GIS-based assessment of groundwater quality for drinking and irrigation purposes in central Iraq

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Abstract In many parts of the world, groundwater is considered to be a key source of fresh water for both the domestic and non-domestic sectors. Where groundwater extraction is implemented, systems to monitor water quality must ensure a safe and sustainable supply. Over the years, Iraq has suffered from surface water quality and supply problems, necessitating groundwater extraction in many regions. This study investigates groundwater quality in a region of central Iraq around Babylon city, covering an area of 5119 km². The data gathered for this study included maps, well locations and water quality data and was sourced from the relevant governmental departments. A base map of the focussed region

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was initially prepared following data collection. The analysed water quality parameters were used as an attribute database to produce thematic maps using a geographical information system (GIS) environment. In this paper, the water quality index (WQI) and the irrigation water quality index (IWQI) were calculated for different groundwater samples using various parameters including the Electrical Conductivity (EC), Cl⁻, HCO3⁻, Na⁺ and pH. Moreover, the groundwater suitability for irrigation purposes has been assessed using indices such as Kelly's ratio (KR), sodium absorption ratio (SAR), residual sodium carbonate (RSC), soluble sodium percentage (SSP) and permeability index (PI). Water quality index maps have been developed using the GIS environment. The obtained results reveal that the groundwater in the study location requires specific treatments to be usable.

Keywords Groundwater quality · Geographic information system · Remote sensing water quality index and irrigation water quality index

Introduction

The availability of useable water is a major global challenge and particularly prevalent in developing countries. Population growth and an increase in industrial activities raise water demand. Simultaneously, these continuous activities increase the mass of wastewater returned to the environment, resulting in the pollution of water bodies and useable water scarcity (Al-Jubouri and



Holmes (2020); Mohammed et al. (2007)). Surface water is strongly connected to groundwater; therefore, any changes to the surface water directly affect groundwater (Bachmann et al., 2019). Moreover, there is an extremely complex relation between rainfall, surface water and water supply contamination. An additional impacting factor is the decrease in river flow due to reduced rainfall which affects ability to dilute effluents and increase pathogen or chemical loading. The water scarcity problem can be managed by treating the contaminated effluents before releasing to the environment in order to conserve the available water bodies. At the same time, alternative water resources could also be sought. The pollutants present in wastewaters have to be identified to be effectively treated and removed. The pollutant type and concentration are the most critical factors determining the treatment method and the process's cost and performance.

Given the issues mentioned above, increasing accessibility to alternative water resources can significantly alleviate water scarcity (Nelly and Mutua, 2016). In Iraq, groundwater resources are gaining interest as an alternative source of water supply due to the prevalence of serious water quality issues. According to Alanbari et al. (2015), this issue will become more severe in the coming years due to a projected drop in the national water supply: expected to be 17.61 BCM (billion cubic meters) in 2025 (whereas currently, demand is roughly approximated to be among 66.8 to 77 BCM). Alanbari et al. (2015) state that the discharges of the Tigris and Euphrates, the main rivers in Iraq, will continue to decrease with time, and both may be entirely dry by 2040. Therefore, there is a real need to use groundwater as an alternative source of water supply. Many countries worldwide have used groundwater as an alternative source for water supply for various reasons, primarily a lack of suitable surface water. For instance, groundwater has been extracted in Europe since at least 1900 for public water supply, industrial use and irrigation (Zektser and Everett, 2004). The total demand for these three uses increased from 40 km³/year in 1900 to 720 km^3 /year in 2000, with demand more than doubling since 1970, (Zektser and Everett, 2004).

The availability of groundwater depends on the topography, surface drainage, geology, slope and land cover. The main factors that influence the water table's elevation determination are the slope and the topographic elevation. Similarly, the drainage pattern has a significant effect as it determines the precipitation rate, which may infiltrate into the ground (Nelly and Mutua, 2016). Rainfall has a vital role in controlling the amount and distribution of groundwater and the permeability of the ground surface. At the sub-surface level, infiltration and retention of the groundwater are heavily influenced by the permeability of the surrounding rock and soil type.

In Iraq, groundwater is becoming essential for potable water supply and irrigation use. Groundwater is also being used to help to maintain the base flow of rivers. The quality of groundwater can be high in many locations due to natural filtration. Groundwater is often colourless, transparent and free from microbial contamination, thus requiring minimal treatment (Singha et al., 2015). Contamination of groundwater results from natural or anthropogenic causes, and it is crucial to highlight that when groundwater is polluted, the water quality may not easily recover when sources of pollutants are stopped.

In some cases, groundwater pollution has been observed in several wells due to the absence of proper sewerage infrastructure (Iraq UN, 2013). Nowadays, pollutant threats have increased due to the number of soluble chemicals resulting from urbanisation, modern agricultural practices and industrial activities. The groundwater properties can be easily affected by the atmosphere, soil, water-rock reactions and external pollutant sources. Previous studies in developing countries have shown that a high percentage of disease is directly related to the low quality of drinking water (Singha et al., 2017), this reveals one of the main motivations of this work.

The present study uses the GIS software ArcGIS 10.5 (ESRI, 2015) to analyse groundwater quality and develop a spatial mapping of the associated data. ArcGIS has been used in other developing countries as a powerful tool for water resourcing and management (Singha et al., 2015). The developed maps are intended to aid decision making on groundwater management at local, regional and national levels. The study focuses on the central Iraqi region of the Babylon Province, a crucial agricultural area for the country, which hosts several population centres. The investigation identifies the groundwater's chemical and microbiological properties and then assesses that water's suitability for various purposes such as drinking and irrigation using different indices. The selected indices used for evaluation of water quality are based on the local government authorities' recommendations. Additionally, the water areas are classified using GIS technology to obtain coloured maps for the water quality indices, enabling the proposal of solutions to treat wastewater efficiently and identification of alternative water consumption sources in the future.

Study area

Babylon (known locally as Babil) Province is situated in central Iraq (the mid-Euphrates Region), approximately 100 km south of Baghdad. The longitudes and the latitudes of the province are located between (44° 2' 42.245" E-45° 2' 2.964" E) and (32° 25' 55.287" N-33° 7' 34.229" N), respectively. The province covers an area of 5119 km². The Euphrates River flows for about 52 km through the study area, 34 km being upstream of the Hindyia barrage which controls the flow into downstream watercourses. Meanwhile, the Tigris River flows for about 112 km through the area, and 20 km of this distance is located east of Babylon City. Not far from the world heritage site of Babylon's ancient city, Al-Hilla is the largest urban centre in the region and is bounded by a group of urban centres (Fig. 1). The population of Babylon province was estimated to be 1,728,000 in 2015. In line with the rest of Iraq, the province is growing with an estimated growth rate of 2.7 (Iraq National Population Commission, 2012). One of the main economic activities for the province is agriculture, and the province is a key national resource in this sector. The regional economy's growth faces many challenges, most notably the continuing occurrence of droughts (IAU, 2010).

In the current study, various locations were selected to evaluate the groundwater quality across the province, and a defined number of wells were selected for investigation. The study area is composed of 4.4% of the Mesopotamia plain, covering about 116,000 km² (Yacoub, 2011). The general topography of this area is characterised by its insignificant gradient from the northwest to the southeast. The highest point is ~62 m above sea level and is located in the northern part, while the lowest point is ~21 m above sea level located in the southern part (Al-Madhlom et al., 2016). In general, the soil in this area has the nature of alluvial, fluvial silty clayey loam consistent with the soil of the whole of the Mesopotamia Plain, all derived from the fluvial depositions of the Euphrates and Tigris Rivers (Yacoub, 2011).

It is essential to highlight that investigating climate change is critical in studies related to shallow groundwater. Different climate elements play a key role in the rainfall rate and humidity that affect the soil's water content. Climate change features such as temperature rise, low rainfall, and dust storm generation significantly affect groundwater quality. The chosen area's climate is characterised by high temperature in summer and a short cold winter period. From November to April, the period is considered



Fig. 1 Location of the study area

the wet season because 90% of the annual rainfall can occur with typical annual precipitation of 175 mm. The annual evaporation rate may reach 2900 mm/year in the Mesopotamia Plain (Yacoub, 2011). The study area's continental climate (desert climate) increases salts' concentration in the water (Madhloom and Al-Ansari, 2018). The study area's location is influenced by the regional tectonic actions that have formed the sedimentary plain's concave fold. Formation of the plane has continued with successive river sediment deposition over the millennia. The sediments in the study area consist of quaternary sediments from the Pliocene to Miocene age. The sediments are characterised by flood plains accumulating in thin layers. Meanwhile, the groundwater movement directions are influenced by several factors, including the topographic setting, precipitation, soil permeability, evapotranspiration, and recharge condition (Buday and Jassim, 1980).

This study is the first of its kind to consider this critical region of Iraq as a whole. Since the agricultural sector is pivotal to the regional economy, water shortage and water quality have a significant impact on both the economy and the population, and in turn, this can extend to the national level. Therefore, finding alternative means to irrigate crops and provide potable water is needed to mitigate the ongoing issue of water scarcity. The flow chart in Fig. 2 outlines the adopted methodology in this study.



Fig. 2 Flow chart outlining adopted methodology

Methodology

Sampling and analysis

The groundwater sampled during the pre-monsoon season of 2015 from 49 selected wells distributed across the province was evaluated. The data domain selection was in line with the approach adopted by previous researchers such as Singha and Pasupuleti (2020) and Alam et al. (2020). The well parameters (e.g. well locations, depths, drilling diameter and drilling purpose) are collected from the Water Resources Ministry in Iraq (Table 1). A spatial variation of the study area's groundwater quality was carried using spatial analysis via ArcGIS. The spatial distribution of the 49 wells is shown in Fig. 3. The Water Resources Ministry in Iraq undertook the laboratory testing of the chemical and microbiological properties. The collection, preservation and chemical analysis was conducted following the American Public Health Association guidelines (APHA, 1998), and see Table 2 for the parameter test results.

Water quality index (WQI)

The WQI is an arithmetical tool used to transform large quantities of water quality data into a single cumulatively derived number and can be utilised to assess and manage groundwater quality (Reyes-Toscano et al., 2020). The WQI was primarily proposed by Brown et al. (1972) and then modified by Backman et al. (1998). The primary outcome of calculating the water quality indices (WQIs) is assessing freshwater suitability for different uses. WQI is mainly based on comparing the measured water quality parameters obtained from the field with regulatory standards (Husain, 1998). However, the WQI cannot be used to substitute the required thorough analysis needed for environmental monitoring and modelling. The advantages of these indices are their ability to represent various variable measurements in a single number. Further, the indices can combine various measurements with various measurement units in a single unit, thus simplifying the results (Zandbergen and Hall, 1998). There are many approaches for quantifying the WQI, such as the formulae, which could be either independent or dependent on water quality standards (Khan et al., 2003).

The evaluation of drinking water quality usually adopts the WQI approach as a reliable measurement and groundwater quality assessment. According to the World Health Organization (WHO) report published in 2004, the WQI illustrates the combined qualitative and measured parameters of the drinking water quality in relation to the WHO recommendations (World Health Organization, 2004). As a result, the WQI can be considered a valuable tool in assessing groundwater quality used for drinking. As this study is conducted in Iraq, it was necessary to review the Iraqi provisions for drinking water and the qualitative parameters' standards (Madhloom and Al-Ansari, 2018). The standard values of most of the parameters have maximum and minimum numbers which are very close to the WHO standards; hence, the WHO standards are applied in this work.

Herein the WQI is obtained following the weighted arithmetic index method (Cude, 2001). The method multiplies the water quality components by a weighting factor and then sums the results using the simple arithmetic mean. To assess the quality of groundwater, the quality rating scale (Q_i) for each parameter is firstly estimated using Eq. 1:

$$Q_{i} = \left[\frac{(V_{\text{actual}} - V_{\text{ideal}})}{(V_{\text{standard}} - V_{\text{ideal}})} \times 100\right]$$
(1)

where Q_i = the quality rating of the *i*th parameter for the total of n water quality parameters,

Vactual	the actual value of the parameter which could
	be achieved from the laboratory test
Videal	the ideal value of the same parameter which
	is gained from a standard table,
Videal	is equal to 7 for the parameter pH, but for
	other parameters it becomes zero.
V _{standard}	the recommended value of the parameter and
	is presented in Table 3.

To determine WQI, the relative weight (Wi) is estimated by a value inversely proportional to the recommended standard (Si) for the corresponding parameter using Eq. 2:

$$Wi = \frac{1}{Si}$$
(2)

where Wi = relative weight of *n*th parameter. Si = standard permissible value of the nth parameter.

Then, the overall WQI is calculated by linear combining of the quality rating with the relative weight according to Eq. 3;

Table 1 Sampling data for the selected well

Well number	Longitude	Latitude	Elevation	Purpose of drilling	Well depth (m)	Drilling diameter (in)
w1	44° 20′ 39″	32° 27′ 57″	15	Private benefit	6	6
w2	44° 25′ 05″	32° 32′ 02″	19	Public benefit	30	5
w3	44° 24′ 49″	32° 32′ 20″	21	Public benefit	12	10
w4	44° 22′ 18″	32° 12′ 32″	15	Public benefit	6	16
w5	44° 25′ 41.0″	32° 15′ 01″	12	Public benefit	6	16
w6	44° 24′ 46.3″	32° 26′ 37.4″	16	Public benefit	-	-
w7	44° 24′ 02.6″	32° 23′ 00″	14	Public benefit	12	12
w8	44° 48′ 34″	32° 26′ 19″	18.4	Public benefit	12	15
w9	45° 01' 18"	32° 20′ 31″	7	Public benefit	12	15
w10	44° 26′ 52″	32° 13′ 10″	14	Public benefit	12	10
w11	44° 44′ 45″	33° 08′ 11″	11	Public benefit	12	15
w12	44° 41' 38"	32° 14′ 46″	17	Public benefit	12	15
w13	44° 36' 44″	32° 11′ 38″	11	Public benefit	12	15
w14	44° 39' 22"	32° 12′ 44″	11	Public benefit	12	15
w15	44° 34' 28"	32° 36' 04"	13	Public benefit	12	15
w16	44° 25′ 47″	32° 40′ 23″		Public benefit	12	15
w17	44° 32' 22"	32° 32′ 41″	10	Public benefit	?	15
w18	44° 42′ 12″	32° 33′ 42″	5	Public benefit	12	15
w19	44° 23′ 47″	32° 38′ 41″	14	Public benefit	12	15
w20	44° 23′ 19″	32° 38′ 27″	13	Public benefit	12	15
w21	44° 25′ 06″	32° 27′ 04″	15	Public benefit	-	-
w22	44° 35′ 49″	32° 14′ 14″	17	Public benefit	12	15
w23	44° 29′ 20″	32° 14′ 34″	11	Public benefit	12	15
w24	44° 27′ 19″	32° 17′ 24″	15	Public benefit	-	-
w25	44° 21′ 48″	32° 22′ 44″	25	Public benefit	12	15
w26	44° 23′ 05″	32° 20′ 30″	16	Public benefit	12	15
w27	44° 18′ 34″	32° 43′ 05″	27	Public benefit	12	15
w28	44° 45′ 42″	32° 24′ 29″	21	Public benefit	12	15
w29	44° 22' 23″	32° 21′ 43″	14	Public benefit	12	15
w30	44° 21′ 47″	32° 22′ 11″	12	Public benefit	-	-
w31	44° 28' 07"	32° 37′ 38″	14	Public benefit	12	15
w32	44° 27′ 49″	32° 37′ 16″	13	Public benefit	12	15
w33	44° 29′ 12″	32° 35′ 15″	20	Public benefit	12	15
w34	44° 42′ 26″	32° 31′ 02″	13	Public benefit	12	15
w35	44° 37′ 42″	32° 34′ 15″	11	Public benefit	12	15
w36	44° 22′ 54.1″	32° 18' 08.5"	21	Public benefit	12	15
w37	44° 37′ 00″	32° 37′ 46″	14	Public benefit	12	15
w38	44° 34′ 17″	32° 39′ 24″	17	Public benefit	15	15
w39	44° 57′ 07″	32° 30′ 23″	15	Public benefit	12	15
w40	44° 59′ 24″	32° 31′ 59″	2	Public benefit	12	15
w41	44° 17′ 03″	32° 43′ 27″	20	Public benefit	12	15
w42	44° 57′ 50″	32° 26′ 33″	13	Public benefit	12	15
w43	44° 56′ 59″	32° 29′ 15″	13	Public benefit	-	-
w44	44° 20′ 22″	32° 31′ 55″	11	Public benefit	12	15
w45	44° 19′ 26″	32° 31′ 50″	16	Public benefit	12	15
w46	44° 18′ 55″	32° 29′ 55″	15	Public benefit	-	-
w47	44° 19′ 47 4″	32° 32' 23 8"	14	Public benefit	12	15
w48	44° 19′ 47 4″	32° 32′ 23.8″	20	Public benefit	12	15
w49	44° 22′ 52″	32° 20′ 54″	11	Public benefit	10	15

$$WQI = \frac{\sum Q_i Wi}{\sum Wi}$$
(3)

For the current paper, the maximum ground WQI suitable for human consumption (for drinking purposes) would be considered a 100 score (Ministry of Environment, 1998). Based on the WQI values, the groundwater quality will be rated, as shown in Table 4.

Irrigation water quality index (IWQI)

Many studies have used the irrigation water quality index (IWQI) that was essentially developed by Meireles et al. (2010). More recently, Abbasnia et al. (2018) used the IWQI to evaluate the water quality when used for irrigation in Iran. There are slight differences between the IWQI method and the WQI based method conducted by the WHO. In order to find the relative weight, the predictable values of each parameter need to be applied. These estimated values have



Fig. 3 Location of the sampling sites

samp
collected
for the
characteristics
Chemical
Table 2

Table 2 Cher	nical ch	laracteristics	for the collecte	ed samples								
Well number	Hq	EC (µs/cm)	TDS (ppm, parts per million)	Na ⁺ (meq/l)	Mg ⁺² (meq/l)	Ca ⁺² (meq/l)	Cl ⁻ (meq/l)	K ⁺ (meq/l)	SO4 ⁻² (meq/l)	HCO3 ⁻ (meq/l)	NO ₃ ⁻ (meq/l)	CO ₃ ⁻² (meq/l)
wl	7.41	8170	5547	34.844	19.917	22.455	29.977	0.2637	36.2544	9.0036	0.1288	0
w2	7.38	1733	1250	8.352	3.5389	4.2914	9.024	0.0512	5.616	1.5744	0.0483	0
w3	7.25	1082	1040	3.9585	5.9256	5.6387	3.0174	0.1536	8.9648	0.8364	0.0322	0
w4	7.15	3160	2409	16.313	8.6415	11.1776	15.002	0.0768	16.016	5.084	0.0966	0
w5	7.3	1128	930	5.655	4.2796	3.992	7.2474	0.0896	7.0096	2.4108	0.0322	ı
w6	7.13	8140	5580	25.274	13.250	17.5149	22.447	3.0976	28.392	8.3804	0.0322	ı
w7	7.12	3690	2590	15.225	9.4645	10.7285	14.946	0.1024	12.48	7.462	0.0483	0
w8	7.31	2530	1830	7.9605	8.1477	7.2854	7.755	0.3584	13.3744	2.3944	0.06118	,
6m	7.14	2030	1418	4.089	5.1026	8.982	4.2018	0.1562	12.7088	1.5908	0.0322	0
w10	7.71	1238	890	3.828	5.5141	5.4391	5.1042	0.128	7.8832	0.8036	0.1288	ı
w11	7.5	2060	1449	7.0035	6.584	6.487	10.490	0.1587	7.0304	2.3944		0.066
w12	7.19	1077	800	6.438	2.3044	2.6447	5.358	0.0154	4.784	1.1152	0.0483	
w13	7.14	2040	1418	4.089	5.1026	8.982	4.2018	0.1536	12.7088	1.5908	0.0322	
w14	7.7	910	800	6.4815	2.3867	2.6946	5.3298	0.0384	4.8048	1.1316	0.0322	
w15	7.5	1949	1450	10.745	1.7283	1.8463	5.076	0.1766	6.4896	3.1816	0.05635	
w16	7.16	3690	2900	17.879	10.618	13.1237	15.313	2.0224	20.5712	7.38	0.11592	
w17	7.31	3030	2188	14.529	7.7362	9.2315	10.716	0.256	14.1648	6.0516	0.0322	,
w18	7.31	2460	1919	5.829	6.8309	6.2375	6.768	0.2816	11.336	1.066		0.099
w19	7.31	1193	006	5.6115	4.115	3.9421	7.2474	0.0794	5.2208	1.0168		0.066
w20	7.1	8840	6190	32.712	18.518	29.441	29.920	0.4352	38.6464	12.3984	0.1127	
w21	7.22	3640	2592	9.222	11.604	12.6746	15.228	0.256	16.4528	3.2964	0.10465	
w22	7.12	937	800	6.4815	2.3867	2.6946	5.3298	0.0384	4.8048	1.1316	0.0322	0
w23	7.2	1406	1095	3.6975	5.4318	3.493	5.3298	0.1152	7.8	1.6072	0.03703	0
w24	7.12	1045	772	2.2185	2.0575	2.0958	2.6226	0.0230	1.8096	2.05	0.0644	0
w25	7.71	16,090	13,200	38.672	34.648	38.9719	47.376	0.1792	47.8816	16.9412	0.2576	
w26	7.2	1631	1300	11.267	3.4566	5.0898	10.096	0.256	8.736	1.2792	0.00161	0
w27	7.4	3000	2221	5.6985	7.5716	6.5369	7.05	0.3584	11.627	1.0824	0.05635	0
w28	7.2	2420	1955	10.005	4.1973	8.0838	6.7116	0.0922	11.482	4.0016	0.0644	
w29	7.6	973	840	6.5685	2.3044	2.7944	5.3862	0.0512	4.8464	1.1152	0.01771	0
w30	7.13	24,500	17,900	91.35	32.591	45.0098	93.201	2.0992	51.9168	20.5328	0.15295	0
w31	7.91	4990	3410	29.624	9.7937	9.4311	17.597	0.3072	24.9184	6.724	0.1449	0
w32	7.13	1340	1080	6.438	2.6336	2.9441	5.2452	0.0666	5.1792	1.0824	0.0322	0
w33	7.7	916	805	6.4815	2.3867	1.6966	5.3298	0.0384	4.8048	1.1316	0.0322	0
w34	7.31	3020	2188	14.529	7.7362	9.2315	10.716	0.256	14.1648	6.0516	0.0322	0
w35	7.2	1327	1260	5.7855	7.1601	6.4371	6.8244	0.384	11.4608	1.1152	0.0322	0
w36	7.2	1375	1260	7.134	6.584	5.8383	8.7702	0.2304	9.5472	1.3776	0.1449	0

Table 2 (cont	tinued)											
Well number	Hq	EC (µs/cm)	TDS (ppm, parts per million)	Na ⁺ (meq/l)	Mg ⁺² (meq/l)	Ca ⁺² (meq/l)	Cl ⁻ (meq/l)	K ⁺ (meq/l)	SO4 ⁻² (meq/l)	HCO ₃ ⁻ (meq/l)	NO ₃ ⁻ (meq/l)	CO ₃ ⁻² (meq/l)
w37	7.15	3440	2400	16.095	8.5592	11.1277	15.002	0.0768	16.0784	5.0676	0.0966	0
w38	7.15	2220	1727	10.919	1.5637	1.7964	4.4556	0.1818	6.4064	3.2636	0.0644	0
w39	7.15	2850	2000	18.662	2.9628	3.5429	10.208	0.3098	10.608	4.428	0.0805	0
w40	7.1	1852	1582	7.917	7.6539	7.5349	12.408	0.1792	8.3616	2.4272	0.0322	0
w41	7.23	3540	2950	18.27	9.9583	13.5229	16.638	4.352	22.1936	4.0672	0.05635	0
w42	7.5	6580	4724	25.578	11.769	17.6147	18.02	1.1264	31.1376	6.7568	0.0966	0
w43	7.41	21,300	14,900	60.422	31.192	46.1076	82.682	1.6896	29.536	20.1064	0.0966	0
w44	7.41	7440	5213	27.318	15.226	18.8622	26.508	0.0307	38.1888	8.4952	0.0322	0
w45	7.25	4940	3674	23.055	11.275	14.2215	18.274	2.304	24.4816	7.7572	0.0322	0
w46	7.32	4980	3910	23.49	13.168	16.7165	20.360	3.0208	27.56	8.364	0.0322	0
w47	7.17	13,400	21,600	101.4	22.056	54.9898	107.13	5.0944	59.904	33.0132	0.1288	0
w48	7.13	21,500	17,450	91.35	32.591	45.0098	93.201	2.0992	51.9168	20.5328	0.1449	0
w49	7.5	1922	1800	14.964	6.7486	6.5369	12.944	0.3328	13.936	1.476	0.0322	0

SO4

NO3

EC

TDS

Parameter

 Table 3
 Water quality

parameter based on

originated based on the irrigation water quality data obtained by the University of California Committee of Consultants (UCCC) and Ayers and Westcot (1994). Using the IWQI model, the dominant parameters must be identified as they play a significant part in assessing irrigation water quality. The dominant parameters include EC, Na⁺, Cl⁻, HCO⁻₃ and SAR. SAR (sodium adsorption ratio) is obtained using Eq. 4:

$$SAR = \frac{Na^{+}}{\sqrt{\frac{Ca^{+2} + Mg^{+2}}{2}}}$$
(4)

After obtaining the required parameters, the accumulation weights (wi) values proposed by Meireles et al. (2010) are then applied. It is worth highlighting that the total value for the five parameters is equal to one (see Table 5) according to Ayers and Westcot, (1994) recommendations.

The water quality measurement parameter (Qi) value and the accumulation witness (Wi) are obtained using each individual parameter value. The criteria suggested by Ayers and Westcot (1994) is adopted during the calculations (see Table 6).

Low Q_i values indicate that the groundwater quality is insufficient; meanwhile, high Q_i values demonstrate

 Table 4
 Water quality index levels adopted in this study based on WHO standards

Water quality index	Description
0–25	Excellent
26–50	Good
51–75	Poor
76–100	Very poor
>100	Unfit for drink

Standards

400

50

1000

500

Table 5IWQI weights(Meireles et al., 2010)

EC 0.211 Na ⁺ 0.204
Na ⁺ 0.204
HCO ⁻ ₃ 0.202
Cl ⁻ 0.194
SAR 0.189
Total 1.0

that the groundwater quality is good. The Qi values are obtained using Eq. 5:

$$Q_{i} = q_{\max} - \left(\frac{\left(x_{ij} - x_{inf}\right) \times q_{iamp}}{x_{amp}}\right)$$
(5)

where q_{max} is the maximum Q_i value for each class, x_{ij} is the observed value of each parameter, x_{inf} is the lower limit value of the class to which the parameter belongs, $q_{i\text{map}}$ is the class amplitude. X_{amp} is the class amplitude to which the parameter belongs.

In this case, the upper limit is treated as the highest value obtained from the water sample analyses and involves assessing X_{amp} for each parameter's last class. Finally, the irrigation water quality index (IWQI) can be determined using Eq. 6:

$$IWQI = \sum_{i=1}^{n} Q_i \times w_i \tag{6}$$

where IWQI is non-dimensional and ranges from 0 to 100; Q_i is the quality measurement of a parameter, *i*th is a function of its concentration (0 to 100). w_i is the normalised weight of the *i*th parameter.

IWQI is classified into five dimensionless parameter classes to determine the irrigation water class suitability (Meireles et al., 2010). As shown in Table 7, this classification was based on the suggested groundwater quality index. Furthermore, Bernardo et al. (1995) has

Table 7	Irrigation	water	quality	index	characteristics	(Meireles
et al., 201	l 0)					

Recommendation Plant	Water use	IWQI
	1000110110	
No toxicity risk	No restriction (NR)	85–100
Avoid salt-sensitive plants	Low restriction (LR)	70–85
Plants which have moderate tolerance to salts	Moderate restriction (MR)	55–70
 Plants with moderate to high tolerance to salts Except for water with low values of CL. Na⁺ and HCO⁻₃ 	High restriction (HR)	40–55
 Plants with high salt tolerance Except for waters with extremely low values of Cl⁻, Na⁺ and HCO₃⁻ 	Severe restriction (SR)	0–40

defined the classes depending on their effect on vegetation, e.g. soil water infiltration reduction, salinity hazard and toxicity.

Several indices have been used globally to evaluate water quality for irrigation (Raihan and Alam, 2008; Sarkar and Hassan, 2006; Talukder et al., 1998; Quddus, 1996; Todd, 1995; Raghunath, 1987). The mineral constituents in the water used for irrigation affect the soil and associated plant growth; therefore, groundwater suitability would largely depend on its mineral constituents. One of the most harmful mineral constituents to plants is salt as it impacts their metabolic rate by limiting the water intake. Furthermore, salts can influence the structure, aeration and the soil's permeability, indirectly impacting the crops' growth. The water suitability for irrigation relies on various aspects regardless of the water properties itself, such as the soil drainage properties, soil type, climate of the area and plants' tolerance to salt (Michael, 1978).

Sodium concentration is another essential factor that needs to be considered in all the classification indices for

Table 6 Qi limiting values (Ayers and Westcot, 1994)

0		, ,			
Na ⁺ (meq/l)	Cl⁻	HCO ⁻ ₃	EC (µs/cm)	SAR (meq/l) ^{0.5}	Q_i
2≤Na ⁺ <3	1≤Cl [−] <4	1≤HCO ₃ [−] <1.5	200≤EC<750	$2 \leq SAR < 3$	85-100
$3 \le Na^+ < 6$	$4 \leq Cl^{-} < 7$	$1.5 \le HCO_3^- < 4.5$	750≤EC<1500	$3 \leq SAR < 6$	60-85
6≤Na ⁺ <9	$7 \leq Cl^{-} < 10$	$4.5 \le HCO_3^- < 8.5$	1500≤EC<3000	$6 \leq SAR < 12$	35-60
$Na^+ < 2 \text{ or } Na^+ \ge 9$	$1 < Cl^{-} \ge 10$	$\text{HCO}_3^- < 1 \text{ or } \text{HCO}_3^- \ge 8.5$	EC<200 or EC≥3000	$2 < SAR \ge 12$	0–35



Fig. 4 WQI spatial distribution map

irrigation water as a result of sodium reactions with soil which can lead to a decrease in permeability (Todd, 1995). Similarly, Kelly's ratio, permeability index (PI), residual sodium carbonate (RSC), soluble sodium percentage (SSP), sodium adsorption ratio (SAR) and the electrical conductivity (EC) can be categorised as crucial indices in the determination of water irrigation suitability (Khan and Abbasi, 2013). Thus, SSP, RSC, PI and KR were calculated in this investigation using Eq. 7 to 10:

$$SSP = \frac{Na^{+*} = \times 100}{Ca^{+2} + Mg^{+2} + Na^{+}}$$
(7)

$$RSC = (CO_3^{-2} + HCO_3^{-}) - (Ca^{+2} + Mg^{+2})$$
(8)

$$PI = \frac{Na^{+} + \sqrt{HCO_{3}^{-}}}{Ca^{+2} + Mg^{+2} + Na^{+}} \times 100$$
(9)

$$KR = \frac{Na^{+}}{Ca^{+2} + Mg^{+2}}$$
(10)

where the concentration of ions is expressed in meq/l.

Eventually, using the GIS environment, the numerical and spatial distribution of the tested parameters was generated from analytical results, and subsequently, the inverse distance weight technique (IDW) was implemented to generate the spatial distribution maps of water quality parameters. The IDW can be defined as a deterministic method used to conduct multivariate interpolation calculations with a known scattered set of points. The cell values can be allocated using a linear-weighted combination set of sample points. In this study, the IDW was selected as it is the most suitable interpolating technique.

Results and discussion

ArcGIS has been used based upon the data of 49 wells across the region in order to generate the required database for groundwater suitability for drinking and



Fig. 5 pH spatial distribution map

irrigation. These databases were then used to create maps of the spatial distribution of all parameters which will be presented and discussed in the next sections of this paper. These maps provide a precise evaluation of groundwater quality and estimate the well's extraction potential with the lowest pollutant contents. It is important to emphasise that the IWQI was calculated using the EC, SAR, Na⁺, CI⁻ and HCO₃⁻ parameters, and the spatial distribution maps were created for each parameter and integrated via ArcGIS/spatial analyst function according to Eq. 5. This integration offers a map of the IWQI index obtained using geostatistical analysis.

Water quality index (WQI)

Virtualising the groundwater quality by location using ArcGIS maps is very important for evaluating the usability of water. The variations of physicochemical characteristics and the WQI of the groundwater in various places around the study area will be presented. The groundwater quality shows variations from well to well, attributed to the surface and subsurface features. Spatial interpolation of ArcGIS using the WQI parameter was utilised for plotting a digital map to describe the suitability of water for



Fig. 6 HCO₃⁻ spatial distribution map

human needs in the study area. These maps represent an efficient tool for managing the water quality and minimising the negative impacts on the ambient environment (Fig. 4).

The results show that around 78% of water samples are potentially suitable for drinking purposes; however, around 22% of the samples are not suitable for human use. As previously discussed, ongoing water shortages in Iraq mean that groundwater resources' demand will rise with time. Thus, more attention is required to avoid groundwater pollution as it can create severe consequences in the future. Irrigation water quality index (IWQI)

Many factors can alter groundwater quality such as the use of fertilisers, malfunction of underground drainage systems, mining activities, disposal of industrial wastes, continuous burning of municipal solid waste (MSW) and presence of landfills. The diversity of water quality in the focussed region is described as follows:

As shown in Fig. 5, the studied water samples' pH value ranged between 7.1 and 7.91. High pH values up to 8.5 in groundwater can be attributed to



Fig. 7 EC spatial distribution map

 Table 8
 Classification of groundwater for irrigation based on electrical conductivity (Singh et al., 2018)

EC (µs/c)m	Water class	Remark on quality
< 250	Low salinity	Excellent
250-750	Medium salinity	Good
750-2250	High salinity	Doubtful
> 2250	Very high salinity	Unsuitable

bicarbonate ions that form the groundwater's main alkaline component (Jerome and Pius, 2010). The

variation of pH between low and high values affects plants' ability to absorb the soil's nutrients. Low pH increases the solubility of manganese and ammonium salts to concentrations that can be harmful to plants.

The results showed that the bicarbonate ion (HCO₃⁻) concentration in the water samples ranged from 0.804–33.013 meq/l; see Fig. 6. Ayers and Westcot (1994) pointed out that bicarbonate's ideal concentration for irrigation purposes is less than 1.5 meq/l. Therefore, all water samples are not appropriate for irrigation.



Fig. 8 Na⁺ spatial distribution map

- Figure 7 shows the spatial distribution from the electrical conductivity (EC) tests. The results revealed considerable differences in the EC values from 910 to 24,500 μ s/cm. This wide range can be attributed to the predominant activities of humans in the region. High EC values can be due to reducing the osmotic plant activity, which interferes with water and nutrient absorption from the soil (Rao, 1986). According to Wilcox's classification of water using EC, adopted by the US Department of Agriculture (Richards, 1954), above 50% of groundwater

sources present in the study region are inappropriate for irrigation purposes (World Health Organization, 1993), while the remaining percentage is doubtful, as shown in Table 8 (Singh et al., 2018).

 Figure 8 shows the spatial distribution map of the sodium content obtained using GIS. The results showed that the percentage of Na⁺ in the water samples was in the range of 2.219 to 101.4 meq/l.



Fig. 9 SAR spatial distribution map

 Table 9
 Classification of groundwater for irrigation based on SAR

SAR values	Water class	Remark on quality
< 10	Low sodium	Excellent
10-18	Medium sodium	Good
18–26	High sodium	Doubtful
>26	Very high sodium	Unsuitable

With regard to the alkalinity of the ground, Fig. 9 shows that the SAR values for the selected water samples ranged from 1.54 to 16.34 meq/l. According to Varol and Davraz (2015), groundwater is inappropriate for irrigation purposes when the SAR value is larger than 18. Table 9 presents the groundwater classification for irrigation based on the SAR (Richards, 1954). Based on the SAR limits in Table 9, all groundwater tested samples confirmed their appropriateness for irrigation.



Fig. 10 CL⁻ spatial distribution map

Table 10 The irrigation water classification for Chloride (Bauderet al., 2003)

Chloride (mg\l)	Effect on Crops
Below 70	Generally safe for all plants
70–140	Sensitive plants show injury
141–350	Moderately tolerant plants show injury
Above 350	Can cause severe problems

Chloride concentration does not influence the soil's physical characteristics; therefore, it is commonly not incorporated in modern water classifications. Table 2 reveals that the chloride ion concentration was in the range of 2.6226–107.1318 meq/l (93-3799 mg/l). Figure 10 shows a relatively high chloride ion concentration in all chosen water samples. Low chloride ion concentrations are essential for plant growth, but chloride ions are toxic to sensitive plants



Fig. 11 IWQI spatial distribution map

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above threshold concentrations, as shown in Table 10.

Based on Table 10, plants were susceptible to low chloride in 5% of the samples and were moderately tolerant in 35% of the samples. Moreover, 60% of the water samples could cause severe problems for plants due to the amount of chloride.

Figure 11 shows the IWQI index map resulting from integrating the parameters mentioned above using ArcGIS analyst extension. According to the IWQI map analysis, there are four classifications of water use restrictions on groundwater suitability for irrigation in the studied area. A total of 20% of groundwater samples belong to the low-restricted category, which could be directly used for irrigation purposes without any treatment process. A total of 50% of groundwater samples belong to the moderate and high restricted category. This category indicates that the groundwater should be suitable for soils with high permeability and without compact layers due to these kinds of soils' capability



Fig. 12 Buffer map around each well

to leach the salts. The remaining percentage of groundwater samples (30%) belong to the severe restriction (SR) category, which is not appropriate for irrigation purposes under normal conditions. The SR category of groundwater can be used for soils with high permeability and excessive water usage to prevent salt accumulation.

It is important to note that a buffer analysis has been performed for the 49 wells, as shown in Fig. 12. This analysis has been conducted to investigate whether some features need to be highlighted in the study area. The buffer analysis was applied with a radius value of 2 km and 5 km, respectively. The results showed that some wells are located within a distance of 2 to 7 km from the river. The reason for this is due to the local perception that the river would be a good source of groundwater for the wells and would improve quality compared to other locations. This illustrates the reason for having a high number of wells near the river (Fig. 12). However, the local perception is not accurate as the soil's hydraulic conductivity in that area is relatively low, according to Al-Madhlom et al. (2016). In



Fig. 13 KR spatial distribution map

reality, as shown in Figs. 4 and 11, the best groundwater quality for irrigation and drinking was in areas located around the central zone of the province. This illustrates that there are many parameters that affect the groundwater quality such as hydraulic conductivity, soil type and the topography, and all of them must be investigated cautiously (Al-Madhlom et al., 2016).

Other indices were used to evaluate the suitability of groundwater for irrigation purposes, namely, permeability index (PI), soluble sodium percentage (SSP), Kelly's ratio (KR) and residual sodium carbonate (RSC). ArcGIS is used to produce maps showing the variation of all the indices mentioned above. The groundwater suitability for irrigation can be evaluated using Kelly's ratio (Kelly, 1951). Kelly's ratio represents the ratio of sodium versus calcium and sodium versus magnesium. According to Reddy (2013), many studies have adopted Kelly's ratio to assess groundwater's suitability for irrigation purposes. When Kelly's ratio is less than or equal to 1, it indicates that water quality is good, when the ratio is higher, the groundwater is not suitable for agricultural purposes because of high alkalinity (Karanth,



Fig. 14 RSC spatial distribution map

1987). Figure 13 shows that 30% of the wells are inappropriate for irrigation use because of the alkalinity, and 70% of the wells contain water with good quality.

When the residual sodium carbonate (RSC) value is less than 1.25 meq/l, it indicates safe water quality, but above 2.5 meq/l, it indicates that the water is generally inappropriate for irrigation uses. When the RSC is in the range of 1.25–2.5 meq/l, it indicates that the water is slightly appropriate for irrigation (Singha and Pasupuleti, 2020). The results shown in Fig. 14 reveal that the RSC values for all samples are less than 1.25 meq/l, indicating that the whole study area is within the safe limit for irrigation.

A classification scheme for rating irrigation waters was proposed by Wilcox (1955) based on soluble sodium percentage (SSP). When the SSP values are higher than 50, it indicates that the water is not safe for irrigation purposes. Nevertheless, when SSP is less than 50, it indicates the water is safe (Richards, 1954). The results presented in Fig. 15 showed that 30% of the study area's water samples are not safe for irrigation purposes, and the remaining percentage (70%) are good quality water.



Fig. 15 SSP spatial distribution map

When the permeability index (PI) value is less than 25, it indicates that the nature of water is unsuitable for irrigation. In contrast, water quality is suitable for irrigation when PI ranges from 25 to 75 and excellent water quality when PI is higher than 75. Figure 16 shows the map of the PI variation of the water samples obtained and shows that the quality of water in the area of study is appropriate for irrigation uses.

As shown in Figs. 15 and 16, the values of the SSP and PI are relatively close. However, the percentage of the suitable groundwater samples for irrigation purposes was quite different due to the differences in the classification standard for both indices. As stated above, the PI results showed that most of the groundwater samples of the study area are appropriate for irrigation use, while the SSP results revealed that about 30% of the water samples are not appropriate for irrigation (Fig. 17). Thus, conducting a comparison between the two indices might be able to provide more information in order to perform an accurate evaluation. Figure 17 shows that 70% of the wells have suitable groundwater quality for irrigation use based on the PI index and the SSP index.



Fig. 16 PI spatial distribution map



Fig. 17 Comparison between SSP and PI

However, the high values of the PI at some wells are accompanied by high SSP values and affect groundwater quality. Although the PI index is in the excellent category at some wells, it is still crucial to investigate the SSP at the same locations. Singh et al. (2015), demonstrate that Ca⁺², Na⁺, Mg⁺² and HCO₃⁻ concentration can influence soil permeability profile. Therefore, these ions are used to evaluate the water PI and quality, as mentioned earlier in Eq. 8. Likewise, Xu et al. (2019) suggested that the high PI values correlate with high bicarbonate and sodium ions in the groundwater. Nevertheless, Singh et al. (2020) pointed out that the percentage of soluble sodium (SSP) is essential for classifying irrigation water in terms of soil permeability. This is due to the fact that sodium-ions in irrigation water can be exchanged with Mg⁺² and Ca⁺² ions from the clay particles leading to reduce soil permeability. The reduction of soil permeability would affect the internal drainage and hardening of soil, which negatively impacts soil quality and seedling emergence. Generally, when the average value of both indices (SSP and PI) varies from 25 to 50, then the quality of the groundwater at that location will be suitable for irrigation in terms of PI and SSP assessment (Fig. 18).

Generally, the dissolved ionic species in the groundwater represent the resultant product that is accomplished due to the weathering of rock-forming minerals and a minor contribution from atmospheric precipitation and anthropogenic activities. Moreover, as a result of the assessment for groundwater quality, it is crucial to remark that there is no significant relationship between groundwater quality and general land use. Therefore, the case study's low water quality could be attributed to mismanagement practices such as poor waste management and poor farm management practices.

Conclusions

In this paper, an assessment of the quality of the groundwater of Babylon province was conducted to explore the feasibility of using it for drinking and irrigation purposes. The data was collected from 49 wells widely distributed across the province. Taking into consideration the research limitations, the following conclusions can be drawn:

- 1- According to WHO standards, around 78% of the samples collected can be used for drinking. The potable groundwater is mainly concentrated in the middle of the province. This would encourage the local government to establish the required infrastructure to invest in groundwater.
- 2- The groundwater of Babylon province could be used for irrigation purposes with precautions to account for the following factors:
- A relatively low pH across the sample region could make the soil harmful to plants by increasing manganese and ammonium salt concentrations' solubility.



Fig. 18 Spatial distribution map between SSP and PI

- In general, the bicarbonate ions (HCO₃⁻) concentration is higher than the ideal concentration (1.5 meq/l.).
- According to the US Department of Agriculture standards, high electrical conductivity was recorded for about 50% of the groundwater samples, making it inappropriate for irrigation purposes.
- High sodium Na⁺ percentage, if used for irrigation, would result in more dense and rigid soil.
- The SAR ratio obtained from the samples indicates a medium to high sodium concentration, which is

appropriate for irrigation but not ideal. Also, SSP values showed that 30% of the water samples tested are not safe for irrigation. Meanwhile, the permeability index (PI) results showed that water quality in the study area is appropriate for irrigation use.

This study is limited to the locations of existing wells and the methods used for data collection. The results can be used to monitor groundwater quality in this region and for comparison across central Iraq since the soil characteristics are geographically reasonably consistent. This study's outcomes can help provide guidelines for the authorities in managing groundwater quality and developing future improvement interventions. Nevertheless, if the authorities decided to use groundwater for irrigation purposes, more detailed testing is required to identify effective treatment processes.

Based on the data presented here, more industrial and agricultural practice regulations are required in this region. The governmental authorities face many challenges implementing new legislations, including the financial support for groundwater extraction and quality control. Challenges also exist at the societal level where the local people's cooperation to adopt new practices requires the government to run education programmes to disseminate the required social awareness. Many countries around the world have an adequate level of knowledge and experience dealing with groundwater. Lessons from other parts of the world where groundwater extraction is successfully run could be adopted as part of the government's national strategy going forward.

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Code availability Not applicable.

Data availability Materials presented in this manuscript are available upon request from the authors.

Compliance with ethical standards

Conflict of interest The authors declare that they have no conflict of interest.

References

- Abbasnia, A., Radfard, M., Mahvi, A. H., Nabizadeh, R., Yousefi, M., Soleimani, H., & Alimohammadi, M. (2018). Groundwater quality assessment for irrigation purposes based on irrigation water quality index and its zoning with GIS in the villages of Chabahar, Sistan and Baluchistan, Iran. *Data in Brief, 19*, 623–631.
- Alam, W., Singh, K. S., Gyanendra, Y., Laishram, R. J., & Nesa, N. (2020). Hydrogeochemical assessment of groundwater quality for few habitations of Chandel District, Manipur (India). *Applied Water Science*, 10(5), 1–13.
- Alanbari, M., Alazzawi, H., Al-Ansari, N., & Knutsson, S. (2015). Environmental assessment of Al-Hilla City wastewater

treatment plants. Journal of Civil Engineering and Architecture, 9(6), 749–755.

- Al-Jubouri, S. M., & Holmes, S. M. (2020). Immobilisation of cobalt ions using hierarchically porous 4A zeolite-based carbon composites: Ion-exchange and solidification. *Journal of Water Process Engineering*, 33, 101059.
- Al-Madhlom, Q., Al-Ansari, N., & Hussain, H. M. (2016). Assessment of groundwater vulnerability in northern Babylon governorate, Iraq. *Engineering*, 8(12), 883–902.
- APHA. (1998). Standard methods for the examination of water and wastewater (1000p). Washington D. C: American Public Health Association.
- Ayers, R. S., & Westcot, D. W. (1994). Water quality for agriculture: Irrigation and drainage paper 29, revision (Vol. 1, pp. 1–130). Rome: Food and Agriculture Organization of the United Nations.
- Bachmann, A., Virginia, T., Al-Obeidi, L. A., and Kilıç, D. T. (2019). *Tigris-Euphrates River Ecosystem*: A status report.
- Backman, B., Bodiš, D., Lahermo, P., Rapant, S., & Tarvainen, T. (1998). Application of a groundwater contamination index in Finland and Slovakia. *Environmental Geology*, 36(1–2), 55– 64.
- Bauder, T. A., Waskom, R. M., & Davis, J. G. (2003). *Irrigation water quality criteria*. Colorado State University Cooperative Extension.
- Bernardo, S., Soares, A. A., & Mantovani, E. C. (1995). Manual of irrigation (488p). Vicosa: UFV.
- Brown, R. M., McClelland, N. I., Deininger, R. A., & O'Connor, M. F. (1972). A water quality index—Crashing the psychological barrier. In *Indicators of environmental quality* (pp. 173–182). Boston: Springer.
- Buday, T., & Jassim, S. Z. (1980). Regional geology of Iraq: Vol. 1. Stratigraphy and Paleogeography, Iraq State Organization for minerals, Baghdad (445pp).
- Cude, C. G. (2001). Oregon water quality index a tool for evaluating water quality management effectiveness 1. *JAWRA Journal of the American Water Resources Association*, 37(1), 125–137.
- ESRI, A. (2015). 10.5: Using ArcGIS spatial analyst. In *Software* User Guide ESRI, USA.
- Husain, T. (1998). Application of the National Water Quality Index prototype in Newfoundland. In *Prepared for state of the environment task group, Canadian Council of Ministers of the Environment.*
- Iraq National Population Commission. (2012). Iraq population situation analysis report.
- United Nations Report for Iraq, (2013). Water in Iraq Factsheet, Iraq Joint Analysis and Policy Unit. https://www. iraqicivilsociety.org//wp-content/uploads/2014/02/Water-Factsheet.pdf ().
- IAU, Babil Governorate Profile (2010), https://reliefweb. int/sites/reliefweb.int/files/resources/CC0511E38D021F8 BC12577EC00311F9E-Full_Report.pdf, ().
- Jerome, C., & Pius, A. (2010). Evaluation of water quality index and its impact on the quality of life in an industrial area in Bangalore, South India. *American Journal of Scientific and industrial research*, 1(3), 595–603.
- Karanth, K. R. (1987). Ground water assessment: Development and management. New Delhi: Tata McGraw-Hill Education.
- Kelly, W. P. (1951). Alkali soils-their formation properties and reclamation (3rd ed.p. 92). New York: Reinhold Publication.

- Khan, F., Husain, T., & Lumb, A. (2003). Water quality evaluation and trend analysis in selected watersheds of the Atlantic region of Canada. *Environmental Monitoring and Assessment*, 88(1–3), 221–248.
- Khan, T. A., & Abbasi, M. A. (2013). Synthesis of parameters used to check the suitability of water for irrigation purposes. *International Journal of Environmental Sciences*, 3(6), 2131–2138.
- Madhloom, H. M., & Al-Ansari, N. (2018). Geographical information system and remote sensing for water resources management case study: The Diyala River, Iraq. *International Journal* of Civil Engineering and Technology, 9(12), 971–984.
- Meireles, A. C. M., Andrade, E. M. D., Chaves, L. C. G., Frischkorn, H., & Crisostomo, L. A. (2010). A new proposal of the classification of irrigation water. *Revista Ciência Agronômica*, 41(3), 349–357.
- Michael, A. M. (1978). *Irrigation: Theory and practice*. Vikas publishing house.
- Ministry of Environment. (1998). System Maintenance of River and Public Water Pollution in Iraq.
- Mohammed, W. T., Abdullah, S. M., & Ghalib, A. M. (2007). Catalytic wet oxidation of phenol in a trickle bed reactor. *Iraqi Journal of Chemical and Petroleum Engineering*, 8(4), 45–52.
- Nelly, K. C., & Mutua, F. (2016). Ground Water Quality Assessment Using GIS and Remote Sensing : A Case Study of Juja Location, Kenya. American Journal of Geographic Information Systems, 5(1), 12–23.
- Quddus, K. G. (1996). Irrigation water quality in some selected villages of Meherpur in Bangladesh. *Bangladesh Journal of Agriculture Science.*, 23, 51–57.
- Raghunath, I. M. (1987). Groundwater (2nd ed.). New Delhi: Wiley Eastern Ltd..
- Raihan, F., & Alam, J. B. (2008). Assessment of groundwater quality in sunamganj Bangladesh Iranian. *Journal of Environmental Health Science and Engineering*, 6(3), 155–166.
- Rao, N. S. (1986). Hydrogeology and hydrogeochemistry of Visakhapatanam, Ph. D Thesis unpublished. Visakhapatnam: Department of Geology, Andhra University.
- Reddy, K. S. (2013). Assessment of groundwater quality for irrigation of Bhaskar Rao Kunta watershed, Nalgonda District, India. *International Journal of Water Resources* and Environmental Engineering, 5(7), 418–425.
- Reyes-Toscano, C. A., Alfaro-Cuevas-Villanueva, R., Cortés-Martínez, R., Morton-Bermea, O., Hernández-Álvarez, E., Buenrostro-Delgado, O., & Ávila-Olivera, J. A. (2020). Hydrogeochemical characteristics and assessment of drinking water quality in the urban area of Zamora, Mexico. *Water*, 12(2), 556.
- Richards, L. A. (1954). Diagnosis and improvement of saline and alkaline soils. In *Agricultural Handbook No.* 60. Washington, D.C.: United States Department of Agriculture.
- Sarkar, A. A., & Hassan, A. A. (2006). Water quality assessment of a groundwater basin in Bangladesh for irrigation use. *Pakistan Journal of Biological Sciences*, 9(9), 1677–1684.
- Singh, K. K., Tewari, G., & Kumar, S. (2020). Evaluation of groundwater quality for suitability of irrigation purposes: A case study in the Udham Singh Nagar, Uttarakhand. *Journal* of Chemistry, 2020.

- Singh, P. K., Atal, S., & Dubey, S. K. (2018). Assessment of shallow groundwater quality for irrigation in parts of Agra District, Uttar Pradesh, India. *Earth Science India*, 11(3), 133–148. https://doi.org/10.31870/esi.11.3.2018.09.
- Singh, S. K., Srivastava, P. K., Singh, D., Han, D., Gautam, S. K., & Pandey, A. C. (2015). Modeling groundwater quality over a humid subtropical region using numerical indices, earth observation datasets, and X-ray diffraction technique: A case study of Allahabad district, India. *Environmental Geochemistry and Health*, 37(1), 157–180.
- Singha, S. S., & Pasupuleti, S. (2020). Hydrogeochemical modeling based approach for evaluation of groundwater suitability for irrigational use in Korba district, Chhattisgarh, Central India. SN Applied Sciences, 2(9), 1–13.
- Singha, S., Devatha, C. P., Singha, S., & Verma, M. K. (2015). Assessing ground water quality using GIS. *International Journal of Engineering Research & Technology*, 4(11), 689–694.
- Singha, S., Pasupuleti, S., Singha, S., & Villuri, V. G. K. (2017). An integrated approach for evaluation of groundwater quality in Korba district, Chhattisgarh using Geomatic techniques. *Journal of Environmental Biology*, 38(5), 865–872. https://doi.org/10.22438/jeb/38/5/MRN-600.
- Talukder, M. S. U., Shirazi, S. M., & Paul, U. K. (1998). Suitability of groundwater for irrigation at Kirimganj Upazila Kishoreganj. *Progressive Agriculture*, 9(1–2), 107– 112.
- Todd, D. K. (1995). *Groundwater hydrology* (3rd ed.625p). New York, U.S.: Wiley and Sons Inc..
- Varol, S., & Davraz, A. (2015). Evaluation of the groundwater quality with WQI (water quality index) and multivariate analysis: A case study of the Tefenni plain (Burdur/ Turkey). *Environmental Earth Sciences*, 73(4), 1725–1744.
- Wilcox, L.V. (1955). Classification and use of irrigation water, USA Dept. Ag. Circ. 696, Washington DC. 19p.
- World Health Organization. (1993). *Guidelines for drinking-water quality*. Geneva: World Health Organization (WHO).
- World Health Organization. (2004). Information products: Water, Sanitation and Health. Geneva: World Health Organization (WHO).
- Xu, P., Feng, W., Qian, H., & Zhang, Q. (2019). Hydrogeochemical characterisation and irrigation quality assessment of shallow groundwater in the central-western Guanzhong basin, China. *International Journal of Environmental Research and Public Health*, 16(9), 1492.
- Yacoub, S. Y. (2011). Stratigraphy of the Mesopotamia plain. Iraqi Bulletin of Geology and Mining, 4, 47–82.
- Zandbergen, P. A., & Hall, K. J. (1998). Analysis of the British Columbia water quality index for watershed managers: A case study of two small watersheds. *Water Quality Research Journal*, 33(4), 519–550.
- Zektser, I. S., and Everett, L. G. (2004). Groundwater resources of the world and their use. *IHP-VI Ser. Groundwater* 6, U. N. Educ., Sci. And cultural organ., Paris.

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