

# **Temporal and spatial trend analysis of surface water quality in the Doce River basin, Minas Gerais, Brazil**

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### **Abstract**

In Brazil, specifcally in the Doce River basin, there is still a great lack of studies on temporal and spatial trends in water quality, since the water quality in the monitoring campaigns is basically evaluated when it comes down to the concentrations of monitored variables. In this sense, the objective of this work was to perform a temporal and spatial trend analysis of water quality data in the Minas Gerais portion of the Doce River basin, Brazil. For this, the Mann–Kendall, seasonal Mann–Kendall and Spearman correlation tests were used in the temporal analysis and the cluster analysis in the spatial analysis. In the analysis of temporal trends, the analyses were performed using the values of the National Sanitation Foundation Water Quality Index (NSFWQI) and the variables that compose it. In the analysis of spatial trends, the stations were evaluated only based on the WQI. With the results of the analysis of temporal trend, it was identifed that most stations did not present a statistically signifcant trend for the WQI. In the stations that presented trends of quality reduction, most of them are in densely populated areas, demonstrating the strong infuence of the poor sanitary conditions of the municipalities to the water quality of the basin. When analyzing the variables that compose the WQI, the results found for nitrate demonstrated that water quality deterioration is also afected by the difuse pollution originating from farming areas. The results for *Escherichia coli* reinforced the impact of the discharge of domestic effluents and demonstrated the absence of a significative trend is still of concern because it can represent the maintenance of a degradation state in the water bodies. In the spatial trend analysis, the CA grouped the monitoring stations into six clusters based on their similarity among the WQI values, and, together with the results of the other analyses, it was verifed that the Caratinga River basin (UGRH5 Caratinga) presented the highest degree of pollution. It was also possible to identify fve stations that can be reallocated or deactivated since they have similarities with other stations located in the same watercourse.

**Keywords** Cluster analysis · Doce River · Monitoring network · Trend analysis · Water pollution

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In the water basins, water quality can be infuenced by a number of natural factors, such as weathering of rocks and soil erosion, as well as by anthropic factors, through agricultural expansion and accelerated population and industrial growth (Bu et al. [2014;](#page-22-0) Barakat et al. [2016;](#page-22-1) Zeinalzadeh and Rezaei [2017\)](#page-25-0). However, the deterioration of water bodies has been aggravated and is becoming a threat to water security around the world (Zhai et al. [2014\)](#page-25-1).

Considering that water quality behavior is a result of the river basin conditions, periodic monitoring campaigns are the frst step toward developing a reliable and adequate database that can be useful for water resource planning and management (Pesce and Wunderlin [2000;](#page-25-2) Simeonov et al. [2003](#page-25-3); Shrestha and Kazama [2007\)](#page-25-4). These campaigns allow the evaluation of the behavior of water quality variables, detecting spatial and temporal variations and support the planning of water resources in the implementation of management instruments, such as water use permits, collection and framing of water bodies in classes of use (ANA [2013a\)](#page-22-2).

In the state of Minas Gerais, Brazil, the monitoring of surface water began to be carried out in 1997 through the "Waters of Minas Project", under the responsibility of the Minas Gerais Water Management Institute (IGAM). In the Minas Gerais portion of the Doce River basin, the network currently has 65 stations in operation and four annual campaigns with quarterly frequency, two complete and two intermediates. In the complete campaigns, carried out every six months, 51 water quality variables are analyzed in common to the set of stations. For the partial campaigns, carried out between the complete campaigns, 19 variables are analyzed in common to the set of stations and to the four monitoring campaigns.

The water quality monitoring campaigns in the Doce River basin are essential, since they allow the identifcation of difuse and punctual contaminations, such as the discharge of domestic effluent without treatment, causing contamination by thermotolerant coliforms; inadequate disposal of solid waste, causing them to be carried to the water bodies; generation of industrial effluents, causing the release of toxic contaminants of diferent kinds; and inadequate soil use, which, thanks to climatic conditions, provides erosion and sediment transport to the water bodies (ECOPLAN-LUME [2010a](#page-23-0)).

Although the monitoring program has demonstrated its importance in providing basic information for the defnition of strategies and the evaluation of the efectiveness of the environmental control system itself, the monitoring campaigns, when carried out for long periods, end up generating an extensive and complex database, which may hinder its analysis and interpretation, often leading to underutilization (Trindade et al. [2017\)](#page-25-5). On the other hand, the existence of historical data series allows a better understanding of the temporal and spatial evolution of water quality and its correlation with natural and anthropic factors.

Among the available methodologies to interpret qualitative data sets, trend analysis studies have demonstrated their potential use as a tool to help water quality management, showing positive, negative or zero trends for the analyzed water quality variables (Yenilmez et al. [2011](#page-25-6); Tabari et al. [2011;](#page-25-7) Elçi and Selçuk [2013](#page-23-1); Sun et al. [2013](#page-25-8); Kisi and Ay [2014](#page-24-0); Chowdhury and Al-Zahrani [2014](#page-23-2); Kurdi et al. [2015;](#page-24-1) Dou et al. [2016;](#page-23-3) Trindade et al. [2017;](#page-25-5) Oliveira et al. [2017b](#page-24-2); Costa et al. [2017;](#page-23-4) Anand et al. [2019;](#page-22-3) Ebadati and Hooshmandzadeh [2019](#page-23-5)). Trend analysis allows the evaluation of long data series, being useful for monitoring the evolution of water quality, as well as to understand the infuence that factors such as changed land use and sources of pollution in the basin exert on water quality, identifying basins with similar characteristics.

Trend analysis in water quality from monitoring networks is rarely published in scientifc journals. This may be partly due to the difculties associated with maintaining long term, stable (consistent over time) and un-interrupted water quality monitoring (Ballantine and Davies-Colley [2014](#page-22-4)). In Brazil, specifcally in the Doce River basin, there is still a great lack of studies on temporal and spatial trends in water quality, since the water quality in the monitoring campaigns carried out by IGAM is basically evaluated when it comes down to the concentrations of monitored variables.

The choice of the Doce River basin as a study area is also justifed by its economic and environmental context since the basin was the target of a major environmental disaster in Brazil. On November 05, 2015, the Fundão tailings dam, operated by Samarco Mineração SA. collapsed. It was in the district of Bento Rodrigues, municipality of Mariana. The dam, classifed as Class III, with high environmental damage potential, was destined to receive and store the waste generated by the iron ore beneficiation activity (IGAM [2017a](#page-24-3)). The dam contained 56.4 million  $m<sup>3</sup>$  of tailings, of which 43 million  $m<sup>3</sup>$  (80% of the total volume) were released into the environment. This amount reached 668 km of rivers and streams of the Doce River basin, in the states of Minas Gerais and Espírito Santo (Carmo et al. [2017\)](#page-23-6), resulting in several impacts on water resources and their uses, such as public supply, irrigation, industrial use, power generation electrical, leisure and fshing, destruction of permanent preservation areas, silting and morphological alterations of water bodies (ANA [2016\)](#page-22-5).

Based on the above, the objective of this work was to analyze the temporal and spatial trend of water quality data in the Minas Gerais portion of the Doce River basin, allowing the identifcation of the water quality variation, the most impacted areas over the years of monitoring and the main monitoring stations to be maintained in the network, which may subsidize management and planning actions aiming the improvement of the monitoring of water quality in the basin.

### **2 Materials and methods**

#### **2.1 Study area**

The study was developed in the Minas Gerais portion of the Doce River basin, Brazil, corresponding to 86% of the total area of approximately  $82,427 \text{ km}^2 \text{ (ANA } 2013b)$  $82,427 \text{ km}^2 \text{ (ANA } 2013b)$  $82,427 \text{ km}^2 \text{ (ANA } 2013b)$ . The Doce River begins in the state of Minas Gerais, in the Mantiqueira and Espinhaço mountains and flows about 879 km until reaching the Atlantic Ocean in the state of Espírito Santo.

In the state of Minas Gerais (Fig. [1](#page-3-0)), the Doce River basin is subdivided into six water resources management units (UGRH) (CBH-Doce [2016a](#page-23-7)). The UGRH has regional identities characterized by physical, sociocultural, economic and political aspects (IGAM [2016](#page-24-4)).

The basin has 98% of its area inserted in the Atlantic Forest biome, one of the most important and threatened in the world. The remaining 2% are Savanna Formation (CBH-Doce [2016b](#page-23-8)). According to Köppen's climatological classification, the predominant climates in the basin are Aw—tropical monsoon zone and Cwa—humid subtropical zone, which is characterized by a dry winter and hot summer, and Cwb – humid subtropical zone, which is characterized by a dry winter and temperate summer (Alvares et al. [2013](#page-22-7)), with average annual precipitation around 1200 mm (Lima et al. [2019](#page-24-5)).

The economic activity in the Doce River basin is quite diversifed, with emphasis on: farming (traditional crops, cofee, sugar cane, animal husbandry); reforestation; the



<span id="page-3-0"></span>**Fig. 1** Division of the Minas Gerais portion of the Doce River basin per UGRH and spatial distribution of IGAM water quality monitoring stations

agribusiness (sugar and ethanol); mining (iron, gold, bauxite, precious stones, etc.); industry (pulp, steel and dairy); trade and support services of industrial complexes; and electricity generation (ECOPLAN-LUME [2010a](#page-23-0)). The region has the largest steel mill complex in Latin America, which is associated with mining and reforestation companies (CBH-Doce [2016b\)](#page-23-8).

#### **2.2 Database and analysis methods used**

The water quality data used in the study came from the water quality monitoring campaigns of the "Waters of Minas Project", where the water quality analyses are carried out by a laboratory accredited by the National Institute of Metrology, Quality and Technology (INMETRO), which regularly participates in analytical quality control (AQC) evaluations and follows standardized methods for water and sewage analysis (APHA et al. [2012](#page-22-8)). Data analysis was performed in each water quality monitoring station and divided into two stages: (a) temporal trend analysis; and (b) spatial trend analysis.

In general, the water quality can refect changes in basins, making rivers good indicators of land use and land cover (Oliveira et al. [2017a;](#page-24-6) Marmontel et al. [2018;](#page-24-7) Shukla et al. [2020;](#page-25-9) Zhang et al. [2020\)](#page-25-10). To support the discussion of the results obtained, the land use and land cover change in the UGRH1 Piranga was also analyzed.

#### **2.3 Temporal trend analysis**

In the analysis of temporal trend, it was decided to analyze monitoring stations with at least 10 years of operation, as was also done in other studies (Ballantine and Davies-Colley [2014](#page-22-4); Dou et al. [2016;](#page-23-3) Trindade et al. [2017;](#page-25-5) Diamantini et al. [2018\)](#page-23-9). Trend analysis over shorter periods tends to provide results infuenced by local hydrological variability and therefore provides misleading results on long-term trends (Howden et al. [2011](#page-23-10)). Therefore, 32 of the 65 water quality monitoring stations located in the Doce River basin (Table [1\)](#page-5-0) were selected and a base period from 2000 to 2017 was adopted.

Due to the period of the database used, the results found were afected by the tailings dam collapse in the municipality of Mariana, since the IGAM historical series include variables sensitive to the impacts resulting from the accident. It is also worth noting that, of the 65 monitoring stations of the IGAM, 14 were afected by the collapse.

For the stations presented in Table [1](#page-5-0), the analyses were performed using National Sanitation Foundation Water Quality Index (NSFWQI), as well as the variables that comprise it: dissolved oxygen, thermotolerant coliforms, hydrogenionic potential, biochemical oxygen demand, nitrate, total phosphorus, water temperature, turbidity, and total solids. The WQI was chosen because it is a globally used index that refects water contamination due to the nine variables considered most representative for the characterization of water quality. The index value varies from 0 to 100, and it is interpreted as follows: excellent (90 < WQI ≤ 100), good (70 < WQI ≤ 90), medium (50 < WQI ≤ 70), bad (25 < WQI  $\leq$  50) and very bad (WQI  $\leq$  25). For the present study, the thermotolerant coliforms were replaced by the *Escherichia coli* from 2013. This fact is due to studies that have shown the species *E. coli* as the only indicator of contamination by thermotol-erant coliforms species in IGAM water quality analyzes (IGAM [2016](#page-24-4)). Studies suggest that *E.coli* is a more reliable indicator of fecal pollution and the occurrence of pathogens in water than total and thermotolerant coliforms (Edberg et al. [2000;](#page-23-11) Leclerc et al. [2001](#page-24-8)).

To evaluate the existence of trends in the historical series, each of the 32 stations was analyzed individually. For this purpose, we used the nonparametric Mann–Kendall tests (MK) (Mann [1945;](#page-24-9) Kendall [1975](#page-24-10)), seasonal Mann–Kendall test (SMK) (Hirsch et al. [1982](#page-23-12)) and Spearman (Gauthier [2001](#page-23-13)). The main reason for the use of nonparametric statistical tests is that they are considered more adequate for data that do not present normal distribution, as it frequently happens in time series (Yue et al. [2002](#page-25-11); Trindade et al. [2017;](#page-25-5) Costa et al. [2017](#page-23-4)).

Since the variables evaluated have a quarterly frequency, it was frst verifed the existence of seasonality among the data using the Kruskal–Wallis (KW) nonparametric test (Kruskal and Wallis [1952](#page-24-11)) at a signifcance level of 5%. For the series that presented significant difference among the quarters ( $p$  value  $< 0.05$ ), the influence of seasonality on the data was taken into account. In these situations, the Seasonal Mann–Kendall test (SMK) was applied, which consists of a variation of the MK test in which seasonality is considered in the data series (Anghileri et al. [2014](#page-22-9)). For the series in which no seasonality was identifed, the MK test was used for temporal trend analysis.

The MK and SMK tests were applied at a signifcance level of 5%, and the non-trend null hypothesis  $(H_0)$  in the series was rejected when *S* was significantly different from zero (*p* value < 0.05), in favor of the alternative hypothesis  $(H_1)$  of data trend. Positive values of the *S* statistics and Kendall's  $\tau$  indicate an elevation trend, negative values indicate a reduction trend, and zero indicates no trend in the series.

<b>Station</b> <sup>a</sup>	Coordinates		<b>UGRH</b>	Water course
	Latitude	Longitude		
<b>RD001</b>	$-20^{\circ}41'18.66''$	–43°18′08.42″	1 Piranga	Piranga River
<b>RD004</b>	$-20^{\circ}47'06.99''$	$-43^{\circ}06'56.99''$	1 Piranga	Xopotó River
<b>RD007</b>	$-20^{\circ}40'18.99''$	$-43^{\circ}05'30.99''$	1 Piranga	Piranga River
<b>RD009</b>	$-20^{\circ}21'00.00''$	$-43^{\circ}19'05.00''$	1 Piranga	Carmo River
<b>RD013</b>	$-20^{\circ}22'59.80''$	$-42^{\circ}54'08.50''$	1 Piranga	Piranga River
<b>RD018</b>	$-20^{\circ}05'53.00''$	$-42^{\circ}37'46.99''$	1 Piranga	Casca River
<b>RD019</b>	$-20^{\circ}01'18.99''$	$-42^{\circ}45'07.99''$	1 Piranga	Doce River
RD021	$-20^{\circ}04'35.77''$	$-42^{\circ}27'58.61''$	1 Piranga	Matipó River
<b>RD023</b>	$-19^{\circ}45'34.99''$	$-42^{\circ}29'06.00''$	1 Piranga	Doce River
RD025	$-19^{\circ}56'21.69''$	$-43^{\circ}10'48.99''$	2 Piracicaba	Piracicaba River
<b>RD026</b>	$-19°50'04.34"$	$-43^{\circ}7'38.431''$	2 Piracicaba	Piracicaba River
RD027	$-19^{\circ}48'36.00''$	$-43^{\circ}14'00.00''$	2 Piracicaba	Santa Bárbara River
<b>RD029</b>	$-19^{\circ}46'00.99''$	$-43^{\circ}02'38.99''$	2 Piracicaba	Piracicaba River
<b>RD030</b>	$-19^{\circ}44'03.75''$	$-43^{\circ}01'41.24''$	2 Piracicaba	Peixe River
<b>RD031</b>	$-19°31'33.86"$	$-42^{\circ}39'28.78''$	2 Piracicaba	Piracicaba River
RD032	$-19°37'11.80''$	$-42^{\circ}48'02.71''$	2 Piracicaba	Piracicaba River
<b>RD033</b>	$-19^{\circ}19'38.93''$	$-42^{\circ}22'32.98''$	2 Piracicaba	Doce River
<b>RD034</b>	$-19°31'48.27"$	$-42^{\circ}36'09.17''$	2 Piracicaba	Piracicaba River
<b>RD035</b>	$-19^{\circ}29'18.99''$	$-42^{\circ}29'38.99''$	2 Piracicaba	Doce River
RD039	$-19^{\circ}13'25.03''$	$-42^{\circ}20'34.68''$	3 Santo Antônio	Santo Antônio River
<b>RD040</b>	$-19^{\circ}01'14.95''$	$-42^{\circ}9'45.529''$	4 Suaçuí	Corrente Grande River
<b>RD044</b>	$-18°53'00.00"$	$-41^{\circ}57'10.00''$	4 Suaçuí	Doce River
<b>RD045</b>	$-18°51'36.19"$	$-41^{\circ}50'01.35''$	4 Suaçuí	Doce River
<b>RD049</b>	$-18^{\circ}34'35.99''$	$-41^{\circ}55'14.00''$	4 Suaçuí	Suaçuí Grande River
<b>RD053</b>	$-18°58'10.19"$	$-41^{\circ}38'49.39''$	4 Suaçuí	Doce River
<b>RD056</b>	$-19^{\circ}43'36.00''$	$-42^{\circ}07'58.99''$	5 Caratinga	Caratinga River
<b>RD057</b>	$-19^{\circ}04'15.78''$	$-41^{\circ}32'39.83''$	5 Caratinga	Caratinga River
<b>RD058</b>	$-19^{\circ}09'58.84''$	$-41°27'35.69"$	5 Caratinga	Doce River
<b>RD059</b>	$-19^{\circ}20'45.72''$	$-41^{\circ}14'19.49''$	6 Manhuaçu	Doce River
<b>RD064</b>	$-20^{\circ}06'59.11''$	$-41^{\circ}55'09.80''$	6 Manhuaçu	Manhuaçu River
<b>RD065</b>	$-19^{\circ}29'51.00''$	$-41^{\circ}10'09.99"$	6 Manhuaçu	Manhuaçu River
<b>RD067</b>	$-19^{\circ}30'20.00''$	–41°00′47.00″	6 Manhuaçu	Doce River

<span id="page-5-0"></span>**Table 1** Water quality monitoring stations located in the Minas Gerais portion of the Doce River basin used in the temporal trend analysis

<sup>a</sup>Stations highlighted in bold refer to those affected by the tailings dam collapse

Then, to corroborate with the results found in the MK and SMK tests, the autocorrelation of the time series was verifed using the Spearman nonparametric test. This test is based on the correlation coefficient (Spearman's  $R$ ), and the trend of data elevation or reduction over time is evaluated. The use of this test was based on the fact that the outliers have little infuence on their results, the collection of samples at regular intervals is not necessary, and the test is simple to apply even in a large dataset (Gauthier [2001\)](#page-23-13).

Similarly, to the MK and SMK tests, correlations with *p* value less than 0.05 (5% signifcance level) were considered signifcant, there was a trend over time for the analyzed variable. In this case, the positive or negative sign of the  $R$  coefficient indicated the rising or decreasing trend, respectively, since it corresponds to the correlation of the values of the variable with time (Trindade et al. [2017\)](#page-25-5).

For monitoring stations, whose results were not significant for the MK or SMK test, but signifcant for the Spearman correlation test, or vice versa, the results were considered as inconclusive with the possibility of elevation, when the values were positive, or reduction when values were negative.

#### **2.4 Spatial trend analysis**

The cluster analysis (CA) was used to verify the spatial trend of monitoring station data in the Minas Gerais portion of the Doce River basin. The CA aimed to group the monitoring stations into clusters, so that the stations within a group are similar to each other, but diferent from other groups (Shrestha and Kazama [2007](#page-25-4); Jung et al. [2016\)](#page-24-12). The CA was used to reliably classify the surface water quality and to allow the guidance of future spatial sampling decisions, reducing the number of stations and the costs (Shrestha and Kazama [2007](#page-25-4); Zhang et al. [2011](#page-25-12); Ajorlo et al. [2013](#page-22-10); Muangthong and Shrestha [2015;](#page-24-13) Jung et al. [2016;](#page-24-12) Barakat et al. [2016](#page-22-1); Calazans et al. [2018a](#page-22-11)).

In order to obtain a more recent characterization of the water quality and to evaluate most of the monitoring stations, the CA was applied considering the data from 2010 to 2017. The time series of the WQI per monitoring station were used as input data and evaluated 64 of the 65 existing monitoring stations, justifed by the fact that the RD011 station has data only from the year 2016 on, which made it impossible to use them in the analysis. In the CA, the hierarchical grouping was applied through the Ward method in the normalized data set, using Euclidean distance as a measure of dissimilarity (linkage distance), as also used in several other studies (Zhang et al. [2011;](#page-25-12) Ajorlo et al. [2013;](#page-22-10) Muangthong and Shrestha [2015\)](#page-24-13).

In order to confrm the results obtained in the CA, the percentage of violation of the water quality variables was calculated according to the limits established by CONAMA Resolution No. 357/2005 (Brasil [2005\)](#page-22-12) and by COPAM/CERH-MG DN No. 01/2008 (Minas Gerais [2008](#page-24-14)). To do this, a preliminary analysis was carried out to flter out the 51 water quality variables common to the stations, which had their limits established in the legislation, which resulted in the selection of 32 variables (Table [2](#page-7-0)).

Since only the Piracicaba River and its tributaries have a framing approved in the scope of the State Water Resources Council (CERH-MG), in the Doce River basin, the class 2 of framing was adopted for the other water bodies, according to CNRH Resolution No. 91/2008 (Brasil [2008](#page-22-13)). Waters with class 2 framing can be used for the following purposes: supply to human consumption, after conventional treatment; protection of aquatic communities; recreation of primary contact, such as swimming, water skiing and diving, according to CONAMA Resolution No. 274/2000 (Brasil [2000](#page-22-14)); irrigation of vegetables, fruit plants and parks, gardens, sports and leisure felds, with which the public may come into direct contact; and aquaculture and fshing. The results of the analysis highlighted the monitoring stations that presented the highest and lowest violation rates according to the limits established by the legislation, which were compared to the CA results.

Variables	
Dissolved aluminium (mg $L^{-1}$ )	Total manganese (mg $L^{-1}$ )
Total arsenic (mg $L^{-1}$ )	Total mercury ( $\mu$ g L <sup>-1</sup> )
Total boron (mg $L^{-1}$ )	Total nickel (mg $L^{-1}$ )
Total cadmium (mg $L^{-1}$ )	Nitrate (mg $L^{-1}$ )
Total lead (mg $L^{-1}$ )	Nitrite (mg $L^{-1}$ )
Free cyanide (mg $L^{-1}$ )	Total ammoniacal nitrogen (mg $L^{-1}$ )
Total chloride (mg $L^{-1}$ )	Dissolved oxygen (mg $L^{-1}$ )
Chlorophyll a $(\mu g L^{-1})$	Hydrogenionic potential
Dissolved copper (mg $L^{-1}$ )	Total selenium (mg $L^{-1}$ )
True color $(mg L^{-1})$	Total dissolved solids (mg $L^{-1}$ )
Total chromium (mg $L^{-1}$ )	Total suspended solids (mg $L^{-1}$ )
Biochemical oxygen demand (mg $L^{-1}$ )	Surfactants (mg $L^{-1}$ )
<i>Escherichia coli</i> (MPN 100 mL <sup>-1</sup> )	Total sulfate (mg $L^{-1}$ )
Total phenols (mg $L^{-1}$ )	Sulfide $(mg L^{-1})$
Dissolved iron (mg $L^{-1}$ )	Turbidity (NTU)
Total phosphorus (mg $L^{-1}$ )	Total zinc (mg $L^{-1}$ )

<span id="page-7-0"></span>**Table 2** Variables of water quality with limits established in the legislation according to the classifcation class

#### **2.5 Land use and land cover change**

The land use and land cover (LULC) maps for the years 2000 and 2017 were compared. The LULC classes were obtained based on the mapping carried out by the MapBiomas Project, collection 4.1 (MapBiomas [2019](#page-24-15)). MapBiomas Project is an initiative that involves a collaborative network of biomes, land use, remote sensing, geographic information system and computer science experts that rely on Google Earth Engine platform and its cloud processing and automated classifers capabilities to generate Brazil's annual land use and land cover time series.

In total, 13 classes of LULC were obtained in both years evaluated. To facilitate the analysis, it was decided to merge some of the LULC classes obtained, totaling nine classes (Table [3](#page-8-0)).

### **3 Results and discussion**

The results of the statistical tests proposed in the methodology were organized in individual worksheets for the WQI and for each water quality variable analyzed. In the KW test, the statistically significant results  $(p<0.05)$  indicated the presence of seasonality among the data. Statistically significant results  $(p < 0.05)$  from the MK, SMK and Spearman correlation tests suggested that there is a trend to change the variable over time. The Kendall's Tau and Spearman's R values indicated the direction of the trend of the variable over time, with positive values indicating elevation and negative values indicating reduction. The results of the temporal trend analyses for the WQI are presented in Table [4](#page-9-0) and defned as:

<span id="page-8-0"></span>

(↑) elevation trend; (↓) reduction trend; (?↑) inconclusive with possibility of elevation; (?↓) inconclusive with possibility of reduction; and (–) no trend.

As can be seen in Table [4](#page-9-0), the Spearman correlation test and MK or SMK were coincident regarding the indication of whether there was a trend in most of the analyzed cases. However, situations were found, as for the WQI, in stations RD007, RD031, RD044 and RD065, in which the tests did not show the same result regarding the signifcance of the temporal trend. For these stations, the WQI trend was considered as inconclusive, with the possibility of elevation or reduction. Among the analyzed stations, it was also observed that only stations RD021 and RD056 do not present seasonality between the values, both of which presented a reduction trend. In other words, water quality may be decreasing in every period of the year analyzed. For a better visualization, the spatial distribution of the WQI temporal trend analysis results for the 32 monitoring stations analyzed in the Minas Gerais portion of the Doce River basin is presented in Fig. [2.](#page-10-0)

As can be seen in Table [4](#page-9-0), 20 of the 32 monitoring stations (62.5%) evaluated did not present a trend in relation to the WQI. However, this scenario should not be understood as good, since the stations that did not present a signifcant trend of elevation and have low values of WQI is an indication that the water quality is presenting a constant degradation.

As a result of the analysis of temporal trend, it was also possible to verify that fve stations presented statistically signifcant temporal trends of WQI elevation: RD009 (UGRH1), RD018 (UGRH1), RD027 (UGRH2), RD040 (UGRH4), RD053 (UGRH4). Among them, only the RD040 station is in a rural area, while the others are downstream from the city of Tumitiringa (RD053) and districts of Monsenhor Horta (RD009), Águas Férreas (RD018) and Santa Rita de Pacas (RD027). For the reduction trend, four stations presented statistically signifcant results: RD021 (UGRH1), RD032 (UGRH2), RD056 (UGRH5), RD064 (UGRH6). Except for the RD032, these stations are in the urban area, downstream of the cities of Raul Soares (RD021), Caratinga (RD056) and Santana do Manhuaçu (RD064).

Based on the results, it is possible to state that the trend to reduce water quality is typical of metropolitan regions and large cities in the countryside, where water bodies present high degradation of water quality resulted from the discharge of domestic

			Station <sup>a</sup> KW test MK test Kendall's Taub	$S^{\rm b}$		SMK test Kendall's Tau <sup>c</sup>	$S^{\rm c}$	Spearman's R	WQI
<b>RD001</b>	0.017				0.289	0.093	57	0.186	
<b>RD004</b>	0.010				0.272	0.097	59	0.189	$\overline{\phantom{0}}$
<b>RD007</b>	0.019				0.225	0.106	65	0.196	21
<b>RD009</b>	0.003				0.011	0.221	135	0.314	$\uparrow$
RD013	0.004				0.622	0.044	27	0.162	$\overline{\phantom{0}}$
<b>RD018</b>	0.000				0.035	0.183	112	0.256	$\uparrow$
<b>RD019</b>	0.000				0.865	$-0.016$	$-10$	0.011	
RD021	0.219	0.001	$-0.264$	$-672$				$-0.384$	$\downarrow$
<b>RD023</b>	0.000				0.325	$-0.087$	$-53$	$-0.057$	
RD025	0.000				0.185	$-0.116$	$-71$	$-0.118$	
<b>RD026</b>	0.013				$\mathbf{1}$	0.002	$\mathbf{1}$	0.056	$\overline{\phantom{0}}$
RD027	0.032				0.008	0.229	140	0.299	$\uparrow$
RD029	0.007				0.596	$-0.047$	$-29$	$-0.068$	
<b>RD030</b>	0.009				0.051	$-0.170$	$-104$	$-0.157$	$\overline{\phantom{0}}$
RD031	0.000				0.051	$-0.170$	$-104$	$-0.204$	$21$
RD032	0.001				0.043	$-0.177$	$-108$	$-0.239$	↓
<b>RD033</b>	0.000				0.677	$-0.038$	$-23$	0.005	
RD034	0.013				0.353	0.082	50	0.135	
<b>RD035</b>	0.000				0.970	$-0.005$	$-3$	$-0.011$	
RD039	0.000				0.733	$-0.031$	$-19$	$-0.004$	-
<b>RD040</b>	0.001				0.010	0.224	137	0.322	$\uparrow$
<b>RD044</b>	0.000				0.021	$-0.201$	$-123$	$-0.150$	$21$
<b>RD045</b>	0.000				0.677	0.038	23	0.062	
RD049	0.000				0.596	$-0.048$	$-29$	0.009	-
<b>RD053</b>	0.000				0.010	0.224	137	0.296	↑
<b>RD056</b>	0.349	0.000	$-0.291$	$-743$				$-0.406$	$\downarrow$
<b>RD057</b>	0.000				0.609	0.046	28	0.068	
<b>RD058</b>	0.000				0.405	$-0.074$	$-45$	$-0.065$	
<b>RD059</b>	0.001				0.211	0.110	67	0.151	
<b>RD064</b>	0.020				0.007	$-0.234$	$-143$	$-0.316$	↓
<b>RD065</b>	0.000				0.078	0.154	94	0.188	
<b>RD067</b>	0.000				0.649	$-0.041$	$-25$	$-0.035$	$\overline{\phantom{0}}$

<span id="page-9-0"></span>**Table 4** Results of the WQI temporal trend analyses in the Minas Gerais portion of the Doce River basin for the 32 monitoring stations analyzed

<sup>a</sup>Stations highlighted in bold refer to those affected by the Fundão tailings dam collapse;

b Values for the MK test

c Values for the SMK test

effluents and industrial effluents. Most municipalities in the basin do not have effluent treatment plants, and when they do, treatment is insufficient. According to a study carried out on sanitary sewage in Brazil, the municipalities of Caratinga and Raul Soares release a total daily load of 3988.0 and 857.4 kg of biochemical oxygen demand, respectively, while smaller municipalities like Piranga and Tumiritinga release much lower values: 328.9 and 234.7 kg of biochemical oxygen demand per day, respectively (ANA



<span id="page-10-0"></span>**Fig. 2** Spatial distribution of the WQI temporal trend analysis in the Doce River basin for the 32 monitoring stations analyzed

[2017](#page-22-15)). This result also corroborates with what was found in other studies carried out in Brazil, which point out the release of non-treated or partially treated domestic effluents as the main cause of water quality reduction (Oliveira et al. [2017b;](#page-24-2) Trindade et al. [2017;](#page-25-5) Costa et al. [2017;](#page-23-4) Calazans et al. [2018a,](#page-22-11) [b\)](#page-22-16). According to Kamal et al. ([2020](#page-24-16)), the status of river water quality, especially in urban areas needs to be monitored closely and frequently due to the steady increase in population and vast development activities in these areas.

For the nine variables of water quality analyzed, the conclusion of the results of the analysis of temporal trend for each station is presented in Table [5.](#page-11-0)

As can be observed in Table [5,](#page-11-0) nitrate presented the worst results, showing an elevation trend in all monitoring stations evaluated. As reported in the literature, nitrate is mainly related to the use of fertilizers in agriculture (Chen et al. [2019;](#page-23-14) Sorando et al. [2019](#page-25-13); Shukla et al. [2020;](#page-25-9) Kamal et al. [2020](#page-24-16)) and irrigation with untreated wastewater (Dhanasekarapandian et al. [2016](#page-23-15)). In a water quality study in the Tahtali basin, Turkey, Elçi and Selçuk ([2013](#page-23-1)) found a reduction trend for nitrate. The authors attributed the improvements in nitrate concentrations to the new fertilizer use regulations. In the Verde River basin, Brazil, Pinto et al. ([2017](#page-25-14)) also found an elevation trend for nitrate in all evaluated water quality monitoring stations, with the results attributed to the carrying of pollutants from the farming and urban areas.

<span id="page-11-0"></span>

<sup>a</sup>Stations highlighted in bold refer to those affected by the tailings dam collapse

<sup>b</sup>BOD: biochemical oxygen demand; *E. coli: Escherichia coli*; P<sub>T</sub>: total phosphorus; NO<sub>3</sub>\_: nitrate; OD: dissolved oxygen; pH: hydrogenionic potential; TS: total solids; T: water temperature; TU: turbidity

When comparing the maps of LULC between the years 2000 and 2017 (Fig. [3\)](#page-12-0), it can be seen that the farming areas represent the highest percentage of LULC with little variation between the years evaluated (Table [6\)](#page-12-1). It is also possible to observe that the urban infrastructure obtained one of the highest percentage values of growth between the years evaluated. In Brazil, many studies have noted the signifcant contribution of increases in urban areas to the water quality deterioration and the alteration of fow regimes (Calijuri et al.



<span id="page-12-0"></span>**Fig. 3** LULC maps for the years 2000 and 2017

Casses of land cover and use	2000		2017		Increase and
	Area $(km^2)$	Cover $(\%)$	Area $(km^2)$	Cover $(\%)$	decrease (km <sup>2</sup> )
Forest formation	17,065.85	23.90	17,298.65	24.22	232.80
Savanna formation	412.16	0.58	383.72	0.54	$-28.44$
Forest plantation	2454.92	3.44	3430.11	4.80	975.19
Non-forest natural formation	430.40	0.60	402.12	0.56	$-28.28$
Farming	49.969.24	69.97	48,587.38	68.04	$-1381.86$
Urban infrastructure	366.96	0.51	521.63	0.73	154.67
Rocky outcrop	443.40	0.62	489.77	0.69	46.37
Mining	4.41	0.01	6.30	0.01	1.89
Water bodies	264.44	0.37	292.10	0.41	27.66

<span id="page-12-1"></span>**Table 6** Changes in land cover and use between 2000 and 2017

[2015;](#page-23-16) Rodrigues et al. [2019\)](#page-25-15). These facts reaffirm the poor sanitary conditions of the basin, demonstrating that, in addition to the contamination from the point pollution (domestic and industrial effluents), the water quality deterioration is also affected by the diffuse pollution originating from farming areas (Yidana et al. [2010](#page-25-16); Dhanasekarapandian et al. [2016;](#page-23-15) Şener et al. [2017;](#page-25-17) Wu et al. [2018](#page-25-18)).

While some contaminants have shown signifcant reduction trends, they may still have concentration levels that exceed the standards set by legislation. This fact can be exemplifed by the *E. coli* variable that, although it was pointed out as one of the most critical variables in the Doce River basin (IGAM [2017b](#page-24-17)), it presented a larger number of stations with a signifcant reduction trend than with elevation trend. However, when evaluating the *E. coli* data at stations RD021 and RD027, they are above the limit established by local legislation most of the time (Fig. [4\)](#page-13-0), regardless of the outcome of the trend analysis.

According to Fraga et al. ([2019](#page-23-17)), this result corroborates most of the studies found in the literature on water quality in Brazilian basins, which highlights the release of domestic effluents without proper treatment as the main source of pollution, with high values of violation mainly for *E. Coli*. Unlike other coliform bacteria, *E. Coli*. are almost exclusively of fecal origin and can be detected in elevated densities in human and animal feces, sewage and water subjected to recent fecal pollution (Hachich et al. [2012](#page-23-18)). Violations found at *E. Coli* levels for stations RD021 and RD027 classify water quality as unsuitable for recreation of primary contact, according to CONAMA Resolution No. 274/2000.

Despite the indicative of poor sanitary conditions, the trends found for BOD, total phosphorus, DO and turbidity are considered good results, since most of them positively favor water quality. It can be seen that the few sewage treatment plants present in the basin manage to reduce a great part of the organic matter and nutrients, such as BOD and total phosphorus, as evidenced by the reduction trend for both variables in the RD035 station, located downstream of the municipality of Ipatinga, in which collects and treats 100% of its domestic efuents (ANA [2017](#page-22-15)). It is also observed that the



<span id="page-13-0"></span>**Fig. 4** Historical *E. coli* data series for RD021 and RD027 stations

turbidity exhibited a trend behavior opposite to TS. Since TS is composed of total suspended solids (TSS) and total dissolved solids (TDS), while turbidity represents basically TSS (Efendi et al. [2015](#page-23-19)), it can be concluded that TS is increasing as a function of TDS.

When comparing the data in Tables [4](#page-9-0) and [5](#page-11-0), it can be seen that most of the stations with a significant reduction trend (RD021, RD032, RD056, RD064), presented a signifcant trend of elevation for *E. coli*, nitrate and TS, which leads to the conclusion that these variables are primarily responsible for the reduction trends of the WQI.

By associating the variables analyzed with the collapsed dam in Mariana, ANA ([2016\)](#page-22-5) reports that turbidity and TS peaks were recorded as the tailings wave moved along the course of the Doce River, with higher values upstream and reduction trend downstream. Tailing particles caused severe changes in the water quality of the Doce River and estuarine region, increasing the turbidity levels in Minas Gerais up to 6000 times (600,000 NTU) higher than the upper limit established by legislation for this parameter (IGAM [2015\)](#page-24-18). Despite the increase in the concentrations of such variables (Fig. [5](#page-14-0)), the dam collapse was not enough to cause signifcant elevation trends in the historical series of the affected stations.

It is important to note that the values shown in Fig. [5](#page-14-0) refer to the water quality campaigns of the "Waters of Minas Project", which did not include measurements right after



<span id="page-14-0"></span>**Fig. 5** Historical series of turbidity and total solids data for the RD019 station, located in the Doce River downstream of the Fundão tailings dam, in Mariana, Minas Gerais state

the collapse of the dam and consequently has lower levels of turbidity and total solids than the one registered by IGAM  $(2015)$ .

For stations that showed an elevation trend for TS, it is possible to state that the result was caused by the defcient soil management in farming and mining, which causes the transport of solids to water (Hatje et al. [2017](#page-23-20); Costa et al. [2017](#page-23-4)). As shown in Table [6](#page-12-1), farming areas correspond to the highest percentage of LULC in the basin. The characteristics of soils and relief lead the basin to a condition of fragility regarded to the susceptibility to erosion, being aggravated by the high percentage of anthropized areas and human activities (ECOPLAN-LUME [2010a](#page-23-0); Oliveira and Quaresma [2017](#page-24-19)). In Table [6](#page-12-1), it is possible to observe that, together with the urban infrastructure, mining obtained one of the highest growth percentages between the years evaluated (2000–2017). According to Hatje et al. [\(2017\)](#page-23-20), small-scale clandestine mining is still present in the basin and contributes to the contamination of the environmental compartments to levels that may have adverse effects on ecosystem services.

Although the dam collapse has no connection with the trends of the water quality variables evaluated, the economic and environmental impacts were disastrous. In addition to turbidity and TS, high levels of arsenic, cadmium, copper, chromium, nickel and mercury in water samples from the Doce River following the dam collapse (IGAM [2015\)](#page-24-18). Based on a survey of 308 cases of mining dam collapses in the world (1915–2016), the Fundão dam disaster can be regarded as the largest technological disaster, considering the volume of tailings released and the geographical extension of environmental damage (Carmo et al. [2017](#page-23-6)), leaving 19 dead, 3 missing and over 600 homeless (Neves et al. [2016](#page-24-20)). The tailings directly hit 135 identifed semideciduous seasonal forest fragments, in a 298 ha of vegetation suppression, located on the banks of Gualaxo do Norte and Carmo Rivers and its tributaries. The tailings also directly hit 863.7 ha of permanent preservation areas associated to watercourses, which were in protected areas, as defned by the federal forest code. Santarém Stream (11.9 km impacted), Gualaxo do Norte River (68.4 km) and Carmo River (24.7 km) were the main rivers and streams completely silted by the tailings. In addition, 294 small creeks were afected by the tailings (Carmo et al. [2017\)](#page-23-6). Forty-one municipalities in the states of Minas Gerais and Espírito Santo were afected, and hundreds of thousands of people were left without access to clean water (Neves et al. [2016\)](#page-24-20) and activities such as fshing and irrigation (Fernandes et al. [2016](#page-23-21)). As a consequence of this, thousands of individual lawsuits and dozens of class actions seeking compensation for material and moral damages were fled at courts in the states of Espirito Santo and Minas Gerais (Losekann et al. [2020\)](#page-24-21).

In the analysis of spatial trend, the CA grouped the 64 monitoring stations of the Minas Gerais portion of the Doce River basin into six clusters, as can be observed in Table [7](#page-16-0). Figure [6](#page-16-1) shows the profile of the clusters based on the analyzed observations.

In order to improve the visualization of the spatial distribution of the stations among the six clusters obtained in the CA, the result of the analysis can also be observed in Fig. [7.](#page-17-0)

Although the CA has grouped the monitoring stations with similar characteristics in relation to the WQI, the clusters have stations grouped in diferent stretches of river and in diferent UGRHs that, in turn, have extensive areas and quite diversifed economic activities. Thus, in the case of the Doce River basin, the choice of the representative stations of each cluster should consider the connectivity between them, their distribution among the UGRHs and the factors that most infuence the variation of water quality. In other words, it is recommended that rivers that have a single monitoring station grouped by the CA should have it prioritized in the network, since timely and more detailed

<span id="page-16-0"></span>



<span id="page-16-1"></span>**Fig. 6** Profle of the six clusters obtained in the Minas region of the Doce River basin

information may be needed. For those who have more than one station in the same cluster it is possible to consider abdicating one of them.

In Table  $8$ , it is possible to observe the grouping of the stations among the clusters, as well as the identifcation of the UGRH and the watercourse in which they are inserted.

As can be observed in Table [8,](#page-18-0) all the stations grouped in cluster 1 are inserted in the UGRH1. It is verifed that only the Piranga and Doce rivers presented more than one station in the same cluster, a fact that, as already mentioned, allows the abdication of some of them.

In the Piranga River, stations RD001 and RD007 are located in the cities of Piranga and Porto Firme, respectively, which, according to ANA ([2017](#page-22-15)), have an index of 0.0%



<span id="page-17-0"></span>**Fig. 7** Map resulting from the CA for the grouping of the 64 monitoring stations in the Minas Gerais portion of the Doce River basin

of collection and treatment of efuents. Despite the precariousness of basic sanitation, the stations did not present a signifcant trend for the WQI (RD001) and an inconclusive trend with possibility of elevation (RD007). Thus, when analyzing all the information, it is not recommended to exclude one of them, since both are susceptible to variations in water quality due to the release of untreated domestic effluent. In the Doce River (RD019 and RD023), the stations did not present a signifcant trend and are in the rural area, which implies minor importance to maintain both stations. However, due to the collapse of the tailings dam in the municipality of Mariana, the maintenance of both stations is extremely important since they are part of the water quality monitoring points that are used by IGAM to assess the impacts caused by the disaster (IGAM [2017a](#page-24-3)).

In cluster 2, there is more than one station in the Piranga, Carmo, Piracicaba and Doce rivers. For grouped stations in Piranga River, RD013 showed no trend for the WQI, while the RD068 station did not possess enough series of data for analysis. Despite the similarity between the water quality data, both stations can be considered as priorities in the monitoring network, since they are very distant from each other, also having stations that were grouped in clusters 1 and 3. In the Piracicaba River, only in the RD075 no trend analysis was performed, and all the others did not present a trend in the WQI data. Despite the nontrend of the data, stations RD025, RD026, RD029, RD034 are in or downstream of the cities of Rio Piracicaba, João Monlevade, Nova Era and Coronel Fabriciano, respectively.



<span id="page-18-0"></span>**Table 8** Grouping of the stations among the clusters and the identifcation of the UGRH and the watercourse in which they are inserted



#### **Table 8** (continued)

Among the cities, only the Rio Piracicaba has 17.2% of its effluents collected and treated, while Coronel Fabriciano, which has the largest number of inhabitants and generates a load of 5775.0 kg BOD day<sup>-1</sup>, this value is 0.0% (ANA [2017\)](#page-22-15). Thus, although they are all allocated in the same watercourse, the stations must be maintained in the monitoring network.

In the Doce River, two of the stations are located upstream (RD044) and downstream (RD045) of the city of Governador Valadares, presenting an inconclusive trend for the WQI with possibility of reduction and non-trend, respectively. Although the city of Governador Valadares has an effluent collection rate of 95.4%, the treatment percentage corresponds to 0.0% (ANA [2017](#page-22-15)). Thus, it is noticed that the two stations were established in order to verify the infuence of the discharge of the Governador Valadares efuent on the water quality of the Doce River. However, because the two stations are grouped in the same cluster, it can be concluded that the objective is not being met or that the impact of the discharge of the municipal effluent in the Doce River is not significant. For both hypotheses, the RD045 station has a priority over RD044, since, because it is located downstream, it allows to continue evaluating the impact of the discharge of the effluents from the municipality of Governador Valadares in the Doce River.

Cluster 3 grouped stations of all UGRHs, presenting the best WQI results among the six clusters obtained and more than one station grouped only in the Manhuaçu and Santo Antônio rivers. In the Manhuaçu River, both stations are in the rural area, with trend analysis only for RD065, which did not present a trend of elevation for the WQI. In the Santo Antônio River, the stations are also located in the rural area, however, none of them had enough data to perform the trend analysis. Therefore, due to the result of no trend of the RD065 station and the lack of information about the other stations makes it difficult to infer about them.

Cluster 4 has stations from the UGRHs 1, 2, 4, 5 and 6. Of the stations located in the Piracicaba River, only RD074 has no trend analysis, while RD031 and RD032 presented an inconclusive trend with the possibility of reduction and reduction trend, respectively. Among the two stations, RD031 is closer to the city of Timóteo, in the metropolitan region of Vale do Aço, considered an important pole of the steel industry in the state of Minas Gerais (ECOPLAN-LUME [2010b\)](#page-23-22). Therefore, the maintenance of RD031 in the monitoring network is more relevant than RD032, since together with RD034, it is possible to

evaluate the infuence of the metropolitan region of Vale do Aço on the water quality of the Piracicaba River. The RD074, besides being far from the others, should be kept in the network as a monitoring measure in bedside areas.

In the Manhuaçu River, the RD064 station is in the city of Santana do Manhuaçu. It presented a reduction trend for the WQI and an elevation trend for the variables *E. coli,* nitrate and total solids, which demonstrates the importance of monitoring water quality on site. The RD095 station is located upstream of RD064 and the efuents from most of the municipalities in the region, including the municipality of Manhuaçu, which has the largest population of UGRH6, with 79,574 inhabitants (IBGE [2010\)](#page-24-22). Although the municipality has a 95.0% domestic effluent collection index, the percentage of treatment corresponds to 0.0% (ANA [2017](#page-22-15)). Therefore, the RD064 station has priority over RD095 since it can better evaluate the impact of the discharge of the effluent from the region on the Manhuaçu River.

In cluster 5, there is more than one station allocated in the Suaçuí Grande e Doce rivers. Among the two stations allocated in the Suaçuí Grande River, only RD049 had enough data for the trend analysis, presenting a nonsignifcant result for the WQI. RD049 is located after the cities of Frei Inocêncio and Mathias Lobato, both with an efuent collection and treatment index of 0.0% (ANA [2017\)](#page-22-15). The RD089 station is located at the mouth of the Suaçuí Grande River and no longer receives the contribution of domestic effluents after the monitoring in RD049, receiving only the contribution of small tributaries and the Itambacuri River, which in turn already has the water quality monitoring being performed in the RD088 station. Therefore, station RD049 has priority over RD089.

For the Doce River, only in the RD083 station was not possible to carry out the trend analysis. The RD053 presented a signifcant elevation trend, while stations RD059 and RD067 presented nonsignifcant results for the WQI. In addition to the elevation trend in WQI, RD053 showed a trend to reduce *E. coli* and total phosphorus, which are considered critical in UGRH4 (ECOPLAN-LUME [2010a;](#page-23-0) IGAM [2017b](#page-24-17)). As can be seen in Fig. [6](#page-16-1), station RD053 is allocated between RD045 (cluster 2) and RD058 (cluster 4), which, as discussed above, was recommended to maintain them in the monitoring network. Thus, when analyzing all the information, it can be concluded that stations RD059 and RD067 should have a higher priority in the monitoring network compared to RD053.

Cluster 6 grouped only the RD056 station, which presented the worst WQI values (Fig. [5](#page-14-0)). The station presented a signifcant reduction trend for WQI and dissolved oxygen, in addition to an elevation trend for total phosphorus and nitrate. The station is in the Caratinga River, downstream of Santa Bárbara do Leste, Santa Rita de Minas and Caratinga municipalities, all of which have an effluent collection and treatment index equal to 0.0%. The municipality of Caratinga is still characterized as the most populous in UGRH6, with approximately 85,239 inhabitants (IBGE  $2010$ ) and a release of 3999.3 kg BOD day<sup>-1</sup> (ANA [2017\)](#page-22-15).

The results found for the RD056 station deserve attention of the water resource management agencies, so that management actions can be carried out to improve the water quality. The establishment of new monitoring stations in UGRH5 can also be performed since other areas of the basin may also be showing signs of deterioration of surface water quality. Figure [8](#page-21-0) shows the violation percentage of the framing class according to the limits established by the legislation for the 64 monitoring stations evaluated in the Minas Gerais portion of the Doce River basin.

As shown in Fig. [8](#page-21-0), the RD056 station was the one with the highest violation index of the framing class for the set of water quality variables analyzed, a result that corroborates with the CA, where the station was grouped in the cluster with the worst WQI values.



<span id="page-21-0"></span>**Fig. 8** Violation percentage of the framing class for the 64 monitoring stations according to the limits established by CONAMA Resolution No. 357/2005 and by COPAM/CERH-MG DN No. 01/2008

Stations with high index of violation of variables can be considered as the most relevant in the network since they indicate areas of degradation and require more monitoring. As for the stations with the lowest violation rates, most of them are grouped in cluster 3, which presented the best results for the WQI (Fig. [6](#page-16-1)), thus corroborating with the CA results.

### **4 Conclusions**

The signifcative trend to reduce water quality is typical of metropolitan regions and large cities in the countryside, where water bodies present high degradation of water quality resulted from the discharge of domestic effluents and industrial effluents (point pollution). In addition to the contamination from the point pollution, the water quality deterioration is also afected by the difuse pollution originating from farming areas.

The absence of a signifcative trend is still of concern because can represent a stagnation in violation percentage values and therefore the maintenance of a degradation state in the water bodies.

The increases in the concentrations of turbidity and total solids after the collapse of the dam it was not enough to cause signifcant elevation trends in the historical series of the afected stations.

The CA results allowed the identifcation of the main monitoring stations to be maintained in the network, thus subsidizing management and planning actions to monitor the water quality in the Minas Gerais portion of the Doce River basin. Stations RD044,

RD032, RD095, RD89 and RD053 can be relocated or deactivated. RD056 has the highest degree of pollution, with priority being given to its maintenance in the monitoring area.

The results of the study demonstrate the potential of using methodologies in the temporal and spatial characterization of the stations monitoring data, which may support planning and management actions in the water quality monitoring network of the Doce River basin.

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### **References**

- <span id="page-22-10"></span>Ajorlo, M., Abdullah, R. B., Yusoff, M. K., et al. (2013). Multivariate statistical techniques for the assessment of seasonal variations in surface water quality of pasture ecosystems. *Environmental Monitoring and Assessment, 185,* 8649–8658. [https://doi.org/10.1007/s10661-013-3201-8.](https://doi.org/10.1007/s10661-013-3201-8)
- <span id="page-22-7"></span>Alvares, C. A., Stape, J. L., Sentelhas, P. C., et al. (2013). Köppen's climate classifcation map for Brazil. *Meteorologische Zeitschrift*. <https://doi.org/10.1127/0941-2948/2013/0507>.
- <span id="page-22-2"></span>ANA. (2013a). Cuidando das Águas—Soluções para melhorar a qualidade dos recursos hídricos, 2nd edn. Brasília
- <span id="page-22-6"></span>ANA. (2013b). Base hidrográfca Ottocodifcada da bacia do rio Doce 1:50.000/1.100.000. In: Agência Nac. Águas. [http://metadados.ana.gov.br/geonetwork/srv/pt/main.home.](http://metadados.ana.gov.br/geonetwork/srv/pt/main.home) Accessed 7 Jul 2018
- <span id="page-22-5"></span>ANA. (2016). Encarte Especial sobre a Bacia do Rio Doce—Rompimento da Barragem em Mariana/MG.
- <span id="page-22-15"></span>ANA. (2017). Atlas Esgotos: Despoluição de Bacias Hidrográfcas. Brasília
- <span id="page-22-3"></span>Anand, B., Karunanidhi, D., Subramani, T., et al. (2019). Long-term trend detection and spatiotemporal analysis of groundwater levels using GIS techniques in Lower Bhavani River basin, Tamil Nadu, India. *Environment, Development and Sustainability*. <https://doi.org/10.1007/s10668-019-00318-3>.
- <span id="page-22-9"></span>Anghileri, D., Pianosi, F., & Soncini-Sessa, R. (2014). Trend detection in seasonal data: From hydrology to water resources. *Journal of Hydrology, 511,* 171–179. [https://doi.org/10.1016/J.JHYDR](https://doi.org/10.1016/J.JHYDROL.2014.01.022) [OL.2014.01.022](https://doi.org/10.1016/J.JHYDROL.2014.01.022).
- <span id="page-22-8"></span>APHA, AWWA, WEF. (2012). Standard methods for the examination of water and wastewater, 22nd edn.
- <span id="page-22-4"></span>Ballantine, D. J., & Davies-Colley, R. J. (2014). Water quality trends in New Zealand rivers: 1989– 2009. *Environmental Monitoring and Assessment, 186,* 1939–1950. [https://doi.org/10.1007/s1066](https://doi.org/10.1007/s10661-013-3508-5) [1-013-3508-5.](https://doi.org/10.1007/s10661-013-3508-5)
- <span id="page-22-1"></span>Barakat, A., El Baghdadi, M., Rais, J., et al. (2016). Assessment of spatial and seasonal water quality variation of Oum Er Rbia River (Morocco) using multivariate statistical techniques. *International Soil and Water Conservation Research, 4,* 284–292. [https://doi.org/10.1016/J.ISWCR.2016.11.002.](https://doi.org/10.1016/J.ISWCR.2016.11.002)
- <span id="page-22-14"></span>Brasil. (2000). Resolução nº 274 do Conselho Nacional de Meio Ambiente (CONAMA). Defi ne os critérios de balneabilidade em águas brasileiras.
- <span id="page-22-12"></span>Brasil. (2005). Resolução nº 357 do Conselho Nacional de Meio Ambiente (CONAMA). Classificação dos corpos de água e diretrizes ambientais para o seu enquadramento, bem como estabelece as condições e padrões de lançamento de efuentes
- <span id="page-22-13"></span>Brasil. (2008). Resolução n° 91 do Conselho Nacional de Recursos Hídricos (CNRH). Dispõe sobre procedimentos gerais para o enquadramento dos corpos de água superfciais e subterrâneos.
- <span id="page-22-0"></span>Bu, H., Meng, W., Zhang, Y., & Wan, J. (2014). Relationships between land use patterns and water quality in the Taizi River basin, China. *Ecological Indicators, 41,* 187–197. [https://doi.org/10.1016/J.ECOLI](https://doi.org/10.1016/J.ECOLIND.2014.02.003) [ND.2014.02.003](https://doi.org/10.1016/J.ECOLIND.2014.02.003).
- <span id="page-22-11"></span>Calazans, G. M., Pinto, C. C., da Costa, E. P., et al. (2018a). The use of multivariate statistical methods for optimization of the surface water quality network monitoring in the Paraopeba river basin, Brazil. *Environmental Monitoring and Assessment, 190,* 491. <https://doi.org/10.1007/s10661-018-6873-2>.
- <span id="page-22-16"></span>Calazans, G. M., Pinto, C. C., da Costa, E. P., et al. (2018b). Using multivariate techniques as a strategy to guide optimization projects for the surface water quality network monitoring in the Velhas river basin,Brazil. *Environmental Monitoring and Assessment, 190,* 726. [https://doi.org/10.1007/s1066](https://doi.org/10.1007/s10661-018-7099-z) [1-018-7099-z](https://doi.org/10.1007/s10661-018-7099-z).
- <span id="page-23-16"></span>Calijuri, M. L., de CastroCosta, J. S. L. S., et al. (2015). Impact of land use/land cover changes on water quality and hydrological behavior of an agricultural subwatershed. *Environmental Earth Sciences, 74,* 5373–5382. <https://doi.org/10.1007/s12665-015-4550-0>.
- <span id="page-23-7"></span>CBH-Doce. (2016a). Deliberação Normativa CBH-Doce nº 51/2016
- <span id="page-23-8"></span>CBH-Doce. (2016b). A bacia do rio Doce. <http://www.cbhdoce.org.br/institucional/a-bacia>
- <span id="page-23-14"></span>Chen, N., Valdes, D., Marlin, C., et al. (2019). Water, nitrate and atrazine transfer through the unsaturated zone of the Chalk aquifer in northern France. *Science of the Total Environment, 652,* 927–938. [https://](https://doi.org/10.1016/J.SCITOTENV.2018.10.286) [doi.org/10.1016/J.SCITOTENV.2018.10.286.](https://doi.org/10.1016/J.SCITOTENV.2018.10.286)
- <span id="page-23-2"></span>Chowdhury, S., & Al-Zahrani, M. (2014). Water quality change in dam reservoir and shallow aquifer: Analysis on trend, seasonal variability and data reduction. *Environmental Monitoring and Assessment, 186,* 6127–6143. <https://doi.org/10.1007/s10661-014-3844-0>.
- <span id="page-23-4"></span>Costa, E. P., Pinto, C. C., Soares, A. L. C., et al. (2017). Evaluation of violations in water quality standards in the monitoring network of São Francisco River basin, the third largest in Brazil. *Environmental Monitoring and Assessment, 189,* 590. [https://doi.org/10.1007/s10661-017-6266-y.](https://doi.org/10.1007/s10661-017-6266-y)
- <span id="page-23-15"></span>Dhanasekarapandian, M., Chandran, S., Devi, D. S., & Kumar, V. (2016). Spatial and temporal variation of groundwater quality and its suitability for irrigation and drinking purpose using GIS and WQI in an urban fringe. *Journal of African Earth Sciences, 124,* 270–288. [https://doi.org/10.1016/j.jafrearsci](https://doi.org/10.1016/j.jafrearsci.2016.08.015) [.2016.08.015.](https://doi.org/10.1016/j.jafrearsci.2016.08.015)
- <span id="page-23-9"></span>Diamantini, E., Lutz, S. R., Mallucci, S., et al. (2018). Driver detection of water quality trends in three large European river basins. *Science of the Total Environment, 612,* 49–62. [https://doi.org/10.1016/J.SCITO](https://doi.org/10.1016/J.SCITOTENV.2017.08.172) [TENV.2017.08.172](https://doi.org/10.1016/J.SCITOTENV.2017.08.172).
- <span id="page-23-6"></span>do Carmo, F. F., Kamino, L. H. Y., Junior, R. T., et al. (2017). Fundão tailings dam failures: The environment tragedy of the largest technological disaster of Brazilian mining in global context. *Perspectives in Ecology and Conservation, 15,* 145–151.
- <span id="page-23-3"></span>Dou, M., Zhang, Y., & Li, G. (2016). Temporal and spatial characteristics of the water pollutant concentration in Huaihe River Basin from 2003 to 2012, China. *Environmental Monitoring and Assessment*. <https://doi.org/10.1007/s10661-016-5503-0>.
- <span id="page-23-5"></span>Ebadati, N., & Hooshmandzadeh, M. (2019). Water quality assessment of river using RBF and MLP methods of artifcial network analysis (case study: Karoon River Southwest of Iran). *Environmental Earth Sciences, 78,* 551. [https://doi.org/10.1007/s12665-019-8472-0.](https://doi.org/10.1007/s12665-019-8472-0)
- <span id="page-23-0"></span>ECOPLAN-LUME. (2010a). Plano Integrado de Recursos Hídricos da Bacia Hidrográfca do Rio Doce - Volume I
- <span id="page-23-22"></span>ECOPLAN-LUME. (2010b). Plano de Ação de Recuros Hídricos da Unidade de Planejamento e Gestão dos Recursos Hídricos - Piracicaba
- <span id="page-23-11"></span>Edberg, S. C., Rice, E. W., Karlin, R. J., & Allen, M. J. (2000). *Escherichia coli*: the best biological drinking water indicator for public health protection. *Journal of Applied Microbiology, 88,* 106S-116S. [https://](https://doi.org/10.1111/j.1365-2672.2000.tb05338.x) [doi.org/10.1111/j.1365-2672.2000.tb05338.x](https://doi.org/10.1111/j.1365-2672.2000.tb05338.x).
- <span id="page-23-19"></span>Efendi, H., & Romanto, W. Y. (2015). Water quality status of Ciambulawung River, Banten Province, based on pollution Index and NSF-WQI. *Procedia Environmental Sciences, 24,* 228–237. [https://doi.](https://doi.org/10.1016/j.proenv.2015.03.030) [org/10.1016/j.proenv.2015.03.030](https://doi.org/10.1016/j.proenv.2015.03.030).
- <span id="page-23-1"></span>Elçi, Ş, & Selçuk, P. (2013). Efects of basin activities and land use on water quality trends in Tahtali Basin, Turkey. *Environmental Earth Sciences, 68,* 1591–1598. <https://doi.org/10.1007/s12665-012-1852-3>.
- <span id="page-23-21"></span>Fernandes, G. W., Goulart, F. F., Ranieri, B. D., et al. (2016). Deep into the mud: Ecological and socioeconomic impacts of the dam breach in Mariana, Brazil. *Natureza and Conservação, 14,* 35–45.
- <span id="page-23-17"></span>Fraga, M. S., da Silva, D. D., Elesbon, A. A. A., & Guedes, H. A. S. (2019). Methodological proposal for the allocation of water quality monitoring stations using strategic decision analysis. *Environmental Monitoring and Assessment, 191,* 776. <https://doi.org/10.1007/s10661-019-7974-2>.
- <span id="page-23-13"></span>Gautheir, T. D. (2001). Detecting trends using Spearman's Rank correlation coefficient. *Environmental Forensics*. [https://doi.org/10.1006/enfo.2001.0061.](https://doi.org/10.1006/enfo.2001.0061)
- <span id="page-23-18"></span>Hachich, E. M., Di Bari, M., Christ, A. P. G., et al. (2012). Comparison of thermotolerant coliforms and *Escherichia coli* densities in freshwater bodies. *Brazilian Journal of Microbiology, 43,* 675–681. [https](https://doi.org/10.1590/S1517-83822012000200032) [://doi.org/10.1590/S1517-83822012000200032.](https://doi.org/10.1590/S1517-83822012000200032)
- <span id="page-23-20"></span>Hatje, V., Pedreira, R. M. A., de Rezende, C. E., et al. (2017). The environmental impacts of one of the largest tailing dam failures worldwide. *Scientifc Reports*. <https://doi.org/10.1038/s41598-017-11143-x>.
- <span id="page-23-12"></span>Hirsch, R. M., Slack, J. R., & Smith, R. A. (1982). Techniques of trend analysis for monthly water-quality data. *Water Resources Research, 18,* 107–121. [https://doi.org/10.1029/WR018i001p00107.](https://doi.org/10.1029/WR018i001p00107)
- <span id="page-23-10"></span>Howden, N. J. K., Burt, T. P., Worrall, F., & Whelan, M. J. (2011). Monitoring fuvial water chemistry for trend detection: Hydrological variability masks trends in datasets covering fewer than 12 years. *Journal of Environmental Monitoring, 13,* 514–521. <https://doi.org/10.1039/c0em00722f>.
- <span id="page-24-22"></span>IBGE. (2010). Censo demográfco 2010. In: Inst. Bras. Geogr. e Estatística. [https://censo2010.ibge.gov.](https://censo2010.ibge.gov.br/resultados.html) [br/resultados.html](https://censo2010.ibge.gov.br/resultados.html)
- <span id="page-24-18"></span>IGAM. (2015). Monitoramento da qualidade das águas superfciais do rio Doce no estado de Minas Gerais. Belo Horizonte
- <span id="page-24-4"></span>IGAM. (2016). Qualidade das águas superfciais de Minas Gerais em 2016. Belo Horizonte
- <span id="page-24-3"></span>IGAM. (2017a). Encarte especial sobre a qualidade das águas do rio Doce após 2 anos do rompimento de barragem de Fundão - 2015/2017. Belo Horizonte
- <span id="page-24-17"></span>IGAM. (2017b). Relatório de Monitoramento das Águas Superfciais nas Bacias Hidrográfcas de Minas Gerais em 2016: Projeto: Sistema de Monitoramento da Qualidade das Águas Superfciais do Estado de Minas Gerais - Águas de Minas. Belo Horizonte
- <span id="page-24-12"></span>Jung, K. Y., Lee, K.-L., Im, T. H., et al. (2016). Evaluation of water quality for the Nakdong River watershed using multivariate analysis. *Environmental Technology and Innovation, 5,* 67–82. [https://doi.](https://doi.org/10.1016/J.ETI.2015.12.001) [org/10.1016/J.ETI.2015.12.001](https://doi.org/10.1016/J.ETI.2015.12.001).
- <span id="page-24-16"></span>Kamal, N. A., Muhammad, N. S., & Abdullah, J. (2020). Scenario-based pollution discharge simulations and mapping using integrated QUAL2K-GIS. *Environmental Pollution, 259,* 113909. [https://doi.](https://doi.org/10.1016/j.envpol.2020.113909) [org/10.1016/j.envpol.2020.113909](https://doi.org/10.1016/j.envpol.2020.113909).
- <span id="page-24-10"></span>Kendall, M. G. (1975). *Rank correlation methods, 4*<sup>a</sup> . Londres: Charles Grifn.
- <span id="page-24-0"></span>Kisi, O., & Ay, M. (2014). Comparison of Mann–Kendall and innovative trend method for water quality parameters of the Kizilirmak River, Turkey. *Journal of Hydrology, 513,* 362–375. [https://doi.](https://doi.org/10.1016/j.jhydrol.2014.03.005) [org/10.1016/j.jhydrol.2014.03.005.](https://doi.org/10.1016/j.jhydrol.2014.03.005)
- <span id="page-24-11"></span>Kruskal, W. H., & Wallis, W. A. (1952). Use of ranks in one-criterion variance analysis. *Journal of American Statistical Association, 47,* 583–621.
- <span id="page-24-1"></span>Kurdi, M., Eslamkish, T., Seyedali, M., & Ferdows, M. S. (2015). Water quality evaluation and trend analysis in the Qareh Sou Basin,Iran. *Environmental Earth Sciences, 73,* 8167–8175. [https://doi.](https://doi.org/10.1007/s12665-014-3975-1) [org/10.1007/s12665-014-3975-1](https://doi.org/10.1007/s12665-014-3975-1).
- <span id="page-24-8"></span>Leclerc, H., Mossel, D. A. A., Edberg, S. C., & Struijk, C. B. (2001). Advances in the bacteriology of the coliform group: Their suitability as markers of microbial water safety. *Annual Review of Microbiology, 55,* 201–234.<https://doi.org/10.1146/annurev.micro.55.1.201>.
- <span id="page-24-5"></span>Lima, R. P. C., Da Silva, D. D., Pereira, S. B., et al. (2019). Development of an annual drought classifcation system based on drought severity indexes. *Anais da Academia Brasileira de Ciências*. [https](https://doi.org/10.1590/0001-3765201920180188) [://doi.org/10.1590/0001-3765201920180188.](https://doi.org/10.1590/0001-3765201920180188)
- <span id="page-24-21"></span>Losekann, C., Dias, T. H., & Camargo, A. V. M. (2020). The Rio Doce mining disaster: Legal framing in the Brazilian justice system. *The Extractive Industries and Society, 7,* 199–208. [https://doi.](https://doi.org/10.1016/j.exis.2019.11.015) [org/10.1016/j.exis.2019.11.015](https://doi.org/10.1016/j.exis.2019.11.015).
- <span id="page-24-9"></span>Mann, H. B. (1945). Nonparametric tests against trend. *Econometrica, 13,* 245–259.
- <span id="page-24-15"></span>MapBiomas. (2019). Coleção 4.1 da Série Anual de Mapas de Cobertura e Uso de Solo do Brasil. [https](https://mapbiomas.org/) [://mapbiomas.org/](https://mapbiomas.org/). Accessed 1 Apr 2020
- <span id="page-24-7"></span>Marmontel, C. V. F., Lucas-Borja, M. E., Rodrigues, V. A., & Zema, D. A. (2018). Efects of land use and sampling distance on water quality in tropical headwater springs (Pimenta creek, São Paulo State, Brazil). *Science of the Total Environment, 622–623,* 690–701. [https://doi.org/10.1016/j.scito](https://doi.org/10.1016/j.scitotenv.2017.12.011) [tenv.2017.12.011.](https://doi.org/10.1016/j.scitotenv.2017.12.011)
- <span id="page-24-14"></span>Minas Gerais. (2008). Deliberação Normativa Conjunta COPAM/CERH-MG nº 01. Dispõe sobre a classifcação dos corpos de água e diretrizes ambientais para o seu enquadramento, bem como estabelece as condições e padrões de lançamento de efuentes, e dá outras providências.
- <span id="page-24-13"></span>Muangthong, S., & Shrestha, S. (2015). Assessment of surface water quality using multivariate statistical techniques: Case study of the Nampong River and Songkhram River, Thailand. *Environmental Monitoring and Assessment, 187,* 548. [https://doi.org/10.1007/s10661-015-4774-1.](https://doi.org/10.1007/s10661-015-4774-1)
- <span id="page-24-20"></span>Neves, A. C. O., Nunes, F. P., de Carvalho, F. A., & Fernandes, G. W. (2016). Neglect of ecosystems services by mining, and the worst environmental disaster in Brazil. *Natureza and Conserva, 14,* 24–27. <https://doi.org/10.1016/j.ncon.2016.03.002>.
- <span id="page-24-19"></span>Oliveira, K. S. S., & Silva da Quaresma, V. (2017). Temporal variability in the suspended sediment load and streamfow of the Doce River. *Journal of South American Earth Sciences, 78,* 101–115. [https://](https://doi.org/10.1016/J.JSAMES.2017.06.009) [doi.org/10.1016/J.JSAMES.2017.06.009](https://doi.org/10.1016/J.JSAMES.2017.06.009).
- <span id="page-24-6"></span>Oliveira, L. M., Maillard, P., & de Andrade Pinto, E. J. (2017a). Application of a land cover pollution index to model non-point pollution sources in a Brazilian watershed. *CATENA, 150,* 124–132. [https](https://doi.org/10.1016/j.catena.2016.11.015) [://doi.org/10.1016/j.catena.2016.11.015.](https://doi.org/10.1016/j.catena.2016.11.015)
- <span id="page-24-2"></span>Oliveira, S. C., Amaral, R. C., Almeida, K. C. B., & Pinto, C. C. (2017b). Qualidade das águas superfciais do Médio São Francisco após a implantação dos perímetros irrigados de Gorutuba/Lagoa Grande e Jaíba. *Eng Sanit e Ambient*. <https://doi.org/10.1590/s1413-41522017136784>.
- <span id="page-25-2"></span>Pesce, S. F., & Wunderlin, D. A. (2000). Use of water quality indices to verify the impact of Cordoba City (Argentina) on Suquia River. *Water Research, 34,* 2915–2926. [https://doi.org/10.1016/S0043](https://doi.org/10.1016/S0043-1354(00)00036-1) [-1354\(00\)00036-1](https://doi.org/10.1016/S0043-1354(00)00036-1).
- <span id="page-25-14"></span>Pinto, C. C., Andrade, S. B., Pinto, É. A., & Oliveira, S. M. A. C. (2017). Análise de tendência de concentrações e cargas de parâmetros físicos, químicos e biológicos da bacia do rio Verde. *Revista Brasileira de Recursos Hídricos*. [https://doi.org/10.1590/2318-0331.0117160030.](https://doi.org/10.1590/2318-0331.0117160030)
- <span id="page-25-15"></span>Rodrigues, A. L. M., Reis, G. B., dos Santos, M. T., et al. (2019). Infuence of land use and land cover's change on the hydrological regime at a Brazilian southeast urbanized watershed. *Environmental Earth Sciences, 78,* 1–13. [https://doi.org/10.1007/s12665-019-8601-9.](https://doi.org/10.1007/s12665-019-8601-9)
- <span id="page-25-17"></span>Şener, Ş, Şener, E., & Davraz, A. (2017). Evaluation of water quality using water quality index (WQI) method and GIS in Aksu River (SW-Turkey). *Science of the Total Environment, 584–585,* 131–144. <https://doi.org/10.1016/j.scitotenv.2017.01.102>.
- <span id="page-25-4"></span>Shrestha, S., & Kazama, F. (2007). Assessment of surface water quality using multivariate statistical techniques: A case study of the Fuji river basin, Japan. *Environmental Modelling and Software, 22,* 464– 475. [https://doi.org/10.1016/j.envsoft.2006.02.001.](https://doi.org/10.1016/j.envsoft.2006.02.001)
- <span id="page-25-9"></span>Shukla, S., Gedam, S., & Khire, M. V. (2020). Implications of demographic changes and land transformations on surface water quality of rural and urban subbasins of Upper Bhima River basin, Maharashtra, India. *Environment, Development and Sustainability, 22,* 129–171. [https://doi.org/10.1007/s1066](https://doi.org/10.1007/s10668-018-0187-y) [8-018-0187-y](https://doi.org/10.1007/s10668-018-0187-y).
- <span id="page-25-3"></span>Simeonov, V., Stratis, J. A., Samara, C., et al. (2003). Assessment of the surface water quality in Northern Greece. *Water Research, 37,* 4119–4124. [https://doi.org/10.1016/S0043-1354\(03\)00398-1](https://doi.org/10.1016/S0043-1354(03)00398-1).
- <span id="page-25-13"></span>Sorando, R., Comín, F. A., Jiménez, J. J., et al. (2019). Water resources and nitrate discharges in relation to agricultural land uses in an intensively irrigated watershed. *Science of the Total Environment, 659,* 1293–1306. [https://doi.org/10.1016/J.SCITOTENV.2018.12.023.](https://doi.org/10.1016/J.SCITOTENV.2018.12.023)
- <span id="page-25-8"></span>Sun, C. C., Shen, Z. Y., Xiong, M., et al. (2013). Trend of dissolved inorganic nitrogen at stations downstream from the Three-Gorges Dam of Yangtze River. *Environmental Pollution, 180,* 13–18. [https://](https://doi.org/10.1016/J.ENVPOL.2013.05.003) [doi.org/10.1016/J.ENVPOL.2013.05.003.](https://doi.org/10.1016/J.ENVPOL.2013.05.003)
- <span id="page-25-7"></span>Tabari, H., Marof, S., & Ahmadi, M. (2011). Long-term variations of water quality parameters in the Maroon River, Iran. *Environmental Monitoring and Assessment, 177,* 273–287. [https://doi.](https://doi.org/10.1007/s10661-010-1633-y) [org/10.1007/s10661-010-1633-y](https://doi.org/10.1007/s10661-010-1633-y).
- <span id="page-25-5"></span>Trindade, A. L. C., Almeida, K. C. B., Barbosa, P. E., & Oliveira, S. M. A. C. (2017). Tendências temporais e espaciais da qualidade das águas superfciais da sub-bacia do Rio das Velhas, estado de Minas Gerais. *Engenharia Sanitária e Ambiental, 22,* 13–24. [https://doi.org/10.1590/s1413-41522016131457.](https://doi.org/10.1590/s1413-41522016131457)
- <span id="page-25-18"></span>Wu, Z., Wang, X., Chen, Y., et al. (2018). Assessing river water quality using water quality index in Lake Taihu Basin, China. *Science of the Total Environment, 612,* 914–922. [https://doi.org/10.1016/j.scito](https://doi.org/10.1016/j.scitotenv.2017.08.293) [tenv.2017.08.293.](https://doi.org/10.1016/j.scitotenv.2017.08.293)
- <span id="page-25-6"></span>Yenilmez, F., Keskin, F., & Aksoy, A. (2011). Water quality trend analysis in Eymir Lake, Ankara. *Physics and Chemistry of the Earth, 36,* 135–140. [https://doi.org/10.1016/j.pce.2010.05.005.](https://doi.org/10.1016/j.pce.2010.05.005)
- <span id="page-25-16"></span>Yidana, S. M., Banoeng-Yakubo, B., & Akabzaa, T. M. (2010). Analysis of groundwater quality using multivariate and spatial analyses in the Keta basin, Ghana. *Journal of African Earth Sciences, 58,* 220– 234. <https://doi.org/10.1016/j.jafrearsci.2010.03.003>.
- <span id="page-25-11"></span>Yue, S., Pilon, P., & Cavadias, G. (2002). Power of the Mann-Kendall and Spearman's rho tests for detecting monotonic trends in hydrological series. *Journal of Hydrology, 259,* 254–271. [https://doi.org/10.1016/](https://doi.org/10.1016/S0022-1694(01)00594-7) [S0022-1694\(01\)00594-7.](https://doi.org/10.1016/S0022-1694(01)00594-7)
- <span id="page-25-0"></span>Zeinalzadeh, K., & Rezaei, E. (2017). Determining spatial and temporal changes of surface water quality using principal component analysis. *Journal of Hydrology: Regional Studies, 13,* 1–10. [https://doi.](https://doi.org/10.1016/j.ejrh.2017.07.002) [org/10.1016/j.ejrh.2017.07.002](https://doi.org/10.1016/j.ejrh.2017.07.002).
- <span id="page-25-1"></span>Zhai, X., Xia, J., & Zhang, Y. (2014). Water quality variation in the highly disturbed Huai River Basin, China from 1994 to 2005 by multi-statistical analyses. *Science of the Total Environment, 496,* 594– 606. [https://doi.org/10.1016/J.SCITOTENV.2014.06.101.](https://doi.org/10.1016/J.SCITOTENV.2014.06.101)
- <span id="page-25-10"></span>Zhang, J., Li, S., & Jiang, C. (2020). Efects of land use on water quality in a river basin (Daning) of the Three Gorges Reservoir Area, China: Watershed versus riparian zone. *Ecological Indicators, 113,* 106226. <https://doi.org/10.1016/j.ecolind.2020.106226>.
- <span id="page-25-12"></span>Zhang, X., Wang, Q., Liu, Y., et al. (2011). Application of multivariate statistical techniques in the assessment of water quality in the Southwest New Territories and Kowloon, Hong Kong. *Environmental Monitoring and Assessment, 173,* 17–27. [https://doi.org/10.1007/s10661-010-1366-y.](https://doi.org/10.1007/s10661-010-1366-y)

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