Biosurfactant-producing strains in enhancing solubilization and biodegradation of petroleum hydrocarbons in groundwater

Hong Liu • Hang Wang • Xuehua Chen • Na Liu • Suriguge Bao

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Abstract Three biosurfactant-producing strains designated as BS-1, BS-3, and BS-4 were screened out from crude oil-contaminated soil using a combination of surface tension measurement and oil spreading method. Thin layer chromatography and infrared analysis indicated that the biosurfactants produced by the three strains were lipopeptide, glycolipid, and phospholipid. The enhancement of solubilization and biodegradation of petroleum hydrocarbons in groundwater employing biosurfactantproducing strains was investigated. The three strain mixtures led to more solubilization of petroleum hydrocarbons in groundwater, and the solubilization rate was 10.5 mg l^{-1} . The combination of biosurfactantproducing strains and petroleum-degrading strains exhibited a higher biodegradation efficiency of 85.4 % than the petroleum-degrading strains (71.2 %). Biodegradation was enhanced the greatest with biosurfactant-producing strains and petroleum-degrading strains in a ratio of 1:1. Fluorescence microscopy images illustrate that the oil dispersed into smaller droplets and emulsified in the presence of biosurfactant-producing strains, which attached to the oil. Thus, the biodegradation of petroleum hydrocarbons in groundwater was enhanced.

H. Liu

School of Resources and Environment Engineering, Jilin Institute of Chemical Technology, Jilin, Jilin 132022, People's Republic of China Keywords Biosurfactant-producing strains · Biodegradation · Petroleum hydrocarbons · Groundwater

Introduction

A great amount of petroleum hydrocarbons inevitably enter the environment as a result of accidents and improper handling in the exploitation, transport, and refinery of crude oil. The petroleum hydrocarbons then migrate into the groundwater system in various ways, such as atmospheric precipitation and soil leaching, thereby leading to groundwater contamination in many countries and regions (Lesage et al. 1997; Vazquez and Mansoori 2000; Wang et al. 2002).

Various existing cleanup technologies for petroleum hydrocarbon treatment can be categorized into three general schemes: chemical, physical, and biological methods (Kermanshahi et al. 2005). Bioremediation has gained much attention and has been applied to groundwater contamination as a potential technology for its numerous advantages, such as high efficiency, relatively low cost, cleanness, and simple application (Holliger 1995; Baveye et al. 1998; Langwaldt and Puhakka 2000). The presence of degrading microorganisms in bioremediation is a key factor for the efficiency of the process (Gallego et al. 2001; Bundy et al. 2002; Kaplan and Kitts 2004). If the degrading microorganism is available, the solubility and bioavailability of petroleum hydrocarbons are other limiting factors (Francy et al. 1991). Existing investigations reported that only 1 to 5 % of contaminants are dissolved in the water

H. Liu · H. Wang · X. Chen · N. Liu (⊠) · S. Bao Key Laboratory of Groundwater Resources and Environment, Ministry of Education, Jilin University, Changchun 130021, People's Republic of China e-mail: liunashr@sina.com

phase, 62 % is present as the free nonaqueous phase [free-product or nonaqueous-phase liquid (NAPL)], and 33 % is adsorbed on the aquifer medium or wrapped in the medium pore because of the capillary force in petroleum-contaminated groundwater (Mackay and Cherry 1989). Surfactants can be used to increase the solubility of NAPL constituents in groundwater (Fountain et al. 1991).

Recent interest in biosurfactants is fueled by their characteristics, such as low toxicity and effectiveness, to reduce interfacial tension between oil and water phases (Zhang and Xiang 2010; Ibrahim et al. 2013). Biosurfactants have been evaluated in the laboratory for their ability to enhance heavy metal ion solubilization, fatsoluble organic contaminant solubilization, and bioavailability by several authors (Mulligan and Eftekhari 2003; Mulligan and Wang 2006; Urum et al. 2003; Banat 1995). Enhanced bioremediation of petroleum hydrocarboncontaminated soil using biosurfactants has been demonstrated at laboratory scale (Lai et al. 2009). However, little data are available on biosurfactants in enhancing the bioremediation of petroleum hydrocarbon-contaminated groundwater, and enhancing petroleum hydrocarbon biodegradation of groundwater in the presence of biosurfactant-producing strains which can metabolize the biosurfactant in situ has not been previously reported.

In this study, biosurfactant-producing strains instead of biosurfactants were added to groundwater. The addition of such strains will ensure the long-term release of biosurfactants and low-cost biosurfactant production because a large amount of organic solvents will be consumed in the extraction of the biosurfactant. Biosurfactant-producing strains were screened and isolated, and the biosurfactants produced by the strains were identified. Petroleum hydrocarbon solubilization and biodegradation in the presence of the biosurfactantproducing strains were also evaluated.

Materials and methods

Strains

Biosurfactant-producing strains and petroleumdegrading strains were screened from crude oilcontaminated soil. The degrading strains were mixed strains acclimatized in low temperature (10 °C) and conserved at 4 °C. These strains were identified as *Acinetobacter* sp. and *Yarrowia* sp.

Culture media

Enrichment medium is composed of the following, in g 1^{-1} distilled water: (NH₄)₂SO₄ 1, KH₂PO₄ 3.4, K₂HPO₄ 4.4, NaCl 1.1, KCl 1.1, MgSO₄·7H₂O 0.2, and yeast extract 0.5. 2 ml of trace element, and 2 ml of 0# diesel oil were supplemented. Fermentation medium composition is composed of the following, in g l^{-1} distilled water: (NH₄)₂SO₄ 1, Na₂HPO₄·12H₂O 1.5, KH₂PO₄ 1.5, NaCl 1, MgSO₄·7H₂O 0.2, EDTA 1, and yeast extract 0.5.2 ml of trace element and 10 ml of 0# diesel oil were supplemented. Mineral salt medium (MSM) is composed of the following, in g 1^{-1} : (NH₄)₂SO₄ 2, NH₄NO₃ 1.2, K₂HPO₄ 1.55, and Na₂HPO₄·12H₂O 0.58. Trace element solution is composed of the following, in g l^{-1} distilled water: ZnSO₄·7H₂O 2.0, CaCl₂·2H₂O 1.0, FeSO₄·7H₂O 5.0, NaMoO₄·2H₂O 0.2, CuSO₄·5H₂O 0.2, $CoCl_2 \cdot 6H_2O$ 0.4, and $MnCl_2 \cdot 2H_2O$ 1.0. The pH values of the media were adjusted to 7, and the media were autoclaved at 121 °C for 20 min.

Isolation and screening of biosurfactant-producing strains

Approximately 5 g of oil-contaminated soil sample was added to the *enrichment medium* and incubated in a rotary shaker at 30 °C and 140 r min⁻¹ for 7 days. Approximately 10 ml of liquid supernatant was transferred to the *enrichment medium* containing 10 ml of 0# diesel oil and incubated under the same condition for another 7 days. After the third subculture, the culture broth was coated on crude oil agar plates. The oil plates were visually inspected for emulsification circles around the colonies, which were indicative of biosurfactant production. The diameter of the emulsification circle depends on the biosurfactant concentration.

The strains were isolated and screened using the modified method of Zhang et al. (2012). In brief, filter papers were uniformly soaked with crude oil and placed on enrichment medium agar plates without carbon source under sterile conditions. Approximately 0.4 ml of the enrichment culture was coated on oil agar plates and cultured for 7 days at 30 °C. The plates were visually inspected for zones of clear oil-dissolved circles, indicating that the biosurfactant was produced. Colonies with large and clear circles were selected, inoculated in *fermentation medium*, and incubated at 30 °C and 180 r min⁻¹ for 4 days. Strains with efficient

Na ⁺	Mg^{2+}	K^+	Ca ²⁺	Cl	NO ₃ ⁻	HPO_4^-	$\mathrm{SO_4}^{2-}$
42,610	8,208	196	38,190	58,215	28,914	6,672	9,637
Fe	Al	Cr	Mn	As	Hg	Cd	Ni
527.4	8.570	48,190	4.88	2.308	Not detected	Not detected	Not detected
Ba	Co	Cu	Zn	Sr	Мо	Se	Pb
123.3	0.029	0.47	0.275	468.5	4.83	0.1089	0.1556

Table 1 Common ions and heavy metals in groundwater ($\mu g l^{-1}$)

biosurfactant-producing capacity were screened out using surface tension measurement and oil spreading method. The surface tension was measured with a JYW-200 surface tension instrument at room temperature (25 °C). The oil-spreading test was done as described by Chandankere et al. (2013). Morphologies of the strains after 24 h of cultivation were observed under a microscope.

Characterization of biosurfactants

To determine the chemical characteristics of the biosurfactants, they were qualitatively analyzed by thin layer chromatography (TLC). Approximately 0.2 ml of the supernatant of the biosurfactant-producing fermentation broth was dissolved in 1 ml of chloroform and subjected to TLC analysis with 65:15:2 (v:v:v) chloroform/methanol/water solution as a developing solvent. The chromogenic reagents were as follows (Zhang et al. 2010): (a) phenol-sulfuric acid reagent, in which 3 g of phenol and 5 ml of sulfuric acid were dissolved in 95 ml of ethanol, and a brown dot blot will appear if glycolipid is present; (b) bromine thyme phenol reagent, in which 40 mg of bromine thyme phenol was dissolved in 100 ml of 0.01 N NaOH, and a blue dot blot will appear if phospholipid is detected; and (c) ninhydrin acetone reagent, in which 0.5 g of ninhydrin was dissolved in 100 ml of acetone, and a red dot blot will appear in the presence of lipopeptide.

Infrared analysis of the purified biosurfactant was conducted followings the methods described by Chandran and Das (2010).

Batch solubility experiments

Experiments were conducted for single-, two-, and three-strain mixtures of BS-1, BS-3, and BS-4. The single-, two-, and three-strain mixtures referred to the single cell suspensions of the three, mixtures of two cell suspensions (1:1, v:v) of the three, and mixtures of the three cell suspensions (1:1:1, v:v:v), respectively. The

optical density (OD) of the aforementioned mixtures was 0.4. Approximately 1 ml of each of the single-, two-, and three-strain mixtures was added to 50 ml of *MSM* supplemented with 10 μ l of 0# diesel. The samples were shaken on a rotary shaker at 10 °C and 120 r min⁻¹ for 7 days. The samples were allowed to stand for 2 h. Then, 10 ml of the lower aqueous phase was drawn using an injector and extracted with 2 ml of hexane. Total petroleum hydrocarbons (TPHs) were analyzed by gas chromatography (GC). The culture solution (0.5 ml) was collected and observed under a microscope before and after 4 days of cultivation.

The *MSM* in this study was prepared with shallow groundwater obtained from Chaoyang district, Changchun. The analytical determination of groundwater was pH 7.3, and electrical conductivity was 320 μ s cm⁻¹. Common ions and heavy metals in groundwater are given in Table 1. Common ions were analyzed using ion chromatograph (Metrohm, Switzerland). Water samples for heavy metals determination were preserved in nitric acid (5 %) within 24 h and then analyzed using ICP-MS (Agilent 7500c, America).

Petroleum hydrocarbons biodegradation

Two sets of biodegradation experiments were performed. In set 1, 100 ml of MSM and 20 μ l of 0# diesel

Table 2 Surface tension and oil spreading diameters

Strains	Surface tension (mN m^{-1})	Oil spreading diameter (cm)
BS-1	27.8	4.0
BS-2	35.6	3.5
BS-3	27.3	4.7
BS-4	26.6	5.2
BS-5	38.4	4.2
BS-6	32.1	4.0

Surface tension of blank fermented liquid was 69.0 ± 0.2 mN m⁻¹ Surface tension of pure water was 71.5 mN m⁻¹



Fig. 1 Morphologies of a Strain BS-1. b Strain BS-3. c Strain BS-4. Bars: (a, b, c) 2 µm

were mixed with 1 ml of cell suspension of degrading strains (OD 0.4). In set 2, 100 ml of *MSM* and 20 μ l of 0# diesel were mixed with 1 ml of cell suspension of degrading and biosurfactant-producing strains (1:1, *v*:*v*; OD 0.4). Both sets were cultivated in a rotary shaker incubator at 10 °C and 120 r min⁻¹ for 7 days. The cultivation solution (10 ml) was collected every day and extracted using 2 ml of hexane. The analysis of TPH was done as Thavasi et al. (2011) and the biodegradation efficiency was calculated as given below.

Biodegradation efficiency (%)= $\left(1-\frac{c_{\rm t}}{c_0}\right) \times 100\%$

where c_t and c_0 are concentrations of TPH after and before cultivation.

Biodegradation experiments with degrading and biosurfactant-producing strains at various ratios were carried out by adding 1 ml of cell suspensions (OD 0.4) of biosurfactant-producing and degrading strains (2:1, 1:1, 2:3, and 1:2, v:v) to 100 ml of MSM supplemented with 20 µl of 0# diesel. These samples were cultivated at 10 °C and 120 r min⁻¹ for 7 days. The cultivation solution (10 ml) was collected after 7 days and extracted using 2 ml of hexane. TPHs were analyzed by GC, and the biodegradation rate was calculated.

Biosurfactant-producing strains in enhancing biodegradation

To examine biosurfactant-producing strains in enhancing biodegradation, 900 μ l of the upper layer of the fermentation broth containing petroleum was collected after 24 h of cultivation in fermentation medium, blended with 90 μ l of phosphate buffer (pH 8.0) and 10 μ l of 50 mM EDTA, and added with CFDA dyestuff to obtain a final concentration of 10 μ M. The dyed samples were cultivated in the dark at 35 °C for 30 min to eliminate unnecessary natural background. Approximately 0.3 ml of the dyed samples was observed under fluorescence microscopy.

Results and discussion

Isolation and screening of biosurfactant-producing strains

As a microorganism can be considered a promising biosurfactant producer when it is able to reduce the



Fig. 2 TLC analysis of purified biosurfactant isolates from a Strain BS-1. b Strain BS-3. c Strain BS-4





surface tension below 40 mN m^{-1} (Olivera et al. 2009), only those isolates lowering culture medium surface tension below this limit were preliminarily screened out (BS-1, BS-2, BS-3, BS-4, BS-5, and BS-6). Surface tension and oil-spreading diameter of fermentation broth produced by the six strains were determined. All the surface tensions were lowered from 69.0 ± 0.2 mN m⁻¹ to less than 40 mN m⁻¹ and the oil expelling circles (ranging from 3.5 to 5.2 cm) were notable and stable (Table 2). Among them, three strains designated as BS-1, BS-3, and BS-4 which were identified as Bacillus sp., Pseudomonas sp. and Micrococcus sp. exhibited significant effect on surface tension reduced to 27.8, 27.3, and 26.6 mN m^{-1} , respectively. Based on this analysis, the three strains were selected for subsequent study. Morphologies of BS-1, BS-3, and BS-4 are shown in Fig. 1.

Chromatography

Thin layer chorography (TLC)

The chemical characteristics of the biosurfactants were evaluated using TLC. The results of TLC analysis (Fig. 2) show that red, brown, and blue dot blots for strains BS-1, BS-3, and BS-4, respectively, were observed after being sprayed with different reagents. According to literature (Bao et al. 2003), this

suggests that biosurfactants produced by the three strains were types of lipopeptide, glycolipid, and phospholipid respectively.

Infrared analysis

As shown in Fig. 3, the absorbance values for each specific band included hydroxyl (OH-) stretching bands at 2,929 cm⁻¹. Stretching bands at 2,928, 2,856, and 1,451 cm⁻¹ confirmed the presence of methylene (-CH2-). The adsorption peaks at 1,658 and 1.657 cm^{-1} were attributed to unsaturated double bonds (C=C). The stretching bands at 1,870 to 1,600 cm^{-1} were attributed to carboxylic acids (C=O), whereas those at 1,340 and 1,361 cm⁻¹ were attributed to cyclic lactone and glycosidic bonds (C-O-C). The adsorption peaks at 1,645 and 1,630 to 1,605 cm^{-1} were carbonyl. These results suggest that all the three biosurfactants were unsaturated fats and aromatic compounds. The stretching bands of BS-1 at 650 to 900 cm⁻¹ were attributed to the presence of ammonium (N-H), which indicates that the biosurfactant was a type of lipopeptide. The ketone and carbonyl stretching bands of BS-3 at 1,735 cm⁻¹ suggest that the biosurfactant was a type of glycolipid. The stretching bands of BS-4 at 1,260 to 1,285 cm^{-1} reveal characteristics of phosphorus, which confirms that the biosurfactant was a type of phospholipid.





Fig. 5 Oil droplets **a** before and **b** after cultivation. *Bars*: (a, b) 50 μm



Solubilization of petroleum hydrocarbons in groundwater

The solubilization of petroleum hydrocarbons in the presence of biosurfactant-producing strains was determined. The results in Fig. 4 suggest that the two-strain mixtures of BS-1 and BS-3 and BS-3 and BS-4, as well as the threestrain mixtures of BS-1, BS-3, and BS-4, exhibited more solubility than the single-strain mixtures. Among them, the three-strain mixture was the best enhancer in terms of TPH solubilization (10.5 mg Γ^{-1}). This result might be due to biosurfactant production by these strain mixtures.

The figures (Fig. 5) before and after 4 days of cultivation illustrate that oil dispersed into smaller droplets, which were more accessible and available for microbes. Thus, the biodegradation of petroleum hydrocarbons in groundwater was enhanced by adding biosurfactantproducing strains. This result was investigated in the following section. Biodegradation of petroleum hydrocarbons in groundwater

The biodegradation efficiency of petroleum hydrocarbons in groundwater with and without the presence of biosurfactant-producing strains are shown in Fig. 6. The data indicate that biosurfactant-producing strains had a significant influence on TPH biodegradation. The TPH biodegradation efficiency was relatively low in preliminary cultivation, but increased after 4 to 5 days of cultivation, followed by a slow increase. Biosurfactant-producing strains combined with petroleum-degrading strains exhibited a higher biodegradation efficiency of 85.4 % after 7 days of cultivation at 10 °C than the petroleumdegrading strains (71.2 %) because of biosurfactant production. This result confirms that the biosurfactantproducing strains could enhance TPH biodegradation in groundwater.





Fig. 7 Biodegradation of petroleum hydrocarbons by strains in different ratios



Figure 7 shows that the biodegradation efficiency of TPH was enhanced the greatest with biosurfactantproducing strains and petroleum-degrading strains in the ratio of 1:1.

Biosurfactant-producing strains in enhancing biodegradation

Biosurfactant-producing strains have a great function in enhancing the biodegradation of petroleum hydrocarbons. As shown in Fig. 8, the oil dispersed into smaller droplets and was emulsified in the presence of the biosurfactant-producing strains, which was advantageous for degrading strains to access and metabolize hydrocarbons in the water phase. On one hand, the droplets greatly increased the opportunity for degrading bacterial cells to come into contact with hydrocarbons. The uptake mechanism of hydrophobic substrate occurs by the direct contact between the hydrocarbon and cell surface (Kumari et al. 2012). On the other hand, cell surface hydrophobicity and affinity between cells and hydrocarbons significantly improved. Thus, TPH biodegradation in groundwater was enhanced in the presence of biosurfactant-producing strains.

Conclusion

This study was performed to investigate biosurfactantproducing strains employed in enhancing the biodegradation of petroleum hydrocarbons in groundwater. Three strains, namely, *Bacillus* sp., *Pseudomonas* sp., and *Micrococcus* sp., which produced lipopeptide, glycolipid, and phospholipid biosurfactants, respectively, were screened from crude oil-contaminated soil. In the presence of the three biosurfactant-producing strains, the oil dispersed into small droplets and emulsified. Petroleum solubilization and biodegradation in groundwater were enhanced. Three-strain mixtures exhibited better effects on petroleum hydrocarbon solubilization in groundwater, and



Fig. 8 Emulsifications of oil at the presence of a Strain BS-1. b Strain BS-3. c Strain BS-4. Bars: (a, b, c) 20 µm

the solubilization rate was 10.5 mg l⁻¹. Biosurfactantproducing strains combined with petroleum-degrading strains degraded 85.4 % of TPH in groundwater, whereas the degrading strains metabolized only 71.2 % after 7 days of cultivation at 10 °C. This result confirms that the biosurfactant-producing strains could enhance TPH biodegradation in groundwater.

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References

- Banat, I. M. (1995). Bioproduction and possible uses in microbial enhanced oil recovery and oil pollution remediation: a review. *Bioresource Technology*, 51, 1–12.
- Bao, M., Mu, B., & Wang, X. (2003). Metabolic process of organisms used in oil recovery. *Chem. Res. Appl.*, 15, 555– 557.
- Baveye, P., Vandevivere, P., Hoyle, B. L., DeLeo, P. C., Sanchez, D., & Lozada, D. (1998). Environmental impact and mechanisms of the biological clogging of saturated soils and aquifer materials. *Crit Rev Environ Sci Tech*, 28, 123–191.
- Bundy, J. G., Paton, G. I., & Campbell, C. D. (2002). Microbial communities in different soil types do not converge after diesel contamination. *Applied Microbiology*, 92, 276–288.
- Chandankere, R., Yao, J., Choi, M. M. F., Masakorala, K., & Chan, Y. (2013). An efficient biosurfactant-producing and crude-oil emulsifying bacterium Bacillus methylotrophicus USTBa isolated from petroleum reservoir. *Biochemical Engineering Journal*, 74, 46–53.
- Chandran, P., & Das, N. (2010). Biosurfactant production and diesel oil degradation by yeast species Trichosporon asahii isolated from petroleum hydrocarbon contaminated soil. *International Journal of Engineering, Science and Technology, 2*, 6942–6953.
- Fountain, J. C., Klimek, A., Beikirch, M. G., & Middleton, T. M. (1991). The use of surfactants for in situ extraction of organic pollutants from a contaminated aquifer. *Haz. Mater.*, 28(3), 295.
- Francy, D. S., Thomas, J. M., Raymond, R. L., & Ward, C. H. (1991). Emulsification of hydrocarbons by subsurface bacteria. *Journal of Industrial Microbiology*, 8, 237–246.
- Gallego, J. R., Loredo, J., Llamas, J. F., Va'zquez, F., & Sa'nchez, J. (2001). Bioremediation of diesel-contaminated soils: evaluation of potential in situ techniques by study of bacterial degradation. *Biodegradation*, 12, 325–335.
- Holliger, C. (1995). The anaerobic microbiology and biotreatment of chlorinated ethenes. *Current Opinion in Biotechnology*, 6, 347–351.
- Ibrahim, M. L., Ijah, U. J. J., Manga, S. B., Bilbis, L. S., & Umar, S. (2013). Production and partial characterization of biosurfactant produced by crude oil degrading bacteria. *International Biodeterioration & Biodegradation*, 81, 28–34.

- Kaplan, C. W., & Kitts, C. L. (2004). Bacterial succession in a petroleum land treatment unit. *Applied and Environmental Microbiology*, 70, 1777–1786.
- Kermanshahi, P. A., Karamanev, D., & Margaritis, A. (2005). Biodegradation of petroleum hydrocarbons in an immobilized cell airlift bioreactor. *Water Research*, 39, 3704–3714.
- Kumari, B., Singh, S. N., & Singh, D. P. (2012). Characterization of two biosurfactant producing strains in crude oil degradation. *Process Biochemistry*, 47, 2463–2471.
- Lai, C. C., Huang, Y. C., Wei, Y. H., & Chang, J. S. (2009). Biosurfactant-enhanced removal of total petroleum hydrocarbons from contaminated soil. *Journal of Hazardous Materials*, 167, 609–614.
- Langwaldt, J. H., & Puhakka, J. A. (2000). On-site biological remediation of contaminated groundwater: a review. *Environmental Pollution*, 107, 187–197.
- Lesage, S., Hao, X., & Kent, S. N. (1997). Distinguishing natural hydrocarbons from anthropogenic contamination in groundwater. *Groundwater*, 35(1), 149–160.
- Mackay, D. M., & Cherry, J. A. (1989). Ground water contamination: pump-and-treat remediation. *Environmental Science* and Technology, 23(6), 630.
- Mulligan, C. N., & Eftekhari, F. (2003). Remediation with surfactant foam of PCP contaminated soil. *Engineering Geology*, 70, 269–279.
- Mulligan, C. N., & Wang, S. (2006). Remediation of a heavy metal contaminated soil by a rhamnolipid foam. *Engineering Geology*, 85, 75–81.
- Olivera, N. L., Nievas, M. L., Lozada, M., Prado, G. D., Dionisi, H. M., & Sineriz, F. (2009). Isolation and characterization of biosurfactant -producing Alcanivorax strains: hydrocarbon accession strategies and alkane hydroxylase gene analysis. *Research in Microbiology*, 160, 19–26.
- Thavasi, R., Jayalakshmi, S., & Banat, I. M. (2011). Application of biosurfactant produced from peanut oil cake by Lactobacillus delbrueckii in biodegradation of crude oil. *Bioresource Technology*, 102, 3366–3372.
- Urum, K., Pekdemir, T., & Gopur, M. (2003). Optimum conditions for washing of crude oil contaminated soil with biosurfactant solutions. *Transactions. Institute of Chemical Engineers*, *81B*, 203–209.
- Vazquez, D., & Mansoori, G. A. (2000). Identification and measurement of petroleum precipitates. *Petro Sci Engineer*, 26, 49–55.
- Wang, Z. D., Li, K., & Fingas, M. (2002). Characterization and source identification of hydrocarbons in water samples using multiple analytical techniques. *Journal of Chromatography A*, 971(1–2), 173–184.
- Zhang, X., & Xiang, T. (2010). Review on microbial enhanced oil recovery technology and development in China. Int. J. Pet. Sci. Technol, 4, 61–80.
- Zhang, X., Li, M., & Xiang, T. (2010). Genetic modification of MEOR bacterium Bacillus licheniformis H strain by low energy ion beam irradiation. *Open Biotechnol*, 4, 14–17.
- Zhang, X. S., Xu, D. J., Zhu, C. Y., Lundaa, T., & Scherr, K. E. (2012). Isolation and identification of biosurfactant producing and crude oil degrading Pseudomonas aeruginosa strains. *Chemical Engineering Journal*, 209, 138–146.