complex, Panrkidh area, Purulia district, West Bengal.
- Bao, Z., Hu, Q., Qi, W., Tang, Y., Wang, W., Wan, P., & Yang, X. J. (2017). Nitrate reduction in water by aluminum alloys particles. *Journal of Environmental Management*, 196, 666–673.
- Bera, A., Mukhopadhyay, B. P., & Barua, S. (2020). Delineation of groundwater potential zones in Karha river basin, Maharashtra, India using AHP and geospatial techniques. *Arabian Journal of Geosciences*. https://doi.org/10.1007/s12517-020-05702-2.
- Bera A, Mukhopadhyay BP, & Biswas S (2021). Aquifer vulnerability assessment of Chaka river basin, Purulia, India using GIS-based DRASTIC model. In Geostatistics and Geospatial Technologies for Groundwater Resources in India. Springer International Publishing. pp. 1–21. https://doi.org/10.1007/ 978-3-030-62397-5.
- Bhattacharya, P. (2016). Analysis of fluoride distribution and community health risk in Purulia district of West Bengal, India. In *Proceedings of the 9th National Level Science Symposium organized by the Christ college* (Vol. 3, pp. 88–92).
- Biswas, S., Mukhopadhyay, B. P., & Bera, A. (2020). Delineating groundwater potential zones of agriculture dominated landscapes using GIS based AHP techniques: A case study from Uttar Dinajpur district West Bengal. *Environmental Earth Sciences*, 79, 302.
- Brindha, K., Rajesh, R., Murugan, R., & Elango, L. (2011). Fluoride contamination in groundwater in parts of Nalgonda District, Andhra Pradesh India. *Environmental Monitoring and Assessment*, 172(1–4), 481–492.
- Carrillo-Rivera, J. J., Cardona, A., & Edmunds, W. M. (2002). Use of abstraction regime and knowledge of hydrogeological conditions to control high-fluoride concentration in abstracted groundwater: San Luis Potosi basin Mexico. *Journal of Hydrology*, 261(1–4), 24–47.
- Chae, G. T., Yun, S. T., Kwon, M. J., Kim, Y. S., & Mayer, B. (2006). Batch dissolution of granite and biotite in water: implication for fluorine geochemistry in groundwater. *Geochemical Journal*, 40(1), 95–102.
- Chakraborty, B., Roy, S., Bera, A., Adhikary, P. P., Bera, B., Sengupta, D., Bhunia, G.S., & Shit, P. K. (2021). Geospatial Assessment of Groundwater Quality for Drinking through Water Quality Index and Human Health Risk Index in an Upland Area of Chota Nagpur Plateau of West Bengal, India. In Spatial Modeling and Assessment of Environmental Contaminants (pp. 327–358). Springer, Cham. https:// doi.org/10.1007/978-3-030-63422-3_19.
- Chakrabarti, S., & Ray, S. (2013). Fluoride contamination in a hard rock terrain: a case study of Purulia district, West Bengal, India. Journal of Chemical, Biological and Physical Sciences JCBPS, 3(4), 2931.

- Chen, J., Wu, H., Qian, H., & Gao, Y. (2017). Assessing nitrate and fluoride contaminants in drinking water and their health risk of rural residents living in a semiarid region of Northwest China. *Expo*sure and Health, 9(3), 183–195.
- Chitsazan, M., Aghazadeh, N., Mirzaee, Y., & Golestan, Y. (2019). Hydrochemical characteristics and the impact of anthropogenic activity on groundwater quality in suburban area of Urmia city Iran. *Environment, Development and Sustainability*, 21(1), 331–351.
- Das, N., Deka, J. P., Shim, J., Patel, A. K., Kumar, A., Sarma, K. P., & Kumar, M. (2016). Effect of river proximity on the arsenic and fluoride distribution in the aquifers of the Brahmaputra floodplains, Assam, northeast India. *Groundwater for Sustainable Development*, 2, 130–142.
- Das, S. (2019). Comparison among influencing factor, frequency ratio, and analytical hierarchy process techniques for groundwater potential zonation in Vaitarna basin, Maharashtra India. *Groundwater* for Sustainable Development, 8, 617–629.
- Datta, P. S., Deb, D. L., & Tyagi, S. K. (1996). Stable isotope (180) investigations on the processes controlling fluoride contamination of groundwater. *Journal of Contaminant Hydrology*, 24(1), 85–96.
- Debernardi, L., De Luca, D. A., & Lasagna, M. (2008). Correlation between nitrate concentration in groundwater and parameters affecting aquifer intrinsic vulnerability. *Environmental Geology*, 55(3), 539–558.
- Dehbandi, R., Moore, F., & Keshavarzi, B. (2018). Geochemical sources, hydrogeochemicalbehavior, and health risk assessment of fluoride in an endemic fluorosis area, central Iran. *Chemosphere*, 193, 763–776.
- Dissanayake, C. B. (1991). The fluoride problem in the ground water of Sri Lanka—environmental management and health. *International Journal of Environmental Studies*, 38(2–3), 137–155.
- Dregne, H. E. (1967). Water quality problems peculiar to arid regions.
- Edmunds, W. M., & Smedley, P. L. (2013). Fluoride in natural waters. In *Essentials of medical geology* (pp. 311–336). Springer, Dordrecht.
- Emenike, C. P., Tenebe, I. T., & Jarvis, P. (2018). Fluoride contamination in groundwater sources in Southwestern Nigeria: Assessment using multivariate statistical approach and human health risk. *Ecotoxicology and Environmental Safety*, 156, 391–402.
- Emenike, C. P., Tenebe, I. T., Omole, D. O., Ngene, B. U., Oniemayin, B. I., Maxwell, O., & Onoka, B. I. (2017). Accessing safe drinking water in sub-Saharan Africa: Issues and challenges in South-West Nigeria. Sustainable Cities and Society, 30, 263–272.
- Fan, A. M. (2011). Nitrate and nitrite in drinking water: a toxicological review. New York: Elsevier.
- Farooq, S. H., Prusty, P., Singh, R. K., Sen, S., & Chandrasekharam, D. (2018). Fluoride contamination of groundwater and its seasonal variability in parts of Purulia district, West Bengal India. Arabian Journal of Geosciences, 11(22), 709.
- Faure, G. (1991). Principles and applications of inorganic geochemistry: a comprehensive textbook for geology students. Macmillan Collier Macmillan. Maxwell Macmillan International.
- Freni, S. C. (1994). Exposure to high fluoride concentrations in drinking water is associated with decreased birth rates. *Journal of Toxicology and Environmental Health, Part A Current Issues*, 42(1), 109–121.
- Frencken, J. E. (1992). Endemic Fluorosis in developing countries, causes, effects and possible solutions. TNO institute for preventive health care.
- Gaciri, S. J., & Davies, T. C. (1993). The occurrence and geochemistry of fluoride in some natural waters of Kenya. *Journal of Hydrology*, 143(3–4), 395–412.
- Gaikwad, S., Gaikwad, S., Meshram, D., Wagh, V., Kandekar, A., & Kadam, A. (2020). Geochemical mobility of ions in groundwater from the tropical western coast of Maharashtra, India: implication to groundwater quality. *Environment, Development and Sustainability*, 22(3), 2591–2624.
- Gibbs, R. J. (1970). Mechanisms controlling world water chemistry. Science, 170(3962), 1088–1090.
- Gizaw, B. (1996). The origin of high bicarbonate and fluoride concentrations in waters of the Main Ethiopian Rift Valley, East African Rift system. *Journal of African Earth Sciences*, 22(4), 391–402.
- Guo, Q., Wang, Y., Ma, T., & Ma, R. (2007). Geochemical processes controlling the elevated fluoride concentrations in groundwaters of the Taiyuan Basin, Northern China. *Journal of Geochemical Exploration*, 93(1), 1–12.
- Grech, P. (1966). Fluorosis in young persons. A further survey in Northern Tanganyika Tanzania. The British Journal of Radiology, 39(466), 761–764.
- Haimanot, R. T., Fekadu, A., & Bushra, B. (1987). Endemic fluorosis in the Ethiopian Rift Valley. Tropical and Geographical Medicine, 39(3), 209–217.
- He, F. J., & MacGregor, G. A. (2008). Beneficial effects of potassium on human health. *Physiologia-plantarum*, 133(4), 725–735.
- Islam, A. T., Shen, S., Haque, M. A., Bodrud-Doza, M., Maw, K. W., & Habib, M. A. (2018). Assessing groundwater quality and its sustainability in Joypurhat district of Bangladesh using GIS

and multivariate statistical approaches. *Environment, Development and Sustainability, 20*(5), 1935–1959.

- Jacks, B., Bhattacharya, A., & Singh, K. P. (2000). High-fluoride groundwaters in India. In Groundwater 2000, Proceedings of the International Conference on Groundwater Research in Kopenhagen (pp. 193–194).
- Jalees, M. I., Farooq, M. U., Anis, M., Hussain, G., Iqbal, A., & Saleem, S. (2020). Hydrochemistry modelling: evaluation of groundwater quality deterioration due to anthropogenic activities in Lahore, Pakistan. *Environment, Development and Sustainability*, 23(3), 3062–3076.
- Jayaprakash, M., Giridharan, L., Venugopal, T., Kumar, S. K., & Periakali, P. (2008). Characterization and evaluation of the factors affecting the geochemistry of groundwater in Neyveli, Tamil Nadu India. *Environmental Geology*, 54(4), 855–867.
- Jenks, G. F. (1967). The data model concept in statistical mapping. *International Yearbook of Cartography.*, 7, 186–190.
- Jha, S. K., Singh, R. K., Damodaran, T., Mishra, V. K., Sharma, D. K., & Rai, D. (2013). Fluoride in groundwater: toxicological exposure and remedies. *Journal of Toxicology and Environmental Health, Part B, 16*(1), 52–66.
- Jiang, S., Su, J., Yao, S., Zhang, Y., Cao, F., Wang, F., & Xi, S. (2014). Fluoride and arsenic exposure impairs learning and memory and decreases mGluR5 expression in the hippocampus and cortex in rats. *PLoS ONE*, 9(4), e96041.
- Kalawapudi, K., Dube, O., & Sharma, R. (2019). Use of neural networks and spatial interpolation to predict groundwater quality. *Environment, Development and Sustainability*, 22(4), 2801–2816.
- Kumar, M., Das, N., Goswami, R., Sarma, K. P., Bhattacharya, P., & Ramanathan, A. L. (2016). Coupling fractionation and batch desorption to understand arsenic and fluoride co-contamination in the aquifer system. *Chemosphere*, 164, 657–667.
- Kumar, M., Goswami, R., Patel, A. K., Srivastava, M., & Das, N. (2020). Scenario, perspectives and mechanism of arsenic and fluoride co-occurrence in the groundwater: a review. *Chemosphere*, 249, 126126.
- Liu, H., Guo, H., Yang, L., Wu, L., Li, F., Li, S., & Liang, X. (2015). Occurrence and formation of high fluoride groundwater in the Hengshui area of the North China Plain. *Environmental Earth Sciences*, 74(3), 2329–2340.
- Logeshkumaran, A., Magesh, N. S., Godson, P. S., & Chandrasekar, N. (2015). Hydro-geochemistry and application of water quality index (WQI) for groundwater quality assessment, Anna Nagar, part of Chennai City, Tamil Nadu India. *Applied Water Science*, 5(4), 335–343.
- Mandal, J., & Sanyal, S. (2019). Geospatial analysis of fluoride concentration in groundwater in puruliya district, west bengal. Space and Culture, India, 6(5), 71–86.
- Marghade, D., Malpe, D. B., & Zade, A. B. (2011). Geochemical characterization of groundwater from northeastern part of Nagpur urban Central India. *Environmental Earth Sciences*, 62(7), 1419–1430.
- Mondal, D., Gupta, S., & Mahato, A. (2013). Fluoride dynamics in the weathered mantle and the saprolitic zone of the Purulia district West Bengal. Advances in Applied Science Research, 4(6), 187–196.
- Moya, J., Phillips, L., Schuda, L., Wood, P., Diaz, A., Lee, R., Chapman, K.(2011). "Exposure factors handbook: 2011 edition." US Environmental Protection Agency (2011).
- Narsimha, A., & Rajitha, S. (2018). Spatial distribution and seasonal variation in fluoride enrichment in groundwater and its associated human health risk assessment in Telangana State, South India. *Human and Ecological Risk Assessment: an International Journal*, 24(8), 2119–2132.
- Narsimha, A., & Sudarshan, V. (2018a). Drinking water pollution with respective of fluoride in the semi-arid region of Basara, Nirmal district, Telangana State India. *Data in Brief*, 16, 752.
- Narsimha, A., & Sudarshan, V. (2018b). Data on fluoride concentration levels in semi-arid region of Medak, Telangana South India. *Data in Brief*, 16, 717.
- Naseem, S., Rafique, T., Bashir, E., Bhanger, M. I., Laghari, A., & Usmani, T. H. (2010). Lithological influences on occurrence of high-fluoride groundwater in Nagar Parkar area, Thar Desert Pakistan. *Chemosphere*, 78(11), 1313–1321.
- Oruc, N. (2003). Problems of high fluoride waters in Turkey (hydrogeology and health aspects). *The short course on medical geology-health and environment. Canberra, Australia.*
- Piper, A. M. (1944). A graphic procedure in the geochemical interpretation of water-analyses. *Eos, Transactions American Geophysical Union*, 25(6), 914–928.
- Prasanth, S. S., Magesh, N. S., Jitheshlal, K. V., Chandrasekar, N., & Gangadhar, K. (2012). Evaluation of groundwater quality and its suitability for drinking and agricultural use in the coastal stretch of Alappuzha District, Kerala India. *Applied Water Science*, 2(3), 165–175.
- Raju, N. J., Dey, S., & Das, K. (2009). Fluoride contamination in groundwaters of Sonbhadra district, Uttar Pradesh, India. *Current science*, 979–985.
- Rao, N. S., & Devadas, D. J. (2003). Fluoride incidence in groundwater in an area of Peninsular India. *Environmental Geology*, 45(2), 243–251.

- Rao, N. S., Marghade, D., Dinakar, A., Chandana, I., Sunitha, B., Ravindra, B., & Balaji, T. (2017). Geochemical characteristics and controlling factors of chemical composition of groundwater in a part of Guntur district, Andhra Pradesh India. *Environmental Earth Sciences*, 76(21), 747.
- Rodell, M., Velicogna, I., & Famiglietti, J. S. (2009). Satellite-based estimates of groundwater depletion in India. *Nature*, 460(7258), 999–1002.
- Saha, A. K. (1997). Assessment and management of groundwater resources of Purulia and West Medinipur districts of West Bengal (Volume 1). Center for study of man and environment (CSME). pp, 1–48.
- Samal, A. C., Bhattacharya, P., Anusaya Mallick, M., Ali, M., Pyne, J., & Santra, S. C. (2015). A study to investigate fluoride contamination and fluoride exposure dose assessment in lateritic zones of West Bengal, India. *Environmental Science and Pollution Research*, 22(8), 6220–6229.
- Schoeller, H. (1967). Geochemistry of groundwater—an international guide for research and practice (Chap. 15, pp. 1–18).
- Shaji, E., Viju, J., & Thambi, D. S. (2007). High fluoride in groundwater of Palghat District, Kerala. Current Science, 240–245.
- Singaraja, C., Chidambaram, S., Jacob, N., Ezhilarasan, E., Velmurugan, C., Manikandan, M., & Rajamani, S. (2016). Taxonomy of groundwater quality using multivariate and spatial analyses in the Tuticorin District, Tamil Nadu, India. *Environment, Development and Sustainability*, 18(2), 393–429.
- Singh, A., Patel, A. K., & Kumar, M. (2020). Mitigating the risk of arsenic and fluoride contamination of groundwater through a multi-model framework of statistical assessment and natural remediation techniques. In Emerging issues in the water environment during Anthropocene (pp. 285–300). Springer, Singapore.
- Singh, G., Kumari, B., Sinam, G., Kumar, N., & Mallick, S. (2018). Fluoride distribution and contamination in the water, soil and plants continuum and its remedial technologies, an Indian perspective–a review. *Envi*ronmental Pollution, 239, 95–108.
- Sivasankar, V., Darchen, A., Omine, K., & Sakthivel, R. (2016). Fluoride: A world ubiquitous compound, its chemistry, and ways of contamination. In *Surface modified carbons as scavengers for fluoride from water* (pp. 5–32). Springer, Cham.
- Spalding, R. F., & Exner, M. E. (1993). Occurrence of nitrate in groundwater—a review. Journal of Environmental Quality, 22(3), 392–402.
- Subba Rao, N. (2003). Groundwater quality: Focus on fluoride concentration in rural parts of Guntur district, Andhra Pradesh India. *Hydrological Sciences Journal*, 48(5), 835–847.
- Thivya, C., Chidambaram, S., Singaraja, C., Thilagavathi, R., Prasanna, M. V., Anandhan, P., & Jainab, I. (2013). A study on the significance of lithology in groundwater quality of Madurai district, Tamil Nadu (India). *Environment, Development and Sustainability*, 15(5), 1365–1387.
- Tiwari, A. K., & Singh, A. K. (2014). Hydrogeochemical investigation and groundwater quality assessment of Pratapgarh district, Uttar Pradesh. *Journal of the Geological Society of India*, 83(3), 329–343.
- Todd, D. K., & Mays, L. W. (1980). Groundwater Hydrology (p. 535). John Willey & Sons. Inc.
- USEPA (US Environmental Protection Agency). "Risk assessment guidance for superfund." *Human Health Evaluation Manual Part A, Interim Final* 1 (1989).
- Vasanthavigar, M., Srinivasamoorthy, K., Vijayaragavan, K., Ganthi, R. R., Chidambaram, S., Anandhan, P., & Vasudevan, S. (2010). Application of water quality index for groundwater quality assessment: Thirumanimuttar sub-basin, Tamilnadu India. *Environmental Monitoring and Assessment*, 171(1–4), 595–609.
- Viswanathan, N., Sundaram, C. S., & Meenakshi, S. (2009). Removal of fluoride from aqueous solution using protonated chitosan beads. *Journal of Hazardous Materials*, 161(1), 423–430.
- WHO, Guidelines for drinking-water quality. (2011), Edition, F. chronicle 38(4), 104-8.
- Zhang, P., Lv, Z., Gao, L., & Huang, L. (2012). A new framework of the unsupervised classification for highresolution remote sensing image. *Telkomnika. Indonesian Journal of Electrical Engineering and Computer Science*, 10(7), 1746–1755.
- Zhang, Y., Wu, J., & Xu, B. (2018). Human health risk assessment of groundwater nitrogen pollution in Jinghui canal irrigation area of the loess region, northwest China. *Environmental Earth Sciences*, 77(7), 273.
- Zolekar, R. B., Todmal, R. S., Bhagat, V. S., Bhailume, S. A., Korade, M. S., & Das, S. (2021). Hydro-chemical characterization and geospatial analysis of groundwater for drinking and agricultural usage in Nashik district in Maharashtra India. *Environment Development Sustainability*, 23, 4433–4452.

Publisher's Note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.