Understanding phenol adsorption mechanisms on activated carbons

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Abstract The interactions between phenol molecules and activated carbons were investigated in order to understand the adsorption mechanism of this aromatic compound. A series of activated carbons with varied chemical composition but similar porous features were synthesized and submitted to phenol exposure from aqueous phase, followed by thermogravimetric analysis and identification of the desorbed species by temperature programmed desorption coupled with mass spectrometry. Based on these experiments, both physi- and chemisorption sites for phenol were identified on the activated carbons. Our results demonstrate that physisorption of phenol depends strictly on the porosity of the activated carbons, whereas chemisorption depends on the availability of the basal planes in the activated carbons. Thus, oxidation of the carbon can suppress the fraction of chemisorbed phenol since the surface functionalities incorporate to the edges of the basal planes; notwithstanding, hydrophilic carbons may present a small but not negligible contribution of chemisorbed phenol depending on the extent of the functionalization. Moreover, these adsorption sites (chemi-) are recovered by simply removal of the surface functionalities after thermal annealing.

Keywords Activated carbon · Phenol · Physisorption · Chemisorption

1 Introduction

Faced with an increasing contamination of water resources, adsorption has become a well-established technique to re-

L.F. Velasco · C.O. Ania (⊠) Instituto Nacional del Carbón, CSIC, P.O. 73, 33080 Oviedo, Spain e-mail: conchi@incar.csic.es move pollutants, activated carbon being the prevailing adsorbent for the purification of water with low pollutant concentration. Currently, the loaded carbon is regenerated offsite by heating or steaming in large industrial facilities. Although thermal reactivation is a highly skilled process that ensures that spent carbon is returned to a reusable quality, there is a major issue concerning the economic costs of the reactivation of the spent material. The efficiency of the different regeneration processes of activated carbons largely depends on the following factors: the porous structure of the carbon and the chemical condition of its surface, the physico-chemical properties of the adsorbent, the methods applied for regeneration, and the conditions under which the regeneration process is conducted (Dranca et al. 2001; Alvarez et al. 2004; Ania et al. 2005a; Salvador and Sanchez-Jimenez 1996).

Phenol is a widespread persistent organic pollutant commonly present in industrial effluents that has been the subject of a number of investigations, since it is a relatively simple and well characterized molecule. Despite the works devoted to the adsorption, desorption and regeneration of phenol using activated carbons, the interactions between activated carbon surfaces and phenol molecule are still rather unknown and controversial (Radovic et al. 2000; Nevskaia and Guerrero-Ruiz 2001; Haydar et al. 2003; Ania et al. 2002; Terzyk 2003).

The aim of this study is to contribute with a novel insight in the understanding of the interactions between phenol molecules and activated carbons with different surface chemistry. Comprehensive knowledge of the adsorption mechanism and identification of the adsorption sites on the carbon surface will contribute to optimize the thermal reactivation process of the carbonaceous adsorbent.

As the first target, we have synthesized carbons with varied surface chemistry by incorporating oxygen functionalities via wet oxidation. Also, the oxidized sample was submitted to a high temperature thermal treatment to remove the created groups; the performance of both samples towards phenol adsorption from aqueous solutions has been investigated and compared with that of the raw material. To identify the nature of phenol adsorption sites (weak interactions or physical adsorption and/or strong interactions or chemisorption), we have characterized the activated carbons before and after phenol exposure by means of different techniques including thermogravimetric analysis, temperature programmed desorption coupled to mass spectroscopy, and gas adsorption. The thermal stability of the phenol bonded to each adsorption site has been explored, and coupled to the needs in the thermal reactivation of the exhausted adsorbent. This has allowed us to throw some light into the understanding of the phenol adsorption mechanism on porous carbons, particularly the role of the activated carbon features (porosity and composition).

2 Experimental

2.1 Materials

A commercial activated carbon—B—obtained from physical activation of coal and supplied by Agrovin SA was used as starting material—particle size fraction of 0.212– 0.710 mm. A second sample with a similar porous structure but different surface chemistry was prepared by wet oxidation at room temperature with a saturated solution of ammonium persulfate in $4N H_2SO_4(1 \text{ g activated carbon:}10 \text{ mL of}$ oxidizing agent, and left stirring overnight). After oxidation, the carbon was washed and dried at 110 °C overnight. The oxidized sample (labeled as BO) was further submitted to a thermal treatment at 800 °C (BOH) under inert atmosphere for 1 hour in order to remove the surface groups and to explore the effect on the chemical, structural and adsorptive features of the activated carbons. Phenol with the highest purity specification was obtained from Aldrich.

2.2 Textural and chemical characterization of the carbons

Textural characterization was carried out by measuring the N₂ adsorption isotherms at -196 °C. Before the experiments, the samples were outgassed under vacuum at 120 °C overnight. The N₂ isotherms were used to calculate the specific surface area, S_{BET} , total pore volume, V_{T} , and pore size distributions which were evaluated by means of the DFT method using a slit pore geometry. Additionally, the volume of pores smaller than 0.7 nm (narrow micropores) was assessed from CO₂ adsorption isotherms at 0 °C with the DR formulism. The materials were further characterized by the determination of the point of zero charge (pH_{PZC})

using a modification of the mass-titration method by Noh and Schwarz (1989). About 0.25 g of activated carbon was dispersed in a suitable volume of distilled water and stored with constant stirring at room temperature until equilibrium was attained (ca. 48 h). Typical activated carbon/water ratios employed were 1, 2, 4, 6, 8, 10, 12, 14 and 16. Once equilibrium had been reached, the pH value was measured using a glass electrode. Newly distilled water was then added in order to obtain the next solid/weight fraction. The plateau in the plot of equilibrium pH versus solid weight fraction corresponded to the pH_{PZC} value of the activated carbon.

2.3 Thermal analysis

Thermal analysis was carried out using a thermogravimetric analyzer from Setaram. Experiments were carried out under an argon flow rate of 50 cm³ min⁻¹, at a heating rate of $15 \,^{\circ}$ C min⁻¹, up to a final temperature of 900 $^{\circ}$ C. For each experiment, about 25 mg of a carbon sample was used. Also, temperature programmed desorption experiments were performed with samples of about 100 mg in a Chemisorb 2750 apparatus from Micromeritics, under the same operating conditions mentioned above. Effluent gases were analyzed by means of a mass quadrupole (Balterzs).

2.4 Phenol adsorption from solution

The saturation of the activated carbons was carried out from batch experiments, putting in contact 1 g of adsorbent with 250 mL of a solution of the organic compound of known initial concentration. The suspensions were shaken at 500 rpm for 24 hours. Afterwards, the samples were filtered and the adsorbate concentration in the supernatant liquid was measured on a Shimadzu 2501 UV-Vis spectrophotometer at the corresponding wavelength. The exhausted activated carbons were washed with a small amount of water to remove the excess of phenol, dried at 60 °C for overnight, and kept in dessicator until use. The samples after phenol exposure will be labeled as the name of the corresponding activated carbon followed by "x" (i.e., Bx accounts for saturated B). Some experiments were also carried out by submitting the saturated and dried B carbon (sample Bx) to an outgassing treatment (vacuum + temperature) at 200 and 350 °C. These samples will be labeled as BxoutT, being T the temperature of the treatment.

3 Results and discussion

3.1 Characterization of the activated carbons

Detailed nanotextural parameters of the studied activated carbons obtained from gas adsorption data (N_2 and CO_2 at

 Table 1
 Main textural parameters determined from gas adsorption data of the pristine (B), oxidized (BO) and annealed carbon (BOH), before and after phenol exposure (series x), and after outgassing at different temperatures (series out)

Sample	$S_{\rm BET}$ $[m^2 g^{-1}]$	V_{TOTAL}^{a} [cm ³ g ⁻¹]	$V_{\text{MICROPORES}}^{\text{b}}$ [cm ³ g ⁻¹]	$V_{\text{MESOPORES}}^{\text{b}}$ [cm ³ g ⁻¹]	W_{0,N_2}^{c} [cm ³ g ⁻¹]	W_{0,CO_2}^{c} [cm ³ g ⁻¹]
В	1031	0.524	0.316	0.089	0.470	0.190
BO	989	0.501	0.306	0.073	0.470	0.233
BOH	1045	0.524	0.319	0.088	0.460	0.188
Bx	404	0.223	0.113	0.047	0.130	0.057
Bxout200	539	0.289	0.153	0.060	0.235	0.072
Bxout350	597	0.318	0.172	0.064	0.243	0.121

^atotal pore volume evaluated at $p/p^0 \sim 0.99$

^bevaluated from DFT method applied to N₂ adsorption data at -196 °C

^cmicropore volume evaluated from the DR method applied to gas adsorption data

 Table 2
 Ultimate analysis and pH of the point of zero charge of the studied carbons

Ultimate analysis (wt.% dry basis)							
	С	Н	Ν	S	0		
В	96.3	0.6	0.7	0.3	2.1	8.9	
BO	90.7	0.7	0.7	0.4	7.5	3.3	
BOH	97.7	0.4	0.1	0.4	1.4	6.6	

-196 °C and at 0 °C, respectively) are shown in Table 1. All the carbons displayed type I nitrogen adsorption isotherms (not shown) according to the BDDT classification (Brunauer et al. 1940), which is characteristic of microporous carbons. It should be noted that the oxidation treatment (sample BO) did not significantly alter the porous features of the as-received carbon, since only a small fall in the apparent surface area and pore volumes was observed. Also, the removal of the surface groups from the oxidized carbon does not seem to have a deep impact on the porosity of the annealed BOH carbon. An exhaustive analysis of the textural properties by means of the CO₂ adsorption isotherms revealed that the oxidation of the parent carbon caused a slight increase in the amount of CO2 adsorbed. The CO2 adsorption isotherm of the parent B carbon was recovered after the thermal annealing at 800 °C (carbon BOH), suggesting that such effect was likely attributed to the functionalization of the carbon surface.

As expected regarding surface chemistry, the oxidation of the activated carbon caused an increase in the oxygen content accompanied by a sharp fall in the pH_{PZC} (Table 2). The mass loss (dry basis) of the raw carbons goes from 2 wt.% for B carbon to 9.5 wt.% in sample BO, confirming the incorporation of surface functional groups upon oxidation. Moreover, the DTG profiles of BO carbon show a peak centered at about 250 °C, corresponding to the decomposition of labile (typically acidic) groups. Above 400 °C, a broad desorption band is obtained, assigned to more stable oxygen-containing surface functionalities (Bandosz and Ania 2006). According to expectations, the thermal treatment at 800 °C of the oxidized carbon caused the complete removal of the surface functionalities (total mass loss below 1 wt.%), rendering a hydrophobic carbon (sample BOH).

3.2 Phenol uptake in activated carbons with varied properties

The effect of the activated carbons features on phenol adsorption mechanism was investigated using un-buffered aqueous solutions and evaluating the phenol retention capacity on each adsorbent. The adsorption capacities for the investigated carbons are 315, 198 and 382 mg g⁻¹ for B, BO and BOH, respectively. It appears that the oxidation of the carbon surface has a large impact on the removal of phenol, which is much larger in the activated carbons of hydrophobic nature (uptake falls by almost 40%). Moreover, the removal of the surface functional groups allows to recover the large capacity of the activated carbon.

These values are in good agreement with the trend observed in the corresponding mass losses after phenol exposure evaluated from the thermal analysis (ca. 9, 5 and 11 wt.% for Bx, BOx and BOHx respectively). This data is also consistent with previous works found in the literature about phenol removal on activated carbons (Moreno-Castilla et al. 1995; Ania et al. 2002; Castillejos-Lopez et al. 2008; Toth et al. 2009; Terzyk 2003), reporting the negative impact of carbon surface oxidation in the overall phenol uptake.

The adsorption performance on a given porous material is traditionally related to the chemical composition and the porous features of the adsorbents. In this sense, according to the similar textural characteristics of the three studied carbons (Table 1), one would expect similar phenol uptakes, or





Fig. 2 TPD-MS profiles after phenol exposure of (i) carbon B, (ii) carbon BO and (iii) carbon BOH. Inset: magnification of the m/z 66, 94 signals

perhaps a somewhat higher uptake on the oxidized carbonwith the highest volume of narrow micropores. Thus based on the experimental data obtained, it is clear that phenol retention cannot be exclusively attributed to the porosity and it should rather be discussed in terms of the different surface chemistry of the studied activated carbons.

3.3 Thermal analysis of the activated carbons after phenol exposure

For this reason we have investigated the samples after phenol exposure by means of thermogravimetric analysis and temperature programmed desorption coupled to mass spectrometry (TPD-MS). The corresponding weight derivatives of the thermal analysis profiles (DTG) as well as some of the mass fragments detected by TPD-MS are shown in Figs. 1 and 2. The most important fragments detected between 25 and 900 °C and their assignments were CH₃⁺ (m/z 15), H₂O (*m*/*z* 18), CO (*m*/*z* 28), O₂ (*m*/*z* 32), CO₂ (*m*/*z* 44), C₅H₅ (*m*/*z* 66), C₅H₁₁ (*m*/*z* 78), C₆H₅O (*m*/*z* 93), C₆H₆O (*m*/*z* 94) and C₁₂H₈O (m/z 168). For clarity purposes, only those fragments corresponding to m/z 18, 66, 78 and 94 are plotted since they provide most useful information about phenol adsorption mechanisms and thermal cracking.

The DTG profiles after phenol exposure allow the different adsorption sites on the studied carbons to be identified (Fig. 1). Three well defined desorption peaks and two wide humps at temperatures above 500 °C can be distinguished in B and BOH samples, as opposed to the oxidized carbon with only two peaks. The first peak at around 100 °C is typically assigned to moisture retained in the porous network of the carbons since phenol exposure is carried out from an aqueous solution. This assignment was confirmed by the m/z 18 signal detected by TPD-MS (Fig. 2).

The peaks at higher temperatures are attributed to phenol retention on different sites of adsorption since they are only detected in the saturated carbons, with the exception of the high temperature humps in BO carbon due to the decomposition of the surface functional groups (Fig. 1). It is generally accepted that the fraction of phenol adsorbed at low temperatures corresponds to physisorption whereas the peaks at higher temperatures belong to chemisorbed species (Moreno-Castilla et al. 1995; Ania et al. 2005b; Castillejos-Lopez et al. 2008; Toth et al. 2009). In good agreement with previous works and based on our DTG data it would seem that phenol is both physi- and chemisorbed in the hydrophobic B and BOH carbons, whereas chemisorption is not detected in the oxidized carbon. However, a deep analysis of the mass to charge fragments by TPD-MS revealed interesting and surprising findings.

An important observation is that when the raw carbon is oxidized the temperature of the first desorption band observed in DTG profiles (physisorbed phenol) does not change. In fact, it remains somewhat the same for the three investigated carbons, which is consistent with their similar textural properties. This result indicates that the presence of the oxygen functional groups at the pores entrance does not hamper the physisorption of phenol. Furthermore, when the functional groups are removed from the carbon surface (sample BOH) the desorption peak still remains in the same position-as the porosity of the carbon remains unaltered after the thermal treatment. At converse, upon removal of the surface functionalities, the high temperature phenol desorption peaks reappear-those detected in the hydrophobic B carbon but not after oxidation-. This result also suggests that there does not exist specific interactions between phenol molecules and the oxygen functionalities (ca. H-bonding and/or anchoring through the CO moiety in carboxylic acids on the carbon surface), which has been a controversial issue in the literature so far (Nevskaia and Guerrero-Ruiz 2001; Haydar et al. 2003).

Following the evolution of the different species throughout the whole temperature desorption range (Fig. 2) it can be seen that phenol is retained on different surface sites. According to literature (Smith 2004), fragmentation of phenol due to ionization in the MS spectrometer yields the fragments 94 ($C_6H_5O^+$) and 66 ($C_5H_5^+$), with 99% relative abundance for m/z 94. During the TPD experiments both fragments appear at the same temperature, thus confirming their formation due to MS cracking of phenol. Additionally, various m/z signals were also observed at temperatures above 500 °C in the profiles of all the saturated carbons (i.e., m/z 78 assigned to C₆H₆⁺). These are due to the thermal cracking of phenol above this temperature (Spielmann and Cramers 1972; Cypres and Betfens 1974), which generates benzene, naphthalene, dibenzofuran, H₂O, CO, H₂ and CH₄, along with a solid residue as main sub-products. Consequently, the intensity of the m/z 94 signal is expected to decrease above 500 °C at expenses of an increase in the mass to charge signals of phenol cracking sub-products.

According to this, the appearance of the m/z 78 (C₆H₆⁺) signal along with an increase in the m/z 15, 28 and 44—assigned to CH₄, CO and CO₂, respectively—in the profiles of all the pre-adsorbed carbons (even in the oxidized sample) has to be linked to the occurrence of high temperature phenol desorption sites. Moreover, the temperature of the m/z 78 peak in carbons B and BOH matches with the desorption humps observed in the DTG profiles (Fig. 1) at high temperatures.

It should also be highlighted that several m/z 18 signals are observed for all the studied carbons, with broad bands which maxima temperature matches with the evolution of the m/z 94, 78 signals. Similar water desorption profiles have been reported for the adsorption of aromatics from aqueous solution (Nevskaia and Guerrero-Ruiz 2001; Castillejos-Lopez et al. 2008), and attributed this finding to the desorption of water associated with the adsorbate in supramolecules. However, one should also take in mind that water is also originated from the thermal cracking of phenol molecules: above 700 °C phenol thermal degradation yields between 6-30 wt.% water (Cypres and Betfens 1974). This would also explain the small hump observed in the m/z18 profiles at temperatures above 700 °C. Another factor to consider is that water may also appear in condensation reactions of phenol to render dibenzofuran (around 400-500 °C). For instance, we have detected the m/z 168 (C₁₂H₈O) signal corresponding to this compound in the desorption profiles of B and BOH. Thus, it may be inferred that water desorption between 200-500 °C may be linked to the water molecules associated to phenol during the adsorption process, while at higher temperatures the contribution due to phenol thermal cracking cannot be neglected.

Comparatively, the nature and number of phenol adsorption sites change with the nature of the activated carbons. For the hydrophobic B and BOH carbons, phenol seems to be adsorbed through physisorption and multiple chemisorption sites, whereas the chemisorption appears to be strongly reduced after oxidation. To throw some light on this issue, we have analyzed saturated B carbon after outgassing at 200 and 350 °C during 3 hours and overnight, respectively. **Fig. 3** Desorption profiles (DTG) before and after phenol exposure of saturated carbon B after outgassing at different temperatures



Fig. 4 TPD-MS profiles after phenol exposure of the saturated B carbon after (**i**) outgassing at 200 °C and (**ii**) outgassing at 350 °C

The DTG and TPD-MS profiles of samples Bxout200 and Bxout350 are shown in Figs. 3 and 4, respectively. Outgassing at 200 °C (sample Bxout200) provoked almost the complete removal of the fraction of physisorbed phenol (Fig. 3), while the chemisorbed fraction remained unaltered: the three bands at temperatures above 300 °C are still detected, as well as the corresponding m/z 94 and 78 signals (the shape of the TPD-MS profiles above 300 °C was not modified with respect to Bx). Analogously, after outgassing at 350 °C only the m/z 94, 78 signals at 500 °C and above are observed, indicating the complete removal of physisorbed phenol as well as a fraction of chemisorbed phenol. It is noteworthy that compared to the removal of the

physisorbed fraction (outgassing at 200 °C for 3 hours), not only a higher temperature but also a much larger time of the outgassing treatment is required to achieve a partial removal of the chemisorbed phenol from the surface of the activated carbon (outgassing at 350 °C for 17 hours). This performance highlights the differences in the nature of the forces involved in physical and chemical adsorption, being the latest considerably stronger.

The removal of a fraction of adsorbed phenol is also confirmed by the textural characterization of Bxout200 and Bxout350 (Table 1); after the outgassing treatment there is an increase in the surface area and micropore volumes, being more notable in the sample treated at 350 °C. The outgassing treatment of the saturated samples assures that all the evolved molecules are withdrawn from the carbon surface preventing any deposition; this demonstrates that the high temperature adsorption sites observed in the outgassed series correspond to phenol uptake from solution, and not to reorganization of the desorbed phenol molecules during the thermal treatment, nor to its degradation sub-products on the porosity of the carbons.

What it's interesting to note is that the outgassing treatment provokes changes in the water desorption profiles, following a similar trend than that observed for signal m/z 94. This confirms that water desorption below 400 °C—where phenol thermal degradation is discarded—is associated to phenol molecules retained on the activated carbon surfaces, as suggested in the literature (Nevskaia and Guerrero-Ruiz 2001).

After oxidation of the pristine activated carbon, the peak corresponding to physisorbed phenol does not change as it was mentioned above (Fig. 1). This was also clearly observed in Fig. 2; however, small peaks in the m/z 94 (Fig. 2, inset) and m/z 78 were also detected at moderate temperatures between 400–600 °C. Despite the intensity of these signals is much lower than in the case of hydrophobic carbon, this demonstrates that there still remains a small but non negligible contribution of chemisorbed phenol in the oxidized carbon. It would seem then that only a part of the chemisorption sites is blocked during the oxidation. In agreement with this finding, the m/z 18 hump above 500 °C is observed for BO carbon (not detected in its corresponding blank in water), confirming either water associated to phenol molecules or water released during phenol thermal cracking.

The negative impact of carbon oxidation on the chemisorption of phenol has been recently reported (Toth et al. 2009), although the authors state that physisorption of phenol is favored in hydrophilic carbons. However, our results demonstrate that physisorption does no depend on the oxidation state of the carbon material but on its porosity. In fact, the mass loss in the saturated carbons accounting for the physisorbed fraction of phenol is 5 wt.% for all the carbons. Only chemisorptions is affected (largely suppressed) by the presence of oxygen-containing groups. In any case, both findings contrast with previous works reporting one single m/z 94 peak in oxidized carbons, where it is concluded that oxygen-containing surface functional groups inhibit the phenol chemisorption on activated carbons (Castillejos-Lopez et al. 2008).

To understand this issue one should take into account that we have carried out a mild oxidation—in order to avoid drastic changes in the porosity of the carbons—, and that the oxygen content in BO carbon is not very large (Table 2). Since the surface functionalities are mainly incorporated at the edges of the basal planes (Donnet 1982), it seems that they have not been fully decorated with the functional groups; thus there would still remain some non-blocked chemisorption sites on the oxidized carbon. This would explain the differences observed with previous works reported in the literature, where unfortunately details on the oxidation extent on the carbons are missing (Castillejos-Lopez et al. 2008).

Moreover, the DTG and TPD-MS of carbon BOH (Figs. 1 and 2) confirm this premise; the high temperature desorption peaks in the DTG profiles reappear in BOH carbon, while the intensity of the m/z 94 and 78 signals above 500 °C— which were partially removed in the oxidized carbon— increases significantly. In fact, the profiles of carbon BOH are similar to those of the as-received B carbon, which demonstrates that the sites for phenol chemisorption are restored by removal of the oxygen functionalities created upon oxidation.

On the other hand, the complete removal of phenol is not achieved even if high temperatures (i.e., 800 °C) are applied; despite the desorption profiles do not reveal the presence of any peak after the high temperature treatment, the porosity of the activated carbons is not fully recovered (ca. 33% fall in surface area and pore volumes after treatment at 700 and 800 °C, respectively). The reason for this is that thermal cracking of phenol gives rise to a solid residue (yield between 30 wt.%) which deposits on the carbon surface and causes pore plugging on the activated carbon, thus limits the thermal regeneration of the exhausted materials (Ania et al. 2005a; Cypres and Betfens 1974). In other words, if regeneration of exhausted activated carbons is carried out at moderate temperatures, there is a fraction of chemisorbed phenol that is not eliminated. In contrast, higher temperatures also provoke a collapse in the porous features of the activated carbons, although as a consequence of the thermal cracking of the phenol molecules forming coke deposits.

Summarizing, it may be concluded then that DTG coupled to TPD-MS allow to attribute the desorption peaks to different phenol adsorption sites: physisorbed phenol (ca., 200 °C), and three chemisorption sites (ca., 400, 550, 700 °C). This also confirms that physisorption of phenol does not depend on the oxidation state of the activated carbons but on their porosity. In contrast, chemisorption seems to be favored in hydrophobic carbons due to interactions of phenol molecules with the edges of the graphene planes in the carbon. Consequently, chemisorbed phenol may be suppressed when these edges are decorated with the surface functionalities, the extension of which depends on the coverage/blocking of the chemisorption sites.

4 Conclusions

This work investigates the mechanisms of phenol adsorption on activated carbons by means of thermal analysis of

the adsorbents after phenol exposure, with the identification of the desorbed species by temperature programmed desorption coupled with mass spectrometry. A series of activated carbons with similar textural properties and different surface chemistry were investigated; these were obtained from a hydrophobic carbon that was oxidized to create oxygencontaining functional groups, and subsequently annealed to remove them. The results demonstrate that physisorption of phenol depends strictly on the porosity of the activated carbons, regardless the oxidation state of the carbon surface. Besides physisorption, phenol molecules interact with the edges of the basal planes in the carbons, showing various chemisorption sites desorbing at temperatures above 400 °C (likely due to different orientations of phenol molecules). Consequently, the incorporation of the surface functionalities on the edges of the basal planes-via carbon oxidation-blocks the chemisorption sites reducing the interactions. However, hydrophilic carbons may present a small but not negligible contribution of chemisorbed phenol depending on the extent of the functionalization.

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