ORIGINAL PAPER

Myeloperoxidase-mediated oxidation of organophosphorus pesticides as a pre-step in their determination by AChE based bioanalytical methods

Tamara Lazarević Pašti · Tatjana Momić · Antonije Onjia · Ljubodrag Vujisić · Vesna Vasić

Received: 24 November 2009 / Accepted: 4 March 2010 / Published online: 31 March 2010 © Springer-Verlag 2010

Abstract In order to improve the sensitivity of assays for inhibitors of the enzyme acetylcholine esterase (AChE), an effective method was developed for the conversion of the organophosphate pesticides (OPs) diazinon, malathion, chlorpyrifos, azinphos-methyl and phorate into more toxic inhibitors. This was accomplished by converting them from the thio form into their oxo form using the enzyme myeloperoxidase. The oxo forms, which are the only products of conversion, were determined by AChE bioassays, using either the free enzyme, or a flow injection analysis manifold with immobilized AChE and spectrophotometric detection. All modified OPs exhibited inhibitory power at ppb levels and within 10 min. The method is considered to represent an excellent means for improving the sensitivity of assays for determination of OPs.

Keywords Organophosphate pesticides · Myeloperoxidase · Oxidation · Bioanalytics · Acethylcholinesterase

Introduction

Among various organic and inorganic pollutants, organophosphorus pesticides (OPs) widely used in agriculture are

Electronic supplementary material The online version of this article (doi:10.1007/s00604-010-0324-2) contains supplementary material, which is available to authorized users.

T. Lazarević Pašti · T. Momić · A. Onjia · V. Vasić Vinča Institute of Nuclear Sciences, PO Box 522, 11001 Belgrade, Republic of Serbia

L. Vujisić

Institute of Chemistry, Technology and Metallurgy, Njegoševa 12, PO Box 473, 11001 Belgrade, Republic of Serbia very dangerous and harmful because of their toxic nature [1]. Their toxic effects are attributed to the irreversible inhibition of the enzyme acethylcholinesterase (AChE) [1-5]. They also inhibit the other enzymes playing an important role in biochemical processes [6]. Moreover, phosphorothionate molecules undergo to chemical transformations after metabolic oxidation which leads to oxon derivatives (phosphates) formation, where the sulfur atom from the thionate group was replaced by an oxygen atom [2, 7, 8]. The products of thio OPs oxidation are more toxic than the original pesticides. However, OPs hydrolysis leads to detoxification, although the products of this process are not completely harmless [9]. Therefore, the close monitoring and removal of OPs from environment is necessary. Several methods, either independent or in conjunction with others, have been used for removal of OPs by adsorption [10–13] including also oxidation with ozone [14], chlorine [8] and biological degradation [15].

There is growing interest in the rapid and accurate monitoring and determination of OPs in environment. Besides complicated and time-consuming conventional tecniques, such as gas chromatography [16] and highperformance liquid chromatography (HPLC) [17-19], and more convenient multisyrige flow injection analysis (MSFIA) [20], electrochemical method [21], various cholinesterase (ChE) based bioanalytical methods and biosensors for in situ detection of OPs were developed [22–26]. The principle of the biosensors is based on the fact that the activity of cholinesterases is inhibited by OPs and the degree of inhibition is dependent on the concentration of anti-ChE in contact with the enzyme. However, oxidized forms of OPs show stronger inhibitory action towards AChE than parental, unoxidized compounds [4] and the limit of their detection depended often on the oxidation pre-step of the OPs containing sample [27-29].

The oxidations of OPs were performed either directly in water samples or in solvents which were diluted in working buffer prior the analysis. For this purpose Nbromosuccinimide (NBS) [30–33] or bromine [28, 29, 34, 35] were introduced as selective and rapid oxidants to enhance the inhibition in determination of thio OPs by biosensors. Moreover, other methods of oxidations were also developed, but were mainly intended for synthesis of oxons from thiophosphates using nitric acid, peroxide, peroxi acids, chlorine, ozone [36–40]. In these studies, the efficiency of oxidation reaction was unknown, and the possibility of enzyme inhibition by certain undesirable products was not addressed.

The metabolic activation of phosphorotioate molecules, which is performed by cytochrome P450 enzymes, yields oxon derivatives. Besides cytochrome P450 systems, peroxidative oxidation of OPs has also been documented with other enzymes, such as horseradish peroxidase [41], chlorperoxidase [42, 43] and soybean lipoxygenase [44]. Enzyme myeloperoxidase (MPO) acts as oxidant enzyme in the process of inflammation and atherogenesis [45]. MPO is relatively nonspecific with respect to its reducing substrates. This enzyme is able to oxidize different substrates among which anilines and phenols [46–48].

In this work the ability of MPO to biocatalyze the transformation of OPs (diazinon, malathion, chlorpyrifos,



Scheme 1 The chemical structures of relevant OP compounds

azinphos-methyl and phorate, Scheme 1) from thio to oxo forms was examined. The aim was to use MPO mediated oxidation of thio OPs to oxo OPs without any preconcentration or extraction step under normal laboratory conditions in order to improve the sensitivity of AChE based bioanalytical assays for their detection.

Experimental

Chemicals and materials

Myeloperoxidase (MPO) from human neutrophils, purified to a purity index (A₄₃₀/A₂₈₀) 0.84 was obtained from Planta Natural Products, Vienna, Austria (www.planta.at). Its concentration was calculated using $\varepsilon_{430}=91~000~\text{M}^{-1}\text{cm}^{-1}$ per heme [49]. Catalase, from bovine liver, acethylcholinesterase (AChE, specific activity 0.28 IU/mg) from electric eel, acethylthiocholine iodide (ASChI), 5,5'-dithio-bis-(2-nitrobenzoic acid) (DTNB) and controlled-pore glass (CPG 240, 80-120 mesh) were purchased from Sigma-Aldrich St. Louis, MO, USA (www.sigmaaldrich.com). Potassiumhydrogenphosphate (K_2 HPO₄×3H₂O), glutaraldehyde, 3-aminopropyltriethoxysilan, 2-hydroxyiminomethyl-1-methylpyridinium iodide (2-PAM) were purchased from Merck KgaA, Germany (www.merck.com). Hydrogen peroxide solutions were prepared daily by diluting a stock solution and the concentration was determined using ε_{240} =43.6 M⁻¹cm⁻¹ [50]. The used pesticides were of at least 93% purity and no further purification of chemicals was performed. Malathion, malaoxon, chlorpyrifos, azinphos-methyl and phorate (OPs), were purchased from Pestinal[®], Sigma-Aldrich, Denmark, whereas diazinon and diazoxon were purchased from Institute of organic industrial chemistry Poland (www. ipo.waw.pl) and Chemical Co (www.the chemco.com). The pesticide working solutions were prepared by dilution of the 1×10^{-3} M stock solutions in ethanol. The final working solutions contained 1 % ethanol to improve the solubility of the pesticides. The pesticide stock solutions were held in refrigerator until used. All chemicals were used without further purification. Deionizated water was used throughout.

0.05 M phosphate buffer, pH 8, was used as a carrier buffer in the FIA experiments. The reactivation of immobilized enzyme was done using 4 mM 2-hydroxyiminomethyl-1-methylpyridinium iodide (pralidoxime, 2-PAM) [25, 26, 51].

Oxidation procedure

The desired concentrations of organophosphates in the range from 10^{-4} – 10^{-7} M were incubated in 50 mM phosphate buffer, pH 6.0, with various concentrations of

MPO in final reaction volume 0.5-1.0 mL. Reaction was started by addition of 50 μ M H₂O₂, and stopped after incubation by adding catalase (100 μ g/mL). The reaction mixtures were centrifuged for 2 min at 10,000 rpm, to separate the supernatant for further analysis. At least three replicates were performed for each compound. The controls without MPO and/or hydrogen peroxide were carried out under the same conditions.

AChE immobilization

55 mg AChE was immobilized on 0.15 g activated controlled pore glass (CPG) according to the procedure described earlier [4, 5, 51]. Briefly, the glass particles activated with 3-aminopropyl-triethoxysilan three-times; afterwards, alkylamino glass was cross-linked with 2.5% glutaraldehyde. The enzyme was dissolved in 3 ml cold phosphate buffer and added to the pre-treated glass in a stream of nitrogen. After filtration, the immobilized enzyme was washed with phosphate buffer and then with cold water. The immobilized beads were stored at 4 °C in phosphate buffer (pH 6.0) until use.

Determination of AChE activity in the presence of OPs and their oxo forms

The inhibition of AChE by the pesticides both, oxidized and unoxidized, was measured using modified Ellman procedure [4, 52] in the absence (control) and presence of inhibitors. The experiments were performed by in vitro exposure of free enzyme (2.5 IU) to inhibitors in final volume 0.65 mL. The incubation time in the presence of inhibitors was 20–60 min, before the reaction was initiated and followed during 8 min. Acetylthiocholine iodide (ASChI) was applied as an enzyme substrate in combination with 5,5-dithio-bis-2-nitrobenzoic acid (DTNB) as a chromogenic reagent. The product 5-thio-2-nitrobenzoate, formed by the reaction of thiocholine (product of the enzymatic reaction) with DTNB, was measured spectrophotometrically at 412 nm (in buffer solution).

Simultaneously, the effect of the selected compounds before and after oxidation on immobilized AChE activity was investigated by using a flow-injection analysis (FIA) manifold [4, 25, 26, 51]. The inhibition of the enzyme was determined by comparing the enzyme activity before and after passage of 200 μ L pesticide solution of desired concentration through the bioanalytical column for a given period of time. The procedure was as follows: the carrier buffer (0.05 M phosphate, pH 8.0) was passed through the sensor system until the base line was stabilized and then the mixture consisting of ASChI and DTNB was injected through 200 μ L injection loop into the carrier stream and the response was recorded [4, 51]. The procedure was repeated at least three times, and the average value of the signal (initial activity) was used for calculations. Then the sample containing the pesticide was injected. The determination of the final enzyme activity was carried out by another injection of the substrate and the remaining enzyme activity was calculated as the ratio of the signal intensities before and after the injection of the sample.

All experiments were performed in triplicate, and the results present mean value \pm SE. Preliminary studies showed that OPs and their products of oxidation, or their combinations, did not interfere with quantification of the yellow product, 5-thio-2-nitrobenzoate.

FIA-AChE bioassay

The flow-injection system consisted of a HPLC pump (Dionex AMP-1), an injection valve (Waters U6K) equipped with a 200 μ L injection loop, detection unit (UV VIS detector, accuracy $\pm 2 \times 10^{-5}$ absorbance units at 412 nm) connected to PC and a bioanalytical reactor (21×3 mm peak column) filled with AChE immobilized on CPG. The carrier buffer (phosphate buffer, pH 8.0) was pumped through a flow-through cell. The flow rates which were used in these experiments were 0.1 mL/min and 0.5 mL/min.

Each bioanalytical column was used for several determinations of pesticide. 13 mg of glass beads with immobilized enzyme was weight before being filled in the bioanalytical column, making the analysis more reproducible, since the initial activity was the same all the time. The biosensor was daily calibrated with 3.4×10^{-6} M malaoxon, which induced about 50% inhibition and was afterwards reactivated with 2-PAM.

When the activity of the enzyme dropped to 90% of the initial value, enzyme was reactivated to its initial activity by five consecutive injections of a reactivator (4 mM 2-PAM). When the reactivation by 2-PAM could not reach the initial enzyme activity, the enzyme was replaced with the fresh one.

Apparatus

The spectrophotometric measurements were performed on Perkin Elmer Lambda 35 UV-Vis spectrophotometer, using 1 cm path length cuvette. pH measurements were performed using Metrohm pH meter, model 713 equiped with glass electrode.

Waters ACQUITY Ultra Performance Liquid Chromatography (UPLC) system coupled with a TUV detector controlled by the Empower software was used. Chromatographic separations were run on an ACQUITY UPLCTM BEH C₁₈, column 1.7 μ m, 100 mm×2.1 mm column (Waters). The analyses of diazinon and diazoxon were done under isocratic conditions with mobile phase consisting of solvent (A) CHCOOH (0.1 vol % in water) and (B) acetonitrile (J.T. Baker) (25:75, v/v). The elutions were monitored at 245 nm. The analysis of malathion and malaoxon were done under isocratic condition with mobile phase consisting of solvent (A) CHCOOH (0.1 vol % in water) and (B) acetonitrile (J.T. Baker) (60:40, v/v). The elutions were monitored at 210 nm. The eluent flow rate was 0.3 mL min⁻¹ and the injection volume was 10 μ L.

GC/MS analysis was performed on Agilent 7890A GC system equipped with 7000A QqQ MS, using DB-5 MS column (30 m×0.25 mm×0.25 μ m). Injection volume was 1 μ L and injector temperature was 250 °C with 10:1 split ratio. Carrier gas (He) flow rate was 1.3 ml/min at 80 °C (constant pressure mode). Column temperature was linearly programmed in a range of 80–300 °C at a rate of 10 °C/min. Transfer line was heated at 280 °C. Mass spectra obtain in EI multiple reactions monitoring (MRM) mode. Collision energy was 15 eV and collision gas was nitrogen. MRM mass spectra were acquired using characteristic transition for standards and samples.

Results and discussion

Reaction of OPs by MPO

In the study of MPO mediated chemical transformation of OPs malathion and diazinon were examined in more details. In the first series of experiments, the OPs reaction with MPO was performed by exposure 1×10^{-5} M water solution to 50 nM MPO during 5 min. The reaction of diazinon with MPO is schematically presented in Scheme 2.

The reaction products were analyzed by UPLC and GC/ MS measurements and chromatograms were compared with those of authentic standards. The identification of the reaction products yielded only one major product—the oxo form of pesticides. In the presence of MPO the sulfur atom was replaced by an oxygen atom in the organophosphorpous pesticides, transforming the phosphorothionate group to an oxon derivative (Scheme 2) [42]. Table 1 presents results obtained by GS/MS analysis of 1×10^{-5} M diazinon and malathion after oxidation. GS-MS/MS mass spectra were acquired in multi reaction mode (MRM) using characteristic transition for standards and samples. Presence of diazoxon and malaoxon in sample was confirmed comparing retention times and MS data of standards obtain under the same conditions. The results were also confirmed by UPLC analysis.

As an example, the UPLC chromatograms obtained for 1×10^{-5} M diazinon before and after incubation with MPO are presented in Fig. 1. Also, UPLC chromatograms of diazoxon that were recorded during 60 min, after the reaction was stopped by catalase, confirmed that there were no other products of conversion, i.e. a further cleavage of oxo form was not observed (Fig. 1) as found in the case of application of some inorganic oxidants [39]. It is worthily to notice that the formation of diazinon hydrolysis product, 2-isopropyl-6-methyl-4-pyrimidinol (IMP), which is usually found due to UV treatment of diazinon [6, 33], was not found in any of tested samples. These results are also in accordance with those obtained when chlorperoxidase was used for biocatalytic oxidation of some OPs. No hydrolysis and halogenated products were found in the chlorperoxidase oxidation of OPs [42, 43]. These oxidations are similar to those performed by cytochromes P450 in in vivo and in vitro systems. The mayor difference is, however, that a further cleavage of oxons, which is typical of P450 catalyzed reaction, was not observed with MPO, as well with chlorperoxidase. Oxidation by inorganic oxidants, such as aqueous chlorine led to hydrolysis of diazoxon, which was much slower process [8]. For oxidation of diazinon and other OPs, previous studies focused on ozonization and showed that the oxidation process was followed by further hydrolysis of diazoxon to IMP and diethyl phosphate [53].

AChE test for detection of OPs oxidation products

In order to find out the optimal conditions for conversion of OPs, from this to oxo form, the influence of MPO in the concentration range from 10–100 nM on 1×10^{-4} , 1×10^{-5} M





Table 1 Identification of OPs and their products of oxidation by GS/MS

m/z	^a RA	^b RT
179, 137, 273	100	13.625
137, 273	35	13.299
125, 127, 173	100	15.462
127	47	14.664
	m/z 179, 137, 273 137, 273 125, 127, 173 127	m/z aRA 179, 137, 273 100 137, 273 35 125, 127, 173 100 127 47

^a Relative abundance (RA)

^b Retention Time (RT)

and 1×10^{-6} M OPs during 1–30 min of incubation was investigated using AChE test. In these experiments the incubation time of oxidized sample with AChE was 20 min. It should be noticed that OPs were 10 fold diluted in AChE test, compared to initial untreated solution samples. The results obtained for diazinon and malathion are presented in Figure S1A,C and S1B,D (figures can be found in the Supporting Information), respectively, as the response of AChE (the percent of activity) to the oxidized and unoxidized OPs. The percent of AChE activity after exposure to ten fold diluted oxidized samples of diazinon and malathion, which were incubated with MPO for 5 min, decreased from 5-100% by increasing MPO concentration, depending also on the initial concentration of pesticide. However, the inhibition of AChE activity was ascribed to oxo form of related pesticides, since the inhibition induced by the same initial concentration (below 1×10^{-5} M) of the untreated parent compounds was almost negligible [4]. Based on the obtained results, 100 nM MPO was found as the optimal concentration to be used in further experiments. The AChE activity in the presence of the oxidized OPs compared to non oxidized samples decreased with the increasing the incubation time, indicating the increasing the concentration of oxo forms. The saturation level of oxo



Fig. 1 UPLC chromatograms of 1×10^{-5} M diazinon before (1) and after (2–5) 5 min of incubation with 50 nM MPO. Oxidized samples (2–5) were followed within 60 min. The time interval between two consecutive chromatograms was 20 min

forms was reached after 10 min incubation of parent compounds with MPO.

In further experiments, the AChE inhibition induced by thio OPs before and after oxidation together with their oxo forms, in the concentration range from 1×10^{-4} M-1× 10⁻⁹ M was evaluated. Results obtained for malathion and diazinon are presented in Fig. 2, as the examples. As can be seen from the inhibition studies, oxo forms exerted for about more than three orders of magnitude stronger inhibitory power, compared to the parent compounds under the same experimental conditions (Fig. 2a and b, curves 1 and 2). Further, OPs in the same concentration range were oxidized by MPO under optimal conditions (10 min incubation with 100 nM MPO). The inhibition of AChE activity obtained using solutions of oxidized OPs is presented in Fig. 2 (curves 3), for comparison. In these experiments the incubation time of OPs with AChE was prolonged to 60 min, in order to decrease the concentration level of OPs which induces the measurable decrease AChE activity (at least 10%).

It is also obvious that after the incubation of OPs with MPO, the inhibition curves were shifted to the lower OPs concentration range, compared to the inhibition curves of



Fig. 2 The concentration dependent inhibition of AChE activity induced by **a** diazinon (1) and **b** malathion (1), before (2) and after (3) exposure to 100 nM MPO during 10 min. The AChE activities are expressed as the mean % activity relative to the corresponding control value in the absence of OPs



Fig. 3 Plot of percent inhibition of AChE activity *vs* initial concentration of diazinon and malathion after 10 min oxidation with 100 nM MPO (incubation time between oxidized OPs and AChE was 60 min)

parent compounds due to the formation of oxo forms of OPs (Fig. 2, curves 3).

Determination of OPs using MPO mediated oxidation and AChE test

For the construction of calibration curves for diazinon and malathion, the assumption was made, that the AChE inhibition was achieved only due to the oxons formed after incubation with MPO. Plot of percent inhibition of AChE activity vs. initial OPs concentration was deduced from inhibition curves for oxidized OPs (Fig. 2, curves 3) and inhibition curves of oxones (Fig. 2, curves 2) and is presented in Fig. 3.

The results presented in Fig. 3 show that the concentrations, which induced measurable AChE activity (at least 10% of control) were below 4×10^{-9} M for malathion, and 2×10^{-7} M for diazinon.

Finally, AChE test was used to determine the concentrations of oxo forms as dependence on MPO concentration

Table 2 Influence of OPs before and after 10 min oxidation with100 nM MPO on AChE activity

OP Initial concentration (M)	Activity (% of control)			
	Unoxidized 1×10^{-5}	Oxidized		
		1×10^{-7}	1×10^{-6}	1×10 ⁻⁵
Diazinon	100	95±5	65±8	20±4
Malathion	100	20±2	0	0
Phorate	100	88±4	45±5	0
Azinphos-methyl	96	81±6	20±7	0
Chlorpyrifos	100	60±5	30±6	10 ± 2
Syntetic mixture	100	75±4	0	0



Fig. 4 FIA signals at flow rate 0.5 mL/min before (control) and after enzyme inhibition by oxidized 1×10^{-7} M OPs

and oxidation time. The obtained results were in agreement with those obtained using conventional analytical techniques, as UPLC. The UPLC calibration graphs for diazinon and diazoxon were constructed (data not shown) to evaluate the adequate concentrations of parent compound and oxo form upon the oxidation of diazinon. Table S1 (table can be found in the Supporting Information) shows the results obtained for evaluation of diazoxon concentration as dependence on MPO concentration. It is obvious that the results obtained from AChE test and UPLC determination for initial diazinon concentration above 1×10^{-6} M are in the range of experimental error. However, diazoxon concentration in the oxidized sample which contained below $1\!\times\!10^{-6}~M$ diazinon before oxidation could not be measured without sample pre-concentration. Moreover, AChE test appeared to be faster and simpler in this case.

The oxidation efficiency as the function of oxidation time was expressed as the oxidation yield based on determined concentrations of oxo forms relative to the initial concentration of the parent compounds. The results for oxidation efficiency as the function of oxidation time

Table 3 Influence of flow rate on the response of sensor enzyme in FIA bioanalytical assay for $1\!\times\!10^{-7}$ M unoxidized and oxidized OPs

Flow rate (mL/min)	Signal intensity (% of control)			
	0.1	0.5	0.1	
OPs	Unoxidized	Oxidized		
Diazinon	100	78±5	70±7	
Malathion	100	65±5	59±2	
Phorate	100	84±4	77±3	
Azinphos-methyl	96	80±3	80±5	
Chlorpyrifos	100	67±3	57±4	
^a Syntetic mixture	100	80±5	72±2	

 $^{a}0.2 \times 10^{-7}$ M of each mentioned OP

for diazinon and malathion, obtained using AChE test and UPLC measurements, are given in Table S2 (table can be found in the Supporting Information). However, the conclusion can be made that the conversion efficiency increased by lowering diazinon concentration. These results can be also compared to those obtained by treatment of several OPs with 10 molar excess of bromine in acetonitrile [29]. This method yielded about 82–100 % oxons. Further, the ozonization of OPs in aqueous solutions led to the complete degradation within 1 h [54]. N-bromsuccinimide was also introduced as rapid and effective oxidant to enhance AChE inhibition in determination of OPs. However, the complete assays could not be performed in less than 2 h [30, 32, 55].

AChE test as described above was also applied to the other OPs (phorate, azinphos-methyl, chlorpyrifos) and their synthetic mixture in tap water, before and after 10 min oxidation with 100 nM MPO. The single initial concentrations were 1×10^{-5} M, 1×10^{-6} M and 1×10^{-7} M, and the synthetic mixtures contained all individual components in the mixture. The results are presented in Table 2. All the oxidized OPs showed the greater percentage inhibition than the corresponding non oxidized samples.

These results show that the chosen method for OPs oxidation can be used to increase the sensitivity of the AChE based enzymatic and sensor analysis of OPs.

The effects of OPs oxidation on the activity of immobilized AChE

The response of the bioanalytical FIA manifold, using immobilized AChE to unoxidized and oxidized OPs was followed in 1×10^{-7} M aqueous solutions of diazinon, malathion, phorate, azinphos-methyl, chlorpyrifos, and also in the synthetic mixture containing 1×10^{-7} M of all components as the sum of their initial concentrations $(0.2 \times$ 10^{-7} M). In the present study, the flow rates were 0.5 mL/ min and 0.1 mL/min. The typical FIA signals obtained for oxidized 1×10^{-7} M malathion, diazinon, phorate, chlorpyrifos and azinphos-methyl at flow rate 0.5 mL/min are presented in Fig. 4. Moreover, the unoxidized 1×10^{-7} M OPs did not induce any inhibition of immobilized AChE (Table 3). After inhibition by each single oxidized OP, AChE was reactivated by 2-PAM, at flow rate 0.1 mL/min to restore the initial AChE activity (control value). As the results presented in Fig. 4 show, the same value of control signals was achieved after each reactivation of sensor enzyme.

As a detection limit, the lowering of the control signal for about 10% was considered. It is worthily to notice, that the lower flow rate induced the higher value of the control signals for about 10%, but the intensity of signals after injection of OP solutions decreased (Table 3). The lower flow rate ensured the prolonged contact time between OPs and sensor enzyme. Based on this finding, the lower flow rate leads to increasing of the FIA manifold detection limit (lowering of the control signal for at least 10 %) for OPs.

The oxidation of OPs for their determination in FIA system was studied by some authors. Chlorperoxidase in citrate buffer was used to oxidize OPs in determination of OPs using bioanalytical assay based on inhibition of AChE which was determined in FIA system by thermal lens spectroscopy [42, 43]. The improved signal was obtained in ionic liquids which were tolerated by the enzyme up to 30% [43]. Bromine was also used as the in situ oxidant for aqueous sample solutions and oxidizes samples were injected directly into FIA system. The efficiency of oxidation was unknown, and the possibility of AChE inhibition by certain undesirable products, including bromine species, was not addressed [28, 29, 34, 35]. The percent of AChE inhibition through use of oxidized OPs increased to between 1.6–18 fold.

The comparison of percent inhibition induced by unoxidized and oxidized 1×10^{-7} M OPs and the standard synthetic mixture, at flow rates 0.5 mL/min and 0.1 mL/min is shown in Table 3. Based on this finding, the lower flow rate leads to decreasing of the detection limit of OPs. It is obvious from Fig. 4 that all investigated OPs at concentration 1×10^{-7} M induced more than 10% inhibition of sensor enzyme activity. Table 3 also presents the detection limits obtained for each single OP at flow rate 0.1 mL/min.

Conclusion

This study demonstrates the applicability of enzyme MPO for the oxidation of OPs as the useful method to improve the sensitivity of AChE-based bioanalytical assays using free or immobilized enzyme. The degree of OPs oxidation depends on the OPs and MPO concentrations, as well as on incubation time between OPs and MPO.

The 10 min incubation time of OPs in phosphate buffer pH 6.0 with 50–100 nM MPO was found to be applicable for OPs oxidation, to produce oxo forms with stronger inhibitory power towards AChE. Since the detection limit of immobilized AChE for OPs after oxidation was below 1×10^{-7} M, the sensitivity of the bioanalytical assay using preoxidation step with MPO is satisfactory. Reactivation of the inhibited enzyme in FIA manifold with 2-PAM restored the enzyme activity almost completely, allowing the repeated use of this sensor for OPs analysis.

The developed procedure might be applied as a convenient screening method for determination of OPs in selected real samples even without complicated clean-up steps, which save time, money and the environment. Acknowledgments The authors are grateful to Dr A. Tomašević from the Pesticide and Environment Research Institute, Belgrade, Republic of Serbia, for providing OPs. We wish to express our gratitude to the Ministry of Science and Environmental Protection of the Republic of Serbia supported this work.

References

- Casida J (2009) Pest toxicology: the primary mechanisms of pesticide action. Chem Res Toxicol 22:609–619
- Kulkarni AP, Hodgson E (1980) Metabolism of insecticides by mixed function oxidase systems. Pharmacol Therapeut 8:379–475
- Sultatos LG (1994) Mammalian toxicology of organophosphorous pesticides. J Toxicol Environ Health 43:271–289
- Krstic D, Colovic M, Bavcon Kralj M, Franko M, Krinulovic K, Trebse P, Vasic V (2008) Inhibition of AChE by malathion and some structurally similar compounds. J Enz Inh Med Chem 23:562–573
- Krstic D, Colovic M, Bavcon Kralj M, Franko M, Krinulovic K, Trebse P, Vasic V (2008) The influence of malathion and its decomposition products on free and immobilized acetylcholinesterase. Russ J Phys Chem A 82:663–668
- Colovic M, Krstic D, Petrovic S, Leskovac A, Joksic G, Savic J, Franko M, Trebse P, Vasic V (2010) Toxic effects of diazinon and its photodegradation products. Toxicol Lett 193:9–18
- Fallscheer H, Cook J (1956) Report on enzymatic methods for insecticides. J Assoc Offic Agric Chem 39:691–697
- Zhang Q, Pehkonen SO (1999) Oxidation of diazinon by aqueous chlorine: kinetics, mechanisms, and product studies. J Agric Food Chem 47:1760–1766
- 9. Shemer H, Linden KG (2006) Degradation and by-product formation of diazinon in water during UV and UV/H_2O_2 treatment. J Hazard Mater 136:553–559
- Gupta VK, Ali I (2001) Removal of DDD and DDE from wastewater using bagasse fly ash, a sugar industry waste. Water Res 35:33–40
- Gupta VK, Jain CK, Ali I, Chandra S, Agarwal S (2002) Removal of lindane and malathion from wastewater using bagasse fly ash–a sugar industry waste. Water Res 36:2483–2490
- Gupta VK, Ali I, Suhas JS, Saini VK (2006) Adsorption of 2, 4-D and carbofuran pesticides using fertilizer and steel industry wastes. J Colloid Interf Sci 299:556–563
- Gupta VK, Ali I (2008) Removal of endosulfan and methoxychlor from water on carbon slurry. Environ Sci Technol 42:766–770
- Beltran FJ, García-Araya JF, Acedo B (1994) Advanced oxidation of atrazine in water–I. Ozonation. Water Res 28:2153–2164
- Kanekar PP, Bhadbhade BJ, Deshpande NM, Sarnaik SS (2004) Biodegradation of organophosphorus pesticides. Proc Indian natn Sci Acd B 70:57–70
- 16. Kouloumbos V, Tsipi D, Hiskia A, Nikolic D, van Breemen R (2003) Identification of photocatalytic degradation products of diazinon in TiO₂ aqueous suspensions using GC/MS/MS and LC/ MS with quadrupole time-of-flight mass spectrometry. J Am Soc Mass Spectrom 14:803–817
- Pérez-Ruiz T, Martínez-Lozano C, Tomás V, Martín J (2005) High-performance liquid chromatographic assay of phosphate and organophosphorus pesticides using a post-column photochemical reaction and fluorimetric detection. Anal Chim Acta 540:383–391
- Abu-Qare A, Abou-Donia M (2001) Simultaneous determination of malathion, permethrin, DEET (N, N-diethyl-m-toluamide), and their metabolites in rat plasma and urine using high performance liquid chromatography. J Pharm Biomed Anal 26:291–299
- 19. Kaur I, Mathur RP, Tandon SN (1997) Parameters affecting the decay of some organophosphorus pesticides: a study by high-

performance liquid chromatography. Biomed Chromatogr 11:22-24

- 20. Ornelas-Soto N, Guzmán-Mar J, López de Alba P, López Martínez L, Barbosa-García O, Cerdà Martín V (2009) Coupled multisyringe flow injection/reactor tank for the spectrophotometric detection of azinphos methyl in water samples. Microchim Acta 167:273–280
- 21. Ly S (2008) Assay of diazinon pesticides in cucumber juice and in the deep brain cells of a live carp. Microchim Acta 163:283–288
- 22. Arduini F, Ricci F, Tuta CS, Moscone D, Amine A, Palleschi G (2006) Detection of carbamic and organophosphorous pesticides in water samples using a cholinesterase biosensor based on Prussian Blue-modified screen-printed electrode. Anal Chim Acta 580:155–162
- Mulchandani A, Mulchandani P, Chen W (1998) Enzyme biosensor for determination of organophosphates. Field Anal Chem Tech 2:363–369
- Trettnak W, Reininger F, Zinterl E, Wolfbeis OS (1993) Fiberoptic remote detection of pesticides and related inhibitors of the enzyme acetylcholine esterase. Sensor Actuat B-Chem 11:87–93
- 25. Pogacnik L, Franko M (1999) Determination of organophosphate and carbamate pesticides in spiked samples of tap water and fruit juices by a biosensor with photothermal detection. Biosens Bioelectron 14:569–578
- Pogacnik L, Franko M (2003) Detection of organophosphate and carbamate pesticides in vegetable samples by a photothermal biosensor. Biosens Bioelectron 18:1–9
- Li PCH, Swanson EJ, Gobas FAPC (2002) Diazinon and its degradation products in agricultural water courses in British Columbia, Canada. Bull Environ Contam Toxicol 69:59–65
- Kumaran S, Tran-Minh C (1992) Determination of organophosphorous and carbamate insecticides by flow injection analysis. Anal Biochem 200:187–194
- Kim YA, Lee HS, Park YC, Lee YT (2000) A convenient method for oxidation of organophosphorus pesticides in organic solvents. Environ Res 84:303–309
- Schulze H, Schmid R, Bachmann T (2002) Rapid detection of neurotoxic insecticides in food using disposable acetylcholinesterasebiosensors and simple solvent extraction. Anal Bioanal Chem 372:268–272
- Barceló D, Lacorte S, Marty JL (1995) Validation of an enzymatic biosensor with liquid chromatography for pesticide monitoring. Trends Anal Chem 14:334–340
- Schulze H, Vorlová S, Villatte F, Bachmann TT, Schmid RD (2003) Design of acetylcholinesterases for biosensor applications. Biosens Bioelectron 18:201–209
- 33. Bavcon Kralj M, Cernigoj U, Franko M, Trebse P (2007) Comparison of photocatalysis and photolysis of malathion, isomalathion, malaoxon, and commercial malathion–Products and toxicity studies. Water Res 41:4504–4514
- Kumaran S, Morita M (1995) Application of a cholinesterase biosensor to screen for organophosphorus pesticides extracted from soil. Talanta 42:649–655
- 35. Lee HS, Kim YA, Cho YA, Lee YT (2002) Oxidation of organophosphorus pesticides for the sensitive detection by a cholinesterase-based biosensor. Chemosphere 46:571–576
- Michalski J, Okruszek A, Stec W (1970) Stereochemistry of oxidation of organophosphorus thiono-compounds and P^{III} compounds by nitric acid and dinitrogen tetroxide. Chem Commun D 1495–1497
- 37. Stec WJ, Okruszek A, Michalski J (1976) Organophosphorus compounds of sulfur and selenium. Stereochemistry of oxidation of thiono- and selenophosphoryl compounds with hydrogen peroxide. J Org Chem 41:233–238
- Herriott AW (1971) Peroxy acid oxidation of phosphinothioates, a reversal of stereochemistry. J Am Chem Soc 93:3304–3305

- 39. Skowronska A, Krawczyk E (1983) A general method for the conversion of thiophosphoryl and selenophosphoryl groups into phosphoryl groups by ozone oxidation. Synthesis 6:509–510
- 40. Kamel A, Byrne C, Vigo C, Ferrario J, Stafford C, Verdin G, Siegelman F, Knizner S, Hetrick J (2009) Oxidation of selected organophosphate pesticides during chlorination of simulated drinking water. Water Res 43:522–534
- Knaak JB, Stahmann MA, Casida JE (1962) Insecticide metabolism in plants, peroxidase and ethylenediaminetetraacetic acidferrous iron-catalyzed oxidation and hydrolysis of parathion. J Agric Food Chem 10:154–158
- 42. Hernandez J, Robledo NR, Velasco L, Quintero R, Pickard MA, Vazquez-Duhalt R (1998) Chloroperoxidase-mediated oxidation of organophosphorus pesticides. Pestic Biochem Phys 61:87–94
- 43. Boškin A, Tran C, Franko M (2009) Oxidation of organophosphorus pesticides with chloroperoxidase enzyme in the presence of an ionic liquid as co-solvent. Environ Chem Lett 7:267–270
- 44. Naidu AK, Naidu AK, Kulkami AP (1991) Role of lipoxygenase in xenobiotic oxidation: parathion metabolism catalyzed by highly purified soybean lipoxygenase. Pestic Biochem Physiol 41:150–158
- 45. Klebanoff SJ (1999) Myeloperoxidase. P Assoc Am Physician 111:383–389
- 46. Shacter E, Lopez RL, Pati S (1991) Inhibition of the myeloperoxidase-H₂O₂-Cl- system of neutrophils by indomethacin and other non-steroidal anti-inflammatory drugs. Biochem Pharmacol 41:975–984

- Kettle AJ, Gedye CA, Hampton MB, Winterbourn CC (1995) Inhibition of myeloperoxidase by benzoic acid hydrazides. Biochem J 308:559–563
- Momic T, Savic J, Vasic V (2009) Oxidation of quercetin by myeloperoxidase. Ress Lett Phys Chem 2009
- Odajima T, Yamazaki I (1970) Myeloperoxidase of the leukocyte of normal blood. I. Reaction of myeloperoxidase with hydrogen peroxide. Biochim Biophys Acta 206:71–77
- Beers RJ, Sizer IW (1952) A spectrophotometric method for measuring the breakdown of hydrogen peroxide by catalase. J Biol Chem 195:133–140
- Pogacnik L, Franko M (2001) Optimisation of FIA system for detection of organophosphorus and carbamate pesticides based on cholinesterase inhibition. Talanta 54:631–641
- Ellman GL, Courtney KD, Andres V jr, Featherstone RM (1961) A new and rapid colorimetric determination of acetylcholinesterase activity. Biochem Pharmacol 7:88–90, IN81, 91-95
- Ohashi N, Tsuchiya Y, Sasano T, Hamada A (1994) Ozonation products of organophosphorous pesticides in water. Jpn J Toxicol Environ Health 40:185–192
- Young K, Chang J-L, Shen Y-S, Lin S-Y (1998) Decomposition of diazinon in aqueous solution by ozonation. Water Res 32:1957–1963
- 55. Bavcon Kralj M, Trebše P, Franko M (2006) Oxidation as a prestep in determination of organophosphorus compounds by the AChE-TLS bioassay. Acta Chim Slov 53:43–51