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Fine‑scale determinants of vertebrate roadkills across a biodiversity hotspot in Southern Spain

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Abstract

An increasing number of wildlife-vehicle collisions occur each year worldwide, which involves extensive economic costs and constitutes one of the main anthropogenic causes of animal mortality. Because of this, there is an urgent need to identify the factors leading to collision hotspots and thus implementing efective mitigation measures. By using a stratifed random sampling survey, we investigated the fne-scale determinants of roadkill probability in small and medium-sized birds and mammals across a country-size region of Southern Spain, Andalusia $(87,000 \text{ km}^2)$, located within a global biodiversity hotspot. During two consecutive seasons, we regularly surveyed 45 road transects of 10 km each and characterized the site-specifc attributes of both roadkill and random points, including traffic density, road design (embankments, medians, fences, roadside vegetation and distance to curves), and adjacent landscape matrix. Based on this information, we investigated variation in collision risk according to landscape and road features, and the life history of the affected taxa. Mortality rates of mammals and birds increased with traffic density, and were also signifcantly afected by the distance to the nearest curve, slope of embankments, height of roadside vegetation, and land use adjacent to roads. Road mortality of both birds and mammals was related to the presence and typology of fences and center medians, so more densely vegetated medians and smaller mesh sizes reduced roadkill probability. Overall, our results indicate that roadkill risk may vary at exceedingly small spatial scales. The information provided by this extensive survey may be used to identify taxa-specifc factors associated to roadkill risk and priority points for action. Our fndings will therefore be relevant for the design of safer roads for both drivers and wildlife through the application of efective mitigation measures.

Keywords Wildlife vehicle collisions · Road mortality · Collision hotspots · Roadkill predictors · Predictive models · Mitigation measures

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Introduction

Roads have multiple ecological impacts, as they can act as barriers by limiting connectivity among populations, contaminate adjacent ecosystems, alter animal behavior, and facilitate dispersal of exotic species, among others (reviewed in Sandberg et al. [1998](#page-16-0); Trombulak and Frissell [2000](#page-16-1); Forman et al. [2003](#page-15-0); Coffin [2007;](#page-15-1) van der Ree et al. [2015b\)](#page-17-0). However, traffic-related mortality due to wildlife-vehicles collisions (WVCs, hereafter) appears to be the most important ecological impact of roads (Trombulak and Frissell [2000;](#page-16-1) Coffin [2007](#page-15-1)).

WVCs are an important traffic safety issue that involves significant monetary costs, primarily due to human injury and material damage (Bissonette et al. [2008](#page-15-2); Huijser et al. [2009\)](#page-16-2), but also high environmental costs (Forman et al. [2003;](#page-15-0) Erritzøe et al. [2003](#page-15-3); Bisson-ette et al. [2008](#page-15-2)). Traffic related mortality is considered one of the most important sources of non-natural mortality in wildlife populations (Forman et al. [2003](#page-15-0); Erritzøe et al. [2003;](#page-15-3) Colino-Rabanal and Lizana [2012\)](#page-15-4). For instance, considering only birds, 27 and 80 millions of fatalities are estimated to occur annually in Europe (Erritzøe et al. [2003\)](#page-15-3) and the United States (Erickson et al. [2005](#page-15-5)), respectively, although actual numbers may be higher. Traffc related mortality may dramatically afect population dynamics (e.g. through diferential incidence into a gender or age class; Madsen et al. [2002;](#page-16-3) Colino-Rabanal and Lizana [2012](#page-15-4)) and constitutes a major threat for endangered species (Mumme et al. [2000](#page-16-4); Gibbs and Shriver [2002](#page-15-6)). Importantly, because of the expansion of the road network and the increase in traffic volume, the ecological impact of roads on wildlife is expected to increase over the next decades in both developed and developing countries (Fulton and Eads [2004](#page-15-7); Meyer et al. [2012](#page-16-5); van der Ree et al. [2015c\)](#page-17-1). Thus, quantifying the impact of roads on wildlife and developing efective mitigation measurements is urgently needed to balance future development requirements and biodiversity conservation (van der Ree et al. [2015c\)](#page-17-1).

Road characteristics have long been recognized as a crucial determinant of roadkills. Factors such as traffic density and velocity, road sinuosity, and the presence of road crosses and elevation changes are frequently associated with collision risk (e.g. Trombulak and Frissell [2000](#page-16-1); Forman et al. [2003](#page-15-0); Clevenger et al. [2003](#page-15-8); Malo et al. [2004](#page-16-6); Seiler [2005;](#page-16-7) Gomes et al. [2009](#page-16-8); Langen et al. [2012;](#page-16-9) Zuberogoitia et al. [2014;](#page-17-2) D'Amico et al. [2015](#page-15-9)). Further, roadside strips of vegetation and land-use adjacent to roads may infuence roadkill risk by determining the presence and movements of animals (e.g. Forman et al. [2003](#page-15-0); Clevenger et al. [2003;](#page-15-8) Malo et al. [2004](#page-16-6); Grilo et al. [2009](#page-16-10); Gunson et al. [2011](#page-16-11)).

Mitigation measures commonly used to reduce animal roadkill include, among others, wildlife crossing structures (e.g. underpasses and overpasses), warning signs, animal detection systems and a variety of fences (reviewed in Glista et al. [2009](#page-16-12); van der Grift et al. [2013\)](#page-17-3). However, none of these measures has been fully efective in preventing WVCs, since their efectiveness strongly depends on the interplay between particular life history traits (e.g. foraging strategy, dispersal or migratory movements) of the species afected by roadkills and environmental factors infuencing collision risk. Given that the implantation of mitigation measures along the entire road network is economically and logistically unfeasible, and that WVCs are typically clustered (Gunson et al. 2011), identifying the factors that increase the risk of collision is essential to implement efective mitigation measures (Gunson et al. [2011](#page-16-11)). Furthermore, recent calls have highlighted the need to conduct additional research that broadens the taxonomic, spatial, and temporal scale of roadkill data sets to optimize the implementation of the mitigation measures (van der Ree et al. [2015c\)](#page-17-1).

In this study, we investigated the fne-scale determinants of roadkill probability in small and medium-sized birds and mammals at a large spatio-temporal scale. Over 22 months, we regularly surveyed 45 road sections of 10 km each distributed across Andalusia (South Spain), an extensive Mediterranean region (87,268 km²) located in a biodiversity hotspot (Myers et al. [2000](#page-16-13)). During the surveys, we characterized site-specifc attributes at both WVC and control points (i.e. randomly-generated points without casualties recorded), using a landscape-level and road level approach (Table [1](#page-3-0)). This information was used to investigate the WVC risk in relation to the species' biology, adjacent land use, and road design based on the predictions presented in Table [1](#page-3-0).

Material and methods

The study was conducted during two consecutive periods (July 2009–June 2010—except September and October—and April 2011–March 2012) in the Autonomous Community of Andalusia, Spain, a region that stretches from the southeast to the southwest of the Iberian Peninsula (Fig. [1](#page-4-0)). The ecosystems in Andalusia, characterized by an extraordinary diversity of species and landscapes, are considered to be highly sensitive to global-change drivers and are thus predicted to experience dramatic biodiversity changes in the next decades (Myers et al. [2000](#page-16-13)).

Briefy, the climate in Andalusia is Mediterranean, but due to a marked interannual variation in rainfall (it may varies from 170 mm/year to more than 1800 mm/year) and a wide elevation range (from sea level to approximately 3500 m.a.s.l.), there is a high diversity in vegetation and landscape conditions (including semiarid zones, forest, mountains or marshland). To capture such environmental diversity, the study region was divided into fve ecoregions, defned as areas characterized by similar landscape characteristics and environmental conditions (GIASA et al. [2006](#page-15-10)). For a detailed description of characteristics and environmental conditions on these ecoregions see Canal et al. [\(2018](#page-15-11)) and GIASA et al. ([2006\)](#page-15-10).

Besides environmental conditions, the selection of the sampling roads aimed at representing the road network in Andalusia. Thus, according to the physic characteristics of the roads (number of lines, speed limit or traffic density), we grouped the surveyed roads into three categories: (1) Highways characterized by a dual carriageway and 120 km/h speed limit; (2) National roads, including all roads belonging to the State Network, Regional and Interregional network except highways and (3) Local roads, all roads belonging to Complementary Regional Network and the Provincial Councils. Both national and local roads are characterized by a single carriageway and a 90 km/h speed limit.

Nine road sections (three replicates per road category) were randomly selected from the road network crossing each of the five ecoregions (3 replicates \times 3 road categories \times 5 ecoregions=45 road sections). For each road, a random number was generated to set the starting point (kilometric point) of the sampling stretch. Overall, we monthly surveyed 450 km along 45 road segments crossing all ecoregions included in Andalusia (Fig. [1](#page-4-0) and Table 1 in Supplementary material).

Monthly surveys were carried out by two experienced observers by driving a vehicle at low speed $(-25-30 \text{ km/h})$ along the shoulder of the road with the emergency lights fashing. The sampling order of the surveyed sections was set at random from month to month and survey session. Roadkilled animals encountered on the paved road or the road verge were identifed at the species level (whenever possible) and its location was recorded using a GPS. All carcasses were removed from roads to avoid duplicating records during posterior surveys. At each point, we recorded site-specifc attributes of roads and their

Zuberogoitia et al. [\(2006](#page-17-4)), [8] Seiler ([2005\)](#page-16-7), [9] Bellis and Graves ([1971\)](#page-15-14), [10] Colino-Rabanal et al. ([2011](#page-15-15)), [11] Van der Ree et al. ([2015a,](#page-17-5) [b](#page-17-0), [c](#page-17-1)), [12] Clevenger and Koci-

olek [\(2013](#page-15-16))

Fig. 1 Situation of the road sections and main ecological units (ecoregions) surveyed during the study period. Surveyed roads (roadkill and control points) are highlighted in black

immediate vicinity, including structures potentially infuencing animal accessibility to roads (see Table [1](#page-3-0) for a description of the measured variables and its predicted efect on WVC). Traffic density of the surveyed roads, defined as the average number of vehicles per day, was obtained from official data at [http://www.fomento.es/MFOM/LANG_CASTE](http://www.fomento.es/MFOM/LANG_CASTELLANO/DIRECCIONES_GENERALES/CARRETERAS/TRAFICO_VELOCIDADES/MAPAS/) [LLANO/DIRECCIONES_GENERALES/CARRETERAS/TRAFICO_VELOCIDADES/](http://www.fomento.es/MFOM/LANG_CASTELLANO/DIRECCIONES_GENERALES/CARRETERAS/TRAFICO_VELOCIDADES/MAPAS/) [MAPAS/](http://www.fomento.es/MFOM/LANG_CASTELLANO/DIRECCIONES_GENERALES/CARRETERAS/TRAFICO_VELOCIDADES/MAPAS/) and [http://www.juntadeandalucia.es/fomentoyvivienda/portal-web/web/areas/](http://www.juntadeandalucia.es/fomentoyvivienda/portal-web/web/areas/carreteras/aforos) [carreteras/aforos.](http://www.juntadeandalucia.es/fomentoyvivienda/portal-web/web/areas/carreteras/aforos) Twenty control points were randomly selected as outlined above (without previous knowledge of roadkill points) within each of the sampling sections $(20 \times 45 \text{ road})$ sections=900 control points) and characterized following the same procedure as for the collision points.

Statistical analyses

We used Generalized Linear Mixed Models (GLMM) to model the probability of WVC in relation to landscape and road features. Separate GLMMs were ftted for birds and mammals (see below) according to the noticeable diferences between their life-history strategies (e.g. spatial ecology or locomotor capacity). Even though the classifcation of roadkill species at the class level might appear simplistic due to major species-specifc diferences in life history traits, such a classifcation may broaden the applicability of the mitigation measures derived from our survey. Further support for the use of a coarse-grained approach comes from the common implementation of similar mitigation measurements for diferent animal groups. For instance, similar measures are applied for ground birds and large terrestrial mammals, whereas the same applies for bats and fying birds (Abbott et al. [2015;](#page-14-0) Kociolek et al. [2015](#page-16-16)). Finally, an analysis of the probability of WVC at a lower taxonomic level (higher functional similarity) including passerines, carnivores and lagomorphs, the groups most afected by roadkills, showed similar results to those found at the class level (see Table 3 in Supplementary material for details).

GLMMs analyzing the probability of WVC separately in birds and mammals included the presence (1) or absence (0, control points) of collision as the dependent variable (binomial distribution and logit link function) and seven explanatory variables as descriptors of each point: road type, distance to the nearest curve and its quadratic term, maximum height of roadside vegetation and its quadratic term, adjacent land use, and type of embankment. Road ID, nested within the ecoregion, was fit as a random factor.

Based on exploratory analyses, the multiple levels of road embankments was reduced to two classes: roads with embankments in any of the road sides instigating birds to fy high above the road (e.g., steep, buried sections) and those allowing animals to fy close to the road surface (e.g. roads sites at ground level). Similarly, for mammals, we reduced the type of road embankments to four classes: roads at ground level, raised, buried and roads with opposing types of embankment at each side (e.g. buried on the right side and raised on the left side).

Note that traffic density was not included in the two models above as the values for roadkill and control points within a given road section would be the same. Thus, to test if the accumulated number of roadkills in a road section was related to traffic density, we used a Pearson correlation.

Special considerations for fences and center medians

Because all highways in Spain are fenced, the efect of fences on vertebrate roadkills could only be investigated using data from secondary and local roads. The latter roads, however, lack median centers and thus, the infuence of this structure on WVC was explored using exclusively data from highways.

Road points were categorized into four classes according to the presence and type of structures preventing the access of wildlife to roads (unfenced points, presence of barbed fences, wire mesh fences and walls; Table [1\)](#page-3-0). The presence and type of barrier (fence/wall) was recorded at both roadsides and the difculty of accessing roads was then determined according to the roadside having the less restrictive type of barrier. Exploratory analyses revealed no diferences between mesh fences and walls in roadkill likelihood, and thus the difficulty of wildlife to access roads was assessed using a threelevel variable: easy (unfenced point), medium (barber fenced) and high (mesh fences and walls) difficulty of access.

Median strips were initially categorized as Jersey barriers and structures with absent, medium or much vegetation (Table [1](#page-3-0)). However, frequency diagrams of the types of median strips revealed that the number of points with Jersey barriers as medians was very small. These points were therefore excluded from the models to avoid a disproportionate infuence of rare categories on model outputs and, consequently, only medians varying in the amount of vegetation cover were analyzed.

The infuence of barriers (fence/walls; except for highways) and medians (only in highways) on roadkill likelihood was analyzed separately for mammals and birds using GLMMs with the same structure as described above.

Model selection

In total, we ftted three GLMMs per vertebrate class analyzing the probability of roadkill in relation to (i) the landscape and road attributes, (ii) the type of barriers and (iii) the amount of vegetation cover in the median strips.

Selection of the fnal models-i.e. containing only statistically signifcant terms- was carried out by sequentially dropping non-signifcant terms from fully saturated models (containing all main efects and interactions) in a hierarchical way, starting with the least signifcant order terms. To confrm whether the inclusion of a predictor was signifcantly informative, we compared the models including and excluding the focal term using Chi square likelihood ratio tests (through maximum likelihood estimations).

We systematically performed model diagnostics statistics to avoid misleading conclusions based on statistical artifacts. Accordingly, we visually checked assumptions about the distribution of residuals through diagnostics plots, and examined collinearity and the existence of infuential data points. To meet statistical assumptions, the distances to the nearest curve and traffic intensity were $log 10$ -transformed. After these transformations, diagnostics analyses did not show obvious deviation from GLMM assumptions.

Our dataset was unbalanced since twenty control points were systematically recorded per road section, whereas the number of recorded roadkill varied among roads (Fig. [2](#page-6-0)). Even though the accuracy of binomial models is robust to unbalanced sampling (Crone and Finlay [2012](#page-15-17)), we repeated the analyses above after creating a balanced dataset (roadkill and control points were randomly selected) to check for consistency between the results based on balanced and unbalanced samples. Because the results obtained using the balanced and raw datasets were similar, we present along the paper the models using the whole dataset to make full use of the available data as suggested by Crone and Finlay [\(2012](#page-15-17)). During our surveys, we found a small fraction of domestic mammals (mostly dogs; see Canal et al. [2018\)](#page-15-11) killed by vehicles. Results from the analyses excluding and including domestic animals were qualitatively similar. For this reason, the results of the analyses including the whole dataset are presented here.

Statistical analyses were carried out in R 3.5.1 (R Core Team [2018](#page-16-17)). For running the GLMMs, we used the packages lme4 (Bates et al. [2014](#page-15-18)), lmerTest (Kuznetsova et al. [2017](#page-16-18)) and Rcpp (Eddelbuettel and François [2011\)](#page-15-19). For a part of the model diagnostics, we used

Fig. 2 Number of recorded bird (left) and mammal (rigth) roadkills in relation to traffic density (average number of vehicles per day) of the roads

the package DHARMa (Hartig [2018](#page-16-19)) and the VIF function from the package car (Fox and Weisberg [2011\)](#page-15-20).

Results

A total of 835 mammals and 555 birds belonging to 19 and 70 species, respectively, were recorded as killed by vehicles during the two study seasons (Tables 1 and 2 in Supplementary material; Canal et al. [2018\)](#page-15-11). 2.8% of roadkills could not be identifed at the species level due to severe damage and/or poor conservation status.

Road mortality analyses revealed common factors associated with the occurrence of roadkills in birds and mammals (Tables [2](#page-7-0) and [3\)](#page-8-0). In both groups, roadkills were related to traffic density (mammals: $R = 0.57$, $P < 0.001$; birds: $R = 0.39$, $P = 0.01$; Fig. [2\)](#page-6-0), the distance to the nearest curve, and the height of the roadside vegetation. For the latter two factors, roadkill probability showed an inverted-U shape, increasing until a maximum distance and height, and decreasing afterwards (Tables [2](#page-7-0), [3](#page-8-0) and Fig. [3](#page-9-0)). Also, in both groups, roadkill risk was afected by the adjacent land use type and the slope of road embankments. Roads crossing forests showed the highest probability of roadkill in birds (Table [2](#page-7-0) and Fig. [4](#page-10-0)), whereas, in mammals, forests and farmlands were the habitats with highest mortality rates (Table 3 and Fig. [4\)](#page-10-0). The presence and type of road embankments also afected roadkill risk. In mammals, the probability of roadkill was lowest in elevated road sites, whereas roads with embankments boosting "high-altitude" fights (e.g. buried roads sections) reduced roadkill probability in birds (Tables [2](#page-7-0), [3](#page-8-0)).

Fixed effects	Estimate	SE	Z	P
Intercept	-3.854	1.014	-3.800	< 0.001
Road category				
National	-0.381	0.204	-1.869	0.062
Local	-0.957	0.229	-4.172	< 0.001
Dist_Curve	1.309	0.388	3.379	0.001
Dist Curve ²	-0.121	0.038	-3.226	0.001
Vegetation	0.503	0.090	5.602	< 0.001
Vegetation ²	-0.056	0.012	-4.560	< 0.001
Land use:				
Farmland	-0.552	0.167	-3.312	< 0.001
Scrubland	-0.873	0.215	-4.069	< 0.001
Urban	-0.556	0.264	-2.107	0.035
Embankments ^a	-0.472	0.189	-2.490	0.013
Random effects	Variance	SD.		
RoadID: Region	0.150	0.387		
Region	0.090	0.300		

Table 2 Relationship between the occurrence of roadkills in birds and the characteristics of the road and adjacent land use

a Simplifed to two levels according road topography while ftting the model: roads hampering birds to fy at low altitude and roads facilitating birds to fly close to the road surface

Fixed effects	Estimate	SE	Z	\boldsymbol{P}
Intercept	-4.539	1.105	-4.107	< 0.001
Road category				
National	-0.669	0.253	-2.643	0.008
Local	-1.499	0.288	-5.205	< 0.001
Dist_Curve	1.062	0.389	2.728	0.006
Dist_Curve ²	-0.097	0.038	-2.572	0.010
Vegetation	0.604	0.091	6.616	< 0.001
Vegetation ²	-0.062	0.012	-5.119	< 0.001
Land use:				
Farmland	0.079	0.167	0.475	0.635
Scrubland	-1.244	0.259	-4.806	< 0.001
Urban	-0.797	0.277	-2.880	0.004
Embankments ^a				
Buried roads	1.342	0.227	5.908	< 0.001
Roads at level	2.402	0.217	11.090	< 0.001
Roads with mixed embank- ments	1.719	0.164	10.468	< 0.001
Random effect	Variance	SD		
RoadID: Region	0.285	0.534		
Region	0.745	0.863		

Table 3 Relationship between the occurrence of roadkills in mammals and the characteristics of the road and adjacent land use

^aSimplified to four levels according road topography while fitting the model: road at ground level, buried, raised and road with mixed embankments (involving roads part buried, part raised and buried-raise

The presence and type of physical structures (fences and walls) preventing access of wildlife to roads reduced the roadkill likelihood in mammals and birds. In both taxa, the number of casualties decreased as the difficulty of accessing roads increased from unfenced points, followed by barber fences and points having mesh wire fences or walls (mammals: estimate (SE) −0.57 (0.225), Z=−2.53, P=0.011; birds: estimate (SE) −0.587 (0.185), $Z=-3.16$, P=0.002; Fig. [5\)](#page-11-0). Roadkill risk also decreased as vegetation cover in the median strips increased (mammals: estimate (SE) -0.330 (0.184), Z= -1.79 , P=0.07; birds: estimate (SE) −0.57 (0.17), Z = −3.345, P < 0.001; Fig. [5](#page-11-0)).

Discussion

Based on a large-scale survey and accurate description of the sampling sites, we have shown that roadkill risk in small and medium-sized birds and mammals may vary at exceedingly small spatial scales and that collision risk is group-specifc. A fne-scale description of the road attributes at both roadkill and random points allowed us to unravel the road characteristics (e.g. steep embankments at roadsides and fences) determining the risk of WCV in birds and mammals. Other factors like the adjacent landscape matrix, the

Fig. 3 Probability of vehicle collision in birds (left) and mammals (right) according to height of roadside vegetation (upper fgures) and distance to the nearest curve (lower fgures). Grey dots are predicted values, the solid line denotes the ftted response of GLMMS and dashed lines show the 95% confdence intervals. Distances to the nearest curve were log transformed to meet normality assumptions

roadside vegetation, and vegetation density in center medians also contributed to determine roadkill probability.

Road related features

Traffic density is one of the most important predictors of roadkills (e.g. Clevenger et al. [2003;](#page-15-8) Seiler [2005;](#page-16-7) Barrientos and Bolonio [2009](#page-15-13); Zuberogoitia et al. [2014](#page-17-2); Gagné et al. [2015\)](#page-15-21), although its infuence on mortality is often non-linear; i.e. mortality peaks occur at intermediate levels of traffic density because animals are reluctant to cross highly transited roads (Madsen et al. [2002;](#page-16-3) Seiler [2005;](#page-16-7) Zuberogoitia et al. [2014](#page-17-2)). In our survey, the number of roadkills was associated with traffic density, but we did not found the expected reduction in mortality at high traffic density, even when we surveyed roads with enormous levels of traffic. At least two factors might explain the lack of a non-linear relationship between WVCs and traffic density. First, there might be a mismatch between the levels of traffic density and animal activity, since traffic density on the surveyed roads may be high only during the day, and many roadkilled species, especially mammals (see Table 2 in the Supplementary material), are most active during the night. Second, although often having

Fig. 4 Probability of vehicle collision in birds (upper) and mammals (bottom) according to type of habitat surrounding the road. Boxplots show the extreme of the lower whisker, the lower hinge, the median, the upper hinge, and the extreme of the upper whisker. Dots are data points that lie beyond the extremes of the whiskers. Only signifcant P-values $(0.05) from post hoc$ Tukey tests are shown

a deterrent effect, traffic noise or lighting may also attract some bird species to roadsides increasing their mortality rates (Blackwell et al. [2015](#page-15-22); Kociolek et al. [2015](#page-16-16)).

The infuence of road topography in the WVC risk was in agreement with our predictions. For birds, collision risk decreased in road sections with steep buried or elevated roadside embankments as opposed to those at the ground level or with soft slopes. Possibly, fat roads enable birds to fy close to the ground while crossing, thereby increasing collision risk, whereas the reverse is likely true if steep embankments are present (Clevenger et al. [2003;](#page-15-8) Kociolek et al. [2015\)](#page-16-16). Note, however, that the efect of topography (as well as that of road characteristics; see below) may be species-specifc and/or conditional on other factors. For example, car lights may dazzle nocturnal birds and increase their susceptibility to WVC or predators (Blackwell et al. [2015;](#page-15-22) Kociolek et al. [2015\)](#page-16-16). Further, scavengers (e.g. raptors) attracted to roads for foraging on roadkilled animals or species typically showing low-fight behaviors (e.g. owls; Massemin and Zorn [1998\)](#page-16-20) may be particularly vulnerable to traffic mortality, and such susceptibility may in turn be increased or diminished by the type of road embankments. For mammals, raised road points and those at the ground level showed, respectively, the highest and the lowest roadkill rates. These results suggest that roads with steep slopes at the roadside may discourage mammals from crossing (Alexander and Waters [2000;](#page-14-1) Clevenger et al. [2003;](#page-15-8) Malo et al. [2004](#page-16-6); Gunson et al. [2011](#page-16-11)). Elevating roads may therefore be a good option to mitigate roadkills of small- and medium-sized mammals, especially when combined with other elements such as fences or crossing structures (Clevenger et al. [2003](#page-15-8); Malo et al. [2004;](#page-16-6) Glista et al. [2009](#page-16-12); Gunson et al. [2011](#page-16-11)).

Vegetation cover in the median strip

Fig. 5 Probability of vehicle collision in birds (left) and mammals (right) according to type of fence (upper fgures) and amount of vegetation cover in the median strips (lower fgures). Boxplots show the extreme of the lower whisker, the lower hinge, the median, the upper hinge, and the extreme of the upper whisker. Dots are data points that lie beyond the extremes of the whiskers

The distance to the nearest curve, as determining the trade-of between improved visibility (reduced WVC risk) and increased velocity (increased WVC risk), was another important predictor of roadkill in mammals and birds. Given that vehicles must decelerate as approaching a curve, the quadratic efect of the distance to the curve on roadkill probability found in our study can be reasonably expected. Non-linear relationship between proximity to the nearest curve and roadkill risk has been previously reported in other studies (Table [1\)](#page-3-0), although the distance with the highest risk of roadkill varies widely among them, possibly due to a number of additional factors (e.g. presence of dense roadside vegetation, type of road and focal species) infuencing the likelihood of roadkill (Malo et al. [2004;](#page-16-6) Ramp et al. [2005;](#page-16-15) Zuberogoitia et al. [2006;](#page-17-4) Grilo et al. [2009](#page-16-10); Gunson et al. [2011\)](#page-16-11).

The presence of roadside barriers (fences with varying mesh sizes and walls) also shaped mortality risk in birds and mammal; so the presence of walls or fences with small mesh size that difcult the access to roads minimized roadkill risk. Our fndings are in agreement with previous works suggesting that, overall, these mitigation measures may be efective in reducing roadkills (Gunson et al. [2011](#page-16-11); van der Grift et al. [2013;](#page-17-3) van der Ree et al. [2015a\)](#page-17-5), but at least two considerations should be taken into account. First, in the case of terrestrial vertebrates (mammals, amphibians and reptiles) fences may act as barriers hampering wildlife (pre-breeding and/or dispersal) movements, thus reducing connectivity between populations (Trombulak and Frissell [2000](#page-16-1); Forman et al. [2003;](#page-15-0) Cofn [2007](#page-15-1)). Second, as reiterated in the literature (Glista et al. [2009;](#page-16-12) van der Grift et al. [2013;](#page-17-3) D'Amico et al. [2015](#page-15-9)), the use of barriers as a mitigate measure to prevent wildlife access to roads should ideally be combined with other measures, such as underpasses and scape structures, to keep permeability between populations and thus avoid the fatal consequences of trap-effects (Colino-Rabanal et al. [2011](#page-15-15); Cserkész et al. [2013](#page-15-23); Zuberogoitia et al. [2014](#page-17-2); van der Ree et al. [2015a\)](#page-17-5).

The infuence of median strips on roadkill risk has been scarcely assessed (Bellis and Graves [1971;](#page-15-14) Clevenger et al. [2003](#page-15-8); Clevenger and Kociolek [2013](#page-15-16)), even when these structures may have a critical effect on WVC (reviewed in Clevenger and Kociolek [2013](#page-15-16)). Medians are usually covered by dense vegetation that may provide relatively undisturbed breeding habitat, food resources (depending on the vegetation composition; Kociolek et al. [2015\)](#page-16-16), and concealment from predators and can therefore attract many animals (Adams [1984;](#page-14-2) Clevenger and Kociolek [2013\)](#page-15-16). Medians may thus increase roadkill risk by increasing wildlife presence and movements around roads (Bellis and Graves [1971;](#page-15-14) Clevenger et al. [2003;](#page-15-8) Clevenger and Kociolek [2013\)](#page-15-16). By contrast, we have found that the roadkill rate of birds and mammals (except lagomorphs; see Supplementary material) decreased as the amount of vegetation cover in the medians increased. Several factors may explain these results. Perhaps, in our study area, the composition and/or structure of vegetation in the median strips are not suitable as a foraging or breeding site. Densely vegetated medians might also function as an obstacle for crossing birds, encouraging them to fy high (see above) and thus avoid potential collisions (Kociolek et al. [2015\)](#page-16-16). In addition, the specifc requirements of the species afected and the synergistic efect of medians and microhabitat attributes might explain the apparent discrepancies between studies. For example, rabbits (the lagomorph most frequently found roadkilled during surveys) predominantly use the roadside vegetation and embankments as a refuge (Planillo and Malo [2013,](#page-16-21) [2018\)](#page-16-22), which might explain the lack of relationship between vegetation cover in the medians and the probability of roadkill in this group. Regardless of the determinants of collision risk, our fndings provide invaluable information about the efects of medians on WVC given the limited knowledge on this topic (Clevenger and Kociolek [2013\)](#page-15-16). Further research (e.g. testing the impact of continuous and discontinuous strips of median cover on diferent vertebrate groups and/or their efect if combined with other crossing structures) is needed to better understand the efect of these linear developments on animal movement and mortality (Clevenger and Kociolek [2013](#page-15-16)).

Landscape features

Bird roadkills were more likely to occur on roads with adjacent wooded areas, perhaps because wooded sites offer lower visibility in relation to more open habitats, as scrubland and farmlands (Clevenger et al. [2003](#page-15-8)). Dense tree cover may at the same time increases bird abundance around roads, as they often use trees as foraging and nesting sites. Indeed, the abundance of a species in the road surroundings was likely a major determinant of their roadkill rate since, although no data on local bird abundances are available (see below), top ranked species recorded in our study are among the most common species in Andalusia (e.g. *Passer domesticus*, *Turdus merula*, *Sylvia atricapilla* or *Erithacus rubecula*). In mammals, the highest rates of fatalities occurred in forested areas, but also in points adjacent to farm areas. These fndings are in agreement with other reports showing that mammal casualties increased in forested areas (Clevenger et al. [2003;](#page-15-8) Malo et al. [2004](#page-16-6); Ramp et al. [2005](#page-16-15); Seiler [2005](#page-16-7); Grilo et al. [2009;](#page-16-10) Gunson et al. [2011\)](#page-16-11). Moreover, the infuence of landscape on the roadkill risk may depend on species-specifc habitat preferences (Gunson

et al. [2011](#page-16-11)) and, in our survey, mammal mortality was dominated by wild rabbits and European hares (see Canal et al. [2018](#page-15-11)), typically associated to open and/or farm areas.

For birds and mammals, roadkill risk increased when the roadside vegetation was either very tall or very short. Short roadside vegetation -or lack thereof- may reduce WVC by increasing the reaction capacity of drivers and animals to dodge the collisions. On the contrary, by providing protection or food, medium-sized vegetation at the roadside such as small trees and shrubs may attract individuals to roads and, subsequently, infuence the probability of roadkill (Barrientos and Bolonio [2009\)](#page-15-13). Small trees and shrubs may also increase collision rates by, for example, favoring low-to-ground-level fights while crossing roads (Clevenger et al. [2003](#page-15-8); Ramp et al. [2006\)](#page-16-14), especially in narrow roads (personal observation) or when central median with scarce or no vegetation are present (see above). Furthermore, the roadside cover in the study area consists of lush plants, such as *Pistacia lentiscus, Rubus ulmifolius, Arbutus unedo*, which contribute to reduce visibility and, consequently, increases roadkill probabilities. In fact, roadside management (e.g. regular cutting and removal of dense vegetation) has proven an efective mitigation measure in diverse carnivores (Trombulak and Frissell [2000;](#page-16-1) Grilo et al. [2009](#page-16-10)) and birds (Kociolek et al. [2015\)](#page-16-16). For non-flying mammals, it is not surprising that the influence of vegetation height on collision risk decreased, after a threshold. In birds, however, tall vegetation should encourage high fights to cross roads, thereby reducing the probability of vehicle collision as vegetation height increases (Clevenger et al. [2003;](#page-15-8) Ramp et al. [2006](#page-16-14)). Not surprisingly, adding fences/walls adjacent to dense vegetation sites has proven an efective mitigation measures in birds and bats (Kociolek et al. [2015\)](#page-16-16). However, as discussed above (see "road related features"), it is important to ensure that those barriers do not entail additional, undesirable impacts on wildlife, such as collision (as may occur if walls are made of clear glass) or insurmountable barriers to movement (Kociolek et al. [2015\)](#page-16-16).

Potential limitations of the study

Field effort in terms of road distance covered and sampling frequency can strongly influence the accuracy of roadkill counts, because roadkills may be clustered in time and space and several biotic (scavengers) and abiotic (rainfall) factors may afect carcass persistence (Guinard et al. [2012;](#page-16-23) Teixeira et al. [2013;](#page-16-24) Barrientos et al. [2018](#page-15-24)). Thus, it is possible that the total number of casualties for some species is underestimated by monthly sampling (Teixeira et al. [2013](#page-16-24)). However, this should not be an issue because an accurate estimate of the number of road casualties is not the primary aim of this study; rather, our goal was to investigate the landscape and road features underlying the probability of roadkill. Our conclusions concerning the determinants of roadkill probability are unlikely to be biased by the sampling strategy, since we randomly alternated the road surveys (i.e. were randomly conducted in relation to the ecoregion, weather, type of road, and their fne-scale characteristics) and, therefore, there is no reason to think that roadkills passed systematically unnoticed at the most risky points, and viceversa.

During the first year of study, survey effort during autumn months was comparatively reduced due to logistic issues. This might have affected the roadkill estimates, for example, by decreasing the detection probability of those species that are most active or abundant during autumn, such as migratory birds. However, despite this potential inaccuracy, autumn peaks of mortality for birds and mammals clearly emerged from our survey and, importantly, the composition and temporal distribution of roadkills (see Canal et al. [2018](#page-15-11)) are in line with those found in surveys conducted at a shorter sampling periodicity (weekly or fortnightly) in the Iberian Peninsula (e.g. Frias [1999](#page-15-25); Grilo et al. [2009](#page-16-10); Garriga et al. [2012;](#page-15-26) Zuberogoitia et al. [2014;](#page-17-2) D'Amico et al. [2015](#page-15-9)). Thus, we are confdent that our results were not qualitatively afected by the lesser survey efort performed during the autumn months of the frst year of study.

Finally, due to the large spatial scale and range of taxa covered by our survey, local estimates of animal abundance and movements could not be obtained for the entire study region, and their potential efects could not be accounted for as suggested by many authors (Fahrig and Rytwinski [2009](#page-15-27); Gunson et al. [2011](#page-16-11); van der Ree et al. [2015b](#page-17-0); D'Amico et al. [2015\)](#page-15-9). Nonetheless, to partially control for this limitation, the characteristics of the landscape (e.g. land use) adjacent to roads were considered in our analyses, as they often infuence animal distribution, abundance, and movements (D'Amico et al. [2015\)](#page-15-9). Future confrmatory studies should explicitly account for these variables when developing models on WVC risk (van der Ree et al. [2015b\)](#page-17-0).

Conclusions

Data on wildlife roadkills were collected at an unusually large temporal (22 moths) and spatial (regional) scale, providing stronger inferences of the patterns detected. Fine-grained characterization of road and adjacent landscape characteristics allowed the identifcation of important factors determining collision risk in small-to medium-sized mammals and birds. Given that roadkill risk may vary at very small spatial scales, we highlight the importance of assessing collision risk based on site-specifc attributes and not uniquely on geographic information systems. Overall, reduced traffic density, steep roadside embankments, and structures hampering road access substantially reduced roadkills. By contrast, the efect of other predictors, such as land use adjacent to roads or the presence of curves, varied between vertebrate groups. It was also evident from our analyses that roadkill risk actually refects the interplay between diferent variables. Hence, we suggest that future studies should focus on assessing the efect of particular predictors in road sections with no or little variation in other infuential factors e.g. by assessing the efect of diferent median designs at sites showing similar roadside and landscape characteristics. Further research addressing the impact of medians on wildlife movement and mortality is urgently required because, despite their widespread use, the actual conservation impact of medians remains unclear.

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