



Northward establishment of the mediterranean mussel *Mytilus galloprovincialis* limited by changing climate

S. A. Lynch · A. Coghlan · B. O.' Leary · E. Morgan · S. C. Culloty

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Abstract Throughout Europe the blue mussel *Mytilus edulis* is dominant in the north while the Mediterranean mussel *Mytilus galloprovincialis* prevails in the south. Studies from the 1970s to the late 2000's documented the northward range expansion of *M. galloprovincialis* in Europe and predicted this trend to continue with climate change. The objectives of this study were to sample predominantly wild mussels (n = 1459) at twenty-four Irish intertidal sites over a seven year period and at three Welsh sites to investigate (a) the abundance and distribution of *Mytilus* spp., identified molecularly by polymerase chain reaction (PCR), and (b) compare with historical observations, at Irish sites where *M. galloprovincialis* was observed to be more abundant than *M. edulis*. Mussels were sampled more than once at certain sites to investigate if there was a temporal effect. The findings of this study indicated that *M. edulis* was consistently the most abundant species, followed by hybrids and *M. galloprovincialis*. At certain sites,

hybrids were detected while *M. galloprovincialis* was absent. This finding may indicate transient *M. galloprovincialis* populations or input from individuals subtidally. One factor that may be of importance was the anomalous cold winter “polar snaps” (2010 and 2011) that occurred during this study. In addition, heavy precipitation events and subsequent increased freshwater loading in bays and estuaries also occurred during this time period. Future warming climate scenarios have been predicted to facilitate the northwards establishment of *M. galloprovincialis*, however, nearshore meteorological extremes may have an impact in its larval settlement, establishment and subsequent reproductive output.

Keywords Mussels · *Mytilus edulis* · *Mytilus galloprovincialis* · Hybrids · Sea surface temperature · Rainfall

Introduction

Due to their widespread distribution, ecological importance and commercial value, *Mytilus* spp. are one of the most studied bivalves (Dailianis 2010). Throughout Europe the dominant species of mussels present are the blue mussel *Mytilus edulis* to the north and the Mediterranean mussel *Mytilus galloprovincialis* to the south. *M. edulis* are boreo-temperate in

S. A. Lynch (✉) · A. Coghlan · B. O.' Leary · E. Morgan · S. C. Culloty
School of Biological Earth and Environmental Sciences, Aquaculture and Fisheries Development Centre, Environmental Research Institute, University College Cork, Cork, Ireland
e-mail: s.lynch@ucc.ie

S. A. Lynch · S. C. Culloty
MaREI Marine and Renewable Energy, University College Cork, Cork, Ireland

their distribution on both coastlines of the Atlantic Ocean and are found intertidally, with a distribution in the eastern Atlantic occurring from Norway (Christiansen 1965) to the border of France and Spain (Sanjuan et al. 1994). *M. galloprovincialis* is native to the Mediterranean, Black and Adriatic Seas and has expanded its range on the Atlantic coast from Ireland to Morocco (Branch and Stephanni 2004). It is listed in the “World’s worst 100 invasive species” where it has outcompeted certain native mussel species and other bivalve species worldwide (Branch and Steffani 2004; Ozer and Guneydag 2014) and is categorized as a non-indigenous species (NIS) in the British Isles (GISD 2017), having first being reported in England in the 1950s (Hepper 1957; Lewis and Seed 1969). *M. galloprovincialis* was first reported to have expanded its range into Ireland in the 1970s (Gosling and Wilkins 1977).

Mytilus galloprovincialis outcompetes the native mussel species due to its faster growth rates, being more tolerant to air exposure and having a higher reproductive output (between 20 and 200% higher output) than the native mussel species (Branch and Steffani 2004). In populations found in the subtidal zones, *M. galloprovincialis* can grow to full size in a short period of time with the growth rate being roughly 50 mm shell length within approximately six to eight months (Page and Hubbard 1987) and is categorized as an invasive species in the British Isles (GISD 2017). Unlike *M. edulis*, which can tolerate a range of temperature changes (5–20 °C) (Zagata et al. 2008), *M. galloprovincialis* has a more constrained temperature range from 10–20 °C (FAO 2017). Lenz et al. (2018) investigated the effect of heat stress in five mussel species, including *M. edulis*, that would occur during transportation in regions with high sea surface temperatures when those species were on ship hulls or in ballast water tanks. That study observed that thermal tolerance was significantly enhanced in *M. edulis* when this species was exposed to a previous heat stress experience. Many studies have shown that *M. edulis* can tolerate wide fluctuations in salinity, desiccation and temperature, which has led to their wide distributional pattern (Tyler-Walters and Seed 2006; FAO 2017). *M. galloprovincialis* is generally cultured at higher salinities (< 34) and mussel production can be affected by low salinity, which kills mussels on the upper 0.5 to 1 m of the ropes (FAO 2017). In Ireland, *M. edulis* is predominantly found at

sheltered sites (Gosling and Wilkins 1981; Gosling et al. 2008) and on wave-exposed Irish shores it intermixes with *M. galloprovincialis* in varying proportions and this may also be seen worldwide (Gosling and Wilkins 1981; Gosling 2004). Gosling et al. (2008) observed that for the sixteen Atlantic sites on the south, west and north Irish coasts *M. galloprovincialis* and hybrids were detected while *M. edulis* was exclusive to east coast sites in the Irish Sea. In that study, it was concluded that *M. galloprovincialis* was increasing, not just at exposed sites but more strikingly at sheltered wave protected locations.

Larval dispersal of both mytilid species is influenced by the speed and direction of surface currents generated by the wind, and *M. galloprovincialis* has successfully established populations worldwide in temperate regions, in particular, where shipping ports are present (Branch and Stephanni 2004; Braby and Somero 2006). Mussel larvae are planktotrophic and can remain in the water column for up to several months before settling out (Bayne 1964; Seed 1969). A “regime shift” caused by climate change during the last century in the north Atlantic (Jones et al. 2001) has resulted in the northward migration of cold-water species of phytoplankton and zooplankton and warm-water species, including the larvae that feed on them, following them. In the North Sea, several demersal fish species have shown northward distributional shifts as a response to global warming (Perry et al. 2005). Lima et al. (2007) evaluated the direction of range shifts of macroalgae spp. based on historical data along the Portuguese coast and observed that cold-water species showed no particular shifting trend while all warm-water species, with shifting distributions, expanded their range northwards. More recently, poleward advances of subtropical species into the European theatre referred to as “African Creep” are occurring in the Eastern Atlantic (Canning-Clode and Carlton 2017), a “mirror” phenomenon of “Caribbean Creep” (Canning-Clode et al. 2011), however, the former study also highlighted that atypical cooling periods may provide a “reset mechanism” that will result in temporary setbacks or rate reductions in range expansions of non-indigenous species.

Hybridization in species that use external fertilization is common and marine hybrid zones are often referred to as extensive and heterogenous due to their large population size, high fecundity and larval

dispersal (Gardner 1997). *Mytilus* spp. populations show extensive hybridisation worldwide (Gosling 1992; Gardner 1994; Saavedra et al. 1996; Toro et al. 2002; Riginos and Cunningham 2005). The European *Mytilus* hybrid zone is spatially complex containing a mixture of pure, hybrid and introgressed individuals spanning over a 1400 km of coastline from France to the north of Scotland but is absent from the Irish Sea (Gosling et al. 1992; Gosling et al. 2008). Evidence of hybridisation and hybrid zones of *M. edulis* and *M. galloprovincialis* in the south west of England and Ireland were first recorded in the 1970s and subsequent studies have further documented this phenomenon (Gosling and Wilkins 1977; Skibinski et al. 1983; Gosling et al. 2008). Gosling et al. (2008) examined the population genetic structure of mussels at twenty sites from 2003 to 2006 on the Irish coast to determine the current range of *M. edulis* and *M. galloprovincialis*. In that study, a higher frequency of *M. galloprovincialis* and hybrids were observed on the Irish Atlantic coast compared to *M. edulis*.

The objectives of this current study were to determine mytilid species composition and abundance at a number of sites around the coast of Ireland (2010–2017) and at several Welsh sites (2011). In this study six Irish sites were revisited more than once and a comparison with the Gosling et al. (2008) was carried out using molecular markers at four sites to determine if any variation has occurred in distribution since that study was carried out, possibly due to changing weather patterns, habitat suitability, mussel fitness, recruitment success and or climate change.

Materials and methods

(1) Mussels sampled in current study

In this study, 1549 mussels were screened from 27 sites around Ireland and Wales with single sample sizes varying from 19 to 106 individuals. Screening took place from 2010 to 2017 with sampling occurring once at most sites, apart from six Irish sites that were sampled more than once (Table 1).

(2) Comparison with previous distribution study

The distribution and abundance of *M. edulis*, *M. galloprovincialis* and hybrids observed in this study were compared with a previous study by

Gosling et al. (2008), which was carried out with the sampling of 780 mussels at twenty sites from all Irish coasts from May 2003 to February 2006. Four sites were common to both studies: Carlingford Lough (Site #19), Carnsore Point (Site #1), Garrettstown (Site #13) and Lough Hyne (Site #14) (Table 1) and for additional comparison between this study and the Gosling et al. (2008) study, all sites were categorised by coast (north, west, south and east) with overall abundance of each *Mytilus* species and hybrids being compared.

(3) Molecular analysis to identify mussel species

Gill tissue (5 mm²) was excised from each mussel (n = 1549) and was stored in 96% ethanol if DNA extraction was not conducted on the same day of sampling. DNA was extracted using the chelex-100 extraction method (Walsh et al. 1991).

PCR was carried out to detect the nuclear DNA markers Me15/Me16 (Inoue et al. 1995). The PCR mastermix was modified slightly to include 5 × green buffer. Amplification was conducted in 25 µl of the reaction mixture containing 14 µl ddH₂O, 5 µl 5 × green buffer (Promega), 2.5 µl of each of the four deoxyribonucleotide triphosphates (dATP, dCTP, dGTP, dTTP), 1.5 µl MgCl₂, 0.5 µl of the primer Me15 (5'-CCAG TATACAAACCTGTGAAGA-3'), 0.5 µl of the primer Me16 (5'-TGTTGTCTTAATGGTTT TAAGA-3'), 1 µl of Taq polymerases and 1 µl of total DNA. The following conditions were used for the PCR in a thermocycler: 94 °C for 30 s, 55 °C for 45 s and 70 °C for 90 s (40 cycles). Electrophoresis of the amplification products was conducted in a 2% agarose gel. The expected product size for *M. galloprovincialis* is 126 bp while in *M. edulis* it is 180 bp. In hybrids, both bands occur simultaneously at 126 bp and 180 bp.

(4) Meteorological data

Irish sea surface temperature data and graphs were sourced from Casal and Lavender (2017) while surface temperature data was sourced from the national meteorological agency Met Eireann (<https://www.met.ie/climate/what-we-measure/temperature>).

A long time series of annual rainfall from twenty-five stations in Ireland was sourced from

Table 1 Mussel sample sites with description of sample month, year and size at each site

Site	Grid reference	Sample date		n
<i>South coast Ireland</i>				
(1) Carnsore point (E)**	52°10'14''N, 6°21'20''W	July 2013	Wild	26
(2) Dunmore east (E)	52°08'60''N, 6°58'59''W	Feb 2011	Wild	30
(3) Annestown (E)	52°08'17''N, 7°18'16''W	Sep 2017	Wild	30
(4) Ballyquin (E)	51°58'28''N, 7°42'13''W	Sep 2017	Wild	30
(5) Stradbally (E)	52°07'28''N, 7°27'54''W	Sep 2017	Wild	30
(6) Ballyvoile (SE)	52°06'20''N, 7°30'38''W	June 2013	Wild	53
(7) Clonea (E)	52°16'42''N, 7°26'16''W	Sep 2017	Wild	30
(8) Dungarvan (SE)*	52°05'17''N, 7°37'31''W	July 2015	Wild	106
		Sep 2017	Wild	30
(9) Helvic head (E)	52°03'10''N, 7°32'22''W	Sep 2017	Wild	30
(10) Whiting bay (E)	51°56'55''N, 7°46'32''W	June 2013	Wild	31
(11) Ballymacoda (S)	51°53'36''N, 7°53'57''W	July 2015	Wild	60
(12) Cork Harbour-Rochespoint (E)*	51°48'34''N, 8°15'16''W	June 2010	Wild	30
		June 2017	Wild	80
		Nov 2017	Wild	60
(12) Cork Harbour-Whitegate (SE)	51°49'36''N, 8°14'15''W	June 2010	Wild	30
(12) Cork Harbour-Whitepoint (S)	51°50'47''N, 8°18'39''W	June 2010	Wild	30
(12) Cork Harbour-Ringaskiddy (S)*	51°50'49''N, 8°19'3''W	June 2010	Wild	30
		Nov 2010	Wild	30
		June 2017	Wild	60
		Nov 2017	Wild	80
(13) Garrettstown (E)*, **	51°38'38''N, 8°35'02''W	June 2013	Wild	30
		May 2017	Wild	60
(14) Lough hyne (S)*, **	51°30'04''N, 9°18'01''W	June 2013	Wild	28
		May 2017	Wild	60
<i>West coast Ireland</i>				
(15) Castlegregory (E)	52°15'39''N, 10°00'41''W	July 2013	Wild	30
(16) Clew bay (SE)	53°48'14''N, 9°42'31''W	Dec 2010	Wild	30
(17) Sligo bay (E)	54°17'56''N, 8°38'25''W	July 2011	Cultured	60
<i>North coast Ireland</i>				
(18) Lough foyle (S)	55°6'29''N, 7°5'38''W	Oct 2010	Wild	19
<i>East coast Ireland</i>				
(19) Carlingford lough (S)*, **	54°03'5''N, 6°11'05''W	July 2010	Cultured	60
		July 2015	Wild	106
(20) Silver strand (S)	52°57'25''N, 6°01'02''W	July 2011	Wild	30
(21) Poulshone strand (S)	52°37'52''N, 6°10'16''W	July 2011	Wild	30
<i>Wales, UK</i>				
(22) The menai, strait (S)	53°17'45''N, 4°03'10''W	May 2011	Cultured	30
(23) Gallows point (S)	53°16'01''N, 4°05'34''W	May 2011	Wild	30
(24) Swansea bay (SE)	51°37'10''N, 3°57'05''W	June 2011	Wild	30
Total N				1549

*Sites where mussels were sampled on more than one occasion

**Gosling et al. (2008) sites

Met Eireann (<https://www.met.ie/climate/what-we-measure/rainfall>) and Noone et al. (2015).

(5) Statistical analysis

Chi-square (χ^2) analysis was carried out using Excel to calculate significant difference in the abundance of each mussel species between (a) sites with varying water quality and (c) sites with varying exposures, with significance determined at $p < 0.05$. Chi-square (χ^2) analysis was also carried between this current study and the Gosling et al. (2008) study, comparing the abundance of each mussel species and their overall abundance on each coast.

Results

(1) *Mytilus edulis*, *Mytilus galloprovincialis* and hybrids of both parent species distribution and abundance at Irish sample sites

In total, 1413 mussels (96.9%) were identified to species by PCR from the Irish study sites. Of this, 898 (63.6%) individuals were *M. edulis*, 235 (16.6%) *M. galloprovincialis* and 280 (19.8%) hybrids were detected along the coast (Table 2 and Fig. 1). Irish sites (#19, #20 and #21) in the Irish Sea were the only sites that were found to have 100% *M. edulis* while all of the Welsh mussels ($n = 90$) were identified as *M. edulis* in the PCR.

Excluding the Irish Sea samples on the east coast, *M. edulis* was the dominant species in eighteen samples on the south, west and north coasts, *M. galloprovincialis* was the dominant species in two samples (Sites #5 and #15) on the south and west coast respectively and hybrids dominated nine samples at sites on the south coast (Table 2). Of the five Irish sites that were sampled more than once and with a time difference of two (Site #8), four (Site #14) and seven (Site #12) years between sampling time points, the status of the dominant species remained the same except for one of the locations Roches Point within Cork Harbour (Site #12), where there was a decrease in *M. edulis* and an increase in hybrids in November 2017. Hybrids were consistently the dominant

species at the marine nature reserve at Lough Hyne (Site #14).

(2)

Comparison of mussel species distribution and abundance in this study with historical study Comparing the results of this study and the Gosling et al. (2008) study by coast abundance and distribution for each of the *Mytilus* spp. and hybrids, a significant difference in *M. edulis* abundance ($p < 0.05$) was observed (Figs. 2 and 3). *M. edulis* was reported to be the dominant species on all coasts in this study, while *M. galloprovincialis* and hybrids were the most abundant in the Gosling et al. (2008) study. A similar abundance in hybrids on the south coast was observed in both studies while no significant difference in *M. galloprovincialis* abundance in both studies was recorded on the west coast.

Over the time points of the two Irish studies there has been an increase in *M. edulis* abundance at certain sites (Dungarvan Site #8, Garrettstown* Site #13 and Carnsore Point* Site #1), while its abundance has remained stable at other sites such as Ringaskiddy (Site #12) and Carlingford Lough* (Site #19). A decrease in *M. edulis* abundance has been observed in more recent years at Roches Point (Site #12 from 2015 to 2017) and at Lough Hyne* (from 2013 to 2017) (Fig. 3). *M. galloprovincialis* abundance has been showing a decreasing trend at Dungarvan, Ringaskiddy, Garrettstown*, (significantly so $p < 0.05$) and at Carnsore Point* (significantly so ($p < 0.05$) while its abundance has started to increase at Lough Hyne*. Hybrid abundance has significantly decreased at Carnsore Point* along with a decrease at Dungarvan and Garrettstown, however, the latter site has shown a slight increase from 2013 to 2017 (+ 8%). Hybrid abundance is consistently increasing over the study time points at Lough Hyne and has significantly increased at Roches Point from 2015 to 2017 (+ 30%).

(3) Meteorological data

Irish waters present a warming trend of $0.3\text{ }^{\circ}\text{C decade}^{-1}$ with warming being more intense in autumn (Fig. 4, Casal and Lavender 2017). A significant latitudinal gradient is present in Irish waters with waters to the south being a lot

Table 2 PCR analysis of mussel samples (No. of individuals/Sample total) from twenty-four Irish and three Welsh sites to determine relative abundance of *Mytilus edulis*, *Mytilus galloprovincialis* and hybrids spatially and temporally

Sample site	Sample date	<i>M. edulis</i>	<i>M. galloprovincialis</i>	Hybrid
Irish sites				
<i>South coast</i>				
(1) Carnsore point	June 2013	31% (8/26)	27% (7/26)	42% (11/26)*
(2) Dunmore east	February 2011	64% (19/30)*	33% (10/30)	3% (1/30)
(3) Annewstown	September 2017	30% (9/30)	30% (9/30)	40% (12/30)*
(4) Ballyquinn	September 2017	32% (10/30)	30% (9/30)	36% (11/30)*
(5) Stradbally	September 2017	19% (6/30)	58% (17/30)*	23% (7/30)
(6) Ballyvoile	June 2013	32% (17/53)	19% (10/53)	49% (26/53)*
(7) Clonea	September 2017	63% (19/30)*	18% (5/30)	19% (6/30)
(8) Dungarvan	April–May 2015	75% (80/106)*	19% (20/106)	6% (6/106)
	September 2017	100% (30/30)*	0% (0/30)	0% (0/30)
(9) Helvic Head	September 2017	31% (9/30)	31% (9/30)	38% (12/30)*
(10) Whiting Bay	June 2013	26% (8/31)	26% (8/31)	48% (15/31)*
(11) Ballymacoda	July 2015	55% (33/60)*	8% (5/60)	37% (22/60)
(12) Cork Harbour-Roches point	June 2010	63% (12/19)*	16% (3/19)	21% (4/19)
	June 2017	66% (53/80)*	22% (18/80)	12% (9/80)
	November 2017	35% (21/60)**	23% (14/60)	42% (25/60)*
(12) Cork Harbour-Whitegate	June 2010	68% (13/19)*	0	32% (6/19)
(12) Cork Harbour-Whitepoint	June 2010	94% (17/18)*	0	6% (1/18)
(12) Cork Harbour-Ringaskiddy	June 2010	80% (16/20)*	5% (1/20)	15% (3/20)
		77% (23/30)*	7% (2/30)	17% (10/60)
	November 2010	71% (43/60)*	17% (5/30)	12% (7/60)
	June 2017	80% (64/80)*	7% (6/80)	13% (10/80)
	November 2017			
(13) Garrettstown	June 2013	50% (15/30)*	40% (12/30)	10% (3/30)
	May 2017	54% (32/60)*	28% (17/60)	18% (11/60)
(14) Lough Hyne**	June 2013	36% (10/28)	14% (4/28)	50% (14/28)*
	May 2017	18% (11/60)	20% (12/60)	62% (37/60)*
<i>West coast</i>				
(15) Castlegregory	July 2013	29% (8/28)	57% (16/28)*	14% (4/28)
(16) Clew bay	December 2010	53% (16/30)*	23% (7/30)	23% (7/30)
(17) Sligo bay	July 2011	86% (52/60)*	7% (4/60)	7% (4/60)
<i>North coast</i>				
(18) Lough Foyle	October 2010	95% (18/19)*	0	5% (1/19)
<i>East coast</i>				
(19) Carlingford Lough	July 2010	100% (60/60)	0	0
	April–May 2015	100% (106/106)	0	0
(20) Silver strand	July 2011	100% (30/30)	0	0
(21) Poulshone strand	July 2011	100% (30/30)	0	0
<i>Welsh sites</i>				
(22) Menai strait	May 2011	100% (30/30)	0	0
(23) Gallows point	May 2011	100% (30/30)	0	0
(24) Swansea bay	June 2011	100% (30/30)	0	0

*Dominant species at each sample site, date and year

**Change in status

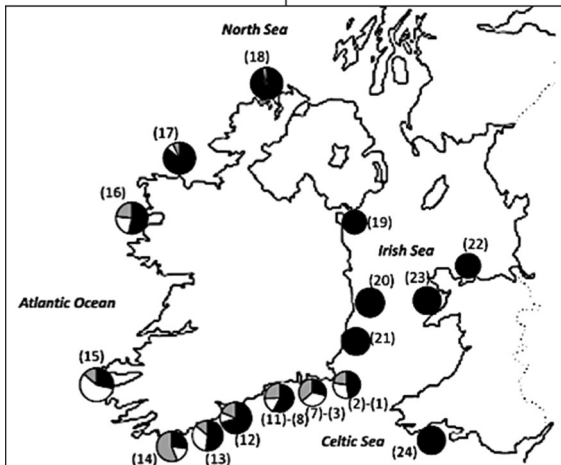


Fig. 1 Map of Ireland and Wales showing the wild *Mytilus edulis* (black), *Mytilus galloprovincialis* (white) and hybrid (grey) abundance and distribution at study sites (study site number in parentheses)

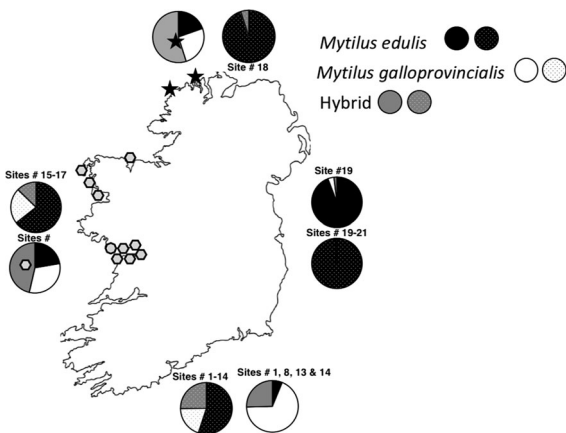


Fig. 2 Comparison of *Mytilus* spp. and hybrid distribution and abundance on each Irish coast in the Gosling et al. (2008) study (block colour fill sites) and in this study (pattern fill)

warmer than waters to the north and a seasonal effect is also present with coldest SST values in March and highest SST values in August.

Mean air temperatures have varied considerably from year to year, however, warming periods have occurred from the late 1980s to the present time (<https://www.met.ie/climate/what-we-measure/temperature>). There has been an increase in the number of days with temperatures over 20 °C, and a decrease in the number of days with temperatures

below 0 °C, however, extreme cold winters did occur in 2010 and in 2011 with air temperatures of -2.5 °C from the long term average (LTA).

Rainfall shows great year to year variability, however, from the 1970s (*M. galloprovincialis* first recorded to be present in Ireland) to 2005, when the Gosling et al. (2008) study was primarily conducted, maximum national annual rainfall on average did not exceed 1290 mm of rainfall (Fig. 5A). Since 2005 national annual rainfall in Ireland has been increasing with up to 1400 mm of rainfall being recorded. Winter 2015/16 was the wettest winter on record, followed by winter 2013/14, and are the only winters where rainfall totals in excess of 500 mm were recorded (Noone et al. 2015) (Fig. 5B). A 30 year running mean of the national annual rainfall has indicated that an increase in average national rainfall of approximately 70 mm over the last two decades has occurred with all seasons showing a small increase in totals over the last few decades (Noon et al. 2015).

Discussion

Mytilus edulis, *M. galloprovincialis* and hybrid forms of both parental species were detected around the coast of Ireland, while *Mytilus trossolus* or hybrids of this species were not detected. *M. edulis* was the dominant species found at most Irish sites with a lower abundance of *M. galloprovincialis* and hybrids. These findings are in stark contrast to the study of Gosling et al. (2008) carried out seven years prior to this study. In the intervening time period (fourteen years) the relative abundance of the parental forms of both mussel species and hybrids around the coast of Ireland has changed with *M. edulis*, the native species, being once again the most frequently observed and abundant species. The slightly higher abundance of hybrids over *M. galloprovincialis* may indicate that these are more suited to survival under Irish conditions and a changing Irish climate than the Mediterranean parent species or it is possible that they are first generation transient hybrids. In the Irish Sea area, *M. edulis* continues to be the only species present between Carlingford Lough and Poulshone Strand along the Irish coast and Galloway Point and Swansea along the Welsh coast. It has been suggested that a thermal front develops during late spring at the boundary between the Celtic Sea and the southern Irish Sea (Brown et al. 2003),

Fig. 3 Comparison of mussel species abundance at Irish sites sampled in this study and in the Gosling et al. (2008) study* including sites that were revisited more than once in this study

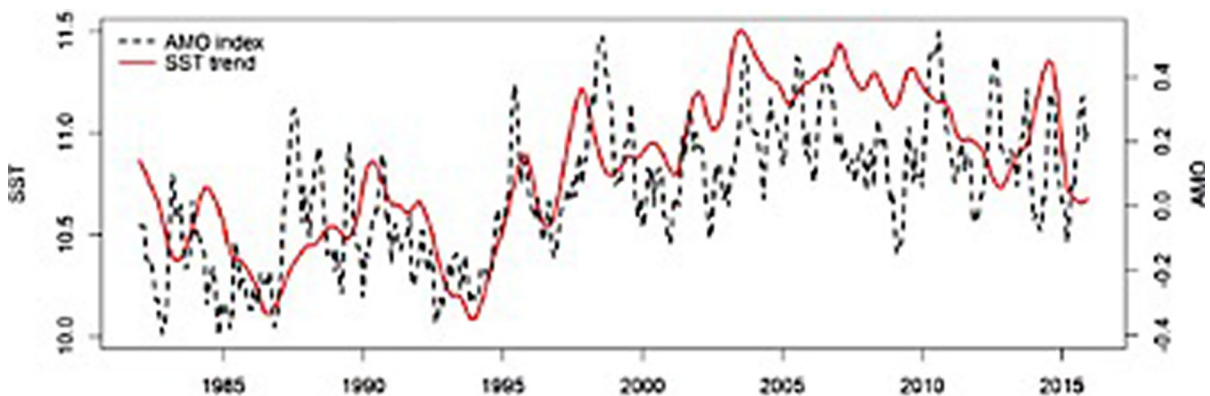
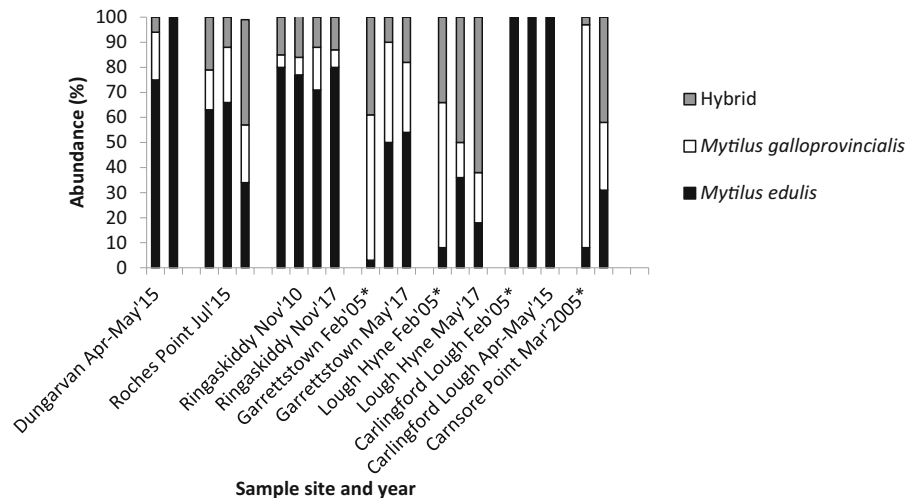


Fig. 4 Sea surface temperature (°C) (SST) in Irish waters showing SST trend and Atlantic Multi-decadal Oscillation (AMO) index sourced and modified from Casal and Lavender (2017)

which prevents the movement of *M. galloprovincialis* into the Irish Sea region.

Mosaicism or spatial genetic patchiness in distribution is a common feature of *Mytilus* hybrid zones (Comesaña and Sanjuan 1997; Rawson and Hilbish 1998) and was observed at most Irish sites in this study. In the Gosling et al. (2008) study, this phenomenon was not observed, however, hybrids and *M. galloprovincialis* only were present at a single site on the west coast of Ireland, with the hybrids outnumbering the parent species possibly indicating hybrid vigour. Hybrids may have lower fitness due to genetic discontinuities (Mayr 1963; Barton and Hewitt 1985) or a lack of “hybrid” habitat (Howard 1986) and mytilid hybrids can be more susceptible to disease (Fuentes et al. 2002), however, certain studies have shown “hybrid vigour” (heterosis) in the first hybrid

generation (F1 hybrid) which may allow hybrids to function over a wider range of environmental conditions than the two parental species (Moore 1977; Arnold 1997). Hybrids were consistently the dominant group at the marine reserve Lough Hyne and this may possibly be due to a number of reasons. One is that fertile hybrids may be reproducing earlier than *Mytilus* spp. and their larvae are settling earlier within the lough. This phenomenon of *Mytilus* hybrids being fertile and reproducing at a faster rate has been observed (Dias et al. 2009). Another reason could be that hybrids grow at a faster rate than either of their parent species, which would mean that there are outcompeting the *Mytilus* spp. for food and space. Additionally, *M. galloprovincialis* are showing signs of recovery in the lough after decreasing in abundance

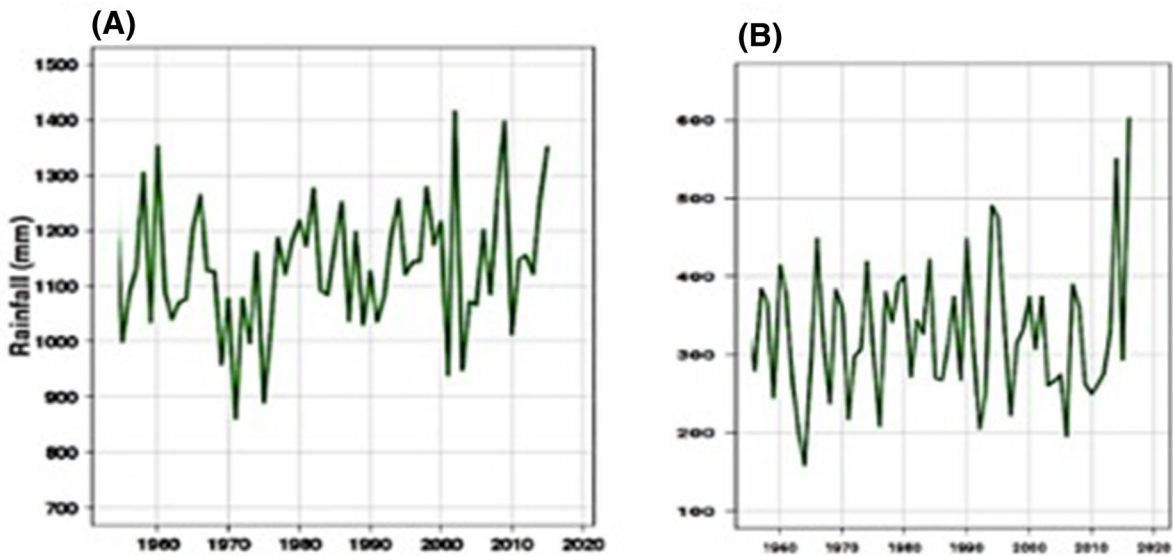


Fig. 5 **A** Mean national annual rainfall for Ireland (25 stations) from the 1960s to 2015 with **B** winter rainfall highlighted. Both graphs sourced and modified from Noone et al. (2015)

from 58% in 2005 to 14% in 2013 with an increase to 20% in 2017.

Irish air temperature recorded by Met Eireann (<https://www.met.ie/climate/what-we-measure/temperature>) during this study and the Gosling et al. (2008) study indicates that two prolonged cold winters occurred in 2010 and in 2011. Although surface seawater temperatures did not deviate dramatically from average temperatures during and shortly after this time period, air temperature, snowfall and ground frosts did. Severe winters have been associated with the complete or almost complete extinction of intertidal populations (Blegvad 1929; Smidt 1944; Kristensen 1957; Beukema 1985), however, *M. edulis* can withstand extreme cold and freezing, surviving when its tissue temperature drops to -10 to -30 °C (Williams 1970; Seed and Suchanek 1992; Loomis 1995). *M. edulis* was relatively unaffected by the severe winter of 1962/63 on the south-east coast of England, with 30% mortality reported and 2% in south Wales (Crisp 1964). Crisp (1964) noted that most mortality was caused by predation on individuals weakened or moribund due to the low temperatures rather than the temperature itself. It is unclear what the minimum temperature tolerance limit is in *M. galloprovincialis* but this mussel species is known to tolerate 7 °C, during upwelling events on the South Africa coastline (Erkom and Van Griffiths 1992). Jansen et al. (2007)

noted that *M. galloprovincialis* populations were more sensitive to cold shock than *M. edulis* populations.

Northward seawater currents make contact with Ireland initially on the south coast before being transported in coastal currents that move around the Irish coast in a clockwise direction (Steele et al. 2013). This would support the higher abundance of *M. galloprovincialis* observed on the south and southwest coasts in this study, as the warmer seawater temperatures experienced in southern Irish waters (Casal and Lavender 2017) would allow a more stable population to remain and possibly self-recruit. Previous studies of mytilid spp. population distribution have focussed on the limiting effects of water temperature on dispersal (Barry et al. 1995; Suchanek et al. 1997; Sagarin et al. 1999). *M. galloprovincialis* prefers warmer waters (Sarver and Foltz 1993; Suchanek et al. 1997) and a range expansion of *M. galloprovincialis* is thought to have occurred from the 1930's to the early 1990's due to increased water temperatures (Geller 1999; Jones et al. 2001; Schneider 2008). However, air temperature and prolonged frosts associated with more frequently occurring “cold snaps” during winter months may have a limiting effect on populations sustaining themselves and outcompeting the native mussel populations.

Of interest, national annual rainfall in Ireland has been increasing with up to 1400 mm of rainfall being

recorded since 2005 and with winters 2013/14 and 2015/16 being the wettest winters on record (Noone et al. 2015). Heavy precipitation events and subsequent increased freshwater loading in bays and estuaries may result in sudden fluctuating and reduced salinity events, which would have an impact on *M. galloprovincialis*. An earlier Irish study carried out by Seed (1974) stated that when salinity was not reduced the invasive *M. galloprovincialis* was the dominant mussel, out competing the native species. *M. galloprovincialis* made up 40% of the population in those locations (Seed 1974). Shurova (2001) observed that the lowering of salinity resulted in an increase in mortality rate and a decrease in numbers, biomass, juvenile recruitment, growth rates and annual production of *M. galloprovincialis* in the Black Sea.

The results of this study would indicate that although *M. galloprovincialis* and hybrids of both parent species are present on the south, west and north coasts of Ireland, the native *M. edulis* is dominant at most sites. Although a warming trend has been observed in sea surface temperatures over several decades, several anomalous cold winters compounded by an increase in rainfall and subsequent run off of freshwater loadings into coastal locations may be having an effect on nearshore habitat suitability for *M. galloprovincialis*. Future predictions of colder winters over Northern Europe, due to climate change and cold polar air streaming down in to Europe (Tang et al. 2013), may be limiting for this Mediterranean species as well as predicted increased precipitation (Noone et al. 2015). Hughes et al. (2015) observed that the eastern limit of distribution of *M. galloprovincialis* in South Africa was also set by unfavourable abiotic conditions. Further sampling of mussels at the Irish sites in the future, nearshore and subtidally, would be of interest to determine if the *Mytilus* spp. population structure is remaining stable over time or if localised factors and meteorological influences impact its composition.

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