

Can climate change jeopardize predator control of invasive herbivore species? A case study in avocado agro-ecosystems in Spain

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Abstract Climate change is one of the most important factors affecting the phenology, distribution, composition and diversity of organisms. In agricultural systems many pests and natural enemies are arthropods. As poikilotherm organisms, their body temperature is highly dependent on environmental conditions. Because higher trophic levels typically have lower tolerance to high temperatures than lower trophic levels, trends towards increasing local or regional temperatures may affect the strength of predator/prey interactions and disrupt pest control. Furthermore, increasing temperatures may create climate corridors that could facilitate the invasion and establishment of invasive species originating from warmer areas. In this study we examined the effect of environmental conditions on the dynamics of an agro-ecosystem community located in southern Spain, using field data on predator/prey dynamics and climate gathered during four consecutive years. The study system was composed of an ever-green tree species (avocado), an exotic tetranychid mite, and two native species of phytoseiid mites found in association with this new pest. We also present a climatological analysis of the temperature trend in the area of study during the last 28 years, as evidence of temperature warming occurring in the area. We found that the range of temperatures with positive per capita growth rates was much wider in prey than in predators, and that relative humidity contributed to explain the growth rate variation in predators, but not in prey. Predator and prey differences in thermal performance curves could explain why natural enemies did not respond numerically to the pest when environmental conditions were harsh.

Keywords Environmental conditions · *Oligonychus perseae* · Phytoseiidae · Community dynamics · Invasive mite

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Introduction

There is increasing evidence that ecosystems are being affected by local or regional warming on all continents (IPCC 2007; Beaumont et al. 2009; Hole et al. 2009). Recent studies have shown that climate change is altering the phenology, distribution, composition and diversity of plant (Morecroft and Keith 2009), bird (Fiedler 2009), mammal (Humphries 2009) and insect (Pelini et al. 2009) communities, as well as of pelagic and plankton marine (Edwards 2009) and intertidal ecosystems (Mieszowska 2009), and coral reefs (Attrill 2009). It is also acknowledged that climate change may become a threat for the expansion and maintenance of crops managed with environmentally friendly techniques, such as biological and conservation pest control. This is because pesticide use may increase with warming due to the formation of climate corridors that may facilitate the spread and establishment of exotic pest species. Also, it has been stated recently that global change is influencing virtually every type of interactions between species (Tylianakis et al. 2008, and references therein). As a result, levels of herbivory have increased, likely because higher trophic levels are more sensitive to warming than lower trophic levels (Cagnolo et al. 2002; Voigt et al. 2003; Preisser and Strong 2004); Last, global warming introduces new uncertainties to the future of agriculture in general, because models of climate change predict an increase of the frequency and intensity of extreme events, such as hurricanes, floods, heat waves or severe drought (Lavalle et al. 2009).

It is predicted that climate warming will vary between regions. According to the analysis of the Intergovernmental Panel for Climate Change, IPCC (2007), summers will tend to be warmer in southern Europe, with an increase in the intensity, duration and number of periods with extreme high temperatures. Within southern Europe, the Mediterranean region is amongst the areas considered highly susceptible to climate change, because it is an area of transition between the Euro-Siberian or temperate and the Saharo-Sindic climatic regions (Benito-Garzon et al. 2008).

In agricultural systems many pests and natural enemies are arthropods. As ectotherm organisms, their body temperature is highly dependent on the ambient temperature. Many of the key processes of arthropods, such as metabolism and development, and factors that determine the intensity of trophic interactions, such as predation rate, movement or speed, are mediated by environmental conditions (Beveridge et al. 2010). It is expected, therefore, that changes in local or regional environmental conditions, especially increases of temperature and periods of extreme heat and drought, may profoundly affect pest control. Indeed, trophic cascades could be dampened and biological control jeopardized (Araújo and Luoto 2007; Hegland et al. 2009), if natural enemy efficiency or life-history is negatively affected by environmental warming. Furthermore, disruption of biological pest control may occur more frequently in agricultural systems affected by exotic invasive pests originating from warmer areas, because they are expected to be more adapted to warming than native natural enemies.

In this study we aimed at identifying whether environmental conditions influenced species abundance and dynamics of an agro-ecosystem community located in southern Spain, using field data on predator/prey dynamics and climate gathered during four consecutive years. We attempted to identify whether different responses to summer environmental conditions between predators and the pest could be detected at the population level in the field. We also present a climatological analysis of the temperature trend over the last 28 years in the area of study as evidence of temperature warming occurring in the area.

Materials and methods

The agro-ecosystem

Avocado (*Persea americana* Mill.; Lauraceae) is an ever-green tree species originating from Mesoamerica and introduced in southern Spain during the sixteenth century by the Spaniards. Since then it has become an economically important crop in the coastal areas of Malaga and Granada, in the region of Andalusia, where, except for almond and olive trees, it is the non-citrus fruit tree that occupies most of the cultivated area (MARM 2011). One of the most eminent characteristics of the avocado cultivation in Andalusia was its excellent phytosanitary status. Indeed, avocado trees in Spain had been exempted from important pests, probably because all the avocado tissues (leaves, seeds, roots and fruits) contain specialized oil cells which accumulate alkaloids and terpenes (Armstrong 1964; Platt-Aloia et al. 1983; Platt and Thompson 1992) with insecticidal, antifeedant, and growth inhibitory activities (Rodríguez-Soana et al. 1997, 1998; Rodríguez-Soana and Trumble 2000). However, this situation has recently been endangered because of the arrival of an exotic pest, the persea mite, *Oligonychus perseae* Tuttle, Baker and Abatiello, a tetranychid species which originated from Mesoamerica.

The persea mite was detected for the first time in commercial avocado crops of the province of Malaga in 2004. It is also described as avocado pest in California (USA), Canary Islands, Costa Rica, Israel, Portugal, Valencia (Vela et al. 2007), and Florida (USA) (Welbourn 2007). This pest builds dense silken nests at the underside of avocado leaves, along the main nerves. Inside the nests adults feed and reproduce, and juveniles undergo development. Feeding causes necrotic spots that can affect up to 90 % of the leaf area (Aponte and McMurtry 1997). Among other functions, nests protect mites against attack from some species of natural enemies, and against adverse environmental conditions (Montserrat et al. 2008; Mori et al. 1999).

Two species of phytoseiid mites have been found in association with this invasive pest in the Andalusian coastal avocado orchards (González-Fernández et al. 2009). *Euseius stipulatus* (Athias-Henriot) is an omnivorous species, not available commercially, that attacks animal prey and also forages on pollen (Bouras and Papadoulis 2005; Ferragut et al. 1987). *Neoseiulus californicus* (McGregor) is a specialist predator of tetranychid mites that is commercially available. Previous studies showed that both species actively hunt and prey on the persea mite, *E. stipulatus* mainly attacking the individuals wandering outside nests, whereas *N. californicus* uses its forelegs to rip the silken walls of the nests to attack the individuals inside the nests (González-Fernández et al. 2009; Montserrat et al. 2008).

The mite predator/prey community dynamics in avocados

During four consecutive years (2006 through 2009), the abundance of phytoseiid mites and the persea mites was recorded during spring and summer, starting at the end of March until the end of September. Two-season samplings were done because yearly population dynamics of phytoseiid mites in avocado trees typically show two population peaks, the first in spring and the second in summer. In spring, when the persea mite is still rare, the omnivore *E. stipulatus* is by far the most abundant species. It has recently been suggested that its numerical response is associated with a non-prey food source, pollen (mainly from olive), which is deposited from the atmosphere on the surface of the avocado leaves (González-Fernández et al. 2009). In summer, the two species of phytoseiids co-occur and

their populations respond numerically to the abundance of the perseia mite (González-Fernández et al. 2009).

Predator and prey populations were monitored in a 1.8 ha avocado orchard (cv ‘Hass’) of ca. 70 avocado trees located at the IHSM La Mayora (Málaga). Ten leaves per tree, covering the whole perimeter of the canopy to account for within-tree abundance variability, at ca. 150 cm height were sampled from 10 to 20 avocado trees every 2 weeks. At each sampling date (a) the number of occupied nests (with mobile stages and/or eggs inside) on the upper margin of the second left vein of the underside of the leaf (UML2, hereafter), and (b) the number of mobile stages of phytoseiid mites on the leaf, was recorded in situ using a field magnifying glass (Ruper 8x). The number of occupied nests on the UML2 is a good estimate of the total number of individuals (mobiles and eggs) per leaf (González-Fernández et al. 2009).

Because phytoseiids in spring probably respond numerically to the abundance of pollen, data of pollen concentration in the atmosphere were obtained from an aerobiology station belonging to the Spanish Aerobiology Network (REA, Red Española de Aerobiología), from 2006 to 2008. This station, which is located 10 km in a straight line from the study area, uses a Hirst type volumetric pollen trap. Data recorded in this station are provided as the number of pollen grains per cubic metre of air and day.

Effect of environmental conditions on predator/prey abundance

Climate data used in the analyses were obtained from an agro-climate station that belongs to the Government of Andalusia. This station was chosen because it is located close to the area of study (ca. 15 km linear distance), and like the avocado orchard positioned close to the sea. The station provided daily records on maximum, minimum and average temperature and relative humidity at a height of ca. 2 m above the ground.

The per capita growth rates between two successive samplings ($r_{t,t+\tau}$) were calculated for both predators and prey as

$$r_{t,t+\tau} = \frac{1}{\tau} \ln \left(\frac{N_{t+\tau}}{N_t} \right),$$

where τ is the time interval (days) between the first and the second sampling event and N_t is the populations size at time t . For each value of $r_{t,t+\tau}$ concurrent values of temperature and humidity were obtained averaging the daily records of temperature and relative humidity during the interval $(t, t + \tau)$. Data were divided into two sets. The first set included all data gathered during the four springs, when pollen in the atmosphere was present and the perseia mite was rare. The second set included all data gathered during the four summer seasons, when pollen in the atmosphere was absent and the perseia mite was present.

The variation of the interval per capita growth rate of predators during spring, and of predators and prey during summer, was analysed by non-linear regression models (Proc NLR, SPSS). The predictor variables were: the first and second order terms of (a) the interval average temperature ($T_{t,t+\tau}$), and (b) the interval average relative humidity ($RH_{t,t+\tau}$). First and second order terms were included because temperature and relative humidity usually have non-linear effects on life history of arthropods; (c) the interval average number of grains of pollen/m³ of air ($Pol_{t,t+\tau}$), when modelling the per capita growth rate of predators in spring; (d) the interval average of the abundance of prey ($Prey_{t,t+\tau}$), when modelling the per capita growth rate of predators in summer; (e) the interval average of the abundance of predators ($Pred_{t,t+\tau}$), when modelling the per capita

growth rate of prey in summer. First, predictor variables were analysed alone to test how much of the variation in $r_{t,t+\tau}$ they could explain. Next, we analysed the whole model (i.e. first and second order term of $T_{t,t+\tau}$ and $RH_{t,t+\tau}$ as well as the interaction between $T_{t,t+\tau}$ and $RH_{t,t+\tau}$, and either $Pol_{t,t+\tau}$, $Prey_{t,t+\tau}$, or $Pred_{t,t+\tau}$) with a backwards stepwise procedure. This method starts with all the variables present, and then it determines, at each step, how ‘insignificant’ the contribution is of a variable in the regression to be removed (threshold at $P = 0.05$). Subsequently, results were used to identify and model the estimated relationship between the interval per capita growth rate, $r_{t,t+\tau}$, as function of the significant predictor variables.

Warming trend in the area of study

Our experimental system included an invasive pest species that is native from Meso-america, where the climate is warmer than in South-Eastern Spain. Also, the area of our study is located in the Mediterranean region, an area classified as highly susceptible to warming and where climate change models predict increasing temperature trends in spring and summer (IPCC 2007). We analyzed whether such a trend existed in the area using monthly temperature data recorded from 1980 to 2008 in a climate station located in the IHSM La Mayora. Data were stored as monthly average, maximum and minimum temperatures. For the analysis, records of temperature were transformed into ‘seasonal temperatures’, averaging the records registered during 3 months. Seasons were defined using the standard meteorological definition, this being: Winter = December + January + February; Spring = March + April + May; Summer = June + July + August; and Autumn = September + October + November. Variables used in the analyses were the average, maximum, and minimum seasonal temperature. Linear regressions were performed, with temperature records as the dependent and time as the independent variables. Because four simultaneous tests were conducted per temperature data set, statistical significance was set at $\alpha/4 = 0.0125$ (Bonferroni method; Sokal and Rohlf 1995).

Results

Predator/prey mite community dynamics

The annual population dynamics of the perseia mite showed an exponential population increase at the beginning of summer followed by a steep decline at the end of summer (Fig. 1c, white dots). The 4-year population dynamics of the predatory mites showed two different patterns. In spring during all 4 years of the study, the phytoseiid populations showed a positive correlation with the abundance of pollen in the atmosphere (Fig. 1b, black dots and green stars, respectively). In summer there was a second phytoseiid population peak, corresponding with a numerical response to the abundance of the perseia mite in 2006 and 2007, but not in 2008 and 2009 (Fig. 1c, black and white dots, respectively). The inspection of the daily records of temperature and relative humidity during the sampling periods revealed that summers in 2008 and 2009 had, relative to those in 2006 and 2007, prolonged periods with high temperatures combined with low relative humidity (Fig. 1a, vertical distance between red and blue dots).

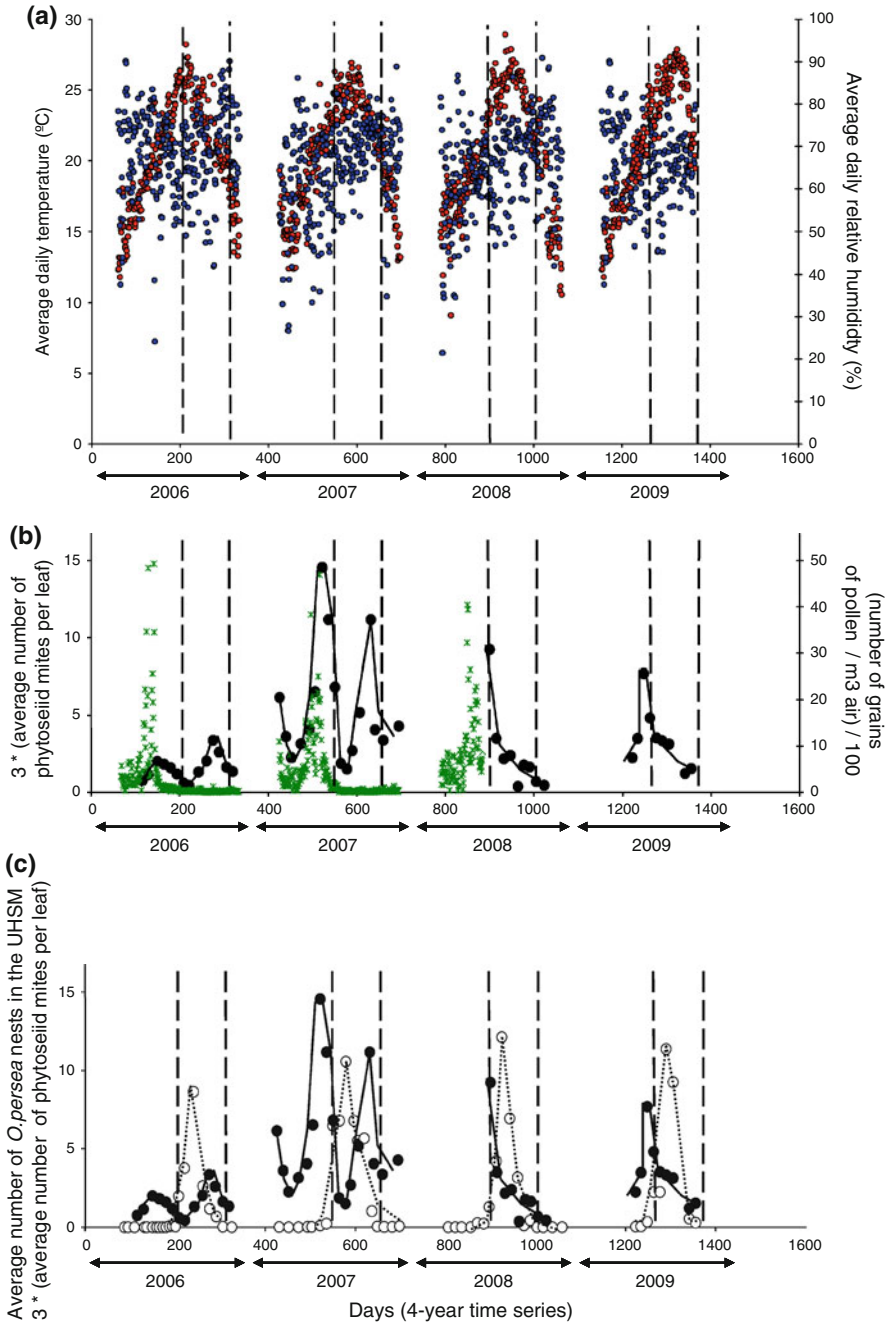


Fig. 1 **a** Daily average temperature (red dots) and relative humidity (blue dots) registered from early spring to late summer, during four consecutive years. **b** Population dynamics of phytoseiid mites (black dots) and grains of pollen in the atmosphere (green stars), during the same period as in **a**. **c** Population dynamics of the phytoseiid mites (black dots) and the persea mite (white dots), during the same period as in **a**. Vertical lines delimit the time period with presence of *Oligonychus perseae* in the avocado trees

Table 1 Results of non-linear regression models using data gathered during *spring*, when pollen in the atmosphere was present and the persea mite was rare. The dependent variable was the per capita growth rate between two successive samplings (i.e. the interval growth rate, $r_{t,t+\tau}$) of *predators*. The predictor variables were the first and second order terms of the interval average temperature ($T_{t,t+\tau}$), the first and second order terms of the interval average relative humidity ($RH_{t,t+\tau}$), and the interval average number of grains of pollen/m³ of air ($Pol_{t,t+\tau}$).

Model	R ² model	F model	P model	d.f.	Predictor variables	Parameter estimates	F	P
Only T effects	0.273	3.387	ns	2.18	Intercept	-0.406	2.168	ns
					$T_{t,t+\tau}$	0.047	2.654	ns
					$T_{t,t+\tau}^2$	-0.001	3.195	ns
Only RH effects	0.218	2.510	ns	2.18	Intercept	1.109	2.798	ns
					$RH_{t,t+\tau}$	-0.035	2.521	ns
					$RH_{t,t+\tau}^2$	0.00026	2.252	ns
Only pollen effects	0.197	4.676	0.043	1.19	Intercept	-0.023	3.793	ns
					$Pol_{t,t+\tau}$	0.010	4.676	0.043
Whole model effects	0.390	5.707	0.012	2.18	Intercept	0.054	2.483	ns
					$Pol_{t,t+\tau}$	0.009	5.134	0.036
					$T_{t,t+\tau}$	Not included		
					$T_{t,t+\tau}^2$	Not included		
					$RH_{t,t+\tau}$	Not included		
					$RH_{t,t+\tau}^2$	Not included		
					$T_{t,t+\tau} * RH_{t,t+\tau}$	-6.3×10^{-5}	5.590	0.029

Predictor variables were first analysed alone to test how much of the variation in $r_{t,t+\tau}$ they could explain, and next they were modelled together with a backwards stepwise procedure

Effect of environmental conditions on predator/prey abundance

When predictor variables were tested alone, only the interval average number of grains of pollen/m³ of air significantly explained the variation of the per capita growth rate of predators in spring (Table 1). When all the variables were included in the model, the abundance of pollen and the interaction between temperature and relative humidity were significant.

With regard to the summer data set, when predictor variables were evaluated alone only the first and second order terms of temperature had an effect on the per capita growth rate of both predator and prey (Tables 2 and 3). Figure 2 depicts the predicted interval per capita growth rate of predator and prey as function of the interval average temperature obtained from the parameter estimates of the regression analysis. Results show that the range of temperatures with positive growth rates was wider in prey than in predators (19.7–29.9 and 20.7–26.9 °C, respectively). When whole models were analysed, the first and second order terms of temperature significantly explained the variation in the interval per capita growth rate of prey and predators (Tables 2 and 3), and the interaction between $T_{t,t+\tau}$ and $RH_{t,t+\tau}$ contributed to the variation of $r_{t,t+\tau}$ in predators (Table 3). Figure 3a, b depict the predicted interval per capita growth rate of predators as function of the average interval temperature and relative humidity obtained from the parameter estimates of the regression analysis. Results show that medium temperatures ($T_{t,t+\tau} \approx 24$ °C) led to positive growth rates over a wide range of relative humidities ($RH_{t,t+\tau} > 50$ %). However, higher and lower temperatures led to positive growth rates only with increasing $RH_{t,t+\tau}$.

Table 2 Results of non-linear regression models using data gathered during *summer*, when pollen in the atmosphere was absent. The dependent variable was the per capita growth rate between two successive samplings (i.e. the interval growth rate, $r_{t,t+\tau}$) of *prey*. The predictor variables were the first and second order terms of the interval average temperature ($T_{t,t+\tau}$), the first and second order terms of the interval average relative humidity ($RH_{t,t+\tau}$), and the interval average of the abundance of predators ($Pred_{t,t+\tau}$).

Model	R ² model	F model	P model	d.f.	Predictor variables	Parameter estimates	F	P	
Only T effects	0.282	5.486	0.010	2.28	Intercept	-5.912	6.667	0.015	
					$T_{t,t+\tau}$	0.497	6.254	0.018	
					$T_{t,t+\tau}^2$	-0.010	5.804	0.023	
Only RH effects	0.178	3.030	ns	2.28	Intercept	-2.171	2.962	ns	
					$RH_{t,t+\tau}$	0.073	3.569	ns	
					$RH_{t,t+\tau}^2$	-0.0006	4.076	ns	
Only predator effects	0.035	1.042	ns	1.29	Intercept	0.001	0.001	ns	
					$Pred_{t,t+\tau}$	0.007	1.042	ns	
Whole model effects	0.282	5.486	0.009	2.28	Intercept	-5.912	6.667	0.015	
					$Pred_{t,t+\tau}$	Not included			
					$T_{t,t+\tau}$	0.497	6.254	0.018	
					$T_{t,t+\tau}^2$	-0.010	5.804	0.023	
					$RH_{t,t+\tau}$	Not included			
					$RH_{t,t+\tau}^2$	Not included			
$T_{t,t+\tau} * RH_{t,t+\tau}$	Not included								

Predictor variable were first analysed alone, to test how much of the variation in $r_{t,t+\tau}$ they could explain, and next they were modelled together with a backwards stepwise procedure

Interval average temperature values below and above 20 and 27.5 °C, respectively, resulted in negative predator growth rates independently of the relative humidity (Fig. 3a, b).

Warming trend in the area of study

Table 4 shows the results of the linear regressions carried out on climate data of each of the four seasons (winter, spring, summer and autumn) and each of the three variables (average, maximum and minimum seasonal temperature) considered, from 1980 to 2008. Results suggest that over the last 28 years temperatures of summer and spring showed an increasing trend (Fig. 4). Temperatures of winter and autumn showed no trends (Fig. 2). These results are in agreement with the predictions of the IPCC (2007).

Discussion

Can environmental conditions jeopardize herbivore population control?

Our results indicate that environmental conditions during summer, in particular high temperatures, had a negative effect on both predators and prey. However, the range of temperatures with positive interval growth rates was much wider in prey than in predators

Table 3 Results of non-linear regression models using data gathered during *summer*, when pollen in the atmosphere was absent. The dependent variable was the per capita growth rate between two successive samplings (i.e. the interval growth rate, $r_{t,t+\tau}$) of *predators*. The predictor variables were the first and second order terms of the interval average temperature ($T_{t,t+\tau}$), the first and second order terms of the interval average relative humidity ($RH_{t,t+\tau}$), and the interval average of the abundance of prey ($Prey_{t,t+\tau}$).

Model	R ² model	F model	P model	d.f.	Predictor variables	Parameter estimates	F	P	
Only T effects	0.261	4.952	0.014	2.28	Intercept	-3.227	8.838	0.006	
					$T_{t,t+\tau}$	0.276	8.612	0.006	
					$T_{t,t+\tau}^2$	-0.0058	8.325	0.007	
Only RH effects	0.114	1.805	ns	2.28	Intercept	-0.574	2.035	ns	
					$RH_{t,t+\tau}$	0.025	1.757	ns	
					$RH_{t,t+\tau}^2$	-0.0002	1.446	ns	
Only prey effects	0.064	1.989	ns	1.29	Intercept	-0.005	0.149	ns	
					$Prey_{t,t+\tau}$	0.004	1.989	ns	
Whole model effects	0.396	5.886	0.003	3.27	Intercept	-3.690	13.146	0.001	
					$Prey_{t,t+\tau}$	Not included			
					$T_{t,t+\tau}$	0.306	12.228	0.002	
					$T_{t,t+\tau}^2$	-0.0067	12.495	0.001	
					$RH_{t,t+\tau}$	Not included			
					$RH_{t,t+\tau}^2$	Not included			
				$T_{t,t+\tau} * RH_{t,t+\tau}$	0.00016	5.990	0.021		

Predictor variable were first analysed alone, to test how much of the variation in $r_{t,t+\tau}$ they could explain, and next they were modelled together with a backwards stepwise procedure

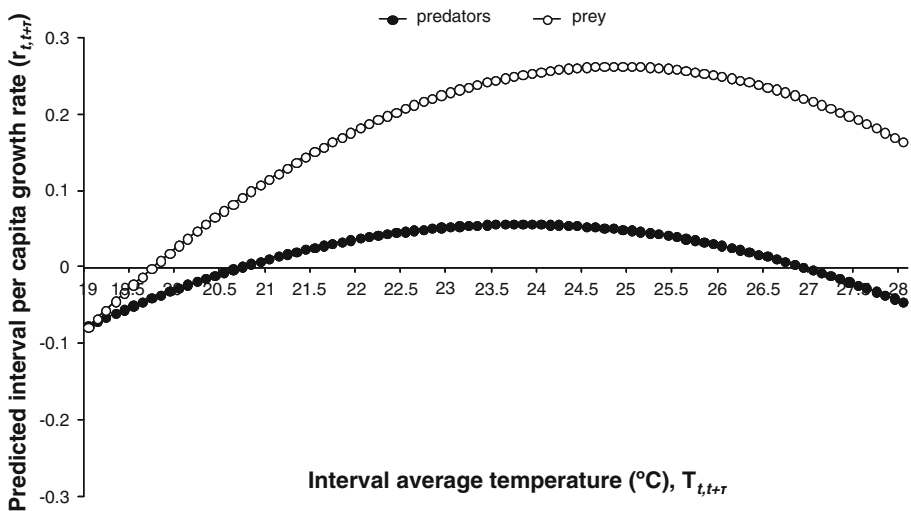


Fig. 2 Predicted interval per capita growth rate ($r_{t,t+\tau}$) of predators (*black dots*) and prey (*white dots*) as function of the interval average temperature ($T_{t,t+\tau}$). Values were obtained from the parameter estimates of the regression analysis

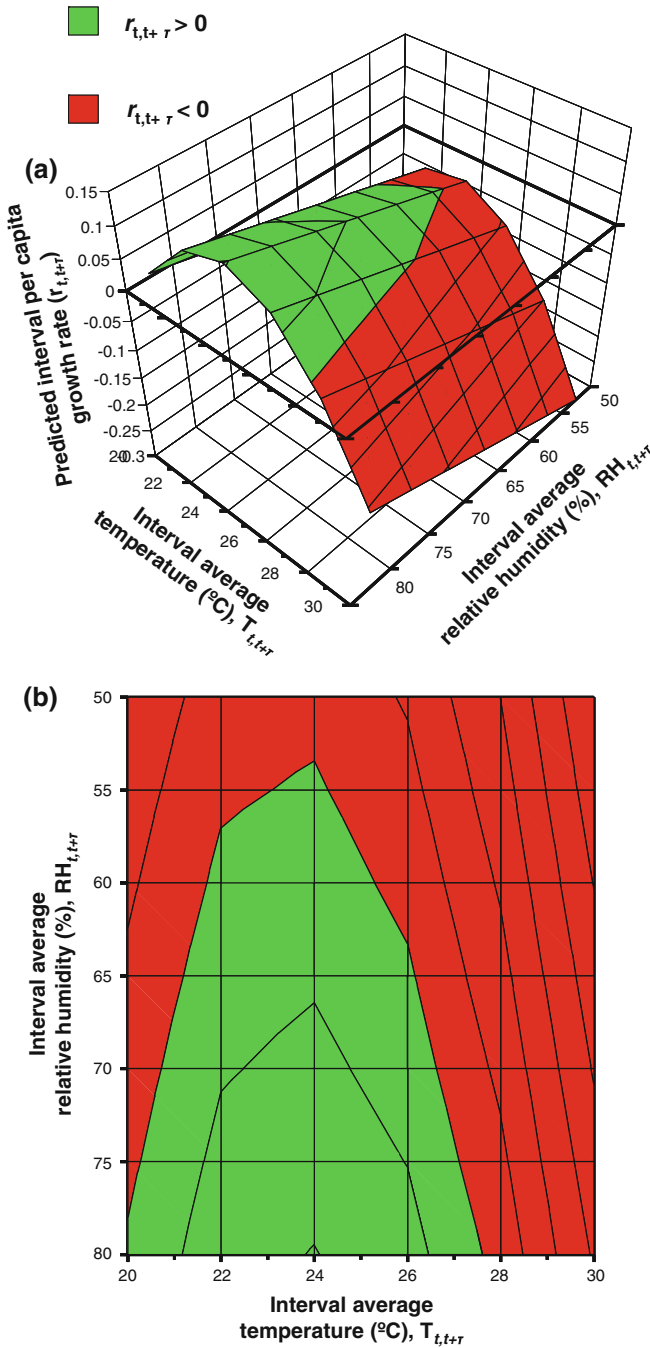


Fig. 3 **a** Three-dimensional and **b** 2-dimensional illustration of the predicted interval per capita growth rate ($r_{t,t+\tau}$) of predators as function of the interval average temperature ($T_{t,t+\tau}$) and the interval average relative humidity (RH $_{t,t+\tau}$). Values were obtained from the parameter estimates of the regression analysis. *Green* (in print: *light grey*) and *red* (*dark grey*) areas stand for positive and negative values of $r_{t,t+\tau}$, respectively

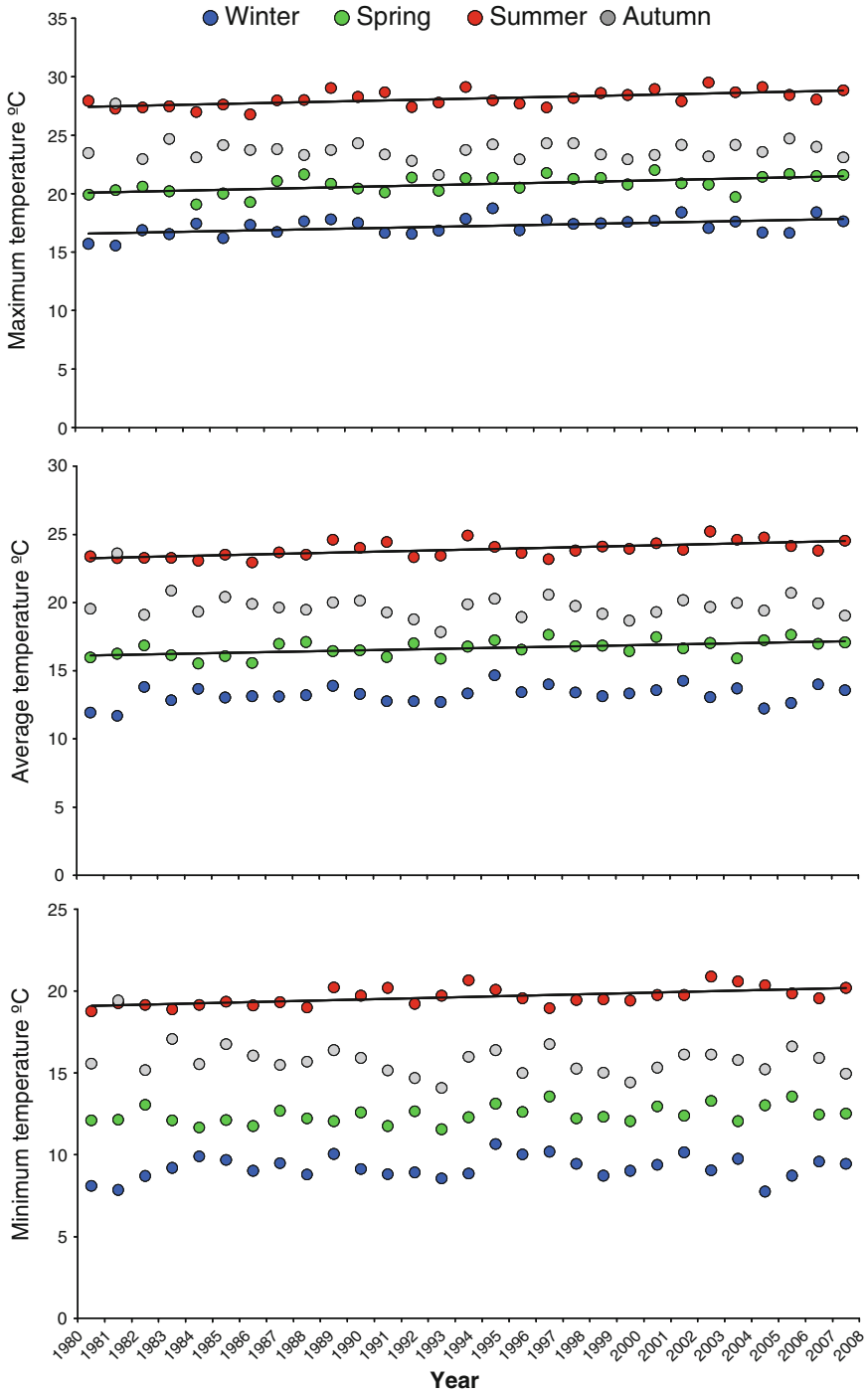


Fig. 4 Linear trends over the seasonal maximum, average and minimum temperature series from 1980 to 2008. Statistically significant regression lines are shown in the figure

Table 4 Value and sign of the slopes obtained from lineal regression performed for each of the four seasons (winter, spring, summer and autumn), and each of the three temperatures (average, maximum and minimum) registered over 28 years, from 1980 to 2008

1980–2008	Temperature	Slopes	Sign	Student's <i>t</i>	<i>P</i>
Winter	Maximum	0.045	+	3.064	<0.005
	Minimum	0.015		0.964	ns
	Average	0.0246		1.724	ns
Spring	Maximum	0.0515	+	3.6061	<0.005
	Minimum	0.0255	+	2.309	ns
	Average	0.0375	+	3.326	<0.005
Summer	Maximum	0.0495	+	3.948	<0.001
	Minimum	0.0395	+	3.851	<0.001
	Average	0.0445	+	4.201	<0.001
Autumn	Maximum	−0.020		−0.884	ns
	Minimum	−0.0355		−1.617	ns
	Average	−0.0275		−1.268	ns

Statistical significance was set at $\alpha/4 = 0.0125$. Degrees of freedom were 27 for all the analysis

(Fig. 2). Furthermore, predator positive growth rates at temperatures below or above ca. 24 °C were only possible when relative humidity was high enough (Fig. 3a, b). Temperature and relative humidity affect life-history parameters of *E. stipulatus* and *N. californicus*. Ferragut et al. (1987) showed that female survival and total reproduction of *E. stipulatus* decreased drastically between 25 and 32 °C, and egg-hatching did not occur when relative humidity was below 60 %. The authors also showed that females that emerged and mated at 32 °C laid no eggs (Ferragut et al. 1987), what could be attributed to oocytes, testes and sperm being injured by heat shock (Krebs and Loeschke 1994; Scott et al. 1997; Wang et al. 2009a, b). Walzer et al. (2007) found that egg-hatching and larval and protonymph survival of *N. californicus* decreased dramatically with decreasing relative humidity, and Nguyen and Amano (2009, 2010) found that the duration of copulation and total reproduction were drastically reduced with temperatures above 30 °C.

The effects of extreme environmental conditions in life-history parameters of the two species could explain why predatory mites did not respond numerically to the presence of the perseia mite in 2008 and 2009. On the one hand, in the avocado growing areas of southern Spain temperatures above 30 °C are common in summer. On the other hand, the summers of 2008 and 2009 were characterized by prolonged periods of high temperatures combined with the so-called 'terral' winds. This is a very dry and hot northern wind that is characteristic of the area of study during summers, due to a foehn effect caused by the mountains in the northern part of the province. Periods of high temperatures combined with low air relative humidity seem to have been more frequent and prolonged in 2008 and 2009 than in 2006 and 2007 (Fig. 1a, vertical separation between blue and red dots). Longer periods of exposure to such extreme environmental conditions may have affected individuals to an extent that translated into observable effects at the population level in the field. Indeed, our results predict that high temperatures combined with low relative humidities will lead to negative growth rates in predators. Climate data used in our analyses were taken from sensors positioned at ca. 2 m above the ground, thereby providing data of the macroclimate. Mites, however, respond to the microclimate of leaves,

which is likely to be less extreme than macroclimate. Nevertheless, macroclimate can be used as a proxy for the microclimate because they are positively correlated (Zhang et al. 2002).

Negative effects of extreme environmental conditions on life-history parameters have been described in many species of arthropod natural enemies that are being used as biological control agents (McMurtry and Scriven 1965; Zhang and Kong 1985; Ferragut et al. 1987; Bakker et al. 1993; Croft et al. 1993; Castagnoli and Simoni 1994; Broufas and Koveos 2001; DeCourcy-Williams et al. 2004; Ferrero et al. 2007, 2010; Hance et al. 2007; Walzer et al. 2007; Momen and Abdel-Khalek 2008, 2007; Vasconcelos et al. 2008; Zhang et al. 2008). Furthermore, thermal performance curves of natural enemies and their prey typically differ, with natural enemies having lower tolerance to higher temperature than prey (Roy et al. 2003; Hance et al. 2007; Stavrinides et al. 2010a, b; results here). Climate change predicts not only increases of local and regional temperatures, a trend that is observed in our area, but also increases in the frequency and duration of extreme environmental conditions. Therefore, one would expect thermal and relative humidity performance curves of natural enemies and prey to further diverge under climate change scenarios. Such divergence may be accentuated when pest species originate from warmer areas, because they are expected to be better adapted to overall higher temperatures. This is probably the case for the pest species in our study, which builds dense silken nests that likely protect them from harsh environmental conditions, as it has been shown in another related species, *Schizotetranychus longus* (Mori et al. 1999). This is certainly a scenario that should be taken into account for the future of biological control. Indeed, extreme environmental conditions can allow pests to escape control by predators, and may be a direct cause of biological control disruption in some agricultural systems (Roy et al. 2003; Stavrinides et al. 2010a, b).

Future directions

Several authors have stated that species interactions should be incorporated into climate-change models to better predict the impact of climate change on natural and agro-ecosystem communities (Wilmers et al. 2007; Tylianakis et al. 2008; Gilman et al. 2010). Certainly, that would be an appropriate approach, because behavioural features of individuals are directly related to environmental conditions, and species interactions are commonly climate-dependent (Beveridge et al. 2010; Gilman et al. 2010). A first step to improve predictions should be to determine how environmental conditions alter the interactions among the species present in a given agricultural community. Environmental conditions can alter foraging behaviour or dispersal capabilities (Skirvin and Fenlon 2003a, b), and affect competitive abilities and predatory capabilities of species (Tylianakis et al. 2008). Knowledge on which interactions between species of the agricultural community are most sensitive to changes in environmental conditions would help to increase the probability of success of biological control strategies. On the one hand, it would provide information on which species of natural enemies are likely to persist during periods of adverse environmental conditions. Subsequently, pest control strategies could focus on methods aimed at improving their establishment, or at artificially increasing their populations. On the other hand, knowledge on the thermal and relative humidity performance curves of natural enemies, relative to those of their prey, could help to optimize predator use. Indeed, we could know when the establishment and the persistence of released natural enemies is likely to fail, and the environmental values against which measures to soften local environmental conditions should be applied.

In this study we showed that local environmental conditions can influence the abundance and population dynamics of pests and their natural enemies. Therefore, they can be a direct cause of biological control disruption. Because climate change models predict an increase of the frequency and intensity of extreme events, such as heat waves, it becomes necessary (a) to understand the relationship between the effect of warming on higher trophic levels and herbivore population outbreaks; (b) to intensify preventive methods, especially those involving control of imported plant material; and (c) to encourage scientific collaboration with research teams of countries where potential future pests are from, in order to develop and establish proper control.

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