

Assessing fire hazard potential and its main drivers in Mazandaran province, Iran: a data-driven approach

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Abstract Fires are a major disturbance to forest ecosystems and socioeconomic activities in Mazandaran province, northern Iran, particularly in the Hyrcanian forest sub-region. Mapping the spatial distribution of fire hazard levels and the most important influencing factors is crucial to enhance fire management strategies. In this research, MODIS hotspots were used to represent fire events covering Mazandaran Province over the period 2000–2016. We applied the ecological niche theory through the maximum entropy (MaxEnt) method to estimate fire hazard potential and the association with different anthropogenic and biophysical conditions, by applying different modeling approaches (heuristic, permutation, and jackknife metrics). Our results show that higher fire likelihood is related to density of settlements, distance to roads up to 3 km and to land cover types associated with agricultural activities, indicating a strong influence of human activities in fire occurrence in the region. To decrease fire hazard, prevention

activities related to population awareness and the adjustment of farming practices need to be considered.

Keywords Structural fire hazard · Maximum entropy · MODIS active fire data · Anthropogenic factors

Introduction

Forests and other wooded land are very important worldwide, providing ecosystem services and refuges for terrestrial biodiversity, influencing employment, economic growth, human livelihoods, and well-being (Earl 1975; Shvidenko and Gonzalez 2005; Nilsson et al. 2011; Kibria et al. 2017). Fires are a major disturbance to forested areas, affecting millions of hectares of land every year, leading to disturbances on biodiversity, affecting the stability of ecosystem functioning and posing significant threats to human lives and livelihoods (Bowman et al. 2009; Özbayoğlu and Bozer 2012; Chandra and Bdishardwaj 2015). Forest vegetation is particularly sensitive to variations in climatic factors, such as extreme drought events (Bottero et al. 2017), which in turn make forests more susceptible to fire (Flannigan et al. 2006; Russo et al. 2017). Besides climate, other factors can strongly influence fire occurrence and spread, such as vegetation and fuel types, topographic conditions, proximity to roads and settlements or human activities (Flannigan et al. 2009; Kasischke et al. 2010; Bowman et al. 2011; Archibald et al. 2013; Chuvieco et al. 2014; Oliveira et al. 2017). Understanding the relationships between the spatial

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distribution of fires and their main drivers is vital to improve fire prevention and mitigation strategies. The influence of biophysical and anthropogenic factors in fire distribution and hazard levels has attracted wide international attention, fostering the development of assessment frameworks, fire danger and risk models based on GIS and the application of novel and advanced techniques worldwide (Chuvieco et al. 2014; Hantson et al. 2015; Šturm and Podobnikar 2017; Tien Bui et al. 2017; Connor et al. 2017; Oliveira et al. 2018). Despite these efforts, fire hazard and risk mapping tools, adapted to the specific features of different fire-prone areas in the globe, are still missing.

Machine learning, statistical techniques and experts' knowledge approaches are commonly used in fire hazard modeling and to analyze fire occurrence patterns at different spatial and temporal scales (e.g., from local to global, and short to long-term analysis (Oliveira et al. 2012; Tien Bui et al. 2016; Satir et al. 2016)). Approaches using presence-absence methods (PA) have been used to analyze the distribution of specific events or elements, based on presence data (ignition, in case of fires) at a certain location (pixel); interpretation about the probability of fire occurrence can be accomplished, for example, by fitting the presence-absence fire data to discriminant analysis, logistic regression and related models (MacKenzie et al. 2002; Adab 2017). Presence-absence methods (PA) have been shown to yield high accuracy in modeling predictive habitat suitability, ultimately determined by the quality of the data (Brotons et al. 2004). Maximum entropy (MaxEnt) is a presence-background method (PB) (Guillera-Aroita et al. 2014) based on the probability distribution of maximum entropy (Phillips et al. 2006) that applies presence-only data with the number of raster covariates to predict probability of suitability for occurrence of a target phenomenon. MaxEnt is particularly suited for modeling with presence-only data and verified absence is not required to fit the model (Renard et al. 2012), which are difficult to confirm because it depends on survey location (MacKenzie et al. 2002). The performance of presence-absence methods will be low if absence data at a particular location is caused by factors not included in the model, such as dispersal limitations, biotic interactions or incorrect assessment (Liu et al. 2005; Pearson et al. 2007).

In Iran, fires are a major disturbance to forest ecosystems (Shafiei et al. 2010; Jafarzadeh et al. 2017) and biomass burning is one of the main drivers of

deforestation and degradation in the northern part of the country (Sobhani and Khosravi 2015). Even though historical forest fire statistics are not completely reliable and the fire data currently available are partial (FAO 2015; Jafarzadeh et al. 2017), the Global Forest Resources Assessment (FAO 2015) accounted for about 3400 ha of forests burned annually, between 2003 and 2012, and ca. 88,000 ha of burned land overall in the country per year. Previous studies were developed to analyze forest fire distribution, hazard and risk levels in Iran, at several scales and using different methodologies. At national level, Eskandari and Chuvieco (2015) obtained a fire danger assessment map based on a framework adapted from other countries, focusing on specific fire ignition and propagation factors; specific areas in the north and west of the country were found to be more likely affected by fires. Other authors applied Geographic Information Systems (GIS), remote sensing techniques and particular statistical analysis to uncover the relations between fires and influencing factors in the provinces of Gilan, Golestan, Ilam or Mazandaram (Mahdavi 2012; Mohammadi et al. 2014; Eskandari and Chuvieco 2015; Najafabadi et al. 2015; Abdi et al. 2016; Adab 2017; Pahlavani and Bigdeli 2017). These studies provided valuable indications on the potential drivers of fire at local and regional levels and on the applicability of different methods. Despite these efforts, the factors contributing to fire occurrence in Iran are still unclear (Eskandari and Chuvieco 2015; Abdi et al. 2018), also because forest studies in the country are still in an early stage; however, the existing studies indicate that anthropogenic factors are likely to have a strong contribution to fire hazard in the country (Adab 2017). Also, the country has not yet built a forest information system (FIS) or an exhaustive database on fire events, hindering the development of coherent fire hazard and risk models (Mahdavi and Naghdi 2007). Important gaps still exist regarding the development of technical and operational tools, based on advanced methodologies, GIS mapping and state-of-the-art scientific knowledge, to improve fire management strategies tailored to the country's needs and applicable to the different phases of fire management (prevention, suppression, mitigation or recovery). These circumstances introduce an obvious degree of uncertainty in the theoretical and applied research on fire modeling in Iran, which can lead to unexpected and imprecise results (De la Riva et al. 2004). Therefore, better information and deeper knowledge concerning fire occurrence distribution,

hazard and risk assessment and influencing factors, is still needed for Iran at different scales and timeframes. Modeling fire hazard levels is essential to provide effective tools for controlling or preventing forest fires. The main purpose of the present study was to improve the understanding of the environmental factors that potentially influence fire hazard levels in the province of Mazandaran, in Iran, which are still largely unknown. Notwithstanding the frequent forest fire activity (Adab et al. 2015), a comprehensive fire occurrence database is still not available for the province. For this reason, the assessment was based on MODIS active fire product. Different environmental covariates were obtained from different data sources with the purpose to provide quantitative estimations of fire hazard potential, measured in association with anthropogenic and biophysical conditions. In this research, we applied the ecological niche theory (EN), classically used for distributional possibilities of species, for predicting the potential geographical distribution of fire hazard, based on the presence of fire events occurred in different areas of the Mazandaran Province. The relative importance of the different environmental factors to fire hazard levels are discussed, in view of the potential implications to improve forest fire prevention and mitigation strategies in the region.

Materials and methods

Study area

The study area corresponds to Mazandaran province which extends over 23,833 km² in the north of Iran and is located along the southern edge of the Caspian Sea (Fig. 1). This area experiences mild and humid climatic conditions (Emadi et al. 2016), with warm summers and cold winters (Peel et al., 2007); mean precipitation values are above the average of the country, about 500 mm annually, increasing in the north of the province (Molavi-Arabshahi et al., 2016). This area belongs to the Hyrcanian forests belt, stretched along the northern slopes of the Alborz mountains; these are mainly composed of temperate deciduous forests and have high species richness and biodiversity levels (Scharnweber et al. 2007; Akhane et al. 2010; Bani Assadi et al. 2015).

Data collection and pre-processing

Fire occurrence data

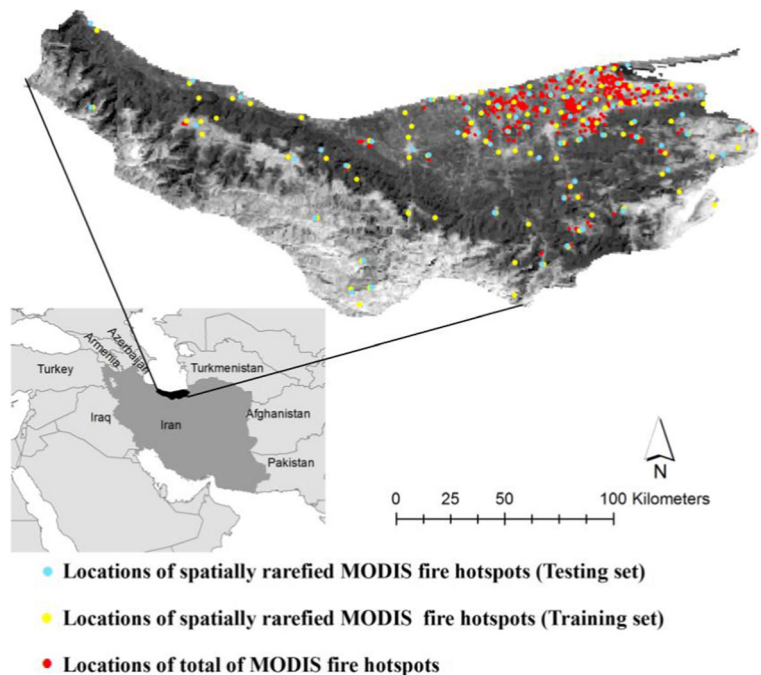
The historical database on forest fire events which is available for Mazandaran Province was coordinated by government institutions and national NGOs, using different fire detection technologies (Jahdi et al. 2016), and subsequently given to the Forests, Range and Watershed Management Organization (Allard 2001). This database has recently integrated information on the location, time, ignition source, and burnt area of each fire ignition, but shows some spatial uncertainty in the fire records, derived from location errors or missing fire events, which may hinder the development of a robust analysis. In order to overcome these issues, MODerate resolution Imaging Spectrometer (MODIS) hotspots (also known as active fire product) were used instead for the study area. MODIS fire hotspots (Collection 6 - C6) used in this study included global daily coverage at 1 km spatial resolution (MOD14A1), which is also a rapid fire detection product for operational near-real time (NRT) use and it is also applied in forest fire-related studies (Justice et al. 2002). MODIS hotspots of all confidence levels were included, because the low confidence hotspots represent extra information which outweighs the potential problem of slightly higher commission errors (Hanson et al. 2013). This dataset has been improved over previous versions (Collection 5 - C5), with a reduced incidence of false alarms caused by small forest clearings and an improved detection of large fires with 1.2% global commission compared to 2.4% in C5 (Giglio et al. 2016).

As the study area is mostly covered by vegetation types that are likely to burn, all MODIS hotspots detected over the period 2000–2016 were retained. In the present study, a total of 726 active fire events were recorded by MODIS sensor in the Mazandaran Province in the period considered, which are integrated as the dependent variable in the model (Fig. 1).

Driving factors

In this research, the term “fire hazard” refers to the probability of a certain location to experience a fire each year (Richards et al. 1999). Fire ignition and spread are a result of the complex interaction between a variety of components. The factors influencing fire hazard levels were categorized into four groups, based upon their

Fig. 1 The location of the study area, Mazandaran province, northern Iran. Labeled red points represent the total number of fire occurrences detected by MODIS sensor prior to spatial filtering ($n = 726$) for the period from November 2000 to July 2016. The source of borders of Middle East countries is from GADM (www.gadm.org). The source of background image is from NASA Earth Observations (NEO) (<https://neo.sci.gsfc.nasa.gov/>)



potential relation with fire occurrence: (1) climatic conditions; (2) topographic conditions; (3) land cover and land use types, and (4) anthropogenic factors, from which we obtained 18 explanatory variables. The technical specifications of each variable are detailed in Table 1 and a brief description of their potential relation with fire occurrence, based on literature, is presented below.

Climatic conditions

Climatic conditions are widely recognized as important factors influencing regional and local fire probability. Annual and interannual climatic variation for precipitation, temperature and drought influences fuel availability and conditions (Schoennagel et al. 2004; Littell et al. 2009) which are linked to the likelihood of fire ignition and to the probability of fire propagation (Archibald et al. 2013; Bedia et al. 2015; Hantson et al. 2015; Moreira et al. 2009; Oliveira et al. 2012; Syphard et al. 2008; Itziar et al. 2015). This is due to the effects of weather parameters on fuels moisture, fuel load and composition, including the amount of fine fuels (e.g., leaves, grasses, forest litter) and fuel-desiccating drought (Schoennagel et al. 2004; MacMillan and Shary 2008). Higher annual precipitation and lower temperatures increase vegetation productivity and

generate higher fuel loads in specific topographic and land cover conditions, such as grasslands and shrublands, which in turn will be available to burn in subsequent drier seasons (Oliveira et al. 2014; Nunes et al. 2016). The relative dryness of fuels is a direct effect of long-term dry conditions and is of greater significance in fire hazard (Renard et al. 2012). Climatic predictors are used in this study to explore their relative importance in driving fire occurrence in the study area (Wimberly and Reilly 2007; MacMillan and Shary 2008; Renard et al. 2012; Alvares et al. 2013; Guo et al. 2017; Tien Bui et al. 2017).

Topographic conditions

Topographic conditions influence fuel type distribution, composition and flammability and also have a big role on local climate variations (Parisien and Moritz 2009; Martínez-Fernández et al. 2013; Oliveira et al. 2014; Modugno et al. 2016). For example, elevation, slope and aspect are linked to plant composition in beech communities in unharvested beech forests in the north of Iran, because of their effects on decreasing air pressure, increasing ultraviolet rays, reduced temperatures and changes in precipitation conditions (Adel et al. 2014).

Forest fires may also be influenced by hydrogeological parameters (Tien Bui et al. 2016). As such, the

Table 1 Variables integrated in the maximum entropy model for assessing fire hazard, including the data source and a brief description. All variables were generated or resampled at 1-km resolution

Category	Code	Description of data	Data source
Explanatory variables			
Climatic conditions	WEI	Wind Exposition Index (dimensionless). Continuous	SAGA Climate and Weather module (Conrad et al. 2015)
	PIS	Potential Incoming Solar radiation (insolation) (kWh/m ²). Continuous	SAGA Lighting module (Conrad et al. 2015)
	AP	Total Annual Precipitation (mm). Continuous	WorldClim, (Fick and Hijmans 2017)
	PDM	Total Precipitation of the Driest Month (mm). Continuous	WorldClim, (Fick and Hijmans 2017)
	AMT	Annual Mean Temperature (°C). Continuous	WorldClim, (Fick and Hijmans 2017)
	MTWM	Maximum Temperature of the Warmest Month (°C). Continuous	WorldClim, (Fick and Hijmans 2017)
	DAH	Diurnal Anisotropic Heating (dimensionless). Continuous	SAGA Climate and Weather module (Conrad et al. 2015)
	DI	Annual Dryness Index computed as annual mean temperature (°C)/annual mean precipitation (mm/year). Continuous	(S. Brown and Lugo 1982)
	Köppen	Köppen climate classification, computed by AP and AMT data that coincides with well-defined vegetation regions around the world (Alvares et al. 2013). Categorical	WorldClim, (Fick and Hijmans 2017)
	Topographic conditions	Elevation	Elevation values obtained from Digital Elevation Model (DEM), at 1-km Resolution. Aggregated from ASTER 30 m Global Digital Elevation Model (GDEM), V2. Continuous
Slope		Slope derived from DEM (%). Continuous	Calculated from GDEM, V2
Aspect		Aspect derived from DEM (°). Categorical	Calculated from GDEM, V2
TWI		Topographic Wetness Index (dimensionless). Continuous	Calculated from GDEM, V2. SAGA Hydrology module (Conrad et al. 2015)
Land cover/vegetation characteristics	LC	Land cover of Mazandaran Province, Globcover2009 V2.3 at 300 m Resolution. Categorical. A total of 14 different categories were retrieved: rainfed croplands (Code 14), mosaic croplands/vegetation (Code 20), mosaic vegetation/croplands (Code 30), closed broadleaved deciduous forest (Code 50), closed needleleaved evergreen forest (Code 70), closed to open mixed broadleaved and needleleaved forest (Code 100), mosaic forest-shrubland/grassland (Code 110), mosaic grassland/forest-shrubland (Code 120), closed to open shrubland (Code 130), closed to open grassland (Code 140), sparse vegetation (Code 150), bare areas (Code 200), water bodies (Code 210), and permanent snow and ice (Code 220).	GLOBCOVER project, European Space Agency (ESA) (Arino et al. 2007)
	DisRo	Mean Euclidean distance to the primary roads (1-km). Continuous	Calculated from digital maps of topographic (The National Cartographic left of Iran (NCC), Iran)
Anthropogenic factors	DisS	Mean Euclidean distance to the settlements, including hamlets, villages, towns and cities. (1-km). Continuous	Calculated from digital maps of topographic (The National Cartographic left of Iran (NCC), Iran)
	DensS		

Table 1 (continued)

Category	Code	Description of data	Data source
Dependent variable	DisRi	Heatmap densities of settlements, including hamlets, villages, towns and cities. (km ²), Epanechnikov kernel shape. Scaled kernel size. Continuous Mean Euclidean distance to the rivers (1-km). Continuous	Calculated from digital maps of topographic (The National Cartographic left of Iran (NCC), Iran) Collected from Forests, Range and Watershed Management Organization (FRWO), Iran)
	Fire occurrence	Fire events Derived from MODIS 1-km MCD14ML (2000-16), Collection 6. Point format shapefiles (.shp)	Fire Information for Resource Management System (FIRMS), National Aeronautics and Space Administration (NASA)

Notes: All climate variables consisted of 30-year means (1970–2000) for WorldClim version 2. Code represents labels for every predictor variable name in the dataset, SAGA GIS (System for Automated Geoscientific Analyses)

Topographic Wetness Index (TWI), a steady-state wetness index that quantifies topographic control on hydrological processes (Sørensen et al. 2006), was included in the analysis. TWI measures the potential soil moisture (Krishna P Vadrevu et al. 2006) which can influence the moisture of dead fuel beds (Matthews 2014). Topographic features also play an important role on the distribution of human activities and the presence of people can increase the probability of fire ignition (Parisien et al. 2016; Nami et al. 2018). Lowland areas with gentle slopes are more attractive to human activities and, therefore, are more densely populated, have a higher concentration of built-up structures and more agricultural fields (Reed et al. 1999; Xu et al. 2006; Armien et al. 2009). TWI and topographic features (elevation, slope and aspect) were derived from a Digital Elevation Model (DEM) at 30 m resolution, resampled to 1 km (Table 1).

Land cover/vegetation

Land cover is a key variable driving the patterns of wildfire occurrence at the landscape level and is used as a proxy of vegetation characteristics. Specific types of land cover (e.g., shrublands or conifer plantations) are considered more fire-prone than others (e.g., wetlands, agricultural areas or recently burned patches) in certain environments, due to differences in fuel load composition and vegetation structure, which subsequently affect flammability levels, fire spread and intensity (Barros and Pereira 2014; Carmo et al. 2011; Grégoire et al. 2013; Moreira et al. 2009; Oliveira et al. 2014; Pereira et al. 2014; San-Miguel-Ayanz et al. 2012; Silva et al. 2009). Land cover also reflects human activities, as is the case of agricultural uses such as annual crops, permanent crops and agro-forestry systems, which are usually less fire-prone than wildland areas (Moreira et al. 2009). For this research, specific land cover types were retrieved (Table 1), according to their potential association with the spatial patterns of fire occurrence in the province.

Anthropogenic variables

Incorporating human-related parameters in fire occurrence analysis is important, since the spatial distribution of fires in different regions of the world is closely related to human presence and activities, although with differing trends (Bistinas et al. 2013; M. Krawchuk and Moritz 2014; Hantson et al. 2015). The data on cause

of ignitions is still unavailable for Iran, but sparse fire studies found that the majority of fires in recent years were human-caused (Mohammadi et al. 2014), whereas lightning is still relevant as an ignition source in certain mountainous areas (Eskandari and Chuvieco 2015). The proximity of populated settlements to forests may increase their susceptibility to fire and the probability of human-caused ignitions (Satoh et al. 2004). As well, socioeconomic settings and the technological development verified in recent decades have led to an overall increase in the number of visitors to forest and woodland ecosystems, due to increased movements of people and vehicles, the spread of tourism and recreational activities and an increased accessibility to forested areas (Guo et al. 2017). Distance to roads reflects the accessibility to forest areas and increases the possibility of fire occurrence (Guo et al. 2017; Armenteras et al. 2017). In the study area, most of the land located near rivers was allocated to traditional home gardens and vegetable gardens, and therefore can contribute to increase tourism activities (Mohammadi et al. 2014). For these reasons, it is of primary importance to include such anthropogenic variables (distance to rivers, distance to primary roads and density of cities) in fire hazard assessment.

Statistical analysis

MaxEnt modeling of fire occurrences

The principle of the MaxEnt model relies on the estimation of the probability distribution of maximum entropy. Occurrences are predicted by finding the distribution that is most spread out (i.e., closest to the uniform), while taking into account a set of constraints (the environmental conditions) of known locations (Chen et al. 2015). MaxEnt iteratively compares the conditional density function of given covariate grids (predictor variables) at presence locations to the marginal density (a background grid consisting of the mean observations) of covariates across the entire study region. Conditional probability is then derived from these two probability densities as sampled from a large number of grids, which gives the relative environmental suitability for presence of a fire for each point in the study area. The MaxEnt represents a list of functional forms (i.e., linear, quadratic, threshold, hinge, product, and categorical) to describe occurrence responses (for fire, in this study) to environmental conditions. Functional forms assist us in fitting highly complex responses by using function

types that cannot be modeled with regression techniques, such as generalized additive models (GAM) (Chen et al. 2015). Linear and quadratic features are used when sample size is low (< 20) (Kumar and Stohlgren 2009). MaxEnt can be trained to fit complex relationships between occurrences and environmental variables, and in this study auto feature type was used to simplify interpretation, since understanding the more complex and multifaceted relationships is still hard to accomplish (Louzao et al. 2012).

Spatial autocorrelation

A spatial correlation may exist between MODIS active fire data, which are collected as points. Ignoring this spatial correlation may yield a biased estimation of the model parameters and misleading inferences; including spatially independent occurrence from neighboring sites in a model can reduce residual error and improve the model estimation (Boria et al. 2014). Therefore, spatial filtering should be conducted before executing the MaxEnt model. Average Nearest Neighbor distances among fire points were calculated to measure the spatial autocorrelation distance between each fire feature and its nearest neighbor's location, and subsequently fire points were rarefied using the Rarefy Occurrence Data at SDMs (Species Distribution Models) tool performed in ArcGIS 10.3 (Brown 2014; Brown et al. 2017).

Regarding presence data, autocorrelation of clustered occurrence points was detected using Moran's I ($I = 0.21$) and Average Nearest Neighbor (z -score -31.2). The expected value of the nearest neighbor distance (7512 m, z -score 3.4) was used in this study to define the spatial resolution of rarefied occurrence data, which corresponds to the average distance between nearest points when these are randomly distributed throughout the study area. From these procedures, 105 points representing fire events were retrieved (training set), to predict fire distribution with the minimum spatial correlation between points (Moran's $I = 0.044$) (Fig. 1). Using the same procedures, 62 points of fire events, out of 621, were selected for the testing set in the fire model (Moran's $I = 0.039$) (Fig. 1).

Output transformations

MaxEnt provides four output transformations (floating point files) for model values in raw, cumulative, logistic and complementary log-log (cloglog). These output

formats are monotonically related to each other, but their scaling properties are totally different from each other, and have different meanings (Baldwin 2009). For many modeling applications, such as the one presented in this study, the probability of fire presence is a relevant result. This probability of presence is given by the logistic and cloglog format. The cloglog gives an estimated value between 0 and 1 of probability of presence, with higher values demonstrating more favorable conditions for fire occurrence. Therefore, cloglog output is more easily interpreted overlapped with the other transformations, when imported into a GIS to map fire likelihood (Kadej et al. 2017).

Model evaluation

In every hazard assessment, it is important to identify how each predictor variable affects the presence of the estimated event, in this case fire, which variable has the largest effect on the model and in what manner these different variables influence fire occurrence likelihood. The relative importance of the predictor variables in the model was tested using three approaches: heuristic (percent contribution), permutation (permutation importance) and jackknife metrics. Caution must be taken in employing heuristic methods because strong multicollinearity among variables can affect the results by adding misleading importance to correlated predictors; a jackknife approach (leave-one-out cross-validation) for evaluating variable importance is recommended, because each variable is excluded at a time when running the model with the remaining variables (Baldwin 2009). The jackknife approach was also used to measure the performance of the individual predictor variables in the model. The model's ability to correctly discriminate the positive (test points) and the negative (pseudo-absence points) cases was evaluated with the receiver operating characteristic (ROC) analyses (Renard et al. 2012). The area under the curve measured by ROC provides a single metric of model performance that is independent of any particular choice of threshold (Phillips et al. 2006). AUC values range between 0.5 and 1, with 1 reflecting a perfect discrimination between presence and absence, and 0.5 indicating a random performance without discriminative power. Predictor variables are also important in what concerns their high training gains when used alone in a model; additionally, a predictor variable can be considered a significant contributor if the training gain is low when the variable

is deleted from the model (Miller et al. 2012). Therefore, in order to compare the importance of the 18 predictor variables (Table 1), AUC statistic was employed and three different models were produced and compared with AUC values: (1) using all predictor variables; (2) using only one variable at a time and (3) systematically dropping out one variable each turn.

The Cohen's kappa (κ) index measures the reliability of a classification model quantitatively, indicating an agreement between the observed and the predicted values (Cohen 1960). Kappa values are interpreted as follows: $\kappa \leq 0$ as indicating no agreement and $\kappa = 0.01$ – 0.20 as none to slight, $\kappa = 0.21$ – 0.40 as fair, $\kappa = 0.41$ – 0.60 as moderate, $\kappa = 0.61$ – 0.80 as substantial, and $\kappa = 0.81$ – 1.00 as almost perfect agreement, using 0.5 as a threshold (Monserud and Leemans 1992). In this study, 105 points were randomly selected from the background, equivalent to the training dataset, for computing Cohen's kappa index via the MedCalc® statistical software.

Model validation

Model validation is a necessary step to assess the predictive performance of a model, considering that the validation error represents an unbiased estimate of a model's predictive power. A common approach to validation is to randomly split the dataset into two portions: a training set and a testing set, thus generating quasi-independent subset data for testing the model (Kumar and Stohlgren 2009; Lezama-Ochoa et al. 2016). The MaxEnt model developed in this research used a training dataset composed of 105 points of presence data. The predictive ability of the model was then assessed by cross-validation without duplicates (Lentz et al. 2008) using the 62 points of presence data set aside initially for testing purposes and, as such, not included in the training phase (Renard et al. 2012). A set of 10,000 random points representing background (or pseudo-absence) points were sampled to create the background distribution which fits the model (Phillips et al. 2006). MaxEnt limits model complexity and, hence regularization was used in this study to reduce model over-fitting through a process for achieving optimal MaxEnt model, similar to the Akaike Information Criterion (AIC) (Anderson and Gonzalez Jr. 2011); further details on this process can be found in (Phillips et al. 2006). ArcMap 10.3® was used to process the outputs of the MaxEnt model, mapping

the predicted fire occurrence likelihood levels in a raster format.

Results

The MaxEnt model generated an ASCII file of predicted fire presence locations, indicating a relative measure of fire proneness of the landscape at 1 km resolution. The cloglog outputs are shown in Fig. 2. The likelihood of a pixel being affected by a fire was expressed within the range of 0 and 1, with higher values representing increased probability and more favorable conditions for fire presence. High fire hazard potential (probability ≥ 0.8) was found in the northeast area of the province and near coastal and urban areas, whereas inland sites, away from the coast and inside mountainous ranges, experienced very low probabilities (probability ≤ 0.1).

According to the relative contributions of the environmental variables to the MaxEnt model (Table 2), Köppen, LC, Elevation and DenS were the strongest predictors of fire distribution, with 21.9, 19.9, 11.2 and 8.4% contribution, respectively. The variables with a lower percentage of contribution were DisRo, Aspect, DisRi and PIS.

The Cohen’s kappa index was 0.6, representing 60% better than random chance, and a slightly better agreement between the maximum entropy model and the training data. Regarding the performance of the MaxEnt model using ROC and AUC, the sensitivity (true positive rate) for the training data was 77.4%, indicating that over 77% of the fire points were correctly classified by

the model, whereas 71.2% of no-fire points were also correctly classified. The overall analysis of the AUC (Fig. 3) shows that the model presents a high goodness-of-fit (AUC = 0.85) with the training dataset, while the AUC is 0.79 when applied to the independent testing dataset, indicating that the model’s predictive power is 79% and low *p* values. Overall, the maximum entropy model performs well on both the training and testing datasets and showed a reasonable overall accuracy.

Figure 4 shows the results of the jackknife test regarding the order of importance of the predictor variables in the model. Density of settlements was the strongest contributor to the MaxEnt model for the study area; if excluded from the model, its overall performance is considerably reduced (Fig. 4). Among all predictor variables, land cover and distance to roads rank second and third in importance, and the remaining variables show similar levels of importance (AUC values). The cumulative contribution of these 18 variables is 79%.

The influence of each predictor variables on fire hazard potential was further investigated by producing response curves. The shape of the response curves can provide some details on the patterns of distribution of fire occurrence, as follows. Figure 5 shows the relationships between the three strongest predictors and fire hazard potential based on jackknife test (Adab et al. 2016). Fire hazard potential is negatively related to distance to roads (DisRo), indicating that fires are more likely to occur closer to roads and urbanized areas. For density of settlements, fire hazard potential reaches a

Fig. 2 Forest fire hazard map for Mazandaran province using the MaxEnt complementary log-log (cloglog) output from environmental variables. The predicted spatial distribution of fire probabilities across the forest landscape, with values ranging from 0 to 1 (grayscale color) is depicted by white to black colors. Black colors show a higher potential fire hazard while the white colors indicate lower potential

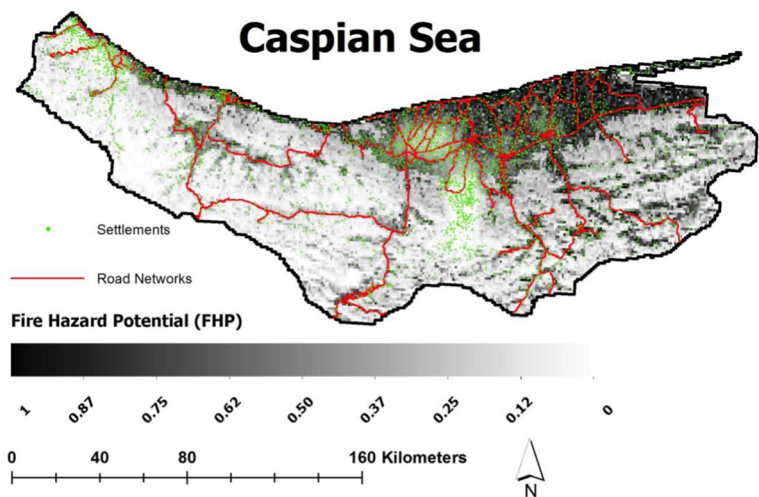
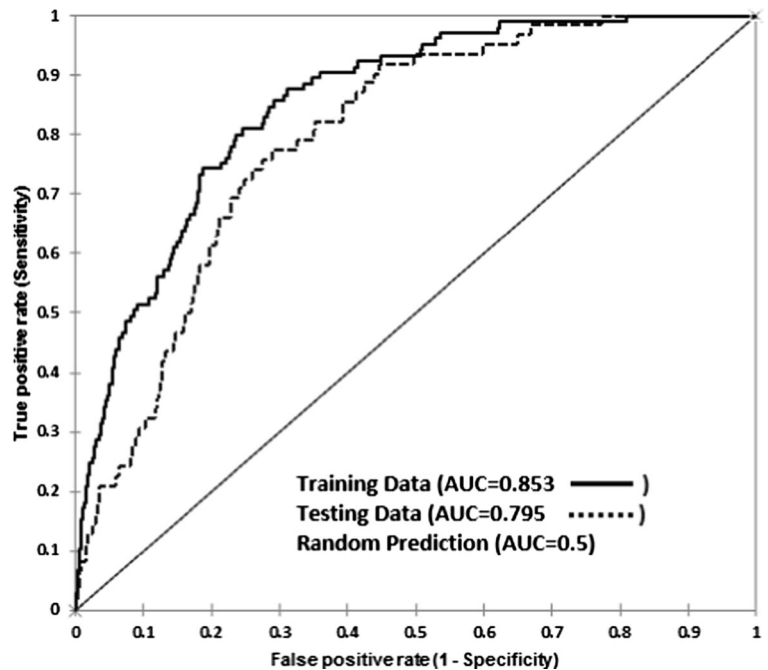


Table 2 Relative contributions of the of predictor variables to the forest fire hazard model. The highest predictive power is Köppen, followed by LC, Elevation, and DenS. For details about variables acronyms, see Table 1

<i>X</i>	Contribution (%)	Permutation importance
Köppen	21.9	5.4
LC	19.6	13.9
Elevation	11.2	4.5
DenS	8.4	13.8
Slope	7.1	3.9
TWI	6.6	1.9
DAH	5.5	8.5
DI	2.8	5.1
WEI	2.8	3.5
PDM	2.6	8.5
MTWM	2.5	0.1
AMT	2.2	13.4
AP	2.2	1.2
DisS	2.2	6.3
DisRo	1.3	6.6
Aspect	0.6	0
DisRi	0.5	2.2
PIS	0.1	0.9

peak and then decreases as the values go up. The probability of fire presence was highest when DenS was 0.26

Fig. 3 The receiver operating characteristic (ROC) curve and area under the ROC curve (AUC) for the maximum entropy. AUC for training data reaches 85% and for testing data is 79% which are higher than 0.5 of a random model



(number of settlements equal to 43 in 1 km²) (Fig. 5a) and DisRo was less than 3000 m (Fig. 5b). The importance of land cover variables indicates that fires are most likely to occur on specific forest types, namely closed to open mixed broadleaved and needleleaved forest (0.86), but also on rainfed croplands (0.83) and in areas with mosaic croplands/vegetation (0.78) (Figs. 5c and 6). On the contrary, fires are less likely to occur on closed broadleaved deciduous forest (0.23) and closed to open shrubland (0.26).

Discussion

Main drivers of fire hazard potential at regional level (Mazandaran province)

The assessment of fire hazard levels in the North of Iran is crucial to further understand the spatial distribution of fire occurrence likelihood and its main influencing factors. We developed a modeling procedure to quantify fire hazard potential at regional level by applying the MaxEnt method, integrating multiple layers of geospatial-based data potentially related to fire occurrence in Mazandaran province and a roadmap is provided for the study area as guidelines for fire hazard reduction.

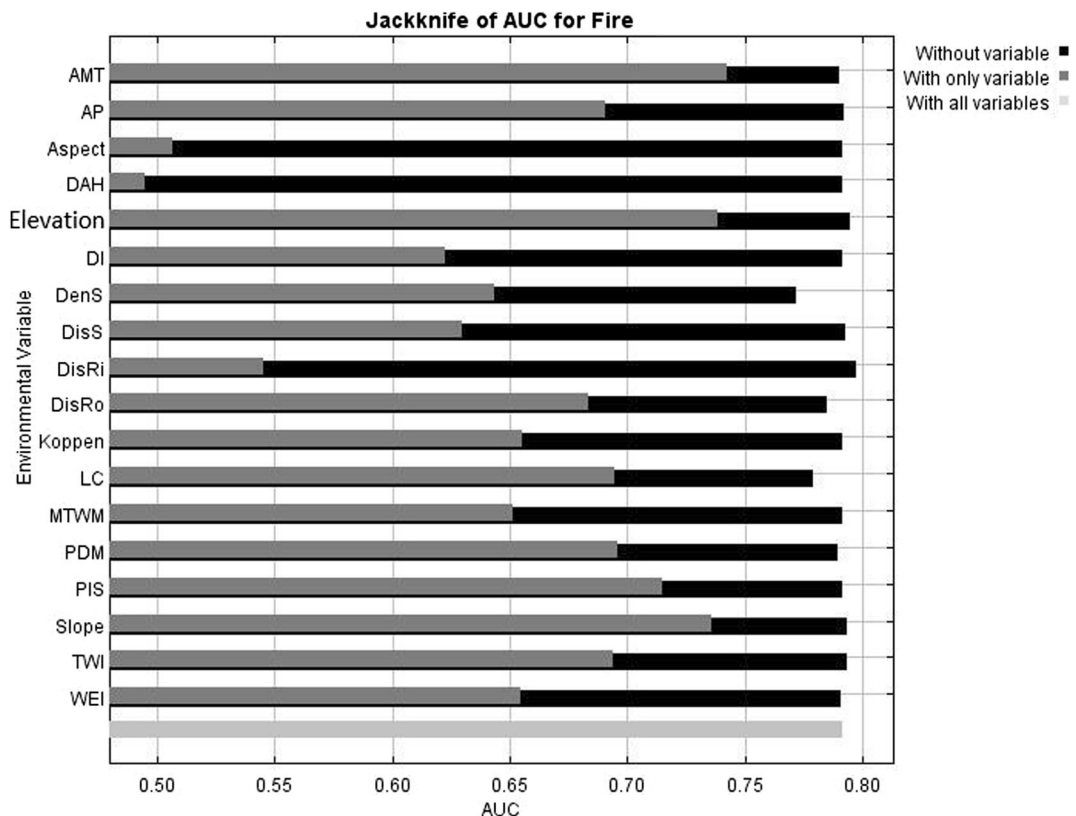


Fig. 4 Estimations of variable importance using Jackknife test for the final MaxEnt model output. For each variable, bars indicate the area under the ROC curve (AUC). The light gray bar represents the full-model AUC when using all variables, the dark gray bar

represents the gain of AUC when the specific variable is used in model and black bars represent the AUC of models without the specific variable, and a lower gain shows that the specific variable has more information that is not present in other variables

Fire hazard potential was highest in the northeast of the province. From the models obtained, with or without jackknife permutation and including single or multiple variables at a time, it was found that anthropogenic variables, particularly density of settlements and specific land cover types related to farming activities, were the most relevant drivers. Regarding the density of settlements, the study area is characterized by many small settlements and villages, with forest and agricultural landscapes neighboring them. Hazard levels in the study area declined when settlements density reached a certain threshold (0.26), (Fig. 5), a trend that is consistent with the results of other studies (Syphard et al. 2007). The probability of man-made ignitions is influenced by population density (Mahdavi 2012; Penman et al. 2013; Oliveira et al. 2014), but the effect in fire occurrence and burned area is positive only up to an intermediate threshold of population density, since the amount of available fuel decreases in more populated areas (M. A. Krawchuk et al.

2009; Bistinas et al. 2013). Another global analysis showed the effect of increasing population on the reduction of fire frequency, except for very sparsely populated parts, where the effect is slightly positive (Knorr et al. 2014).

Among all land cover types existing in the study area, closed to open mixed broadleaved and needleleaved forest, rainfed croplands, and mosaic croplands/vegetation (i.e., cropland (50–70%) and grassland/shrubland/forest between 20 and 50%) significantly influenced fire hazard potential. The contribution of the closed to open mixed broadleaved and needleleaved forest to fire hazard is likely related to their specific fuel load and composition, particularly the bulk of fine fuels produced (Agee et al. 1976; Small and Bush 1985; Stohlgren 1988; Dimitrakopoulos 2002). Also, the higher presence of tourism activities due to recent improved accessibility to these forests could increase the possibility of human-caused ignitions (Caldararo 2002; Ajin et al. 2016).

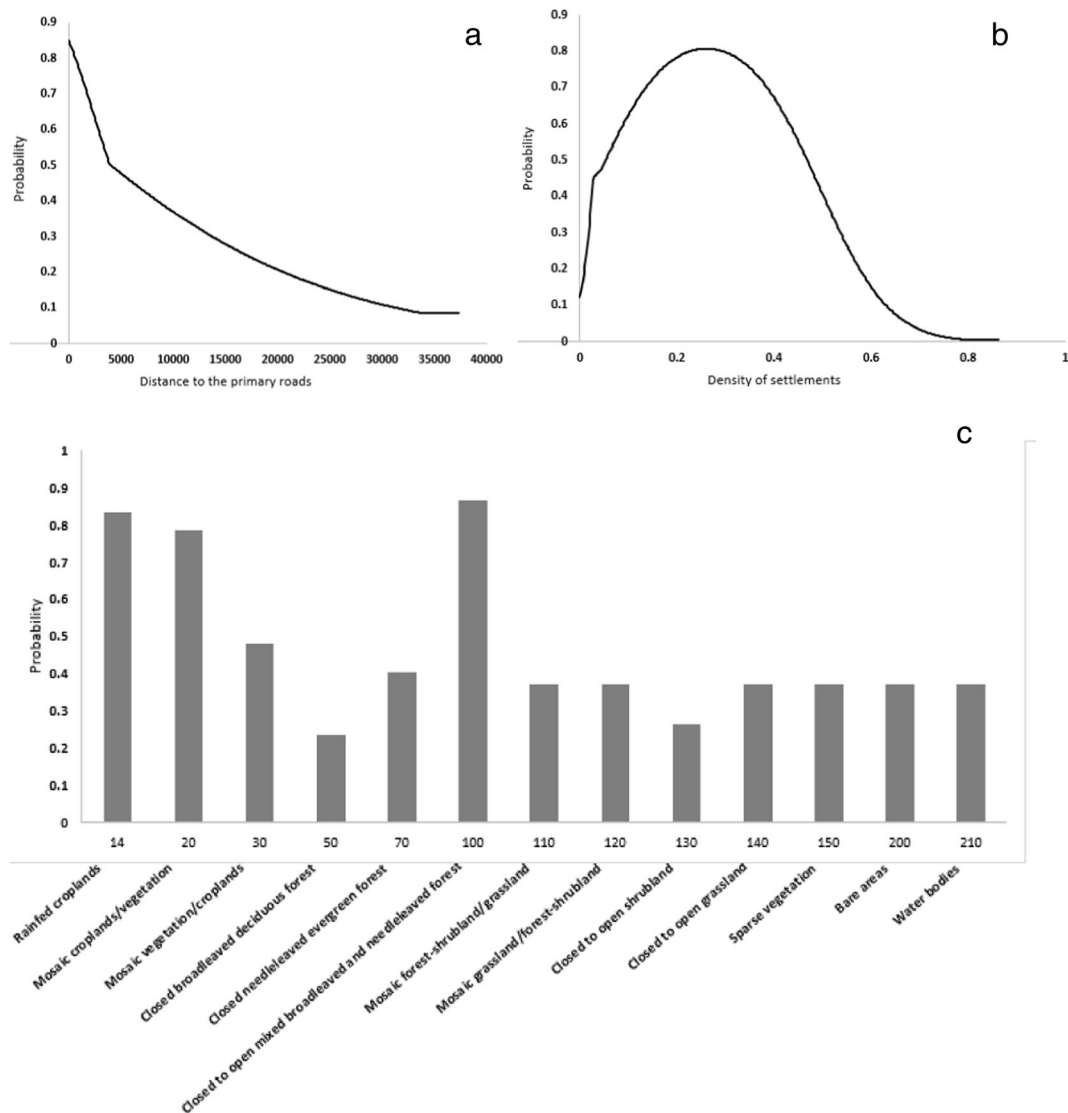


Fig. 5 Relationships between third highest predictors and the probability of presence of fire occurrence in Mazandaran province that included only the corresponding variable. **a** Distance from roads (meter) (DisRo). **b** City/settlement densities (scaled size)

(DenS). **c** Landcover. Black lines are the response curves of fire occurrence to predictors. The y-axis shows the logistic probability of fire presence; higher value of logistic probabilities indicates higher fire hazard

The presence of croplands/natural vegetation as a relevant variable indicates the potential influence of ignitions related to agricultural activities, which can spread to other vegetated areas nearby, a pattern also found in Southern Europe and the Middle East (San-Miguel-Ayaz et al. 2012; Ganteaume et al. 2013; Giglio et al. 2013). In Golestan Province in Iran, Adab (2017) has also argued that the mosaic croplands/vegetation class provides the greatest amount of fuel to burn, followed by rainfed cropland. Indeed, fire is widely used as a tool for shifting cultivation and converting

forest to agricultural land in many countries (Ketterings and Bigham 2000; Ruiz-Mirazo et al. 2012). In total, 321 and 171 hotspots were detected on mosaic croplands/vegetation and rainfed croplands, respectively, which cover over 600,000 ha of land in Mazandaran province. At the global level, as evaluated from MODIS data, agricultural fires accounted for ~9% of the global annual fires (Korontzi et al. 2006), with most biomass burning occurring in savannas (48–51%), and cleared areas used for agriculture showing the highest mean probability of fire occurrence, which increased even

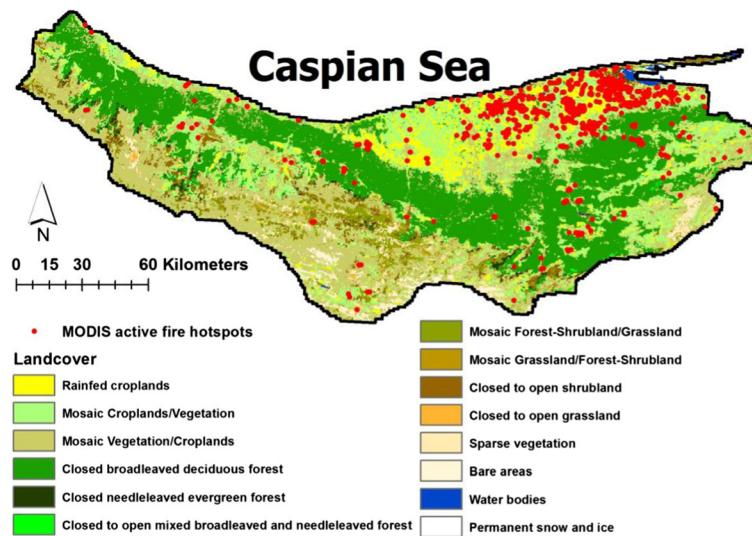


Fig. 6 Map showing the distribution of fire-affected area into land cover types. 14—rainfed croplands, 20—mosaic croplands/vegetation, 30—mosaic vegetation/croplands, 50—closed broadleaved deciduous forest, 70—closed needleleaved evergreen forest, 100—closed to open mixed broadleaved and needleleaved forest,

110—mosaic forest-shrubland/grassland, 120—mosaic grassland/forest-shrubland, 130—closed to open shrubland, 140—closed to open grassland, 150—sparse vegetation, 200—bare areas, 210—water bodies, 220—permanent snow and ice (without presence of MODIS active fire data)

further under drought conditions (Devisscher et al. 2016). Particularly in Iran, previous studies have shown that land cover/land use (LULC) is a relevant parameter for predicting fire occurrence in the Caspian Hyrcanian forest ecoregion, and also in arid and semi-arid ecosystems (Bashari et al. 2016; Adab 2017; Nami et al. 2018). This is due to a longer dry period each year that increases the amount of dry fuel available to burn, thus increasing the probability of an ignition turning into a large fire (Agee 1996).

Climatic variables, represented by Köppen categories (Humid Temperate Cfa), showed the highest importance in the overall model (without jackknife permutation); annual mean temperature alone provided evidence of increased fire occurrence by 74% in Mazandaran Province, affecting all classes of land cover, likely related to the effects on moisture content on live and dead fuels (Adab et al. 2016). Climate parameters such as temperature and moisture play a substantial role in vegetation growth and, hence, in fuel loads (Dwyer et al. 2000) that can determine the potential to start and spread a fire during a dry period (Dwyer et al. 2000). The northeastern side of the study area is uniformly covered by humid warm-temperate climate and temperature is highest in July, coinciding with the areas where more fuel accumulates. Other studies have found that climate with long periods of moisture stress contributed to fuels dryness

conditions (Dwyer et al. 2000) which can sustain fire occurrence in arid and semi-arid ecosystems in Iran (Bashari et al. 2016).

In this study, distance to roads showed a low contribution in the first model (without jackknife permutation), contrarily to what was found in the other models (with jackknife permutation), where DisRo ranked second in importance, likely due to the association with the presence of people, as ignition agents, and their accessibility to vegetated areas. In the latter, fire hazard potential levels were higher at shorter distances to roads, but started to level off at 4 km; this is consistent with the results of other studies (Romero-Calcerrada et al. 2008; Penman et al. 2013; Rodrigues and de la Riva 2014; Devisscher et al. 2016). Contrarily, other research found that distances to roads did not contribute significantly to fire hazard, as found by (Renard et al. 2012) in the Western Ghats of India the dual influence of distance to roads in fire hazard seems to show strong regional variations, a pattern that should be further explored in future research (Zhang et al. 2016).

Several studies have identified that fires are closely correlated with topography (e.g., elevation, aspect and slope) (Carmo et al. 2011; Martínez-Fernández et al. 2013; Oliveira et al. 2014; Satir et al. 2016), opposite to what was found in this research. Our findings indicate that the topographic features of northern Iran do not

seem to influence fire hazard levels, a pattern that was also found in other studies done for Iran (Abdi et al. 2016; Adab 2017). Overall, the importance of different explanatory variables on forest fire hazard highly depends on the characteristics of the study area and on the scale of analysis (Arpaci et al. 2014).

Model performance

Fire hazard potential was analyzed based on a presence-only data and MaxEnt was found a suitable method, considering that absence data is not required to fit the model (Renard et al. 2012). Massada et al. (2013) argued that a presence-absence modeling approach may be more suitable in areas with long-term fire records and where only a small part of the area can sustain a fire. We used MaxEnt, a popular method in ecological niche models (ENMs) because fire is presence-only data and this method has previously shown higher predictive accuracy than generalized linear models (GLMs), General Algebraic Modeling System (GAMs), BIOCLIM, or Genetic Algorithm for Rule Set Production (GARP) in other fields such as biodiversity assessment and species distribution mapping (Hernandez et al. 2006; Franklin 2010). MaxEnt is able to cope well with sparsely, irregularly sampled data and minor location errors (Kramer-Schadt et al. 2013). Even though MaxEnt has outperformed in comparison with other methods (Renard et al. 2012), it may lead to over-fitting if many input variables with too few occurrence records are integrated in the model (Kramer-Schadt et al. 2013). In the absence of comprehensive and complete national or regional databases, the use of MODIS hotspots is a reasonable option, because MODIS thermal sensors provide a considerable long-term database collection with minor location errors (1000 m² in size). Global commission error (false alarms) for the MOD14 Collection 6 daytime was 1.2%, compared to 2.4% in Collection 5 and Middle East commission error rates for MOD14 Collection 6 was 2% (Giglio et al. 2016). The commission errors (false alarms) for the Middle East are usually associated with perimeters of hot, arid, and sparsely vegetated or barren regions most commonly associated with deserts (Giglio et al. 2016). Although caution is needed regarding the use of worldwide datasets, given the potential error in terms of spatial location (Parente et al. 2016), MODIS active fire data and MERIS GlobCover product, both with medium spatial resolution (1 km or 500 m) have been widely

used to test and validate fire hazard models at different scales, when local or higher resolution data is lacking or when data are only available for partial time periods (Krishna PRASAD Vadrevu and Justice 2011; Renard et al. 2012; Eskandari and Chuvieco 2015). The use of the MaxEnt methodology allowed to develop models based on datasets that contain only fire presence data (Elith et al. 2006; Arpaci et al. 2014), an approach that requires additional work to further examine the consistency of global databases in comparison with national or regional ones, to ensure that the analysis of the spatial distribution of forest fires in the study area is not hindered by data limitations and by the characteristics of fire ignition and area burned registry.

Practical implications to fire management

In this study, fire hazard potential maps show where a fire is most likely to occur in Mazandaran province. Overall, our findings revealed that land cover types associated with agricultural activities, represented by rainfed croplands and mosaic croplands/vegetation, and the human presence evidenced in settlements density and distance to roads, are the most important factors influencing fire likelihood in the study area. These findings can contribute for defining fire prevention and mitigation strategies adjusted to the province's fire conditions. These strategies may include zoning regulations defined by legal instruments, to restrict the intensity of urban development in hazardous areas, as well as a timely allocation of public resources and support systems in areas that are more fire-prone (Eshliki and Kaboudi 2012; Mirzaei 2013). Additionally, fuel reduction measures over fire-prone areas should be a priority, mainly surrounding settlements and corridors in hazardous zones (Curt and Delcros 2010; Marchal et al. 2017). The implementation of specific fuel management measures, particularly changing vegetation structure and creating fuel discontinuities around buildings and other built structures, can help reduce fire hazard in these areas and mitigate potential fire impacts. Their implementation will have to be evaluated according to land ownership, available public resources and existing legal regulations.

Furthermore, considering the influence of agricultural land uses in fire hazard levels in the regions, strategies to adjust agricultural practices should also be considered, such as training farmers to improve the use of fire and increase safety measures. As well, prevention

measures focusing on public awareness, to reduce ignitions and to improve population preparedness in case a fire occurs, are of utmost importance; in fact, current international initiatives stress the need for a stronger community participation in disaster preparedness, mitigation and recovery, regardless of the phenomenon (fires, earthquakes, floods) that occurs (UNISDR 2015).

Conclusions

This paper showed how biophysical and anthropogenic predictors interact to estimate potential fire hazard at the landscape level in the north of Iran. The maximum entropy method was successfully used to analyze forest fire distribution in Mazandaran Province and to display areas of higher fire hazard. We found that density of human settlements, land cover types associated with farming activities and distance to roads were the most influencing factors of high hazard fire potential. Annual mean temperature was also found relevant in one of the models tested, whereas topographic conditions do not seem to influence fire hazard in the province. Considering the relevance of agricultural land use types and settlements density in fire hazard distribution, strategies to improve people's awareness and communities preparedness should be prioritized. Further work is needed regarding the integration of other variables potentially associated with fire likelihood in the region, such as evapotranspiration, dead and live fuel moisture or the location of tourism sites that are not yet available in a compatible format and scale. Additional research is also required to further examine the consistency of global datasets for modeling fire hazard at local and regional levels. This study aimed to improve the understanding of fire drivers in Mazandaran province, which can contribute to provide better mapping tools and further enhance decision-support systems used by fire managers and other operational organizations.

References

- Abdi, O., Kamkar, B., Shirvani, Z., Teixeira da Silva, J. A., & Buchroithner, M. F. (2016). *Spatial-statistical analysis of factors determining forest fires: A case study from Golestan, Northeast Iran* (pp. 1–14). *Natural Hazards and Risk: Geomatics*. <https://doi.org/10.1080/19475705.2016.1206629>.
- Abdi, O., Kamkar, B., Shirvani, Z., Teixeira da Silva, J. A., & Buchroithner, M. F. (2018). Spatial-statistical analysis of factors determining forest fires: A case study from Golestan, Northeast Iran. *Geomatics, Natural Hazards and Risk*, 9(1), 267–280. <https://doi.org/10.1080/19475705.2016.1206629>.
- Adab, H. (2017). Landfire hazard assessment in the Caspian Hyrcanian forest ecoregion with the long-term MODIS active fire data. *Natural Hazards*, 87(3), 1807–1825. <https://doi.org/10.1007/s11069-017-2850-2>.
- Adab, H., Devi Kanniah, K., & Beringer, J. (2016). Estimating and up-scaling fuel moisture and leaf dry matter content of a temperate humid forest using multi resolution remote sensing data. *Remote Sensing*, 8(11), 961.
- Adab, H., Kanniah, K. D., Solaimani, K., & Sallehuddin, R. (2015). Modelling static fire hazard in a semi-arid region using frequency analysis. *International Journal of Wildland Fire*, 24(6), 763–777. <https://doi.org/10.1071/WF13113>.
- Adel, M. N., Pourbabaei, H., & Dey, D. C. (2014). Ecological species group—environmental factors relationships in unharvested beech forests in the north of Iran. *Ecological Engineering*, 69, 1–7.
- Agee, J. K. (1996). *Fire ecology of Pacific Northwest forests*. Washington, D.C: Island press.
- Agee, J. K., Wakimoto, R. H., & Biswell, H. H. (1976). Fire and fuel dynamics of Sierra Nevada conifers. *Forest Ecology and Management*, 1, 255–265.
- Ajain, R., Loghin, A.-M., Vinod, P., & Jacob, M. (2016). Forest fire risk zone mapping using RS and GIS techniques: A study in Achankovil forest division, Kerala, India. [original article]. *Journal of Earth, Environment and Health Sciences*, 2(3), 109–115. <https://doi.org/10.4103/2423-7752.199288>.
- Akhani, H., Djamali, M., Ghorbanalizadeh, A., & Ramezani, E. (2010). Plant biodiversity of Hyrcanian relict forests, N Iran: An overview of the flora, vegetation, palaeoecology and conservation. *Pakistan Journal of Botany*, 42, 231–258.
- Allard, G. B. (2001). The fire situation in Islamic Republic of Iran. *Global Forest Fire Assessment 1990-2000* (pp. 495). Rome, Italy: FAO, Forestry Department.
- Alvares, C. A., Stape, J. L., Sentelhas, P. C., de Moraes, G., Leonardo, J., & Sparovek, G. (2013). Köppen's climate classification map for Brazil. *Meteorologische Zeitschrift*, 22(6), 711–728.
- Anderson, R. P., & Gonzalez Jr., I. (2011). Species-specific tuning increases robustness to sampling bias in models of species distributions: An implementation with Maxent. *Ecological Modelling*, 222(15), 2796–2811.
- Archibald, S., Lehmann, C. E. R., Gómez-Dans, J. L., & Bradstock, R. A. (2013). Defining pyromes and global syndromes of fire regimes. *Proceedings of the National Academy of Sciences*, 110(16), 6442–6447. <https://doi.org/10.1073/pnas.1211466110>.
- Arino, O., Gross, D., Ranera, F., Leroy, M., Bicheron, P., Brockman, C., et al. (2007). GlobCover: ESA service for global land cover from MERIS. In *Geoscience and Remote Sensing Symposium, 2007. IGARSS 2007. IEEE International*, (pp. 2412–2415): IEEE.
- Armenteras, D., Barreto, J. S., Tabor, K., Molowny-Horas, R., & Retana, J. (2017). Changing patterns of fire occurrence in proximity to forest edges, roads and rivers between NW Amazonian countries. *Biogeosciences*, 14(11), 2755–2765.
- Armién, A. G., Armién, B., Koster, F., Pascale, J. M., Avila, M., Gonzalez, P., et al. (2009). Hantavirus infection and habitat

- associations among rodent populations in agroecosystems of Panama: Implications for human disease risk. *The American Journal of Tropical Medicine and Hygiene*, 81(1), 59–66.
- Arpaci, A., Malowerschnig, B., Sass, O., & Vacik, H. (2014). Using multi variate data mining techniques for estimating fire susceptibility of Tyrolean forests. *Applied Geography*, 53, 258–270. <https://doi.org/10.1016/j.apgeog.2014.05.015>.
- Baldwin, R. A. (2009). Use of maximum entropy modeling in wildlife research. *Entropy*, 11(4), 854–866.
- Bani Assadi, S., Kaboli, M., Etemad, V., Ghadiri Khanapostani, M., & Tohidifar, M. (2015). Habitat selection of cavity-nesting birds in the Hyrcanian deciduous forests of northern Iran. *Ecological Research*, 30(5), 889–897. <https://doi.org/10.1007/s11284-015-1293-z>.
- Barros, A. M. G., Pereira, J. M. C. (2014). Wildfire Selectivity for Land Cover Type: Does Size Matter? *PLoS ONE*, 9(1), e84760. <https://doi.org/10.1371/journal.pone.0084760>.
- Bashari, H., Naghypour, A. A., Khajeddin, S. J., Sangoony, H., & Tahmasebi, P. (2016). Risk of fire occurrence in arid and semi-arid ecosystems of Iran: An investigation using Bayesian belief networks. *Environmental Monitoring and Assessment*, 188(9), 531. <https://doi.org/10.1007/s10661-016-5532-8>.
- Bistinas, I., Oom, D., Sá, A. C., Harrison, S. P., Prentice, I. C., & Pereira, J. M. (2013). Relationships between human population density and burned area at continental and global scales. *PLoS One*, 8(12), e81188.
- Boria, R. A., Olson, L. E., Goodman, S. M., & Anderson, R. P. (2014). Spatial filtering to reduce sampling bias can improve the performance of ecological niche models. *Ecological Modelling*, 275, 73–77.
- Bottero, A., D'Amato, A. W., Palik, B. J., Bradford, J. B., Fraver, S., Battaglia, M. A., & Asherin, L. A. (2017). Density-dependent vulnerability of forest ecosystems to drought. *Journal of Applied Ecology*, 54(6), 1605–1614. <https://doi.org/10.1111/1365-2664.12847>.
- Bowman, D. M. J. S., Balch, J., Artaxo, P., Bond, W. J., Cochrane, M. A., D'Antonio, C. M., DeFries, R., Johnston, F. H., Keeley, J. E., Krawchuk, M. A., Kull, C. A., Mack, M., Moritz, M. A., Pyne, S., Roos, C. I., Scott, A. C., Sodhi, N. S., & Swetnam, T. W. (2011). The human dimension of fire regimes on earth. *Journal of Biogeography*, 38(12), 2223–2236. <https://doi.org/10.1111/j.1365-2699.2011.02595.x>.
- Bowman, D. M. J. S., Balch, J. K., Artaxo, P., Bond, W. J., Carlson, J. M., Cochrane, M. A., D'Antonio, C. M., DeFries, R. S., Doyle, J. C., Harrison, S. P., Johnston, F. H., Keeley, J. E., Krawchuk, M. A., Kull, C. A., Marston, J. B., Moritz, M. A., Prentice, I. C., Roos, C. I., Scott, A. C., Swetnam, T. W., van der Werf, G. R., & Pyne, S. J. (2009). Fire in the earth system. *Science*, 324(5926), 481–484. <https://doi.org/10.1126/science.1163886>.
- Brotons, L., Thuiller, W., Araújo, M. B., & Hirzel, A. H. (2004). Presence-absence versus presence-only modelling methods for predicting bird habitat suitability. *Ecography*, 27(4), 437–448.
- Brown, J. L. (2014). SDMtoolbox: A python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *Methods in Ecology and Evolution*, 5(7), 694–700.
- Brown, J. L., Bennett, J. R., & French, C. M. (2017). SDMtoolbox 2.0: The next generation python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *PeerJ*, 5, e4095.
- Brown, S., & Lugo, A. E. (1982). The storage and production of organic matter in tropical forests and their role in the global carbon cycle. *Biotropica*, 14, 161–187.
- Caldararo, N. J. S. o.t. T. E. (2002). Human ecological intervention and the role of forest fires in human ecology. 292(3), 141–176.
- Carmo, M., Moreira, F., Casimiro, P., & Vaz, P. (2011). Land use and topography influences on wildfire occurrence in northern Portugal. *Landscape and Urban Planning*, 100(1–2), 169–176.
- Chandra, K., & Bhardwaj, A. K. (2015). Incidence of forest fire in India and its effect on terrestrial ecosystem dynamics, nutrient and microbial status of soil. *International Journal of Agriculture and Forestry*, 5(2), 69–78.
- Chen, F., Du, Y., Niu, S., & Zhao, J. (2015). Modeling forest lightning fire occurrence in the Daxinganling mountains of northeastern China with MAXENT. *Forests*, 6(5), 1422–1438.
- Chuvieco, E., Martínez, S., Román, M. V., Hantson, S., & Pettinari, M. L. (2014). Integration of ecological and socio-economic factors to assess global vulnerability to wildfire. *Global Ecology and Biogeography*, 23(2), 245–258. <https://doi.org/10.1111/geb.12095>.
- Cohen, J. (1960). A coefficient of agreement for nominal scales. *Educational and Psychological Measurement*, 20(1), 37–46.
- Connor, C. D. O., Calkin, D. E., & Thompson, M. P. (2017). An empirical machine learning method for predicting potential fire control locations for pre-fire planning and operational fire management. *International Journal of Wildland Fire*, 26(7), 587–597. <https://doi.org/10.1071/WF16135>.
- Conrad, O., Bechtel, B., Bock, M., Dietrich, H., Fischer, E., Gerlitz, L., et al. (2015). System for automated geoscientific analyses (SAGA) v. 2.1. 4. *Geoscientific Model Development Discussions*, 8(2), 2271–2312.
- Curt, T., & Delcros, P. (2010). Managing road corridors to limit fire hazard. A simulation approach in southern France. *Ecological Engineering*, 36(4), 457–465.
- De la Riva, J., Pérez-Cabello, F., Lana-Renault, N., & Koutsias, N. (2004). Mapping wildfire occurrence at regional scale. *Remote Sensing of Environment*, 92(3), 363–369.
- Devisscher, T., Anderson, L. O., Aragão, L. E., Galván, L., & Malhi, Y. (2016). Increased wildfire risk driven by climate and development interactions in the Bolivian Chiquitania, Southern Amazonia. *PLoS One*, 11(9), e0161323.
- Dimitrakopoulos, A. P. (2002). Mediterranean fuel models and potential fire behaviour in Greece. *International Journal of Wildland Fire*, 11(2), 127–130. <https://doi.org/10.1071/WF02018>.
- Dwyer, E., Grégoire, J.-M., & Pereira, J. M. C. (2000). Climate and vegetation as driving factors in global fire activity. In J. L. Innes, M. Beniston, & M. M. Verstraete (Eds.), *Biomass burning and its inter-relationships with the climate system* (pp. 171–191). Dordrecht: Springer Netherlands.
- Earl, D. E. (1975). *Forest energy and economic development*. Clarendon Press.
- Elith, J., Graham, C. H., Anderson, R. P., Dudík, M., Ferrier, S., Guisan, A., et al. (2006). Novel methods improve prediction of species' distributions from occurrence data. *Ecography*, 29(2), 129–151.

- Emadi, M., Shahriari, A. R., Sadegh-Zadeh, F., Jalili Seh-Bardan, B., & Dindarlou, A. (2016). Geostatistics-based spatial distribution of soil moisture and temperature regime classes in Mazandaran province, northern Iran. *Archives of Agronomy and Soil Science*, 62(4), 502–522. <https://doi.org/10.1080/03650340.2015.1065607>.
- Eshliki, S. A., & Kaboudi, M. (2012). Community perception of tourism impacts and their participation in tourism planning: A case study of Ramsar, Iran. *Procedia - Social and Behavioral Sciences*, 36, 333–341. <https://doi.org/10.1016/j.sbspro.2012.03.037>.
- Eskandari, S., & Chuvieco, E. (2015). Fire danger assessment in Iran based on geospatial information. *International Journal of Applied Earth Observation and Geoinformation*, 42, 57–64.
- FAO (2015). Global forest resources assessment 2015—desk reference. (pp. 253). Rome: FAO/UN.
- Fick, S. E., & Hijmans, R. J. (2017). WorldClim 2: New 1-km spatial resolution climate surfaces for global land areas. *International journal of climatology*, n/a-n/a, 37, 4302–4315. <https://doi.org/10.1002/joc.5086>.
- Flannigan, M. D., Amiro, B. D., Logan, K. A., Stocks, B., & Wotton, B. (2006). Forest fires and climate change in the 21st century. *Mitigation and Adaptation Strategies for Global Change*, 11(4), 847–859.
- Flannigan, M. D., Krawchuk, M. A., de Groot, W. J., Wotton, B. M., & Gowman, L. M. (2009). Implications of changing climate for global wildland fire. *International Journal of Wildland Fire*, 18(5), 483–507.
- Franklin, J. (2010). *Mapping species distributions: Spatial inference and prediction*. Cambridge: Cambridge University Press.
- Ganteaume, A., Camia, A., Jappiot, M., San-Miguel-Ayanz, J., Long-Fournel, M., & Lampin, C. (2013). A review of the main driving factors of forest fire ignition over Europe. *Environmental Management*, 51(3), 651–662.
- Giglio, L., Randerson, J. T., & Werf, G. R. (2013). Analysis of daily, monthly, and annual burned area using the fourth-generation global fire emissions database (GFED4). *Journal of Geophysical Research: Biogeosciences*, 118(1), 317–328.
- Giglio, L., Schroeder, W., & Justice, C. O. (2016). The collection 6 MODIS active fire detection algorithm and fire products. *Remote Sensing of Environment*, 178, 31–41. <https://doi.org/10.1016/j.rse.2016.02.054>.
- Grégoire, J.-M., Eva, H. D., Belward, A. S., Palumbo, I., Simonetti, D., & Brink, A. (2013). Effect of land-cover change on Africa's burnt area. *International Journal of Wildland Fire*, 22(2), 107.
- Guillera-Arroita, G., Lahoz-Monfort, J. J., & Elith, J. (2014). Maxent is not a presence-absence method: A comment on Thibaud et al. *Methods in Ecology and Evolution*, 5(11), 1192–1197.
- Guo, F., Su, Z., Wang, G., Sun, L., Tigabu, M., Yang, X., et al. (2017). Understanding fire drivers and relative impacts in different Chinese forest ecosystems. *Science of the Total Environment*, 605, 411–425.
- Hantson, S., Lasslop, G., Kloster, S., & Chuvieco, E. (2015). Anthropogenic effects on global mean fire size. *International Journal of Wildland Fire*, 24(5), 589–596. <https://doi.org/10.1071/WF14208>.
- Hantson, S., Padilla, M., Corti, D., & Chuvieco, E. (2013). Strengths and weaknesses of MODIS hotspots to characterize global fire occurrence. *Remote Sensing of Environment*, 131, 152–159.
- Hernandez, P. A., Graham, C. H., Master, L. L., & Albert, D. L. (2006). The effect of sample size and species characteristics on performance of different species distribution modeling methods. *Ecography*, 29(5), 773–785.
- Itziar, R. U., Gonzalo, Z., Joaquín, B., José, M. G., Jesús San, M.-A., Andrea, C., et al. (2015). Fire activity as a function of fire-weather seasonal severity and antecedent climate across spatial scales in southern Europe and Pacific western USA. *Environmental Research Letters*, 10(11), 114013.
- Jafarzadeh, A. A., Mahdavi, A., & Jafarzadeh, H. (2017). Evaluation of forest fire risk using the Apriori algorithm and fuzzy c-means clustering. *Journal of Forest Science*, 63(8), 370–380.
- Jahdi, R., Salis, M., Darvishsefat, A. A., Alcasena, F., Mostafavi, M. A., Etemad, V., Lozano, O. M., & Spano, D. (2016). Evaluating fire modelling systems in recent wildfires of the Golestan National Park, Iran. *Forestry: An International Journal of Forest Research*, 89(2), 136–149. <https://doi.org/10.1093/forestry/cpv045>.
- Justice, C., Giglio, L., Korontzi, S., Owens, J., Morisette, J., Roy, D., et al. (2002). The MODIS fire products. *Remote Sensing of Environment*, 83(1), 244–262.
- Kadej, M., Zajac, K., Smolis, A., Tarnawski, D., Tyszecka, K., Malkiewicz, A., et al. (2017). The great capricorn beetle *Cerambyx cerdo* L. in south-western Poland—The current state and perspectives of conservation in one of the recent distribution centres in Central Europe. *Nature Conservation*, 19, 111–134. <https://doi.org/10.3897/natureconservation.19.11838>.
- Kasischke, E. S., Verbyla, D. L., Rupp, T. S., McGuire, A. D., Murphy, K. A., Jandt, R., Barnes, J. L., Hoy, E. E., Duffy, P. A., Calef, M., & Turetsky, M. R. (2010). Alaska's changing fire regime—Implications for the vulnerability of its boreal forests this article is one of a selection of papers from the dynamics of change in Alaska's boreal forests: Resilience and vulnerability in response to climate warming. *Canadian Journal of Forest Research*, 40(7), 1313–1324.
- Ketterings, Q. M., & Bigham, J. M. (2000). Soil color as an indicator of slash-and-burn fire severity and soil fertility in Sumatra, Indonesia. *Soil Science Society of America Journal*, 64(5), 1826–1833.
- Kibria, A. S., Behie, A., Costanza, R., Groves, C., & Farrell, T. (2017). The value of ecosystem services obtained from the protected forest of Cambodia: The case of Veun Sai-Siem Pang National Park. *Ecosystem Services*, 26, 27–36.
- Knorr, W., Kaminski, T., Armeth, A., & Weber, U. (2014). Impact of human population density on fire frequency at the global scale. *Biogeosciences*, 11(4), 1085–1102. <https://doi.org/10.5194/bg-11-1085-2014>.
- Korontzi, S., McCarty, J., Loboda, T., Kumar, S., & Justice, C. (2006). Global distribution of agricultural fires in croplands from 3 years of moderate resolution imaging spectroradiometer (MODIS) data. *Global Biogeochemical Cycles*, 20(2).
- Kramer-Schadt, S., Niedballa, J., Pilgrim, J. D., Schröder, B., Lindenborn, J., Reinfelder, V., Stillfried, M., Heckmann, I., Scharf, A. K., Augeri, D. M., Cheyne, S. M., Hearn, A. J., Ross, J., Macdonald, D. W., Mathai, J., Eaton, J., Marshall,

- A. J., Semiadi, G., Rustam, R., Bernard, H., Alfred, R., Samejima, H., Duckworth, J. W., Breitenmoser-Wuersten, C., Belant, J. L., Hofer, H., & Wilting, A. (2013). The importance of correcting for sampling bias in MaxEnt species distribution models. *Diversity and Distributions*, *19*(11), 1366–1379.
- Krawchuk, M., & Moritz, M. (2014). Burning issues: Statistical analyses of global fire data to inform assessments of environmental change. *Environmetrics*, *25*(6), 472–481.
- Krawchuk, M. A., Moritz, M. A., Parisien, M.-A., Van Dorn, J., & Hayhoe, K. (2009). Global pyrogeography: The current and future distribution of wildfire. *PLoS One*, *4*(4), e5102.
- Kumar, S., & Stohlgren, T. J. (2009). Maxent modeling for predicting suitable habitat for threatened and endangered tree *Canacomyrica monticola* in New Caledonia. *Journal of Ecology and the Natural Environment*, *1*(4), 094–098.
- Lentz, D. L., Bye, R., & Sánchez-Cordero, V. (2008). Ecological niche modeling and distribution of wild sunflower (*Helianthus annuus* L.) in Mexico. *International Journal of Plant Sciences*, *169*(4), 541–549.
- Lezama-Ochoa, N., Murua, H., Chust, G., Van Loon, E., Ruiz, J., Hall, M., et al. (2016). Present and future potential habitat distribution of *Carcharhinus falciformis* and *Canthidermis maculata* by-catch species in the tropical tuna purse-seine fishery under climate change. [original research]. *Frontiers in Marine Science*, *3*(34). <https://doi.org/10.3389/fmars.2016.00034>.
- Littell, J. S., McKenzie, D., Peterson, D. L., & Westerling, A. L. (2009). Climate and wildfire area burned in western US ecoregions, 1916–2003. *Ecological Applications*, *19*(4), 1003–1021.
- Liu, C., Berry, P. M., Dawson, T. P., & Pearson, R. G. (2005). Selecting thresholds of occurrence in the prediction of species distributions. *Ecography*, *28*(3), 385–393.
- Louzao, M., Delord, K., Garcia, D., Boué, A., & Weimerskirch, H. (2012). Protecting persistent dynamic oceanographic features: Transboundary conservation efforts are needed for the critically endangered balearic shearwater. *PLoS One*, *7*(5), e35728.
- MacKenzie, D. I., Nichols, J. D., Lachman, G. B., Droege, S., Andrew Royle, J., & Langtimm, C. A. (2002). Estimating site occupancy rates when detection probabilities are less than one. *Ecology*, *83*(8), 2248–2255.
- MacMillan, R., & Shary, P. (2008). *Geomorphometry: Concepts, software, applications*. Elsevier Science. Chap. Landforms and Landforms elements in geomorphometry.
- Mahdavi, A. (2012). Forests and rangelands? Wildfire risk zoning using GIS and AHP techniques. *Caspian Journal of Environmental Sciences*, *10*(1), 43–52.
- Mahdavi, A., & Naghdí, R. (2007). Information and data flow analysis for forestry sector in Iran as a basic requirement for designing a forest information system (FIS). *Caspian Journal of Environmental Sciences*, *5*(2), 147–153.
- Marchal, J., Cumming, S. G., & McIntire, E. J. (2017). Land cover, more than monthly fire weather, drives fire-size distribution in southern Québec forests: Implications for fire risk management. *PLoS One*, *12*(6), e0179294.
- Martínez-Fernández, J., Chuvieco, E., & Koutsias, N. (2013). Modelling long-term fire occurrence factors in Spain by accounting for local variations with geographically weighted regression. *Natural Hazards and Earth System Sciences*, *13*(2), 311–327.
- Massada, A. B., Syphard, A. D., Stewart, S. I., & Radeloff, V. C. (2013). Wildfire ignition-distribution modelling: A comparative study in the Huron–Manistee National Forest, Michigan, USA. *International Journal of Wildland Fire*, *22*(2), 174–183.
- Matthews, S. (2014). Dead fuel moisture research: 1991–2012. *International Journal of Wildland Fire*, *23*(1), 78–92.
- Miller, R. H., Masuoka, P., Klein, T. A., Kim, H.-C., Somer, T., & Grieco, J. (2012). Ecological niche modeling to estimate the distribution of Japanese encephalitis virus in Asia. *PLoS Neglected Tropical Diseases*, *6*(6), e1678.
- Mirzaei, R. (2013). *Modeling the socioeconomic and environmental impacts of nature-based tourism to the host communities and their support for tourism : perceptions of local population, Mazandaran, north of Iran*. Universitätsbibliothek, Gießen.
- Modugno, S., Balzter, H., Cole, B., & Borrelli, P. (2016). Mapping regional patterns of large forest fires in wildland–urban interface areas in Europe. *Journal of Environmental Management*, *172*, 112–126. <https://doi.org/10.1016/j.jenvman.2016.02.013>.
- Mohammadi, F., Bavaghar, M. P., & Shabaniyan, N. (2014). Forest fire risk zone modeling using logistic regression and GIS: An Iranian case study. *Small-scale Forestry*, *13*(1), 117–125. <https://doi.org/10.1007/s11842-013-9244-4>.
- Monserud, R. A., & Leemans, R. (1992). Comparing global vegetation maps with the kappa statistic. *Ecological Modelling*, *62*(4), 275–293.
- Moreira, F., Vaz, P., Catry, F., & Silva, J. S. (2009). Regional variations in wildfire susceptibility of land-cover types in Portugal: Implications for landscape management to minimize fire hazard. *International Journal of Wildland Fire*, *18*(5), 563–574.
- Najafabadi, A. T. P., Gorgani, F., & Najafabadi, M. O. (2015). Modeling forest fires in Mazandaran Province, Iran. *Journal of Forestry Research*, *26*(4), 851–858. <https://doi.org/10.1007/s11676-015-0107-z>.
- Nami, M., Jaafari, A., Fallah, M., & Nabiuni, S. (2018). Spatial prediction of wildfire probability in the Hyrcanian ecoregion using evidential belief function model and GIS. *International journal of Environmental Science and Technology*, *15*(2), 373–384.
- Nilsson, K., Sangster, M., & Konijnendijk, C. C. (2011). Forests, trees and human health and well-being: Introduction. In *Forests, trees and human health* (pp. 1–19): Springer.
- Nunes, A., Lourenço, L., & Meira, A. C. (2016). Exploring spatial patterns and drivers of forest fires in Portugal (1980–2014). *Science of the Total Environment*, *573*, 1190–1202.
- Oliveira, S., Félix, F., Nunes, A., Lourenço, L., Laneve, G., & Sebastián-López, A. (2018). Mapping wildfire vulnerability in Mediterranean Europe. Testing a stepwise approach for operational purposes. *Journal of Environmental Management*, *206*, 158–169. <https://doi.org/10.1016/j.jenvman.2017.10.003>.
- Oliveira, S., Oehler, F., San-Miguel-Ayán, J., Camia, A., & Pereira, J. M. (2012). Modeling spatial patterns of fire occurrence in Mediterranean Europe using multiple regression and random forest. *Forest Ecology and Management*, *275*, 117–129.

- Oliveira, S., Pereira, J. M. C., San-Miguel-Ayanz, J., & Lourenço, L. (2014). Exploring the spatial patterns of fire density in southern Europe using geographically weighted regression. *Applied Geography*, *51*, 143–157. <https://doi.org/10.1016/j.apgeog.2014.04.002>.
- Oliveira, S., Zêzere, J. L., Queirós, M., & Pereira, J. M. (2017). Assessing the social context of wildfire-affected areas. The case of mainland Portugal. *Applied Geography*, *88*, 104–117.
- Özbayoğlu, A. M., & Bozer, R. (2012). Estimation of the burned area in forest fires using computational intelligence techniques. *Procedia Computer Science*, *12*, 282–287.
- Pahlavani, P., & Bigdeli, B. (2017). Providing the fire risk map in forest area using a geographically weighted regression model with gaussian kernel and modis images, a case study: Golestan Province. *International Archives of the Photogrammetry, Remote Sensing & Spatial Information Sciences*, 42.
- Parente, J., Pereira, M. G., & Tonini, M. (2016). Space-time clustering analysis of wildfires: The influence of dataset characteristics, fire prevention policy decisions, weather and climate. *Science of the Total Environment*, *559*(Supplement C), 151–165. <https://doi.org/10.1016/j.scitotenv.2016.03.129>.
- Parisien, M.-A., Miller, C., Parks, S. A., DeLancey, E. R., Robinne, F.-N., & Flannigan, M. D. (2016). The spatially varying influence of humans on fire probability in North America. *Environmental Research Letters*, *11*(7), 075005.
- Parisien, M.-A., & Moritz, M. A. (2009). Environmental controls on the distribution of wildfire at multiple spatial scales. *Ecological Monographs*, *79*(1), 127–154.
- Pearson, R. G., Raxworthy, C. J., Nakamura, M., & Townsend Peterson, A. (2007). Predicting species distributions from small numbers of occurrence records: A test case using cryptic geckos in Madagascar. *Journal of Biogeography*, *34*(1), 102–117.
- Penman, T. D., Bradstock, R. A., & Price, O. (2013). Modelling the determinants of ignition in the Sydney Basin, Australia: Implications for future management. *International Journal of Wildland Fire*, *22*(4), 469–478. <https://doi.org/10.1071/WF12027>.
- Pereira, M. G., Aranha, J., & Amraoui, M. (2014). Land cover fire proneness in Europe. *Forest Systems*, *23*(3), 598.
- Phillips, S. J., Anderson, R. P., & Schapire, R. E. (2006). Maximum entropy modeling of species geographic distributions. *Ecological Modelling*, *190*(3), 231–259.
- Reed, D. J., Spencer, T., Murray, A. L., French, J. R., & Leonard, L. (1999). Marsh surface sediment deposition and the role of tidal creeks: Implications for created and managed coastal marshes. *Journal of Coastal Conservation*, *5*(1), 81–90.
- Renard, Q., Pélissier, R., Ramesh, B. R., & Kodandapani, N. (2012). Environmental susceptibility model for predicting forest fire occurrence in the Western Ghats of India. *International Journal of Wildland Fire*, *21*(4), 368–379. <https://doi.org/10.1071/WF10109>.
- Richards, S. A., Possingham, H. P., & Tizard, J. (1999). Optimal fire management for maintaining community diversity. *Ecological Applications*, *9*(3), 880–892. [https://doi.org/10.1890/1051-0761\(1999\)009\[0880:OFMFC\]2.0.CO;2](https://doi.org/10.1890/1051-0761(1999)009[0880:OFMFC]2.0.CO;2).
- Rodrigues, M., & de la Riva, J. (2014). An insight into machine-learning algorithms to model human-caused wildfire occurrence. *Environmental Modelling & Software*, *57*, 192–201.
- Romero-Calcerrada, R., Novillo, C., Millington, J., & Gomez-Jimenez, I. (2008). GIS analysis of spatial patterns of human-caused wildfire ignition risk in the SW of Madrid (Central Spain). *Landscape Ecology*, *23*(3), 341–354.
- Ruiz-Mirazo, J., Martínez-Fernández, J., & Vega-García, C. J. J. o. e. m. (2012). Pastoral wildfires in the Mediterranean: Understanding their linkages to land cover patterns in managed landscapes. *98*, 43–50.
- Russo, A., Gouveia, C. M., Páscoa, P., DaCamara, C. C., Sousa, P. M., & Trigo, R. M. (2017). Assessing the role of drought events on wildfires in the Iberian Peninsula. *Agricultural and Forest Meteorology*, *237–238*(Supplement C), 50–59. <https://doi.org/10.1016/j.agrformet.2017.01.021>.
- San-Miguel-Ayanz, J., Rodrigues, M., de Oliveira, S. S., Pacheco, C. K., Moreira, F., Duguay, B., et al. (2012). Land cover change and fire regime in the European Mediterranean region. In *Post-fire management and restoration of southern European forests* (pp. 21–43): Springer.
- Satir, O., Berberoglu, S., & Donmez, C. (2016). Mapping regional forest fire probability using artificial neural network model in a Mediterranean forest ecosystem. *Geomatics, Natural Hazards and Risk*, *7*(5), 1645–1658. <https://doi.org/10.1080/19475705.2015.1084541>.
- Satoh, K., Song, W., & Yang, K. T. (2004). A study of forest fire danger prediction system in Japan. In Proceedings. 15th International Workshop on Database and Expert Systems Applications, 2004., 30 Aug.-3 Sept. 2004 (pp. 598–602). doi:<https://doi.org/10.1109/DEXA.2004.1333540>.
- Schamweber, T., Rietschel, M., & Manthey, M. (2007). Degradation stages of the Hyrcanian forests in southern Azerbaijan. *Archiv für Naturschutz und Landschaftsforschung*, *46*(2), 133–156.
- Schoennagel, T., Veblen, T. T., & Romme, W. H. (2004). The interaction of fire, fuels, and climate across Rocky Mountain forests. *BioScience*, *54*(7), 661–676. [https://doi.org/10.1641/0006-3568\(2004\)054\[0661:TIOFFA\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2004)054[0661:TIOFFA]2.0.CO;2).
- Shafiei, A. B., Akbarinia, M., Jalali, G., & Hosseini, M. (2010). Forest fire effects in beech dominated mountain forest of Iran. *Forest Ecology and Management*, *259*(11), 2191–2196.
- Shvidenko, A., & Gonzalez, P. (2005). Forest and woodland systems. In R. Hassan, R. Scholes, & N. Ash (Eds.), *Ecosystems and human well-being: Current state and trends, vol 1. Findings of the condition and trends working group of the Millennium Ecosystem Assessment* (pp. 587–621). Washington, DC: Island Press.
- Silva, J. S., Moreira, F., Vaz, P., Catry, F., & Godinho-Ferreira, P. (2009). Assessing the relative fire proneness of different forest types in Portugal. *Plant Biosystems - An International Journal Dealing with all Aspects of Plant Biology*, *143*(3), 597–608.
- Small, R. D., & Bush, B. W. (1985). Smoke production from multiple nuclear explosions in nonurban areas. *Science*, *229*(4712), 465–469. <https://doi.org/10.1126/science.229.4712.465>.
- Sobhani, A., & Khosravi, H. (2015). Assessing environmental sensitivity areas to desertification in north of Iran. *Current World Environment*, *10*(3), 890–902.
- Sørensen, R., Zinko, U., & Seibert, J. (2006). On the calculation of the topographic wetness index: evaluation of different methods based on field observations. *Hydrology and Earth System Sciences*, *10*(1), 101–112.
- Stohlgren, T. J. (1988). Litter dynamics in two Sierran mixed conifer forests. I. Litterfall and decomposition rates. *Canadian Journal of Forest Research*, *18*(9), 1127–1135.

- Šturm, T., & Podobnikar, T. (2017). A probability model for long-term forest fire occurrence in the Karst forest management area of Slovenia. *International Journal of Wildland Fire*, 26(5), 399–412. <https://doi.org/10.1071/WF15192>.
- Syphard, A. D., Radeloff, V. C., Keeley, J. E., Hawbaker, T. J., Clayton, M. K., Stewart, S. I., & Hammer, R. B. (2007). Human influence on California fire regimes. *Ecological Applications*, 17(5), 1388–1402. <https://doi.org/10.1890/06-1128.1>.
- Syphard, A. D., Radeloff, V. C., Keuler, N. S., Taylor, R. S., Hawbaker, T. J., Stewart, S. I., & Clayton, M. K. (2008). Predicting spatial patterns of fire on a southern California landscape. *International Journal of Wildland Fire*, 17(5), 602.
- Tachikawa, T., Hato, M., Kaku, M., & Iwasaki, A. (2011). Characteristics of ASTER GDEM version 2. In Geoscience and remote sensing symposium (IGARSS), 2011 IEEE international, (pp. 3657–3660): IEEE.
- Tien Bui, D., Bui, Q.-T., Nguyen, Q.-P., Pradhan, B., Nampak, H., & Trinh, P. T. (2017). A hybrid artificial intelligence approach using GIS-based neural-fuzzy inference system and particle swarm optimization for forest fire susceptibility modeling at a tropical area. *Agricultural and Forest Meteorology*, 233, 32–44. <https://doi.org/10.1016/j.agrformet.2016.11.002>.
- Tien Bui, D., Le, K.-T. T., Nguyen, V. C., Le, H. D., & Revhaug, I. (2016). Tropical forest fire susceptibility mapping at the cat Ba National Park Area, Hai Phong City, Vietnam, using GIS-based kernel logistic regression. *Remote Sensing*, 8(4), 347.
- UNISDR, U. Sendai framework for disaster risk reduction 2015–2030. In *3rd United Nations World Conference on DRR, 2015*: UNISDR Sendai, Japan.
- Vadrevu, K. P., Eaturu, A., & Badarinath, K. (2006). Spatial distribution of forest fires and controlling factors in Andhra Pradesh, India using spot satellite datasets. *Environmental Monitoring and Assessment*, 123(1–3), 75–96.
- Vadrevu, K. P., & Justice, C. (2011). Vegetation fires in the Asian region: Satellite observational needs and priorities. *Global Environmental Research*, 15(1), 65–76.
- Wimberly, M. C., & Reilly, M. J. (2007). Assessment of fire severity and species diversity in the southern Appalachians using Landsat TM and ETM+ imagery. *Remote Sensing of Environment*, 108(2), 189–197.
- Xu, D., Shao, G., Dai, L., Hao, Z., Tang, L., & Wang, H. (2006). Mapping forest fire risk zones with spatial data and principal component analysis. *Science in China Series E: Technological Sciences*, 49, 140–149.
- Zhang, Y., Lim, S., & Sharples, J. J. (2016). Modelling spatial patterns of wildfire occurrence in south-eastern Australia. *Geomatics, Natural Hazards and Risk*, 7(6), 1800–1815.