



Phenological shifts compensate warming-induced drought stress in southern Siberian Scots pines

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Abstract

Global climate change impacts the functioning and productivity of forest ecosystems at various spatiotemporal scales across a wide range of biomes. Although summer temperatures are considered the main driver of boreal tree growth, the importance of soil moisture availability is likely to rise with decreasing latitude and increasing warming. Here, we combine dendrochronological measurements with evidence from tree growth modeling and remote sensing to quantify the effect of climate on phenology and productivity of Scots pines (*Pinus sylvestris* L.) in southern Siberia. Between 1960 and 2017, pine ring widths along a latitudinal transect from 53° to 56°N were mainly controlled by the availability of summer soil moisture. This finding challenges the common belief that summer temperatures are the predominant growth control in boreal forests. Moreover, we show that earlier growing season onsets can compensate for warming-induced drought stress. Despite the phenotypic plasticity of Scots pines to adapt to warmer and drier conditions, we speculate that predicted climate change will likely exceed the species' physiological tolerance in much of Eurasia's forest-steppe by the end of the twenty-first century.

Keywords Climate change · Drought stress · Forest decline · Phenology · Tree growth · Vaganov–Shashkin model

Introduction

The Russian boreal forest represents the world's largest continuous forest biome, accounting for circa 50% of all boreal forests (Frankina et al. 1997) and storing an estimated 46–67 petagrams of carbon in its living biomass (Houghton et al. 2007; Frankina et al. 1996). However, predicted climate change would probably affect these forest ecosystems (Allen et al. 2015, 2010; Choat et al. 2012; Kharuk et al. 2017) by modifying the phenology and productivity of trees, compromising the forests' capacity to sequester carbon from the atmosphere (Bradley et al. 1999; Trumbore et al. 2015), and thus affecting the global carbon cycle negatively (Bradshaw and Warkentin 2015; Girardin et al. 2016).

The significant warming trend documented across Siberia during the last decades (Groisman et al. 2012) affects different forest types' geographical distribution and growth (Tchebakova et al. 2016). For example, the risk of climate change impacts on forest ecosystems is considerable in southern Siberia, where tree growth during the short but hot summers tends to be more constrained by soil moisture availability (Arzac et al. 2018, 2021; Tabakova et al. 2020). Therefore, the predicted warming and temperature-induced

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drought (Anisimov et al. 2008; Shvidenko et al. 2012) will influence the phenology and physiology of trees, affecting forest health (Frank et al. 2015; Tchebakova et al. 2016).

Plant phenological patterns are sensitive to climate variability (Menzel et al. 2020; Piao et al. 2019). Current warming trends will likely result in earlier starts and delayed ends of the growing seasons (Jeong et al. 2011; Zhang et al. 2004). An extensive literature body reports increasing growing season length of boreal forests in recent time (Fu et al. 2015; He et al. 2018a, 2018b; Menzel et al. 2006; Spinoni et al. 2015; Yang et al. 2017; Zhou et al. 2003). Hence, geographical regions experiencing above-average warming, as Russian boreal forests and forest-steppe, are of high interest for understanding the response of main forest tree species to changing climate conditions.

Long-term xylem phenology studies are time-consuming and usually restricted in space and temporal scale. Therefore, process-based models and remote sensing tools might facilitate plant phenology research (He et al. 2018a; Yang et al. 2017). For example, tree radial growth (wood formation) models, such as the Vaganov–Shashkin model (VS-model; Vaganov et al. 2006), can simulate tree growth and phenology in response to daily variability in both air temperature and soil moisture (Anchukaitis et al. 2020; He et al. 2018a; Shishov et al. 2016; Tyckov et al. 2019; Vaganov et al. 2006). At the same time, vegetation indices like the EVI (Enhanced Vegetation Index; Huete et al. 1999), based on photosynthetic activity changes, leaf biomass and canopy structure (Duchemin et al. 1999; Reed et al. 1994) are used to estimate intra-annual changes in canopy greenness and annual forest productivity and phenology at high

spatiotemporal resolution. Nevertheless, remote sensing datasets (e.g., EVI) are constrained to temporal series no earlier than the year 2000, whereas tree growth simulations could be extended over longer time periods.

We combined evidence from dendrochronological measurements, tree growth modeling and remote sensing at three sites with different water balance levels to increase our understanding of Scots pine xylem phenology and growth variability in southern Siberia. More specifically, this study aims to: (i) compare VS-model phenological simulations against MODIS (Moderate Resolution Imaging Spectroradiometer) satellite observations, (ii) evaluate inter-annual changes in pine secondary growth (xylem phenology) via the VS-model over the 1960–2017 period, and (iii) link the timing of phenological phases with spring temperatures as the trigger of the growing season onset. We hypothesized that (i) water availability will be the predominant climatic factor constraining pine growth even in the wettest site, and (ii) warmer springs conditions and the plastic responses of pine cambial activity will compensate the expected ring width reductions due to summer drought stress by shifting pine phenology to earlier onsets after snowmelt at the driest site.

Materials and methods

Study area

Three Scots pine (*Pinus sylvestris* L.) sites with different water balance levels were sampled along a latitudinal transect (56–53°N) in southern Siberia (Krasnoyarsk Krai,



Fig. 1 Study area. Stars represent the location of the sampling sites. M, moderate (Pogorelsky bor); W, wet (Stolby); D, dry (Shushensky bor)

Table 1 Sites characterization (climate data 1960–2017) and summary of statistics calculated for the period 1960–2017 for *Pinus sylvestris* residual ring width chronologies

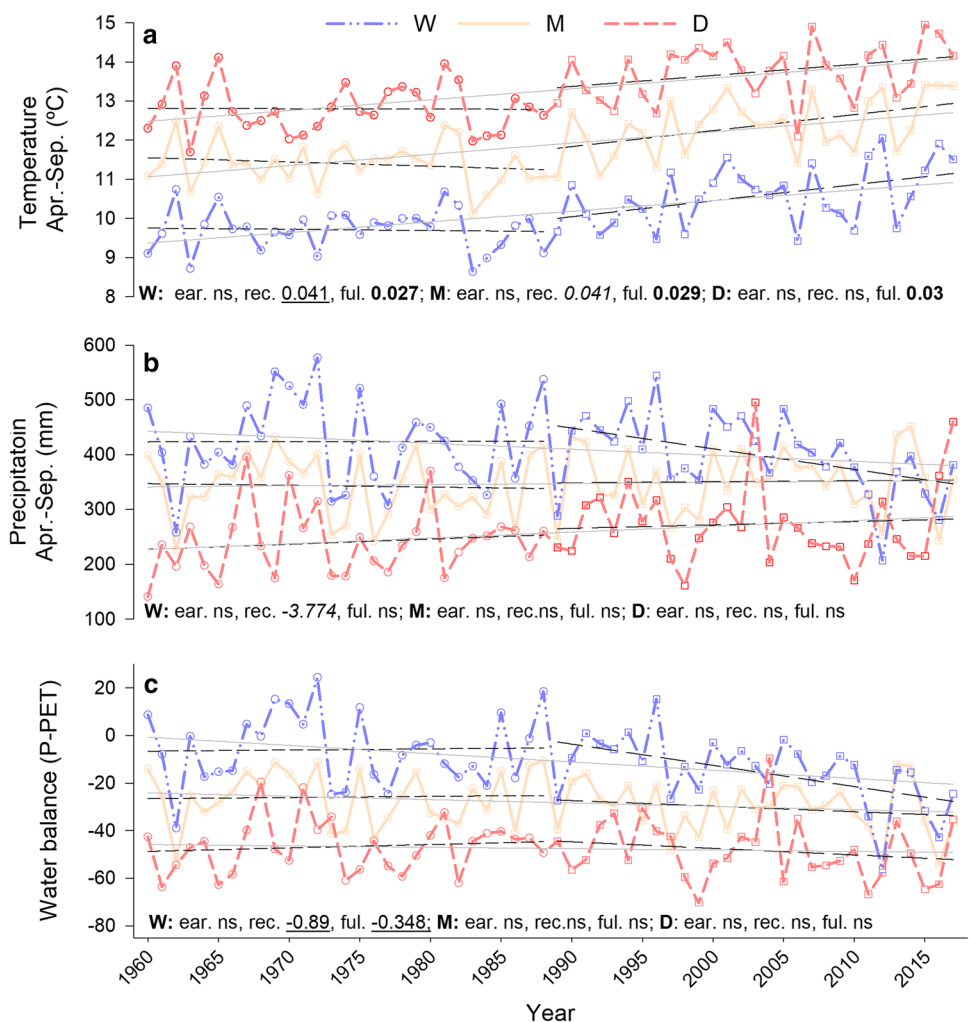
	Wet	Moderate	Dry
Altitude (m asl)	538	164	266
Mean annual <i>T</i> (°C)	0.1	1.3	1.5
<i>T</i> mean spring	0.4	1.9	3.1
<i>T</i> mean summer	15.3	16.8	18.1
Total annual <i>P</i> (mm)	668	537	308
Total spring <i>P</i>	136	106	43
Total summer <i>P</i>	246	215	183
Meteo station	Stolby	Krasnoyarsk	Abakan
No. trees	20	17	20
Age ± SD	116 ± 15	101 ± 13	85 ± 13
RW (mm) ± SD	1.00 ± 0.18	1.80 ± 0.44	2.02 ± 0.75
<i>ms_x</i>	0.133	0.108	0.165
<i>r_{bt}</i>	0.382	0.383	0.461
EPS	0.921	0.909	0.945
AC1	0.250	0.575	0.614

ms_x, mean sensitivity, *r_{bt}*, mean correlation between trees, *EPS* expressed population signal, *AC1* first order autocorrelation

Russia; Fig. 1). The northern site was located in the experimental plot “Pogorelsky bor” within the northern forest-steppe (hereafter moderate site; 56° 22' N, 92° 57' E). The central site in the National Park “Krasnoyarsk Stolby” in the mid-mountain taiga (hereafter wet site; 55° 54' N, 92° 43' E). The southern site in the National Park “Shushensky bor” in the transition from the southern taiga into the forest-steppe (hereafter dry site; 53° 18' N, 91° 58' E). The moderate and dry sites are mono-specific *P. sylvestris* forest stands. In contrast, pine trees co-occur with *Larix sibirica* Ledeb., *Betula pendula* Roth, and *Abies sibirica* Ledeb. at the wet site.

The transect is characterized by a cold continental climate (Kottek et al. 2006), with mean annual temperatures ranging from 0.1 °C at the wet site to 1.5 °C at the dry site, and the total annual precipitation ranging from 308 mm at the dry site to 668 mm at the wet site (Table 1). Daily climate data (1960–2017) were obtained from the nearest meteorological station to the moderate and dry sites (www.meteo.ru) and from a local station in the National Park “Krasnoyarsk Stolby” at the wet site. During the studied period, mean

Fig. 2 Climate conditions (temperature (a), precipitation (b) and water balance (c)) during April–September over the period 1960–2017. Water balance (WB) was calculated as $WB = P - PET$, where *P* represents the precipitation and *PET* the potential evapotranspiration. Trends are shown for the full period (ful., solid lines) and the early (ear., short-dashed lines) and recent (rec., long-dashed lines) periods. Regression coefficients (b1) in bold type are significant at $P < 0.001$, coefficients underlined are significant at $P < 0.01$ and coefficients in italic are significant at $P < 0.05$ (non-significant coefficients are not shown). W, wet site; M, moderate site; D, dry site



annual temperature has increased at an average rate of 0.3 °C per decade ($R^2 = 0.25$; $P < 0.001$), warming faster at the moderate and wet sites in recent time (Fig. 2a). In recent times, the total annual precipitation has decreased at the wet site without significant trends at the moderate and wet sites (Fig. 2b). Water balance ($WB = P - PET$), where P is the precipitation, and PET is the potential evapotranspiration calculated by the Thornthwaite equation (Thornthwaite 1948), was estimated for the 1960–2017 period. Lower WB values correspond to the dry site, but only significant decreasing trends were found at the wet site (Fig. 2c).

Sampling design and tree-ring measurements

Twenty mature dominant or co-dominant trees were sampled at each site during autumn 2018. Two wood cores per tree were taken at breast height using a 5-mm diameter increment borer. Cores were polished using a belt sanding machine with successive sandpaper (up to 600 grit) until a smooth and flat surface was obtained.

Wood cores were scanned at 3200 dpi with an Epson Perfection V800 scanner (Epson, Japan). Tree-ring width (RW) was measured on the scanned cores using Coorecorder version 9.3 (Cybis Elektronik & Data AB, Sweden). The accuracy of cross-dating was verified using COFECHA (Grissino-Mayer 2001), and each raw RW series was standardized using ARSTAN (Cook and Holmes 1996). Spline functions with a 50% frequency cut-off at 32 years were fitted individually to the raw series. Residual chronologies were computed using the bi-weight robust average. The approach is flexible enough to reduce any of the non-climatic lower frequency variances while at the same time preserving the high-frequency, inter-annual climate information (Cook and Peters 1981), which we consider relevant for the further assessment of the past five decades.

Simulations of tree growth and phenological dynamics

Tree growth and phenological key dates were simulated by the VS-Oscilloscope (VS-O) Ver. 1.32 (Shishov et al. 2016; Tychkov et al. 2019), the visual parametrization interface of the environmental block of tree-ring growth of the Vaganov–Shashkin model available at <http://vs-genn.ru/>. The model simulates daily tree growth rates (Gr) based on two pre-defined growth-limiting functions, given the temperature and soil moisture conditions occurring every day of the year. Thus, daily tree growth is defined by the limitation of the growth rates dependent on temperature or soil moisture and weighted by the photoperiod determined by the site's latitude (Tychkov et al. 2019; Vaganov et al. 2006). Consequently, the $Gr = \min(Gr_W, Gr_T) \cdot Gr_E$, where Gr_W , Gr_T and Gr_E are the daily partial growth rates based on soil

moisture, temperature and solar irradiation, respectively. The Gr_E depends on the latitude (Gates, 2012), whereas Gr_W and Gr_T are nonlinear functions scaled between 0 and 1 (Fig. A1).

To assess the effect of local conditions on growth, the model requires a calibrated parametrization by fitting the simulated RW chronology against the observed one. Therefore, the evaluation of the VS-model is based on the values of Pearson's correlation coefficient (r) and the Gleichläufigkeit synchrony (g_{lk}) between the simulated and actual RW chronologies (Shishov et al. 2016). As the output of the model, values of daily growth rates (Gr_W , Gr_T and Gr), simulated RW indices and annual estimations of the date for the start (SoS), end (EoS) and length (LoS) of the growing season are provided. By definition, the SoS occurs when the daily temperature is greater than or equal to the value of the minimum temperature required for tree growth (T_{min}), and the temperature sum for the previous ten days reaches some critical level (T_{beg}). The EoS is defined as the last day within a year when the growth rate's value is greater than the critical growth rate (V_{cr}) and the temperature sum is no longer higher or equal T_{beg} (Vaganov et al. 2006). Finally, the LoS is estimated as the number of days between SoS and EoS.

In this study, the VS-model was parameterized for each site over the 1960–2017 period. Residual RW chronologies and daily climate data were used as input data. The calibration against the residual RW chronologies was performed from 1960 to 1988 and verification from 1989 to 2017.

Validation of phenology simulations against remote sensing

Simulated phenological series were compared to derived data from the Moderate Resolution Imaging Spectroradiometer (MODIS) for a 14-year period (2001–2014) at each site. Annual estimations of vegetation phenology were obtained from the MODIS Land Cover Dynamics data product MCD12Q2 (Ver. 005; Ganguly et al. 2010). The data set includes the timing for the onset of four phenological phases (greenup, maturity, senescence and dormancy) with 500 m-ground resolution. The phenophase transition dates were based on logistic models (Zhang et al. 2006), by the estimation of the curvature-change rate of the Enhanced Vegetation Index (EVI), computed from composed eight-day MODIS Nadir Bidirectional Reflectance Distribution Function (BRDF; Schaaf et al. 2002) and the Adjusted Land Surface Reflectance (NBAR-EVI) data fitted with a sigmoidal function. The seasonal onset of continuous EVI increment corresponds to annual vegetation greenup and the seasonal onset of EVI minima corresponds to vegetation dormancy dates, associated with SoS and EoS, respectively (Zhang et al. 2006).

To test differences in mean simulated phenological dates (SoS, EoS and LoS) compared to derived MODIS datasets (and due to gaps in the MODIS dataset), we applied a one-way ANOVA (data source as a fixed factor). In addition, to compare inter-annual patterns, Pearson's correlations and linear regressions were performed.

Pine growth and phenology response to climate

Pine growth response to climate was assessed using Pearson's correlations between monthly climatic data (temperature, precipitation and water balance) and RW residual chronologies for the 1960–2017 period. Correlation coefficients were calculated with a temporal window comprising from the previous year July to September of the current growing season. In addition, we grouped the climatic variables in climatic seasons for the current growing year: winter (December–February), spring (March–May), summer (June–August) and autumn (September–November), as well during critical months for growth (May–July). Finally, climate correlations were re-assessed by splitting the study period into two equally long 29-year intervals, 1960–1988 and 1989–2017.

Linear regression functions were applied to analyze the link between phenological dynamics and spring temperatures over the 1960–2017 period. Linear regressions were also used to evaluate trends in the simulated pine phenological series (SoS and EoS) over the full and the early/late periods. Finally, since the SoS is defined by temperature in the model, we evaluated how spring temperature increments might affect the start of the growing season in the future. We forecasted potential changes in SoS by extrapolating current temperature to possible increments of 4 to 5 °C as predicted by the end of the century in southern Siberia (Groisman et al. 2012).

Results

Climate response of radial growth

Pine climate sensitivity was consistent at all the sites during the growing season (Fig. 3a). Secondary growth was favored by late spring and early summer precipitation, with earlier signal (May, $r=0.47$; $P<0.001$) at the dry site and later signals (June) at the moderate and wet sites ($r=0.32$; $P<0.01$ and $r=0.26$; $P<0.05$, moderate and wet sites, respectively). The aggregated effect of spring (March–May) precipitation showed a significant positive impact on growth at the dry site. Precipitation from previous growing season (August and November) also favored tree growth at the dry site ($r=0.33$; $P<0.01$ and $r=0.25$; $P<0.05$, respectively). Whereas winter precipitation (previous December) showed a negative effect at the wet site ($r=-0.36$; $P<0.01$).

Warm April and previous October temperatures enhanced tree growth at the moderate and wet sites ($r=0.32$ and $r=0.36$; $P<0.01$, respectively). On the other hand, May and June temperatures exerted a negative effect at the dry site ($r=-0.35$; $P<0.01$ and $r=-0.23$; $P<0.05$), with the aggregated effect of May–July temperatures amplifying the negative effect ($r=-0.41$; $P<0.001$).

Pearson's correlations between RW residual chronologies and water balance (P-PET) suggested a dependence on water availability at different temporal windows among the studied sites. Thus, the moderate and wet sites responded to a shorter water availability window (May, $r=0.34$; $P<0.01$ and $r=0.22$; $P<0.05$, respectively). Whereas the dry site responded for a longer temporal window (late spring and mid-late summer), with the highest signal in August ($r=0.37$; $P<0.01$). Previous December signals mimic the response to precipitation at the moderate and wet sites.

Pine climate sensitivity decreased in the recent period (1989–2017) compared to the early period (1960–1988). As a result, temperature and precipitation signals from late spring and summer declined and became not significant, except at the dry site where the negative effect of May–July and summer temperatures remained (Fig. 3b & c). Interestingly, the impact of water availability negatively increased, with the response to water balance becoming significant during July–August at the dry site. Moreover, the aggregated effect of water balance consecutive months showed a shift from the spring to the summer (and extended to the autumn) from the early to the recent period, with a maximal signal reached during the summer (0.52 ; $P<0.01$).

Tree growth model simulations

Tree growth simulations were significantly correlated ($P<0.05$) with the residual RW chronologies at all the sites for both the calibration and the verification period (Fig. A2, see also Table A1 for model parameters). However, at the wet site, the performance of the model was suboptimal. Correlation coefficients for the calibration period (1960–1988) ranged from $r=0.31$ at the wet site (suboptimal performance of the model) to $r=0.66$ at the moderate site, and the synchrony (glk, Gleichläufigkeit) between series from 66% at the wet site to 69% at the moderate and dry sites. Correlations for the verification period (1989–2017) ranged from $r=0.37$ at the wet site to $r=0.51$ at the moderate site, and a glk between 64% at the wet site to 75% at the dry site.

According to VS-model simulations, the growing season length averaged from 132 ± 14 (mean \pm SD) at the wet site to 155 ± 12 days at the dry site over the 1960–2017 period (Table 2). The SoS was earlier at the dry site (DOY 121 ± 9) and later at the wet site (DOY 142 ± 12). The model highlighted a growth limitation by soil moisture from early summer (DOY 152 ± 9) to early autumn (DOY

Table 2 Modeled phenology of *P. sylvestris* by VS-model for the period 1960–2017

	SoS DOY (mean ± SD)	EoS DOY (mean ± SD)	LoS DOY (mean ± SD)
<i>W</i>	142 ± 12	273 ± 8	132 ± 14
<i>M</i>	133 ± 16	277 ± 8	144 ± 17
<i>D</i>	121 ± 9	276 ± 7	155 ± 12

SoS starting of growing season, *EoS* ending of growing season, *LoS* length of the growing season, *W* wet site, *M* moderate site, *D* dry site

Table 3 Vegetation phenology estimated from MODIS (MCD12Q2) dataset for the sampling sites during the common period 2001–2014 and modeled phenology of *P. sylvestris* by VS-model for the same period

	Site	SoS DOY (mean ± SD)	EoS DOY (mean ± SD)	LoS DOY (mean ± SD)
MODIS	<i>W</i>	130 ± 10	268 ± 9	137 ± 15
	<i>M</i>	128 ± 10	263 ± 12	136 ± 12
	<i>D</i>	117 ± 6	277 ± 5	161 ± 5
Modeled	<i>W</i>	144 ± 12	262 ± 25	118 ± 22
	<i>M</i>	123 ± 11	283 ± 7	161 ± 10
	<i>D</i>	118 ± 11	280 ± 7	162 ± 11

SoS starting of growing season, *EoS* ending of growing season, *LoS* length of the growing season, *W* wet site, *M* moderate site, *D* dry site

254 ± 11), whereas the importance of temperature goes from autumn to mid-spring (Fig. A3).

Validation of phenology simulations against remote sensing

MODIS-derived phenological dataset showed a longer growing season length at the dry site (161 ± 5; mean ± SD) and shorter at the moderate site (136 ± 12) for the 2001–2014 period (Table 3). During the same period, the VS-model simulated a longer growing season length at the dry site (162 ± 11) and shorter at the wet site (118 ± 22). One-way ANOVA did not reveal significant differences in mean SoS at any site ($P > 0.05$). However, mean EoS was significantly different at the moderate and wet sites ($F = 25.9$; $P < 0.05$ and $F = 5.3$; $P < 0.05$, respectively) and LoS at the moderate site ($F = 1.9$; $P < 0.05$). Pearson’s correlations between simulated and MODIS-derived SoS were positive at all the sites but only significant at the wet site ($r = 0.46$; $P < 0.05$). Whereas EoS correlations coefficient were significant only at the dry site ($r = -0.75$; $P < 0.001$; Fig. A4).

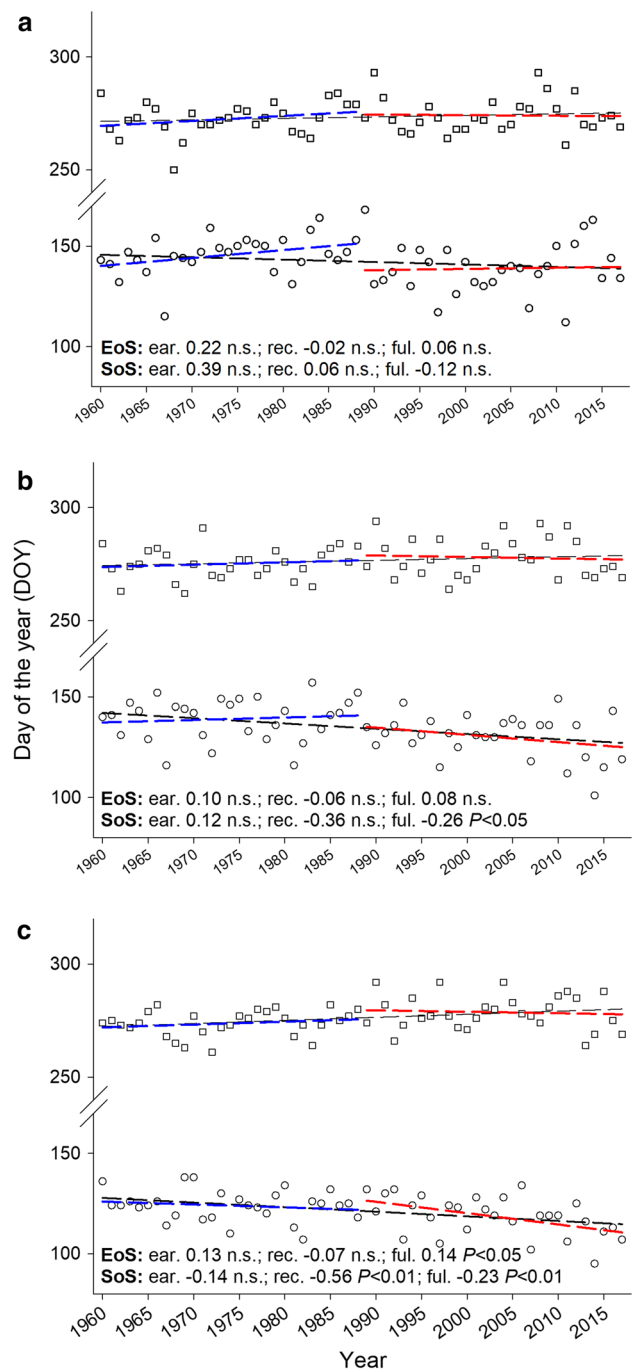


Fig. 4 Phenological trends simulated by the VS-model for the starting of the growing season (circles) and ending of the growing season (squares) and linear regressions for the full period (1960–2017, black-dashed lines), early period (1960–1988, blue-dashed lines) and recent period (1989–2017, red-dashed lines). Regression coefficients (b1) and significance are shown in the figures. Wet site (a), moderate site (b), dry site (c)

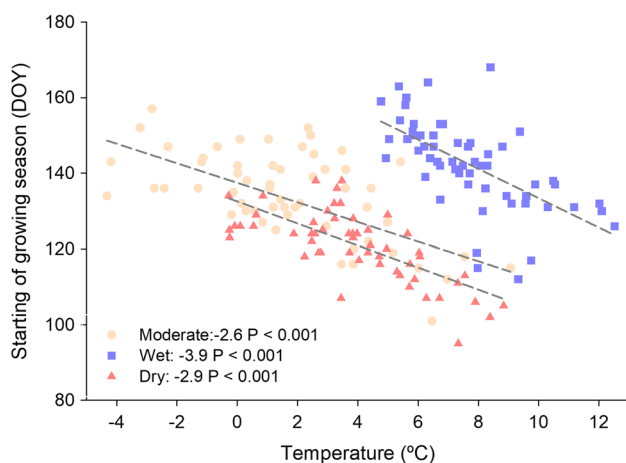


Fig. 5 Linear regression between the simulated starting of the growing season (SoS) with mean April (moderate and dry sites) and May (wet site) temperature for the period 1960–2017. Trends (gray-dashed lines) and regression coefficients (b1) and significance are shown in the figure

***Pinus sylvestris* phenological dynamics and response to climate**

Tree growth simulations revealed a trend of earlier SoS at all the sites over the 1960–2017 period (Fig. 4). Linear regressions suggested the advance of SoS at a rate from 2.3 days/decade at the dry site ($R^2=0.18$; $P<0.01$) to 2.6 days/decade at the moderate site ($R^2=0.15$; $P<0.05$). The dry site is the only showing a significant delay of EoS ($R^2=0.10$; $P<0.05$) at a rate of 1.4 days/decade. No significant trends were found at the wet site. When phenological trends were evaluated in the early and recent periods, the SoS advance increased to 5.6 days/decade from 1989 to 2017 at the dry site (Fig. 4).

Linear regression between simulated SoS and spring temperature over the full period (Fig. 5) suggested a strong response to April temperature at the moderate ($R^2=0.23$; $P<0.001$) and the dry ($R^2=0.49$; $P<0.001$) sites, whereas the wet site responded on May temperature ($R^2=0.39$; $P<0.001$). In this sense, a rise in spring temperature supposes the advance of SoS on an average rate of 3.1 days/°C along the studied transect. The VS-model also forecasted the lengthening of *P. sylvestris* growing season linked to rising temperatures by the end of the current century. Temperature increments between 4 and 5 °C might suppose an advance of SoS from 12 to 16 days at the moderate site and 16 to 17 days at the dry site.

Discussion

***Pinus sylvestris* growth and climate response**

Contrary to the common assumption that temperature is the primary driver of boreal forest growth (Nemani et al. 2003), pine secondary growth was mainly limited by water availability at the three sites. Precipitation and water balance signals echo previous findings in the region (Antonova and Stasova 1993; Arzac et al. 2018, 2021; Babushkina et al. 2019; Fonti and Babushkina 2016; Shah et al. 2015; Tabakova et al. 2020), highlighting the relevance of water availability for tree growth even at the coldest and wettest site. The shift in the timing and strength of climate sensitivity among the sites reflects an alleviation of drought stress with increasing latitude. In contrast, the positive response to water balance and precipitation (linked to a negative response to temperature) at the dry site suggests a strong limitation by temperature-induced drought (Arzac et al. 2021), being an adequate soil moisture level necessary for cambial reactivation and tree growth in drought-prone areas (Zeng et al. 2020). Our results also showed a decrease in the climate sensitivity to temperature and precipitation during recent years (1989–2017), in agreement with previous observations in the region (Arzac et al. 2018). However, the response to water balance increased at the dry site during the summer months in recent times, suggesting favorable conditions at the beginning of the growing season (spring) but harshened later due to summer drought. This condition may be exacerbated in future due to climate change, limiting tree growth. Differences among the sites might also be influenced by photoperiod (Körner and Basler 2010) and increasing radiation limitation due to cloudiness. Therefore, higher temperatures and lower radiation limitations (by lower cloudiness) might induce maximal growth at the driest site, despite the robust climatic control.

***Pinus sylvestris* phenology and climate**

The advance of SoS during the last five decades is in line with previous observations and simulations in northern latitudes (Jeong et al. 2011; Piao et al. 2015; Schwartz et al. 2006; Shen et al. 2015). Moreover, the average rate of 2.5 days/decade detected over the studied sites mimics the magnitude reported by previous studies based on model simulations (e.g., Yang et al. 2017) and remote sensing (e.g., Myneni et al. 1997) in boreal regions. In this sense, warmer springs induce a faster snow melting, increasing water and nutrients availability and promoting an earlier onset of the cambial activity (Dye and Tucker 2003; Kirilyanov et al. 2003; Peng et al. 2010; Zeng et al. 2020). Despite the robust

climatic control at the dry site, it also showed a longer growing season and tree growth, with clear dependence of SoS to spring warming (April). Therefore, our results might suggest that earlier growing seasons start at the dry site, due to warm springs, compensate for much of the expected growth reduction by low water availability levels later in the growing season (e.g., Shiklomanov et al. 2013).

VS-model simulations based on increasing spring temperatures projected earlier growing seasons start by the end of the century. Since the model defines the SoS based on temperature, it is possible to forecast potential realistic changes in the SoS. However, the limitation in daily climate forecast for the region difficult the prediction of growth rates and EoS dependent on temperature and soil moisture (as well as increasing drought limitation). Nevertheless, phenology changes linked to increasing temperature might substantially affect forest functioning and global carbon dynamics in future (Keenan et al. 2014; Piao et al. 2019). Longer vegetation seasons in global ecosystems (Zhu et al. 2016) could trigger photosynthetic activity and carbon assimilation (Keenan et al. 2014; Zhang et al. 2020), increasing forest productivity (Nemani et al. 2003; Schwartz et al. 2006; Zhang et al. 2020).

Pinus sylvestris has shown high physiological adaptability to contrasting environmental conditions. However, the warming-induced drought might also lead to future water deficits affecting forest health in the area (Arzac et al. 2018; Kharuk et al. 2013a, 2013b; Tabakova et al. 2020). Therefore, if the induced drought exceeds pine tolerance, the Russian boreal forest structure might be affected by forest-steppe expansion over taiga (Soja et al. 2007; Tchebakova et al. 2016). Thus, leading to a decrease in forest cover and potentially reducing the productivity and carbon assimilation capacity of the taiga ecosystem (Gustafson et al. 2011; Shvidenko et al. 2012). Therefore, precipitation and snowmelt/snowpack will be determining factors to keep the soil moisture balance for tree growth and phenology in future, as is evidenced by the increasing response to water balance at the dry site in recent time (Arzac et al. 2018, 2019; Babushkina et al. 2015; Grippa et al. 2005; Shen et al. 2015).

VS-model simulations and remote sensing

Some uncertainties might arise from the use of model simulations in combination with remote sensing observations. For example, the available MODIS dataset (2001–2014) might not be robust enough to evaluate temporal trends (Yang et al. 2017). In addition, the presence of gaps in the data set because of cloudiness might also induce errors when comparing both approaches. Observed differences in EoS at the moderate and the wet sites could respond to EVI's limited capacity to reproduce growing dynamics in the evergreen

needleleaf forest (Wu et al. 2014). A limitation exacerbated in mixed forests as at the wet site where the fraction of photosynthetically active radiation from deciduous species (*L. sibirica* and *B. pendula*) might overlap with *P. sylvestris*. In addition, a suboptimal performance of the model at the same site might respond to non-climatic factors, like the impact of insect outbreaks in the area (Safonova et al. 2019). However, their impact has been detected only in the last ten years.

Although the comparison of inter-annual patterns between modeled and remote sensing phenological series by Pearson's correlation was poor, mean SoS values were comparable, highlighting the model's potential to reconstruct phenological dynamics over long spatiotemporal scales (e.g., Jevšenak et al. 2020), particularly for the determination of the SoS (Tumajer et al. 2021). Nevertheless, the reconstruction of EoS and LoS dynamics based on the model might be uncertain, limiting the capacity of the model to reconstruct phenological dynamics beyond SoS. The VS-O has shown limitations determining the EoS compared to direct phenological observations (e.g., Butto et al. 2020; Jevsenak et al. 2020; Tumajer et al. 2021). A potential source of the model's poor performance may be the overly simple statistical approach to estimate the EoS, based on values of integral growth rate and a mechanistic parameter (V_{cr} and T_{beg}) which are constant for every year of simulation, and not considering non-climatic factors as hormones concentration and non-structural carbohydrates dynamics (Hartmann et al. 2017; Furze et al. 2019).

Conclusions

The combination of multiple data sources allowed us to disclose the effect of climate on *Pinus sylvestris* growth and fill gaps in xylem phenology (SoS dynamics) studies over extended spatiotemporal scales. Our findings suggest that *P. sylvestris* secondary growth is plastic enough to adapt to a certain threshold of harsh environmental conditions, adjusting its growth strategy to favorable climatic windows. The release of summer drought stress by earlier SoS, triggered by spring warming, might compensate for tree growth limitation by low water levels at the dry site later in the growing season. However, predicted rising temperatures and drought events in the southern Siberian forest-steppe might exceed the species' physiological tolerance, compromising tree growth and phenology. Altogether ultimately affecting the vitality, dynamics and distribution of Siberian forests, with critical ecological consequences on the global scale.

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Declarations

Conflicts of interest The authors have no conflicts of interest.

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