



# Synthesis and evaluation of tannin-thiosemicarbazide-formaldehyde resin for selective adsorption of silver ions from aqueous solutions

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## Abstract

To anchor chelating groups with nitrogen and sulfur atoms on tannin, a novel adsorbent (tannin-thiosemicarbazide-formaldehyde resin) was prepared through Mannich reaction by using blank wattle tannin, thiosemicarbazide and formaldehyde as raw materials. The adsorption behaviour of  $\text{Ag}^+$  from aqueous solution on the resin was evaluated via batch adsorption experiments. Fourier transform infrared spectroscopy (FT-IR) and elemental analysis were applied to verify the successful immobilization. The surface morphology, thermal stability and pore structure of the resin were also characterized. The results showed that the adsorption isotherm of  $\text{Ag}^+$  for the resin was described well by the Freundlich model.  $\text{Ag}^+$  adsorption equilibrium was achieved within 180 min, and the kinetic data were better fitted by the pseudo-second-order kinetic equation than by the pseudo-first-order and intraparticle diffusion equations. The adsorption capacity first increased and then stabilized with increasing pH (ranging from 1.0 to 7.0), and the resin exhibited high selectivity towards  $\text{Ag}^+$  in relation to  $\text{Pb}^{2+}$ ,  $\text{Cd}^{2+}$ ,  $\text{Ni}^{2+}$  and  $\text{Ca}^{2+}$ . After three regeneration and reuse cycles, the adsorption capacity reached 1.68 mmol/g (84.0% of removal efficiency). Based on the experimental results and findings from various characterization techniques, the mechanism of  $\text{Ag}^+$  adsorption onto the resin could be attributed to inner-sphere complexation and chelation between  $\text{Ag}^+$  and multiple electron-rich atoms (N, O, and S), in which S atoms played the most important role.

**Keywords** Tannin · Adsorption · Silver · Thiosemicarbazide · Mechanism

## 1 Introduction

Silver is a precious metal as well as a heavy metal and has been widely adopted in the electroplating, automotive, electronics, aerospace, shipbuilding, decoration, and other industries due to its excellent gloss, ductility, conductance and thermal conductivity [1, 2]. As a precious metal, silver is a nonrenewable resource, and stored silver is very rare on Earth. According to the conditions in the solar energy industry, it is expected that silver shortages will begin in 2030

with the depletion of Ag resources [3]. As a heavy metal, silver ions are very difficult to biodegrade, and can accumulate hundreds of times under the biological amplification in the food chain, posing potential risks to microorganisms, plants and humans, thereby causing damage to and impacting ecosystems [4–7].

From environmental protection, human health and sustainable resource development perspectives, there is an urgent need to effectively extract and recover  $\text{Ag}^+$  from industrial wastewater. Numerous techniques have been employed for separating  $\text{Ag}^+$  from wastewater, such as hydrometallurgy [8], electrochemical deposition/reduction [9–11], chemical precipitation [12], solvent extraction [13], ion exchange [14], membrane filtration [15, 16], and adsorption [17–19]. Among these techniques, adsorption is commonly applied and studied due to its the advantages of notable simplicity, adaptability and effectiveness and is regarded as the most attractive approach for removing contaminants from wastewater [20].

In recent years, chemically modified adsorbents from various agricultural byproducts used as starting materials

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have indicated enormous potential for selective adsorption of precious metals owing to their widespread availability, cost-effectiveness, biodegradability, sustainable utilization and ecological adaptability [2, 19, 21–23]. Tannins are the most abundant source of natural aromatic compounds extracted from leaves, roots, bark, seeds and wood [24]. Due to the presence of many adjacent phenolic hydroxyl groups in their molecules, tannins are considered excellent candidates for manufacturing biosorbents for removing heavy metals and recovering precious metals [25, 26]. However, tannins are water-soluble compounds that undergo a process of immobilization/insolubility before being used as adsorbents. Insolubility of tannins can be achieved through cross-linking reactions with formaldehyde [27, 28], and their immobilization can be achieved by modifying water-insoluble matrices with tannins [29–36].

According to the hard–soft acid–base (HSAB) principle [37], the phenolic hydroxyl groups in tannin molecules are hard bases, so the affinity of tannin-based bioadsorbents for soft acids such as precious metals is expected to be low. To improve the affinity of tannin-based bioadsorbents, further chemical modifications have been implemented. Gurung et al. modified persimmon tannin gel with tetraethylenepentamine, bithiourea and N-aminoguanidine to adsorb  $\text{Au}^{3+}$ ,  $\text{Pd}^{2+}$  and  $\text{Pt}^{4+}$  [38–40]. Xiong et al. modified tannin gel with diethylamine for adsorbing precious metals from simulated wastewater [41]. Although these tannin based adsorbents containing chelating groups with N and S atoms exhibit a remarkable adsorption performance for precious metals, there are also certain drawbacks in the preparation process, such as solvent toxicity, notable time consumption, low grafting efficiency and complex synthesis processes. For example, the anchoring of tetraethylenepentamine onto persimmon tannins involves a three-step modification process. First, tannins are gelled with formaldehyde in basic media for 12 h. Then, the obtained tannin gel is subjected to an intermediate chloromethylation step in concentrated hydrochloric acid for 12 h. Finally, under sodium carbonate catalysis, the chloromethylation product reacts with tetraethylenepentamine in a heterogeneous system of dimethylformamide for 24 h [38]. Therefore, it is necessary to design and develop new approaches for anchoring chelating groups with N and S atoms onto tannins.

In this study, a tannin-thiosemicarbazide-formaldehyde resin (BTC) resin containing N and S atoms was synthesized rapidly through one-step Mannich reaction in an aqueous solution and applied for  $\text{Ag}^+$  recovery. Characterizations were performed to validate the reaction and elucidate the characteristics of the synthesized resin. To achieve the desirable adsorption capacity for  $\text{Ag}^+$ , the influences of the pH, contact time, initial concentration of  $\text{Ag}^+$  and ionic strength were investigated in detail by batch experiments. The adsorption kinetics and isotherms were also studied

through different models. The selectivity of BTC resin for  $\text{Ag}^+$  adsorption was also explored in solutions containing multiple metal ions. Finally, the mechanism of  $\text{Ag}^+$  adsorption onto the resin was determined by various characterization instruments.

## 2 Experimental methods

### 2.1 Materials and chemicals

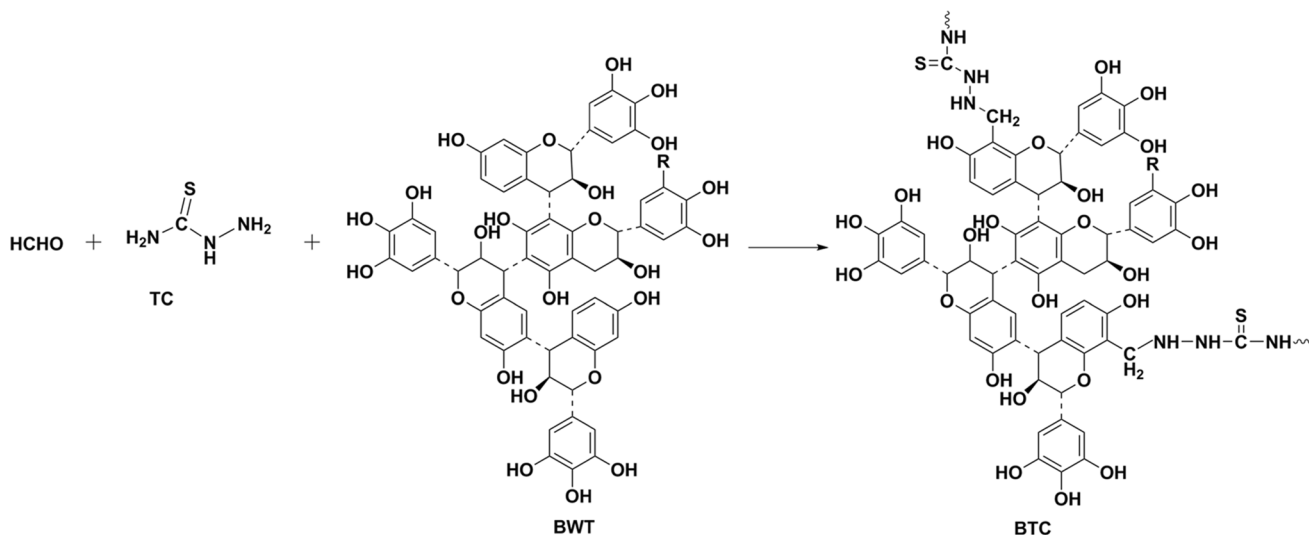
Blank wattle tannin (BWT) was obtained from Guangxi Wuming tannin extract factory, China. Thiosemicarbazide (TC), 0.1 mol/L of  $\text{AgNO}_3$  solution and 37 wt% of formaldehyde solution were purchased from Shanghai Titan Technology Co., Ltd., China. All other chemicals were of analytical grade without further purification.

### 2.2 Synthesis of tannin-thiosemicarbazide-formaldehyde resin

Tannin-thiosemicarbazide-formaldehyde (BTC) resin was prepared through the Mannich reaction and the proposed reaction is shown in Fig. 1. Briefly, 3 g of blank wattle tannin (BWT) was dissolved in 50 mL of pure water, followed by the addition of 2.0 g of thiosemicarbazide (TC) to the solution. After the pH was adjusted to the desired value, the temperature of the reaction was increased to 80 °C, and 2.5 mL of formaldehyde (37 wt%) was added to the flask within 30 min. The reaction was performed for 5 h. Subsequently, the sample was separated from the mixture by centrifugation and washed three times with deionized water and anhydrous ethanol. Afterwards, the obtained product was dried under vacuum at 40 °C for 24 h.

### 2.3 Characterization

Fourier transform infrared spectroscopy (FT–IR) of the samples was performed using an FT–IR spectrometer (FT–IR, Nicolet 6700). Thermogravimetric analysis (TGA) was performed with a TGA instrument (Netzsch, STA 449C) within the temperature range of 30–800 °C at a rate of 10 °C/min under the protection of nitrogen. The C, H, N and S contents in BWT and BTC were determined with an elemental analyser (EA, Vario Macro Cube, Elementar). The specific surface area and pore size were measured on a specific surface area tester (ASAP2020) through  $\text{N}_2$  adsorption–desorption isotherms. The surface morphology and elemental distribution of the samples were observed by scanning electron microscopy and energy dispersive spectroscopy (SEM–EDS, FEI QUANTA 200). X-ray photoelectron spectroscopy (XPS, EscaLab 250Xi) was employed to measure the surface



**Fig.1** Synthesis route for the preparation of BTC sorbent

state and electronic structure of the BTC resin before and after Ag<sup>+</sup> adsorption, and the obtained spectra were fitted using XPSPEAK4.1 software. X-ray diffraction (XRD) patterns of the BTC resin before and after Ag<sup>+</sup> adsorption were recorded by an XRD instrument (D8-Advance). Zeta potential values were obtained with a nanoparticle size and zeta potential analyser (Litesizer 500).

**2.4 Adsorption experiments**

Batch adsorption experiments were performed to evaluate the adsorption behaviour of Ag<sup>+</sup> onto the BTC resin. In typical experiments, 50 mg of the BTC resin was mixed with 50 mL of silver ion solution and continuously shaken at 150 rpm and 25 °C. The effects of the pH, initial Ag<sup>+</sup> concentration, contact time, coexisting metal ions and ionic strength were investigated by varying one variable while maintaining all other parameters fixed. The influence of the pH was studied by adjusting the solution pH by either HNO<sub>3</sub> or NaOH to between 1.0 and 7.0. The mixture with a Ag<sup>+</sup> concentration of 2 mmol/L of was shaken in a thermostatic shaker for 24 h at 25 °C. Adsorption isotherm experiments were conducted with Ag<sup>+</sup> concentrations ranging from 1 to 6 mmol/L at pH 5.0 for 24 h to reach adsorption equilibrium. In the kinetics study, samples were withdrawn at periodic time intervals from the reaction mixture with 4 mmol/L Ag<sup>+</sup>. The effect of coexisting metal ions was investigated by mixing 2 mmol/L Pb<sup>2+</sup>, Ni<sup>2+</sup>, Cd<sup>2+</sup> and Ca<sup>2+</sup> with 2 mmol/L Ag<sup>+</sup> at pH 2. The effect of the ionic strength was evaluated by adding solutions with different concentrations of NaNO<sub>3</sub> ranging from 0 to 0.5 mol/L, with 4 mmol/L Ag<sup>+</sup>. The residual Ag<sup>+</sup> concentration after adsorption was determined by using an atomic absorption spectrophotometer (WFX 200). The adsorption

capacity (*q<sub>t</sub>*) of Ag<sup>+</sup> on the BTC resin can be calculated according to Eq. (1):

$$q_t = (C_0 - C_t)V/m \tag{1}$$

where *C*<sub>0</sub> and *C*<sub>*t*</sub> (mmol/L) are the initial concentration and concentration of the adsorbate after time *t*, respectively, *q<sub>t</sub>* (mmol/g) is adsorption capacity after time *t*, *m* (g) is the weight of the BTC used and *V* (L) is the volume of the solution.

**3 Results and discussion**

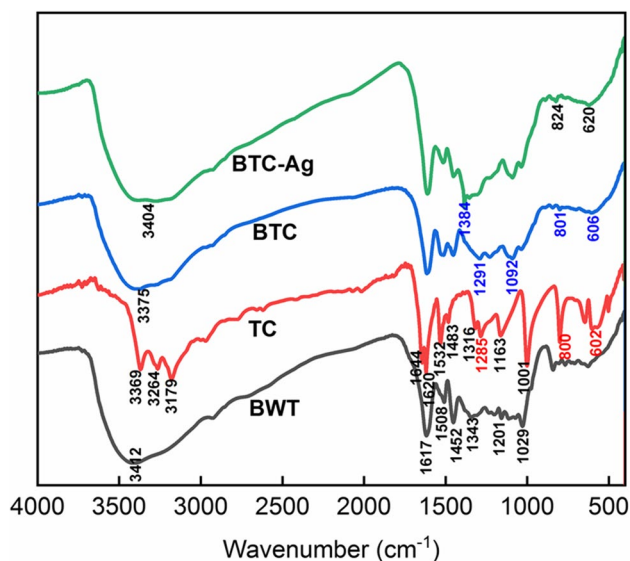
**3.1 Structure characterization**

Table 1 provides the elemental analysis results for C, N, O and S in BWT and BTC. The N and S contents in BTC were 10.65% and 7.282% respectively, which far exceeded those in BWT. In addition, the C value of BTC decreased after chemical modification due to the lower C content in TC. These findings confirmed that TC was successfully attached to BWT.

Figure 2 shows the FT-IR spectra of BWT, TC and BTC. In the spectrum of BWT, the absorption band at 3412

**Table 1** Elemental analysis of BWT and BTC

Sample	Element content (%)			
	C	H	N	S
BWT	49.19	4.938	0.36	0.718
BTC	44.88	4.934	10.65	7.282



**Fig.2** FT-IR spectra of BWT, TC, BTC and BTC-Ag ( $\text{Ag}^+$  adsorbed by BTC)

$\text{cm}^{-1}$  corresponded to O–H stretching vibrations, the peaks at 1617, 1508 and 1452  $\text{cm}^{-1}$  represented characteristic adsorption bands of the benzene ring; the peaks at 1343  $\text{cm}^{-1}$  were related to O–H bending vibrations, and the peaks at 1201 and 1029  $\text{cm}^{-1}$  could be attributed to the C=C–O stretching vibrations and C–O–C stretching vibrations, respectively [42, 43]. Regarding TC, the absorption peaks at 3369, 3264 and 3179  $\text{cm}^{-1}$  corresponded to N–H stretching vibrations, and those at 1644 and 1620  $\text{cm}^{-1}$  corresponded to  $\text{NH}_2$  bending vibrations. The stretching vibration bands of N–C=S(I), N–C=S(II) and N–C=S(III) were observed at 1532 and 1483  $\text{cm}^{-1}$ , 1316, 1285 and 1163  $\text{cm}^{-1}$ , and 1001  $\text{cm}^{-1}$ , respectively [44, 45]. The peaks at 800 and 600  $\text{cm}^{-1}$  could be attributed to the out-of-plane vibrations and wagging vibrations of N–H. After chemical modification, the spectral structures of BTC and BWT were similar but exhibited a slight shift in wavenumbers, which demonstrates that the core structure of BWT remained unchanged.

In addition, the spectrum of BTC contained new peaks at approximately 1291 and 1092  $\text{cm}^{-1}$ , which were assigned to the  $-(\text{C}=\text{S})-\text{N}-$  moiety of TC [46, 47]. The N–H out-of-plane vibration and wagging vibration peaks also appeared at 801 and 606  $\text{cm}^{-1}$ , respectively [48]. These observations revealed that the BTC adsorbent was successfully prepared.

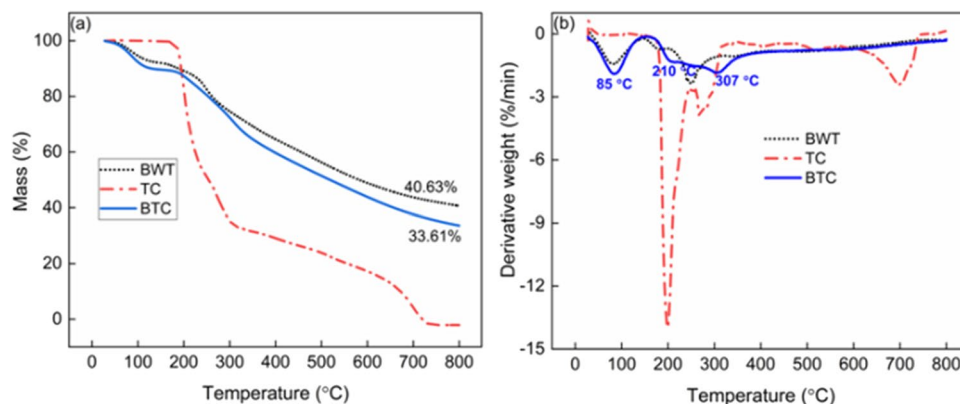
The thermal stabilities of BWT, TC and BTC were compared, and the results are shown in Fig. 3. BTC showed three major steps of transient weight loss at temperatures of (1) 85, (2) 210 and (3) 307  $^{\circ}\text{C}$ , corresponding to (1) moisture evaporation, (2) decomposition of the TC moiety on BTC and (3) BWT degradation, respectively [49]. Comparatively, the residue weight of BWT was greater than that of BTC. These results indicated that TC was successfully anchored on BWT.

The surface morphology of the prepared BTC adsorbent was characterized by SEM, and the results are shown in Fig. 4(a). The BTC adsorbent exhibited a spherical porous structure with a rough surface and uneven size. The pore structure and morphology are crucial parameters for an adsorbent to adsorb metal ions [50, 51], so a Brunauer–Emmett–Teller (BET) test was conducted, and the nitrogen adsorption–desorption curve is shown in Fig. 4(b). The BTC curve is a typical type IV curve, as defined by the International Union of Pure and Applied Chemistry (IUPAC) [52]. The BET surface area, total pore volume and average pore size were 67.34  $\text{m}^2/\text{g}$ , 8.069  $\text{nm}$  and 0.1358  $\text{cm}^3/\text{g}$ , respectively, suggesting that the BTC resin was not a totally porous adsorbent.

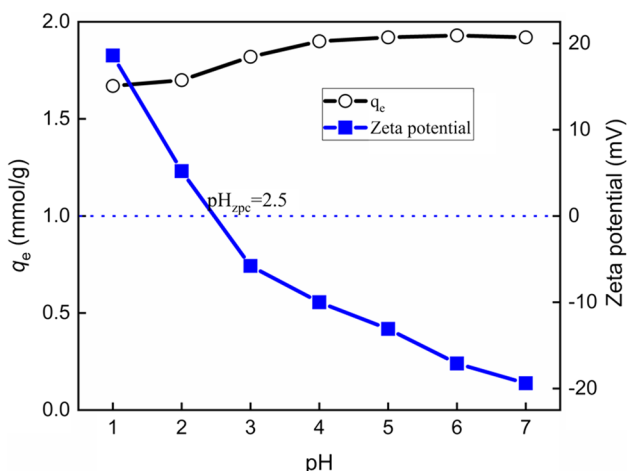
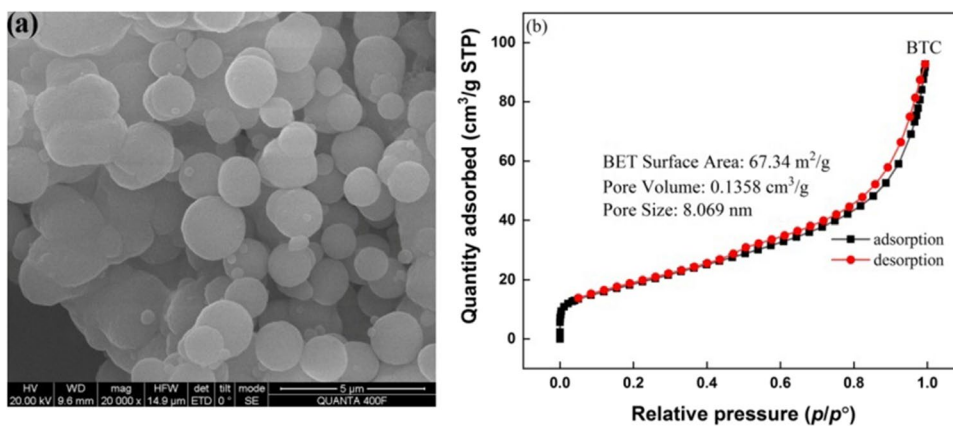
### 3.2 Influence of the pH on $\text{Ag}^+$ adsorption onto the BTC resin

It is widely acknowledged that the pH is a main parameter for the metal ion adsorption, because it not only affects the metal species but also influences the surface charge of the adsorbent [53]. The influences of the pH on the adsorption capacity and zeta potential of BTC were investigated at 2  $\text{mmol/L}$  of  $\text{AgNO}_3$  for 24 h. The pH was adjusted

**Fig.3** (a) TGA and (b) DTG curves of BWT, TC and BTC



**Fig.4** (a) SEM and (b) nitrogen adsorption–desorption curve of BTC



**Fig.5** Influence of pH on adsorption capacity ( $C_0$ : 2 mmol/L of  $Ag^+$ ; sorbent dosage: 1 g/L; T:  $25 \pm 1$  °C; time: 24 h) and zeta potential of BTC

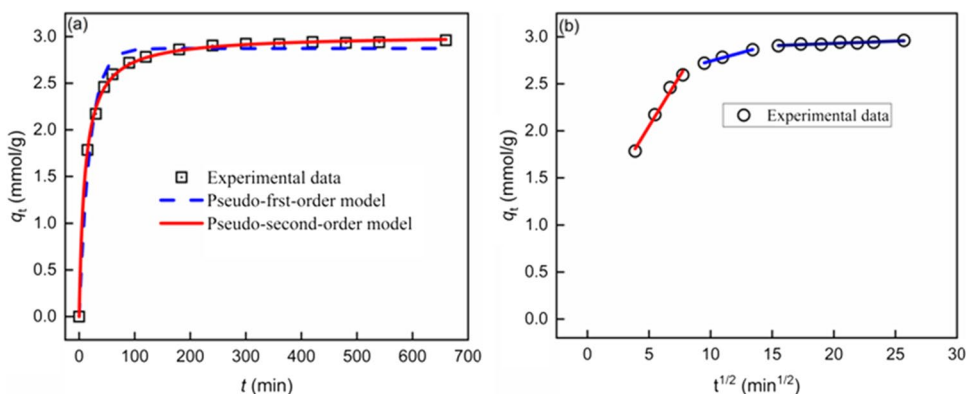
over the range of 1.0 to 7.0 to avoid the possibility of silver precipitation [54], and the results are shown in Fig. 5. The adsorption capacity for  $Ag^+$  first increased and then stabilized with increasing pH, and the maximum removal

rate was greater 95% at pH 4–7. The  $pH_{ZPC}$  of BTC was measured at 2.5, indicating that the surface of BTC was positively charged at pH values less than 2.5 due to protonation of the functional groups. The positively charged BTC resin was unfavourable for the adsorption of cationic silver because of electrostatic repulsion. When the pH was increased from 3.0 to 4.0, the functional groups on the surface of BTC dissociated, and the surface was mainly negatively charged, which was promoted  $Ag^+$  adsorption, thus increasing the adsorption capacity. However, with further increasing pH, the  $Ag^+$  adsorption capacity changed nonsignificantly even if the zeta potential continued to decrease.

### 3.3 Influences of the contact time and adsorption kinetics

The influence of the contact time on the  $Ag^+$  adsorption capacity is shown in Fig. 6(a). The adsorption capacity increased rapidly during the first 60 min, then slowed and reached equilibrium at 180 min. Pseudo-first-order (Eq. 2), pseudo-second-order (Eq. 3) and intraparticle diffusion models (Eq. 4) [35] were employed to fit the experimental data.

**Fig.6** (a) Influence of contact time on adsorption capacity ( $C_0$ : 4 mmol/L of  $Ag^+$ ; sorbent dosage: 1 g/L; T:  $25 \pm 1$  °C) and fitting by Pseudo-first-order and Pseudo-second-order models; (b) fitting by Intraparticle diffusion model



$$q_t = q_e(1 - e^{-k_1 t}) \quad (2)$$

$$q_t = \frac{q_e^2 k_2 t}{1 + q_e k_2 t} \quad (3)$$

$$q_t = k_3 \sqrt{t} + C \quad (4)$$

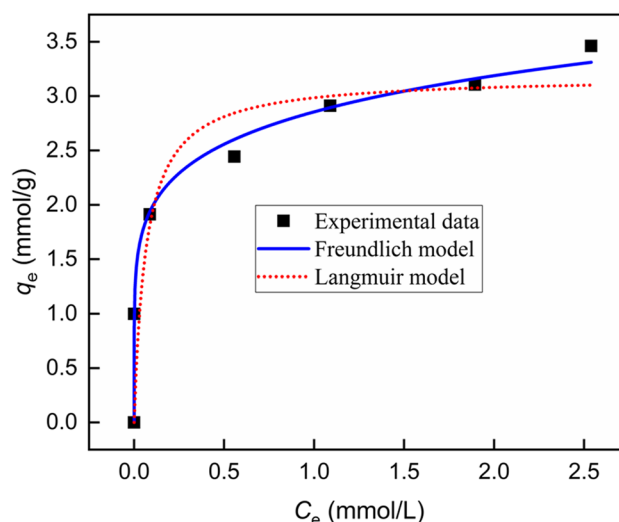
where  $q_e$  and  $q_t$  (mmol/g) are the amount of solute adsorbed at equilibrium and at any time  $t$  (min) per unit weight of adsorbent respectively,  $k_1$  ( $\text{min}^{-1}$ ) is the pseudo-first-order rate constant,  $k_2$  (mmol/g min) is the pseudo-second-order rate constant and  $k_3$  (mmol/g  $\text{min}^{0.5}$ ) is the intraparticle diffusion rate constant.

The fitting results and calculated kinetics parameters are shown in Fig. 6(a) and Table 2, respectively. In contrast to the pseudo-first-order model, which overestimated the adsorption capacity for  $\text{Ag}^+$  at the initial stage and underestimated it at the equilibrium stage, the pseudo-second order model was more suitable for describing the kinetics of  $\text{Ag}^+$  adsorption onto BTC because of the greater correlation coefficient and closer equilibrium adsorption capacity between the fitting and experimental values, revealing that chemisorption was the rate-controlling step.

The adsorption kinetic data were also described by an intraparticle diffusion model, and the adsorption process was separated into three steps, as shown in Fig. 6(b). The first initial fast step corresponded to the external surface adsorption within the initial 60 min, the subsequent slow step represented intraparticle diffusion during the adsorption time ranging from 60 to 180 min, and the final equilibrium step after 240 min reflected equilibrium adsorption. Obviously, the fitting result of the first linear section deviated from the origin, indicating that  $\text{Ag}^+$  adsorption on BTC was not completely governed by intraparticle diffusion [35, 55].

### 3.4 Influences of the initial $\text{Ag}^+$ concentration and adsorption isotherm

To evaluate the maximum adsorption capacity, the change in the adsorption capacity with different initial  $\text{Ag}^+$  concentrations was investigated. As shown in Fig. 7, the amount of



**Fig. 7** (a) Influence of initial  $\text{Ag}^+$  concentration ( $C_0$ : 1–6 mmol/L of  $\text{Ag}^+$ ;  $\text{pH}_0$ : 5.0; sorbent dosage: 1 g/L;  $T$ :  $25 \pm 1$  °C; time: 24 h) and fitting by Langmuir and Freundlich models

$\text{Ag}^+$  adsorbed increased with increasing initial  $\text{Ag}^+$  concentration. The Langmuir (Eq. 5) and Freundlich (Eq. 6) models [35] were applied to describe the experimental data of  $\text{Ag}^+$  adsorption onto the BTC resin.

$$q_e = \frac{q_{m,L} k_L C_e}{1 + k_L C_e} \quad (5)$$

$$q_e = k_F C_e^{1/n} \quad (6)$$

where  $q_m$  and  $q_e$  (mmol/g) are the maximum adsorption capacity and equilibrium amount of  $\text{Ag}^+$  adsorbed per unit weight of the BTC, respectively;  $C_e$  (mmol/L) is the final equilibrium concentration of  $\text{Ag}^+$ ;  $k_L$  (L/mmol) is the Langmuir constant;  $k_F$  (mmol/(g (L/mmol) $^{1/n}$ )) and  $n$  represent the Freundlich constants.

The fitting curves and calculated parameters obtained from these two models are shown in Fig. 7 and Table 3. Notably, the Freundlich model is widely used to indicate that the adsorption occurs on heterogeneous surfaces, while the Langmuir model is based on the assumption that adsorption occurs on homogeneous monolayer surfaces with identical adsorption. As indicated in Table 3, the Freundlich model exhibited a greater correlation coefficient than the Langmuir

**Table 2** Kinetics parameters for the  $\text{Ag}^+$  adsorption on BTC

Kinetic models	$q_{e,cal}$	Constant	$R^2$
Pseudo-first-order	2.8375	$k_1=0.05145$	0.981
Pseudo-second-order	3.0177	$k_2=0.03146$	0.999
Intraparticle diffusion	—	$k_{3(1)}=0.9815$ $C_1=0.2138$	0.982
		$k_{3(2)}=2.3822$ $C_2=0.0360$	0.986
		$k_{3(3)}=2.8323$ $C_3=0.0049$	0.848

**Table 3** Parameters calculated from Freundlich and Langmuir models for  $\text{Ag}^+$  adsorption on the BTC

Langmuir model			Freundlich model		
$Q_{max}$	$k_L$	$R^2$	$k_F$	$n$	$R^2$
3.1825	15.1236	0.839	2.8553	6.289	0.992

model, indicating that the BTC surface is heterogeneous and serves as another factor that favours  $Ag^+$  adsorption. In addition, the value of  $n$  was greater than 1, indicating that the adsorption process easily proceeded in aqueous medium.

To illustrate the excellent performance of BTC, the maximum adsorption capacity ( $Q_{max}$ ) values in Table 4 were compared to those of other sorbents reported in the literatures. According to the  $Q_{max}$  values, BTC can be considered an effective sorbent compared with other sorbents for  $Ag^+$  removal.

The standard Gibbs energy change ( $\Delta G^0$ ) of  $Ag^+$  adsorption onto BTC was determined by using Eq. 7 and Eq. 8 [60].

$$k_c = \frac{k_F \rho}{1000} \left( \frac{10^6}{\rho} \right)^{\left(1 - \frac{1}{n}\right)} \tag{7}$$

$$\Delta G^0 = -RT \ln k_c \tag{8}$$

where  $k_c$  is the equilibrium constant;  $k_F$  (mmol/(g (L/mmol)<sup>1/n</sup>)) represents the Freundlich parameter;  $n$  is an empirical constant of Freundlich model;  $R$  is the universal gas constant (8.314 J/mol K),  $T$  is the temperature (K) and  $\rho$  is the density of pure water (1.0 g/mL).

The negative Gibbs energy change (-14.24 kJ/mol) at 298 K represents the thermodynamic feasibility and spontaneity of  $Ag^+$  adsorption onto BTC.

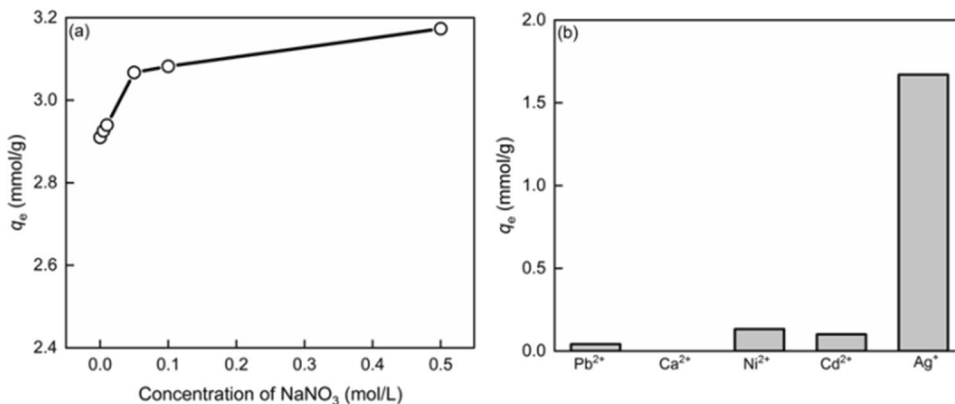
### 3.5 Influence of the ionic strength on the adsorption capacity

The effect of the ionic strength on  $Ag^+$  adsorption was investigated with  $NaNO_3$  concentrations ranging from 0.005 to 0.5 mol/L. As shown in Fig. 8(a), the adsorption capacity for  $Ag^+$  gradually increased with increasing  $NaNO_3$  concentration. Generally, if an outer-sphere surface complex is formed between the adsorbent and adsorbate by ion exchange or electrostatic forces, the adsorption capacity decreases with increasing ionic strength, while for inner-sphere surface complexes, the adsorption capacity increases with increasing ionic strength or is independent of the ionic strength [61, 62]. Therefore, the observed positive influence of the ionic strength suggested that  $Ag^+$  adsorption onto BTC entailed an inner-sphere adsorption mechanism through a coordinate-covalent bond between the ligand and  $Ag^+$ , which is consistent with earlier research of Dong et al. [53].

**Table 4** Comparison of the adsorption capacity with the various sorbent

Sorbent	$Q_{max}$ (mmol/g)	Equilibrium time (min)	Ref
Melamine- formaldehyde- thiourea resin	0.951	65	[18]
Multi-thiolated cage mesoporous monoliths	1.67	20	[56]
Superparamagnetic carbon	0.57	60	[57]
2-Mercaptobenzimidazole magnetic chitosan	2.02	1920	[23]
polymer functionalized by 3-amino-1,2,4-triazole-5-thiol and glutaraldehyde	4.16	400	[58]
$Ag^+$ imprinted chitosan gel beads	0.83	2880	[59]
BTC	3.18	180	This work

**Fig.8** (a) Effect of ionic strength on  $Ag^+$  adsorption onto BTC ( $C_0$ : 2 mmol/L of  $Ag^+$ ;  $pH_0$ : 2.0; sorbent dosage: 1 g/L;  $T$ :  $25 \pm 1$  °C; time: 24 h); (b) Selective adsorption of BTC for  $Ag^+$  ( $C_0$ : 4 mmol/L of  $Ag^+$ ; sorbent dosage: 1 g/L;  $T$ :  $25 \pm$  °C; time: 24 h)



### 3.6 Adsorption selectivity

Considering the complexity of the real aquatic environment, the selectivity of the BTC resin for  $\text{Ag}^+$  adsorption was evaluated in a mixed solution containing  $\text{Ag}^+$ ,  $\text{Pb}^{2+}$ ,  $\text{Ca}^{2+}$ ,  $\text{Ni}^{2+}$ , and  $\text{Cd}^{2+}$  at the same ratios. Figure 8(b) shows that the  $\text{Ag}^+$  removal rate of the BTC resin exceeded 95%, while the removal rates for the other metal ions were less than 10%, revealing that BTC exhibited high selectivity for the  $\text{Ag}^+$  adsorption from the mixed solution.

### 3.7 Regeneration and reuse study of the BTC

To regenerate and reuse BTC, a desorption experiment was conducted by immersing BTC-Ag in a 0.5 mol/L thio-urea–0.1 mol/L  $\text{HNO}_3$  mixed solution for 1 h and washing with ultrapure water for five times [45, 48]. The regenerated BTC was reused during the next cycle. After three regeneration and reuse cycles, the adsorption capacities were 1.91, 1.82 and 1.68 mmol/g, corresponding to removal efficiencies of 95.5%, 91.1% and 84.1%, respectively. Therefore, the BTC adsorbent possessed favourable stability and reusability for  $\text{Ag}^+$  recovery from wastewater.

### 3.8 Adsorption mechanism

The variations in the surface morphology and element distribution on the BTC resin after  $\text{Ag}^+$  uptake were characterized by SEM–EDS. As shown in Fig. 9(a), the spherical morphology of BTC remained unchanged after  $\text{Ag}^+$  adsorption.

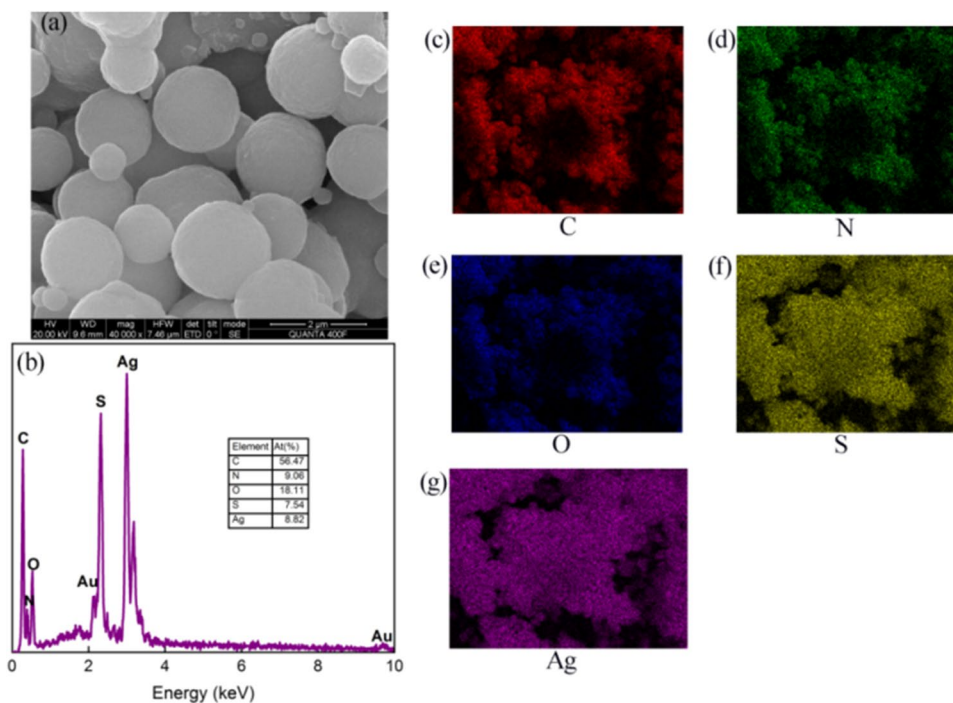
The elements on BTC after  $\text{Ag}^+$  uptake were examined by EDS (Fig. 9(b)), and the EDS spectrum revealed that the main constituents were C, O, N, S and Ag, demonstrating successful  $\text{Ag}^+$  adsorption onto the BTC surface. The corresponding atomic percentages were 56.47%, 18.11, 9.06%, 7.54% and 8.82%, respectively. The atomic percentage ratio of S to Ag was less than 1, indicating that other functional groups participate in  $\text{Ag}^+$  adsorption.

The distributions of the C, O, N, S and Ag on the BTC-Ag surface were also explored and are shown in Fig. 9(c–g). Elemental Ag exhibited consistency with S, providing evidence that S atoms on the BTC played an important role in  $\text{Ag}^+$  adsorption. According to the soft-hard acid–base (SHAB) theory,  $\text{Ag}^+$  (a soft metal ion) exhibits a greater affinity for hard bases with donor atoms in the order of  $\text{O} < \text{N} < \text{S}$  [37, 46, 63, 64].

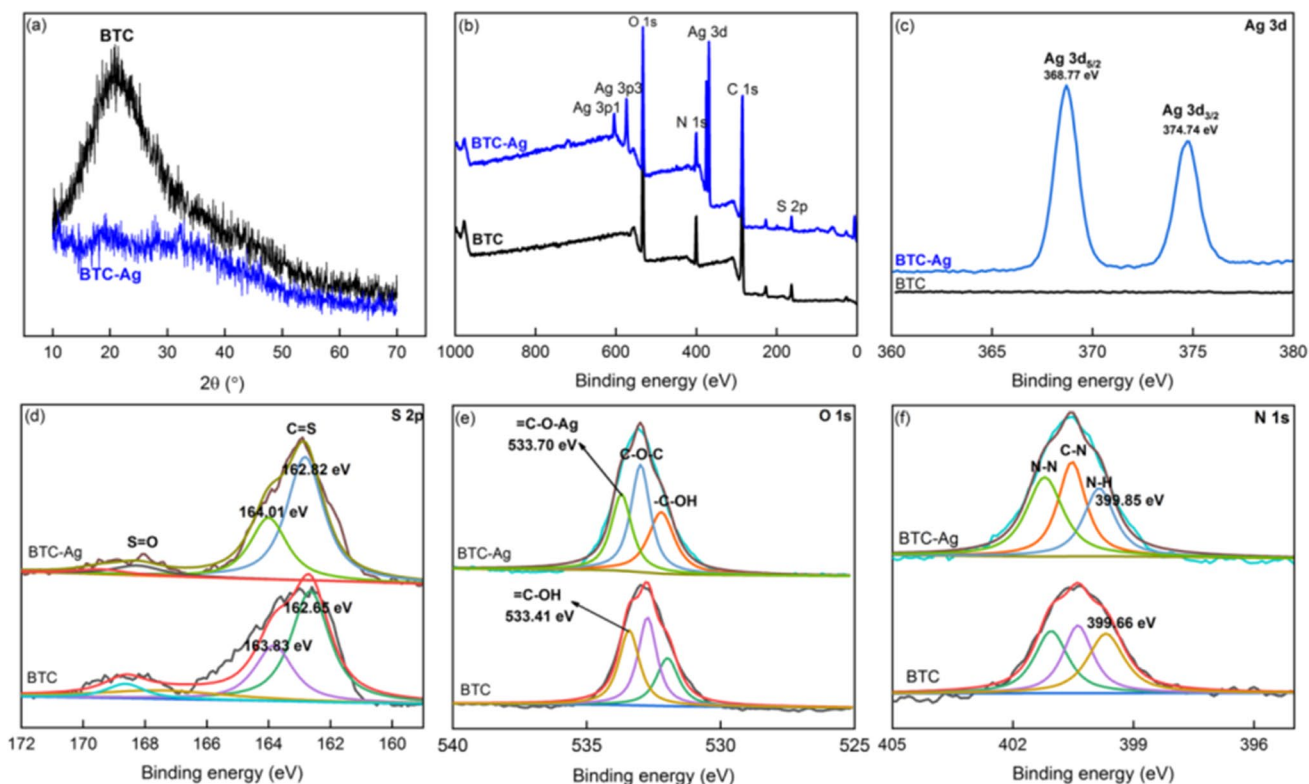
To examine the adsorption mechanism, the TF-IR spectrum of BTC after  $\text{Ag}^+$  adsorption (BTC-Ag) was obtained. As shown in Fig. 2, compared to the spectrum of the BTC, a new peak at  $1384\text{ cm}^{-1}$  related to  $\text{Ag-S}$  stretching vibration was observed in the spectrum of BTC-Ag [65]. In addition, the peak attributed to phenolic hydroxyl groups became narrower and shifted from  $3375$  to  $3404\text{ cm}^{-1}$ . The peaks attributed to N–H out-of-plane vibrations and wagging vibrations shifted from  $801$  and  $606\text{ cm}^{-1}$  to  $824$  and  $620\text{ cm}^{-1}$ , respectively. These observations indicated that O, S and N atoms on the BTC surface participated in  $\text{Ag}^+$  adsorption.

XRD analysis of the BTC resin before and after  $\text{Ag}^+$  adsorption was performed, and the results are shown in Fig. 10(a). BTC exhibited an amorphous structure with a

**Fig.9** (a) SEM, (b) EDS and (c–g) elements distribution of BTC after  $\text{Ag}^+$  adsorption







**Fig. 10** (a) XRD patterns of BTC and BTC-Ag ( $\text{Ag}^+$  adsorbed by BTC);(b) XPS survey spectra of BTC and BTC-Ag; (c) higher resolution of Ag 3d, S 2p, O 1s and N 1s spectra of BTC and BTC-Ag

**Table 5** Relative atomic ratios of BTC and BTC-Ag measured by XPS

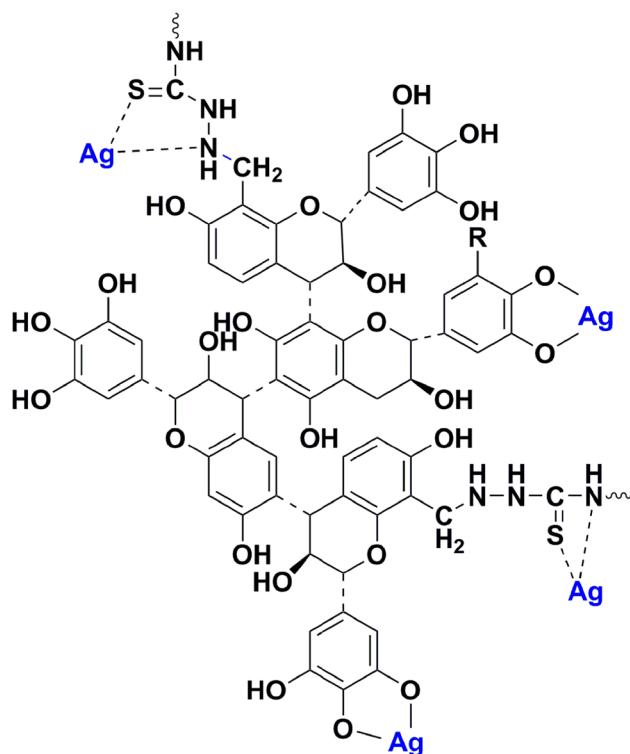
Samples	C(%)	O(%)	N(%)	S(%)	Ag(%)
BTC	63.02	22.91	10.73	3.36	0
BTC-Ag	58.47	22.44	11.19	3.46	4.43

broad hollow peak at  $2\theta = 22^\circ$ , which indicates the presence of polar functional groups [66]. After  $\text{Ag}^+$  adsorption, no new peak was observed, and the intensity of the peak at  $2\theta = 22^\circ$  substantially decreased. This could be ascribed to the formation of complexes between the phenolic groups on the BTC surface and silver ions, which reduces the polarity of the functional group [66].

To better assess the mechanism by which  $\text{Ag}^+$  could chelate with the functional groups on BTC, XPS measurements of BTC before and after  $\text{Ag}^+$  adsorption were performed. Figure 10(b) and Table 5 provide the XPS survey spectra and corresponding atomic ratios of the raw BTC and BTC-Ag measured by XPS. Before  $\text{Ag}^+$  adsorption, four peaks at 284.8, 533.07, 400.58 and 162.93 eV corresponded to C 1s, O 1s, N 1s and S 2p, respectively, were observed in the spectrum of fresh BTC, further illustrating the successful introduction of TC into BWT. After

$\text{Ag}^+$  adsorption, a new peak at a binding energy of 368.7 eV related to the representative peak of Ag 3d appeared. The notable signals in the deconvolution of Ag 3d orbital (Fig. 10(c)) provided evidence of  $\text{Ag}^+$  adsorption onto the BTC (Table 5), which is consistent with the SEM-EDS results. In addition, the shift in the electron binding energy in the high-resolution XPS spectra indicated a change in bonding around an atom. Figure 10(d), (e) and (f) show the S 2p, O 1s and N 1s XPS spectra, respectively, of BTC and BTC-Ag. The S 2p XPS spectra were deconvoluted into two distinct different component peaks, which could be attributed to S=O and C=S. The binding energies of C=S at 162.82 and 164.01 eV in pristine BTC shifted to 162.65 and 163.83 eV, respectively, after  $\text{Ag}^+$  adsorption, suggesting a possible chemical change of in S once  $\text{Ag}^+$  was adsorbed onto BTC [50]. Similar phenomena were further confirmed by shifts in the binding energies of O 1s and N 1s before and after  $\text{Ag}^+$  adsorption, revealing that N, O and S participated in  $\text{Ag}^+$  adsorption, which is well consistent with the FT-IR results.

As described in this section, the mechanism of  $\text{Ag}^+$  adsorption onto the as-prepared BTC resin, as shown in Fig. 11 could be attributed to the chelation between  $\text{Ag}^+$  and multiple electron-rich atoms (N, O, and S), in which the S atom played the most important role.



**Fig. 11** Suggested mechanism of  $\text{Ag}^+$  adsorption onto BTC

## 4 Conclusions

A tannin-based adsorbent (BTC) with nitrogen and sulfur atoms was successfully fabricated through the Mannich reaction. The BTC adsorbent exhibited high adsorption capacity and selectivity for  $\text{Ag}^+$  from solutions in the presence of other metals, such as  $\text{Pb}^{2+}$ ,  $\text{Ni}^{2+}$ ,  $\text{Ca}^{2+}$  and  $\text{Cd}^{2+}$ . The adsorption isotherm of  $\text{Ag}^+$  for BTC was described well by the Freundlich model, and the kinetic data of  $\text{Ag}^+$  adsorption onto the resin were suitably fitted by the pseudo-second-order model. Furthermore, batch experiments and various characterizations revealed that the interaction mechanism between the BTC resin and  $\text{Ag}^+$  involved inner-sphere complexation and chelation between  $\text{Ag}^+$  and multiple electron-rich atoms (N, O, and S). Among the multiple electron-rich atoms, the S atom played the most important role. In conclusion, this work described a one-step method in an aqueous solution for developing a tannin based adsorbent containing nitrogen and sulfur atoms for silver recovery.

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**Authors' contributions** X.B. Sun developed the concept, designed the experiments and drafted the manuscript. S.Y. Yin performed the experiments and acquired the data. Y.H. You reviewed the manuscript.

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## Declarations

**Competing interests** The authors declare no competing interests.

**Conflict of interest** The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

**Ethical approval** Not applicable.

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