

Bioavailability assessment of phosphorus and metals in soils and sediments: a review of diffusive gradients in thin films (DGT)

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Abstract This paper provides an overview of the principle and latest development of the diffusive gradients in thin films (DGT) technology and its applications in environmental studies with a focus on bioavailability assessment of phosphorus and metals in sediments and soils. Compared with conventional methods, DGT, as a passive sampling method, has significant advantages: in situ measurement, time averaged concentrations and high spatial resolution. The in situ measurement avoids artificial influences including contamination of samples and sample treatment which may change the forms of chemicals. The time averaged concentration reflects representative measurement over a period of time. The high-resolution information captures the biogeochemical heterogeneity of elements of interest distributed in

microenvironments, such as in the rhizosphere and the vicinity of the sediment-water interface. Moreover, DGT is a dynamic technique which simultaneously considers the diffusion of solutes and their kinetic re-supply from the solid phases. All the advantages of DGT significantly promote the collection of “true” information of the bioavailable or labile forms of chemicals in the environment. DGT provides potential for applications in agriculture, environmental monitoring and the mining industry. However, the applications are still at the early testing stage. Further studies are needed to properly interpret the DGT-measured results under complex environmental conditions, and standard procedures and guideline values based on DGT are required to pave the way for its routine applications in environmental monitoring.

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Introduction

The diffusive gradients in thin films (DGT) technology provides a novel approach for the in situ measurement of the labile forms of chemical elements, such as phosphorus (P), sulphur (S), arsenic (As) and metals in waters, sediments and soils (Davison and Zhang 1994, 2012). It was invented in Lancaster by Bill Davison and Hao Zhang in 1993 (DGT Research Ltd. 2014; Davison and Zhang 1994). The simple device uses a hydrogel

binding layer impregnated with Chelex resin or other binding agents to accumulate ions. The binding layer is overlain by a diffusive layer of hydrogel and a filter. Ions have to diffuse through the filter and diffusive layer to reach the binding layer. It is the establishment of a constant concentration gradient in the diffusive layer that forms the basis for measuring chemical element concentrations in solution quantitatively. Deployment time can vary from 1 h to 3 months provided the binding resin is not saturated. The measurement is independent of ionic strength if the range is within 10 mM–1 M. For Chelex-100 as the binding resin, the measurement is independent of pH in the range of 5–8.3. The effect of temperature can be predicted from the known temperature dependence of the diffusion coefficient (Zhang and Davison 1995).

DGT can be used for many different purposes (DGT Research Ltd. 2014), including the following:

- In situ measurements
- Monitoring (time averaged concentrations)
- Speciation (labile inorganic and/or organic species)
- Bioavailability (effective concentration)
- Fluxes in sediments and soils
- Kinetic and thermodynamic constants
- High spatial resolution measurements (sub-mm)
- 2D concentration images

This paper provides an overview of DGT theory and its recent development and applications in bioavailability assessment with comments on its future development. The main purpose of this review is to promote the use of DGT; thus, the scope is restricted to the general concepts and applications without technical details which are available elsewhere (Davison and Zhang 1994, 2012; DGT Research Ltd. 2014). For the applications, this review focuses on the uses of DGT in bioavailability assessment of phosphorus and metals in sediments and soils, while an extensive review was performed on a similar topic in water (INAP 2012).

Theory of DGT

The fundamental theory behind DGT is Fick's first law of diffusion. For deployment, the unit is immersed in water, inserted into sediments or placed in close contact with wet soils. The labile forms of chemical elements diffuse through the filter and diffusive gel and are then

adsorbed on the binding gel. The process is illustrated in Fig. 1, which is explained as follows.

The principles of DGT applied in water, sediments and soils are similar (Davison and Zhang 2012). When the unit is deployed, there is a diffusive boundary layer (DBL) between the diffusive layers (including both diffusive gel and filter membrane) and the solution. Within a few minutes, a steady state linear concentration gradient is established between the solution and the binding gel (Harper et al. 1998). The flux J of an ion through the diffusive gel is given by Fick's first law of diffusion (Eq. 1):

$$J = D \frac{dC}{dx} \quad (1)$$

where D is the diffusion coefficient and dC/dx is the concentration gradient. Assuming the diffusion coefficients in the diffusive layers and DBL are the same, the concentration gradient can be calculated as (Eq. 2):

$$\frac{dC}{dx} = \frac{C - C'}{\Delta_g + \delta} \quad (2)$$

where C is the concentration of an ion in solution, C' is the concentration at the boundary between the binding gel and the diffusive gel, Δ_g is the combined thickness of the diffusive gel layer and filter layer and δ is the boundary layer thickness. If the free ions are in rapid equilibrium with the binding agent, with a large binding constant, C' is effectively zero, provided the binding agent is not saturated. In well-stirred solutions, the boundary layer thickness δ is negligibly small. Therefore, Eq. 1 can be simplified as Eq. 3:

$$J = D \frac{C}{\Delta_g} \quad (3)$$

During the period of deployment, when the binding gel is not saturated, the flux can also be calculated based on the amount of accumulated mass (M) of an ion, exposure area (A) of the DGT unit and deployment time (t):

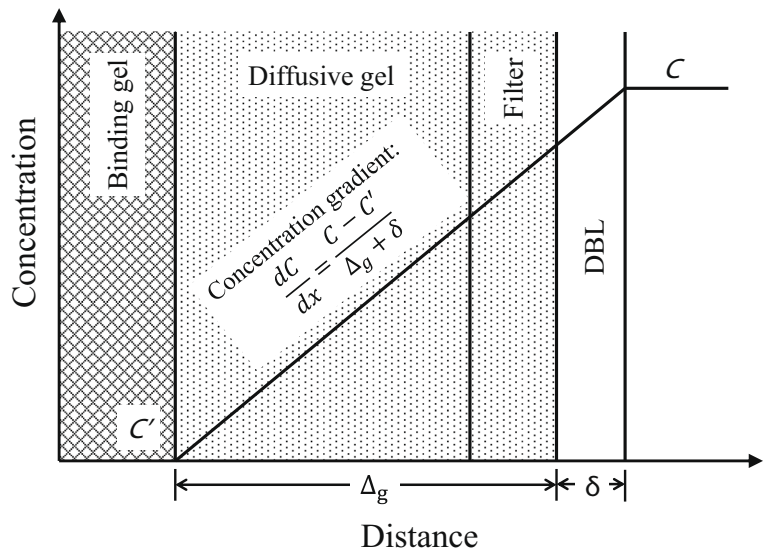
$$J = \frac{M}{At} \quad (4)$$

Combining Eqs. 3 and 4, the concentration C of an ion in the solution can be calculated as Eq. 5.

$$C = \frac{M \Delta_g}{DtA} \quad (5)$$

In this equation, the accumulated mass M can be measured directly in the binding-gel layer by drying it

Fig. 1 Simplified schematic cross section of a DGT (redrawn based on Davison and Zhang (1994) and Zhang and Davison (1995))



and using a beam technique such as laser ablation ICP-MS or, in the case of radionuclides, by direct counting. However, it is often obtained via elution of the ion from the binding gel into solution for routine measurements to minimize the cost. The accumulated mass can be calculated using the following Eq. 6.

$$M = \frac{C_e(V_g + V_e)}{f_e} \tag{6}$$

where C_e is the concentration of the ion in the elution solution, V_e is the volume of the elution solution, V_g is the volume of the gel and f_e is the elution efficiency. Diffusion coefficient D is related to the property of the diffusive gel and temperature. A list of diffusion coefficients for the most commonly used polyacrylamide gels cross-linked with an agarose derivative (APA) has been measured (Zhang and Davison 1999), and their tabulated values for different temperatures and elements are available from the dgtresearch.com website (DGT Research Ltd. 2014).

It needs to be noted that there are some assumptions for Eq. 5, such as the omissions of the concentration on binding layer C' and DBL thickness δ . The first omission is not regarded as problematic. For the influence of omitting DBL thickness δ in Eq. 5, it was suggested that the error caused would be much less than the typical experimental errors (e.g. 10 %) (Davison and Zhang 2012). The assumptions hold for most conditions, but there are exceptions and challenges, especially in systems with complex heterogeneous ligands (Davison and Zhang 2012).

Development of DGT

Since its invention 20 years ago, DGT technology has been developing rapidly. Significant developments have been made in two aspects of the technology: new binding gels for various analytes and 2D high-resolution measurements.

The analytes that can be measured by DGT are determined by the binding agent in use (Table 1). It is known that the Chelex resin can take up trace metals, as it contains paired iminodiacetate ions which act as chelating groups in binding polyvalent metal ions. The binding agent for the first DGT was Chelex resin (Davison and Zhang 1994; Zhang and Davison 1995). After that, the ferrihydrite gel with strong affinity for phosphorus was used to measure labile phosphorus (Zhang et al. 1998) and silver iodide was included in the gel to take up sulphide (Teasdale et al. 1999). The radioactive caesium was adsorbed by a gel containing ammonium molybdophosphate (Murdock et al. 2001). Recently, the Zr oxide gel was developed to measure phosphorus and inorganic arsenic with high capacities (Ding et al. 2010; Sun et al. 2013, 2014). It needs to be mentioned that the DGT method has recently been developed to measure organic compounds (Chen et al. 2012) which is beyond the scope of this review.

Besides measurements of individual chemical elements, techniques for simultaneous measurements of multiple elements have been developed. Two separate gels of silver iodide and Chelex-100 were used together to measure sulphide and metals

Table 1 Development of binding gels for measurements of various dissolved chemicals

Analyte(s)	Binding gel	Binding capacity or max. concentration	Reference
Metals	Chelex-100 resin	5×10^{-4} M Cd, 10-h deployment	Zhang and Davison (1995)
Phosphorus	Ferrihydrite (Fe oxide, based on $\text{Fe}(\text{NO}_3)_3 \cdot 9\text{H}_2\text{O}$)	$\sim 2 \mu\text{g P cm}^{-2}$ (0.4 mg L^{-1}), 24-h deployment	Zhang et al. (1998), Ding et al. (2010)
Sulphide	Silver iodide (AgI)	$1 \mu\text{mol cm}^{-2}$ ($62.6 \mu\text{mol L}^{-1}$, 15 h)	Teasdale et al. (1999)
Radioactive caesium	Ammonium molybdophosphate	2.3 mg of stable Cs (equivalent to 7 GBq ^{137}Cs or 108 GBq ^{134}Cs)	Murdock et al. (2001)
Phosphorus	Zr oxide (based on $\text{ZrOCl}_2 \cdot 8\text{H}_2\text{O}$)	$>100 \mu\text{g P cm}^{-2}$ ($>20 \text{ mg L}^{-1}$), 24-h deployment	Ding et al. (2010)
Sulphide (simultaneous measurements with metals)	AgI in gel (combined use with Chelex-100 resin)	N/A	Motelica-Heino et al. (2003)
Arsenic, phosphorus and metals	Titanium dioxide (Metsorb)	$\sim 12 \mu\text{g P cm}^{-2}$ (2.5–5 times higher than ferrihydrite)	Bennett et al. (2010), Panther et al. (2010, 2013)
Phosphorus and cations	Ferrihydrite and Chelex-100 (mixed binding layer, MBL)	$12 \mu\text{g P}$ on gel (2.52 cm^2 , $850 \mu\text{g L}^{-1}$; $75 \mu\text{g Mn}$ in gel (3 mg L^{-1})), 24-h deployment	Mason et al. (2005)
Phosphorus and sulphur	Zr oxide-silver iodide (AgI)	$\sim 80 \mu\text{g P cm}^{-2}$ ($>20 \text{ mg L}^{-1}$), 24-h deployment	Ding et al. (2012)
Potassium and phosphorus	Mixed Amberlite and ferrihydrite	$450 \mu\text{g K}$ on gel (2.54 cm^2); 1.2 mg P L^{-1} , 8-h deployment	Zhang et al. (2013)
Phosphorus and iron	Zr oxide-silver iodide (AgI)	$90 \mu\text{g P cm}^{-2}$ and $75 \mu\text{g Fe cm}^{-2}$	Xu et al. (2013)
Arsenic	Zr oxide	$159 \mu\text{g As(III)}$ and $434 \mu\text{g As(V)}$ in freshwater per device	Sun et al. (2014)

in sediments (Motelica-Heino et al. 2003). A mixed binding layer (MBL) containing a mixture of ferrihydrite and Chelex-100 was developed to measure phosphorus and cations (Mason et al. 2005). A titanium dioxide gel-assembled DGT has been used to simultaneously measure arsenic, phosphorus and metals (Bennett et al. 2010; Panther et al. 2010, 2013). The hydrous zirconium oxide (Zr oxide) has been combined with silver iodide to measure both phosphorus and sulphide (Ding et al. 2012) and combined with Chelex to measure phosphorus and iron (Xu et al. 2013). The mixed Amberlite and ferrihydrite gel has been recently developed to measure potassium and phosphorus (Zhang et al. 2013). Recently, it has been found that the capacities of the Zr oxide DGT for As in freshwater and seawater were 5–19 times and 3–13 times more than those reported for the commonly used ferrihydrite and Metsorb DGTs, respectively (Sun et al. 2014).

Another significant development in DGT is the 2D high-resolution measurement, which provides new evidences for the micro-scale biogeochemical heterogeneity of sediments (Stockdale et al. 2009). The scales have generally reached sub-millimetre level using various

technologies, including proton-induced X-ray emissions (Davison et al. 1997), computer-imaging densitometry (Teasdale et al. 1999; Devries and Wang 2003; Jezequel et al. 2007; Robertson et al. 2008; Ding et al. 2013), laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS) (Warnken et al. 2004; Stockdale et al. 2008; Santner et al. 2012; Stahl et al. 2012) and 2D slicing (Ding et al. 2011, 2012) (Table 2). Besides measurements for a single analyte in 2D high resolution, simultaneous measurements of multiple analytes have received attention, which enable the investigation of the relationships among multiple elements during their transport processes in pore waters and from the solid phases, e.g. Zn, Mn, Fe and As (Davison et al. 1997), iron and sulphur (Jezequel et al. 2007; Robertson et al. 2008), phosphorus, vanadium and arsenic (Stockdale et al. 2008, 2010), phosphorus and sulphide (Ding et al. 2012) and iron and phosphorus (Xu et al. 2013).

Development of new binding gels enables DGT to analyse diverse analytes in the environment, especially those of environmental significance, and to collect high-resolution information on a 2D level. The simultaneous measurements of multiple analytes are helpful to capture their coherent dynamics between pore waters and solid

Table 2 Development of 2D high-resolution analysis using DGT at the sub-millimetre scales

Analyte(s)	Resolution	2D acquisition method	Reference
Zn, Mn, Fe and As	0.1 mm	Proton-induced X-ray emissions	Davison et al. (1997)
Sulphide	0.238 mm	Computer-imaging densitometry	Teasdale et al. (1999)
Sulphide	≤0.4 mm	Computer-imaging densitometry	Devries and Wang (2003)
Trace metals	0.1 mm	Laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS)	Warnken et al. (2004)
Iron and sulphur	0.13 mm	Computer-imaging densitometry	Jezequel et al. (2007)
Phosphorus, vanadium and arsenic	0.333 mm	Laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS)	Stockdale et al. (2008, 2010)
Iron and sulphide	1.0 mm	Computer-imaging densitometry	Robertson et al. (2008)
Phosphorus	0.45 mm	2D slicing of binding gel	Ding et al. (2011)
Phosphorus and sulphide	0.45 mm (P), 0.169 mm (S)	2D slicing of binding gel (P), computer-imaging densitometry (S)	Ding et al. (2012)
Phosphorus	0.05 mm×0.333 mm	Laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS)	Santner et al. (2012)
Trace metals	0.2 mm	Laser ablation inductively coupled plasma mass spectrometry (LA-ICP-MS)	Stahl et al. (2012)
Phosphorus	0.42 mm	Computer-imaging densitometry	Ding et al. (2013)

phases, and the 2D high-resolution measurements provide novel information of their micro-scale heterogeneity. Further development is needed to increase the range of analytes measured both individually and simultaneously, and easy and cheap methods are needed to make the 2D high-resolution measurements more accessible. One of the main issues should be the comparability of the results from different binding gels. Standardized DGT methods are required to ensure that the DGT results measured using different binding gels and under different conditions can be meaningfully compared. The capacity of a binding gel layer for DGT response is a key limiting factor for a wide application of DGT. Taking the complex background in natural environments into consideration, it is vitally important to develop and apply high-capacity DGT techniques to ensure a reliable and robust measurement especially under challenging environments.

Applications of DGT in bioavailability assessment

Applications of DGT in soil P assessment

Phosphorus (P) is an essential element for plant growth and is also added to animal feed (Bomans et al. 2005). In soils, phosphorus exists in different forms: associated

with soil particles, adsorbed on Fe-Al oxides or Ca-carbonates, incorporated in organic matter and, to a much lesser extent, in soluble form dissolved in the soil solution (Bomans et al. 2005). Since only a small portion of P is available for plant uptake, it is often amended via fertilizer application, and P becomes one of the most expensive nutrients in agriculture (van Raij et al. 2002). Economic benefit can be achieved via accurate assessment of P status in soils to maximize fertilizer efficiency (Mason et al. 2010).

It is recognized that the total concentrations of P in soils fail to provide sufficient information for the assessment of its bioavailability. However, the established soil testing methods for assessment of available phosphorus (P), such as Colwell (Colwell 1963) and Olsen (Olsen et al. 1954), have been shown to fail to reliably predict plant P requirements over a range of soil types (e.g. calcareous, acidic with high iron or aluminium) (Holdford et al. 1985; McBeath et al. 2005; Mason et al. 2010). Recently, it has been demonstrated that the DGT method predicted plant responsiveness to applied P more accurately than Colwell P (Burkitt et al. 2002; Moody 2007; Mason et al. 2010, 2013). In an experiment comparing the various soil test methods for wheat response to applied P, it has been found that, compared with DGT, the Colwell, Olsen and resin methods are more likely to measure more non-labile forms of P in soils (Mason et al. 2013).

Since the development of a binding gel for P analyses (Zhang et al. 1998), it has been applied for soil P assessment in numerous studies, with successful demonstrations of plant yield response to P fertilizers (Tandy et al. 2011; Six et al. 2012) and more specifically for wheat (McBeath et al. 2007; Mason et al. 2010), tomato (Menziez et al. 2005) and maize (Six et al. 2013) responses to P fertilizers. Mason et al. (2008) demonstrated that for analysis of P in soils, the Fe oxide binding gel is highly specific to P and less subjected to anionic interferences compared to the resin method. However, conflicting results for the performance of DGT for the assessment of plant availability for P have been reported for rice (Six et al. 2013) and pasture (Burkitt et al. 2011). However, Burkitt et al. (2011) acknowledged that their conclusion was not firm due to a number of issues identified in their study (e.g. low P fertilizer effectiveness, possible solubilization of less labile P forms by pasture root exudates and potential increase in the amount of P adsorbed during soil storage).

The DGT technology has provided an effective way of assessing P availability for plants; thus, it provides a more precise recommendation for P fertilizer application in nutrient management (Zhang et al. 2013). Further work needs to consider a wide range of plant types, soil types and climate conditions. This would allow the establishment of guideline values for P fertilizer application based on the DGT measurements.

Applications of DGT in sediment P assessment

Eutrophication has been one of the major threats to water quality worldwide, with the basic cause in most cases likely to be excess P inputs (McGarrigle et al. 2010). However, even though external sources are being regulated, eutrophication of rivers and lakes has become a major and persistent problem, e.g. in Ireland (Taylor et al. 2012). With regulated inputs, the contribution of P from the internal sources of sediments becomes important. An accurate assessment of P status in sediments should consider the diffusive process of pore water P coupled with its kinetic resupply from the solid phase. Traditional chemical extraction methods can hardly satisfy such a requirement, as they are developed based on operationally defined response to chemical reagents rather than on a true reflection of P lability (Condrón and Newman 2011). Due to its in situ feature, DGT technology is responding to the kinetic solid-solution

interaction of P rather than a pseudo-equilibrium between extractant and soil/sediment (Davison and Zhang 2012).

The purpose of using DGT in the measurement of sediment P includes three aspects, measurement of dissolved reactive phosphorus (DRP) concentration, observation of heterogeneous distribution of DRP on a small scale and assessment of sediment reactivity in resupplying DRP. Zhang et al. (1998) first applied DGT to measure the concentration of DRP in sediments at a vertical spatial resolution of 1 mm. A steep increase of DRP was observed with increasing sediment depth, reflecting a risk of sediment P to be released to overlying water. The concentration gradients of DRP were frequently obtained with DGT especially in the vicinity of the sediment-water interface (SWI) in later studies (Monbet et al. 2008; Pichette et al. 2009; Ding et al. 2010; Santner et al. 2010; Xu et al. 2012), implying that DGT is potentially a new tool in the assessment of internal P load. The feasibility needs a further study, including establishment of a quantitative relationship between the DGT-measured concentration and the flux of DRP from sediment.

A distinct heterogeneous distribution of DRP in sediments on a small scale was observed from DGT measurement at sub-millimetre and 2D levels (Ding et al. 2011, 2012, 2013). This observation discovered small, discrete spots (termed microniche with diameters of 1 to several mm) with elevated P concentrations in a sediment profile. The formation of the sulphide microniche previously reported has been attributed to strong decomposition of enriched organic matter in localized zones (Stockdale et al. 2009). The bioturbation of the tubificid worm significantly increased the chemical heterogeneity of DRP, whose effects were much more pronounced than observed at a routine spatial resolution (~1 cm) (Ding et al. 2011). Simultaneous measurements of DRP with dissolved sulphide further observed their simultaneous release in a microniche with a diameter of ~3 mm and in locally aggregated zones with a length over 1 cm. This new phenomenon was attributed to the simultaneous reductions of Fe(III) and sulphate and the associated release of Fe-bound P in sediments (Ding et al. 2012). However, the depletion of DRP at the sulphidic microniche was also observed in a sediment profile from a productive lake (Stockdale et al. 2008). The inverse relationships between DRP and dissolved sulphide imply complex biogeochemical processes controlling their cycling.

As DGT is a dynamic technique which can demonstrate the resupply of sediment solids, the measurement of DRP with DGT is especially useful in the characterization of sediment P reactivity. An easy way is to compare the DGT-measured DRP concentration to the real concentration of DRP in pore water. Their ratio (R) close to 1 indicates highly reactive sediment, whereas $R \ll 1$ indicates inert sediments. The R values collected in several studies were mostly less than 0.5 except for a profile from Lake Wellington, which reflected a limited resupply of DRP from solid reactive P pools (Monbet et al. 2008; Ding et al. 2010; Xu et al. 2012). The surface sediments generally had lower values of R compared to deeper sediments, which is further demonstrated by higher sediment response time (T_c) and lower partitioning coefficient (K_d) (Monbet et al. 2008). The R values in the sediment of an algal-dominated region of Lake Taihu were lower than those in a macrophyte-dominated region (Ding et al. 2010). Capping of polluted sediments with Fe- and Al-rich soil also decreased R values because the capped sediments had larger K_d values and greater adsorption capacity (Q_{max}) (Xu et al. 2012).

Recently, a promising technique was developed for the high-resolution imaging of labile phosphorus (P) in sediments in combination with DGT (Ding et al. 2013). This technique was based on the surface coloration of the Zr oxide binding gel using the conventional molybdenum blue method following the DGT uptake of P to this gel. The accumulated mass of the P in the gel was then measured using computer-imaging densitometry (CID). The use of this technique has observed small, discrete spots with elevated P concentrations in a sediment profile. The high-resolution 2D measurement of P in sediment profiles thus provides novel information for the characterization of P in sediments, which deserves more attention.

The current development of DGT for sediment P assessment is still mainly at the laboratory level with limited field applications. Field applications under various benthic environments are needed, and the link between DGT-measured P results and eutrophication mechanisms and trophic levels should be properly established in the future.

Applications of DGT in the bioavailability assessment of metals in soils

It is well known that the total concentrations of metals in soils bear little information of their availability to plants

(Nolan et al. 2003; Degryse et al. 2009). Traditionally, extraction methodologies, including neutral salts (such as CaCl_2 (Novozamsky et al. 1993) and NaNO_3 (Sanka and Dolezal 1992)), diluted or mild acids (0.1 M HCl (Baker and Amacher 1982)) and organic extractants including ethylene-diamine-tetra-acetic acid (EDTA) (Ure et al. 1993) and diethylenetriaminepentaacetic acid (DTPA) (Lindsay and Norvell 1978), are applied to assess the bioavailability of metals in soils, but they are operationally defined and have been found with only limited/varying success (McLaughlin et al. 2000; Menzies et al. 2007).

It has been recognized that, in conventional methods of testing soil solution, metal speciation may change during sampling and extraction and the kinetics of metal resupply from solid phase to solution are not considered (Hooda and Zhang 2008). Meanwhile, the bioavailability of metals in a given soil is dependent on both their concentrations in the soil solution and their rate of transport through the soil (Hooda and Zhang 2008). DGT measures directly the mean flux of labile species in soils to the device during the deployment. Therefore, it provides a novel and promising approach for the measurement of bioavailable metal concentrations in soils (Zhang et al. 2001). The effective concentration, C_E , measured by DGT has been shown to give a better correlation to plant uptake than any other measurement (Zhang et al. 2001; Song et al. 2004; Nolan et al. 2005; Koster et al. 2005). The main reason is that DGT mimics the main mechanism of plant uptake by lowering the concentration locally and inducing diffusive supply and release from the solid phase (Lehto et al. 2006). It automatically accounts for all soil properties, including pH and organic matter content.

DGT assessment of bioavailability is most accurate under conditions where the diffusive transport of an element from soil to the plant roots is rate-limiting for its uptake (Degryse et al. 2009). If plant uptake is slow, competitive cations may affect the plant uptake, while they have no effect on DGT flux, whereas labile complexes do not contribute to the plant uptake but they are measured by DGT. When the plant has little affinity for the element or when the supply is large and the plant uptake is saturated, the DGT flux and plant uptake may not correlate well (Degryse et al. 2009). In mining tailings when the pH is low, competition with other cations that are present at very high concentrations

may hinder the accumulation of metals by the chelating resins (Conesa et al. 2010). Another important factor affecting DGT deployment in soils is moisture. Soil should be moist enough (field capacity or above) in order to maintain contact with the membrane (Hooda et al. 1999). Based on an experiment dealing with Cd, the DGT flux reached the highest level when the soil moisture content was at the maximum water holding capacity (MWHC) level (Hooda et al. 1999). Therefore, when assessing bioavailability using the DGT results, soil moisture should be considered and preferably consistent soil moisture is considered for comparison and assessment purposes.

The applications of DGT in the bioavailability assessment of metals in soils have received wide attention in recent years (Table 3), with both positive (Emily et al. 2012; Liu et al. 2012; Tatiana Garrido and Jorge

Mendoza 2013; Senila et al. 2012) and negative (Muhammad et al. 2012; Williams et al. 2012) results. Researches have also provided cautionary comments or conditions on the use DGT for metal bioavailability assessment; for example, DGT prediction of plant availability was best in anaerobic compared to aerobic soils (Mundus et al. 2012). The DGT technique was fairly predictive of bioavailability in the greenhouse conditions but not in the field (Agbenin and Welp 2012). The DGT approach was demonstrated to be a dynamic in situ measuring technique that can be used as a surrogate of bio-indicators if the cationic correction is taken into account (Ferreira et al. 2013).

Besides plant bioavailability assessment, a recent study assessed metal bioavailability to earthworms using DGT, showing that DGT can be used to mimic the earthworm uptake of heavy metals in contaminated

Table 3 Examples of recent studies on plant availability assessment of soil metals using DGT

Element	Species/soils	Findings	Ref.
Mn	Different Scandinavian agricultural soils	It is unlikely that the DGT techniques can be used to accurately predict the plant availability of Mn in crop production. DGT prediction of plant availability was best in anaerobic compared to aerobic soils. However, DGT can be a valuable tool in understanding factors affecting Mn availability.	Mundus et al. (2012)
Cd, Cu, Pb and Zn	<i>Sorghum bicolor</i> ; <i>Lactuca sativa</i>	The DGT technique was fairly predictive of bioavailability in the greenhouse but not in the field.	Agbenin and Welp (2012)
Pb, Zn, Cu and Cd	<i>Triticum aestivum</i>	DGT and leachate concentrations in combination with bioassays in undisturbed soil cores can be used to account for metal bioavailability in soil.	Emily et al. (2012)
Cd, Zn	<i>Taraxacum officinale</i> and <i>Plantago lanceolata</i>	DGT did not perform better than soil solution, as Cd and Zn uptake in the plant was not limited by diffusion.	Muhammad et al. (2012)
Cd, Zn	Rice	Due to the inherent heterogeneity of the rice rhizosphere soils, deployment of DGT in dried and homogenized soils offers the best possibility of a soil screening tool.	Williams et al. (2012)
Cu, Zn and Cd	Various vegetables	DGT can be used for the general evaluation of the risks associated to soil contamination with Cu, Zn and Cd in field conditions.	Senila et al. (2012)
Hg	Rice (<i>Oryza sativa</i>)	DGT can predict the bioavailability of methyl mercury in rice paddy soil, and the DGT method can provide a quantitative description of the rate of uptake of this bioavailable methyl mercury.	Liu et al. (2012)
Zn and Cd	<i>Salix smithiana</i> clone	Single extraction with the DGT value of Cd was not able to predict shoot Cd removal on the tested soils.	Puschenreiter et al. (2013)
Cu	Soil solution	The DGT allowed a representative estimation of the amount of Cu available in the soil.	Tatiana Garrido and Jorge Mendoza (2013)
Cu	<i>Fontinalis antipyretica</i>	The DGT approach was demonstrated to be a dynamic in situ measuring technique that can be used as a surrogate of bio-indicators if the cationic correction is taken into account.	Ferreira et al. (2013)

soils (Bade et al. 2012). The DGT results were found highly correlated with physiologically based extraction test (PBET) results for As, Pb, Cu and Zn in contaminated sites (Bade et al. 2013). It has also demonstrated that DGT results reasonably predict monomethylmercury uptake by clams from the aqueous phase and provide the basis for application of the DGT device as a surrogate for a sentinel organism for monitoring bioavailable mono-methylmercury (Clarisse et al. 2012).

For the applications in bioavailability assessment of metals in soils, DGT has provided a promising approach. Further work is required to have more plant species tested and to have more metals analysed under various field conditions including soil and climate conditions.

Conclusion

Based on the concepts of diffusion and absorption, DGT is now an established technique quantitatively measuring concentrations of a wide range of analytes in waters, sediments and soils. It provides an effective way for the assessment of P in soils for nutrient management, P in sediment for eutrophication assessment and metals in soils for bioavailability assessment.

DGT works well under the assumption of the establishment of a constant diffusive gradient during deployment, and the results rely on the feature of the binding gels. Further development is needed to have more analytes measured both individually and simultaneously, and easy and cheap methods are needed to have the 2D high-resolution measurements performed. The current applications of DGT are still mainly for research purposes. Standardized methodologies, especially under complicated field conditions, are needed to render the results derived from DGT more meaningful, so that guideline values from DGT for environmental monitoring and management could be established in the future.

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