**RESEARCH PAPER** 



# The seasonal changes of core bacterial community decide sewage purification in sub-plateau municipal sewage treatment plants

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#### Abstract

The decline of sewage purification efficiency in winter is a frequent problem in sub-plateau municipal sewage treatment plants (MSTPs). Understanding the links between activated sludge (AS) bacterial community and sewage purification is crucial for exploring the cause of this problem. In this study, Illumina high-throughput sequencing technology was applied to investigate the seasonal changes of AS bacterial community in sub-plateau MSTPs. The sequencing result indicates that the bacterial community OTU number, diversity, and relative abundance in winter are significantly lower than that in summer samples. The discriminant linear effect size analysis (LEfSe) reveals that *Proteobacteria* and *Chloroflexi* members were enriched in summer AS, while *Actinobacteria* and *Firmicutes* were enriched in winter AS. The results indicate that different core bacterial community assembly was developed in summer and winter, respectively. The changes in bacterial community may be the reasons for the lower sewage purification efficiency in winter. Furthermore, redundancy analysis (RDA) shows that temperature and dissolved oxygen (DO) are the principal factors that drive the seasonal changes in the core bacterial community diversity, richness and structure in sub-plateau MSTPs. Thus, the sub-plateau AS selects for a unique community assembly pattern and shapes the particular AS ecosystem. These results expand previous understanding and provide insight into the relationship between bacterial community and performance of sub-plateau MSTPs.

Keywords Sub-plateau MSTPs · Activated sludge · Core bacterial community · Sewage purification

# Introduction

The sub-plateau region is defined as the transition zone from the Tibet plateau to the loess plateau in northwest China, where the average altitude ranges from 1000–2500 m and

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environmental ecosystems are fragile. Therefore, the local municipal sewage treatment plants play an important role in protecting the aquatic ecological system, especially as the headstream of Yangtze river and Yellow river. However, the efficiency of sewage purification in sub-plateau municipal sewage treatment plants (MSTPs) is clearly lower than that in plain MSTPs, and obviously decreased in winter season [13, 28].

Recent studies showed that microbial community structure and diversity were major factors related to nutriment removal and process stability in engineered microbial systems [8, 10, 21, 26]. While activated sludge (AS) bacterial community structure in different regions has significant geographical difference [9, 29]. It may be assumed that the local environmental variables such as the influent microbiota, temperature, altitude, dissolved oxygen (DO), and operational processes had remarkable effect on the AS microbial community structure [19, 21]. Previous study has shown that influent wastewater was a source of AS bacterial community populations, but there was a significant difference between influent and AS [25]. In addition, other research indicated that the environmental and operational factors play a greater role in shaping AS community than influent [16]. For example, DO concentration significantly affected the microbial community structure and the efficiency of nutrients removal [14, 23], temperature also evidently affected the microbial community structure, activity, diversity, and process performances [5], and altitude was an important factor influencing denitrifying bacteria community [19]. Like that, we can be supposed that the unique AS microbial community presents in subplateau MSTPs and the efficiency of sewage purification links with them. And now, no relevant literature focused on AS microbial community structure and functions in sub-plateau MSTPs. This study attempts to clarify the AS bacterial community structures and seasonal changes in sub-plateau MSTPs, which would be beneficial to explore the reasons for the poor performance in winter, and guide to develop the appropriate operation strategy for sub-plateau MSTPs.

In this study, 24 AS samples were collected from four full-scale sub-plateau MSTPs. Bacterial community diversity, structure, and relative abundance between summer and winter AS samples were measured by Illumina MiSeq sequencing. The objectives of this study were (1) to reveal the core AS bacterial community in sub-plateau MSTPs, (2) to compare the differences in bacterial community diversity, richness, and structure between summer and winter AS samples, (3) to unravel the correlation between bacterial community and sewage purification efficiency.

## Materials and methods

### **MSTP introduction and sample collection**

AS samples were collected from the aeration tanks of the four full-scale MSTPs, which basic information is listed in Table 1. During the sampling period, the MSTPs were in good condition and working in a steady state, though a slight sludge foaming was observed in the aeration tanks in winter. Triplicate samples for each tank were collected in summer and winter, respectively. Each AS sample was collected using a 500-mL sterilized polyethylene bottle, then centrifuged at 5000 rpm for 15 min and the sediments stored at -80 °C before DNA extraction. At the same time, influent and effluent quality indices, including chemical oxygen demand (COD), ammonium (NH<sub>4</sub><sup>+</sup>-N), and total phosphorus (TP), were measured by total carbon and nitrogen analyzer (Multi, N/C2100, Analytik Jena) and water quality analyzer (DR3900, Hach, American). The data are shown in Table 2.

#### **DNA extraction and Illumina MiSeq sequencing**

A total of 24 samples were collected, and about 5 g of each AS sample was used to extract genomic DNA by the Fast DNA<sup>®</sup> SPIN Kit for Soil (MP Biomedicals, CA, USA).

Table 1	Characteristics of four s	sampling sub-plateau MSTPs in northwest China											
MSTP	Treatment process <sup>a</sup>	Location	Altitude	$T(^{\circ}C)$		DO <sup>b</sup> (mg/L	(	SRT (d		HRT (h		MLSS (r	ng/L)
			(m)	Sum <sup>c</sup>	Win	Sum	Win	Sum	Win	Sum	Win	Sum	Win
AN	A <sup>2</sup> /O	Lanzhou Anning $(36.11^{\circ} \text{ N}, 103.67^{\circ} \text{ E})$	1540	$17 \pm 4$	$11 \pm 3$	$2.4 \pm 0.2$	$1.8 \pm 0.4$	11	14	6	10	3886	2357
XG	A <sup>2</sup> /O	Lanzhou Xigu (36.10° N, 103.75° E)	1578	$18\pm4$	$12 \pm 3$	$2.2\pm0.3$	$1.7 \pm 0.5$	10	12	12	14	4962	3520
ΖT	CAST	Lanzhou Chengguan (36.08° N, 103.87° E)	1532	$19\pm 3$	$13 \pm 4$	$2.1 \pm 0.3$	$1.9\pm0.3$	12	15	12	13	5470	2789
XN	$A^{2}/O + JSBC$	Xining Dongcheng (36.58 $^{\circ}$ N, 101.87 $^{\circ}$ E)	2186	$12 \pm 4$	$7\pm 2$	$1.9 \pm 0.3$	$1.6\pm0.3$	15	17	11	14	3199	1964

<sup>1</sup>Treatment process abbreviation: A<sup>2</sup>/O anaerobic/anoxic/oxic, CAST cyclic activated sludge technology, JSBC Joint System Biological Contact

"DO dissolved oxygen concentration, SRT sludge retention time, HRT hydraulic retention time, MLSS mixed liquid suspended solids

Sum summer, Win winter

Table 2 Influent, effluent quality, and nutriments removal efficiency in winter and summer seasons

MSTP	Sample code	Season	Influent (mg/L)			Effluent (mg/L)			Removal efficiency (%)		
			COD	NH4 <sup>+</sup> -N	ТР	COD	NH4 <sup>+</sup> -N	ТР	COD	NH4 <sup>+</sup> -N	ТР
AN	ANW	Win	$362 \pm 89$	17.4±10.3	$4.4 \pm 1.2$	$58 \pm 15$	3.17±1.1	$0.99 \pm 0.3$	83.9	81.8	77.6
	ANS	Sum	$346 \pm 105$	$23.2 \pm 5.1$	$4.2 \pm 1.3$	$53 \pm 16$	$2.25 \pm 1.4$	$0.76 \pm 0.2$	84.6	90.3	81.9
XG	XGW	Win	$371 \pm 113$	$15.9\pm8.2$	$3.6 \pm 1.7$	$57 \pm 11$	$2.86 \pm 0.8$	$0.77 \pm 0.2$	84.6	82.0	78.7
	XGS	Sum	$323 \pm 140$	$25.5\pm6.2$	$3.7 \pm 1.7$	$38 \pm 10$	$2.13 \pm 1.0$	$0.70 \pm 0.3$	86.9	91.6	81.2
ZT	ZTW	Win	$384 \pm 107$	$18.3 \pm 7.7$	$4.6 \pm 1.4$	49 <u>±</u> 13	$2.24 \pm 1.0$	$1.07 \pm 0.3$	87.2	87.8	76.7
	ZTS	Sum	363 ± 116	$26.4 \pm 5.4$	$4.3 \pm 1.5$	$34 \pm 10$	$1.91 \pm 1.2$	$0.88 \pm 0.2$	90.6	92.8	79.6
XN	XNW	Win	$352 \pm 148$	$13.6 \pm 6.3$	$3.1 \pm 2.3$	$40 \pm 13$	$4.35 \pm 2.6$	$0.89 \pm 0.4$	88.6	68.0	71.3
	XNS	Sum	$355 \pm 124$	$20.1 \pm 3.3$	$3.5 \pm 1.8$	$38 \pm 12$	$2.81 \pm 1.2$	$0.70\pm0.2$	89.3	86.0	80.0

After performing the assay for concentration and purity DNA, the V3–V4 regions (~465 nucleotides) of bacterial 16S rRNA gene were amplified with primers 338F (5'-ACTCCTACGGGAGGCAGCAG-3') and 806R (5'-GGA CTACHVGGGTWTCTAAT-3') [6]. Finally, the purified PCR amplicons were paired-end sequenced on the Illumina MiSeq PE300 platform (Illumina, USA) in Majorbio Co., Ltd (Shanghai, China). All 16S rDNA sequences have been deposited into the NCBI Sequence Read Archive (SRA) database with Bioproject accession number PRJNA551365.

#### 16S gene sequence analysis

Raw sequences were quality filtered and clustered into operational taxonomic units (OTUs) by USEARCH 7.0 (https:// drive5.com/uparse/) [3, 7]. Then the OTUs were classified into exact bacterial taxonomy against the SILVA database (https://www.arb-silva.de) and Ribosomal Database Project (RDP, https://rdp.cme.msu.edu/) by Bayesian classifier at 70% threshold. Based on the OTUs information, estimators of community diversity (Shannon and Simpson) and community richness (Chao and Ace), principal co-ordinates analysis (PCoA), and redundancy analysis (RDA) were performed using the online Majorbio I-Sanger Cloud Platform (www.i-sanger.com). Partial least squares discriminant analysis (PLS-DA) was used to find resultant variables between sample groups. LEfSe method uses the Kruskal-Wallis test to identify features with significant differences between winter and summer samples, and LDA to evaluate the effect of size on each feature [9, 20]. An LDA threshold score of 4.0 and a significant  $\alpha$  of 0.05 was used to detect biomarkers.

#### Statistical analysis

All statistical analyses were performed on SPSS 23.0 (SPSS Inc., Chicago, USA). The p value < 0.05 was statistically significant. Student's t test was used to compare the differences in bacterial diversity and functional genera between winter and summer samples. One-way analysis of variance

(ANOVA) and Tukey's HSD comparisons were used to determine statistically significant differences among different groups. Correlations among core bacterial community richness, diversity, and environmental variables were considered as statistically significant if p value < 0.05 and the Spearman's rank correlation coefficients (SRCCs)  $\geq$  0.5 or  $\leq -0.5$ . All data are presented as the means of triplicate samples (n=6) and expressed as the mean $\pm$  SE. The relevant column charts reflecting community structures were drawn by Microsoft Excel 2016.

#### **Results and discussion**

# The core AS bacterial communities in sub-plateau MSTPs

The qualified 16S rRNA gene sequences were aligned against the SILVA database and affiliated into different taxonomic levels at 97% similarity. A total of 46 phyla were identified from the AS samples. The dominated bacterial phyla (relative abundance > 1%) account for 93.98–99.25% of the total community abundance (Fig. 1a). Among them, Proteobacteria was the most abundant members (26.10-34.62%), followed by Actinobacteria (13.74–27.78%), Bacteroidetes (18.72–21.04%), Chloroflexi (7.67%-18.52%), and Firmicutes (4.06-6.88%) in summer samples. But in winter samples, Actinobacteria was the most abundant members (31.39-41.59%), followed by Proteobacteria (18.23-22.94%), Bacteroidetes (11.14-27.80%), Firmicutes (7.64–11.83%), and Chloroflexi (3.92–8.82%). The rest of the 35 phyla (relative abundance < 1% in each sample) were defined as "other phyla" accounting for 0.75-6.02%. It is clear that the relative abundance of bacterial phylum community varied greatly between the summer and winter samples, but the core bacterial phylum community structure remained stable to some extent. Furthermore, the core bacterial phylum community structure was shared with various geographical AS samples from the plain, plateau regions

Fig. 1 Bacterial community composition of core bacterial phyla (>1%) a classified genera (top 30). b Note: ANS-samples from AN MSTP in summer, XGS—samples from XG MSTP in summer, ZTS-samples from ZT MSTP in summer, XNS—samples from XN MSTP in summer, ANW-samples from AN MSTP in winter, XGW-samples from XG MSTP in winter, ZTW—samples from ZT MSTP in winter, XNW—samples from XN MSTP in winter, refer to Table 2 in detail



b

			1			_			
			<u> </u>				- _		
1	1	5	56	4	1633	0	9	0	C10-SB1A
	1318	523	319	1	18	6	45	283	Run-SP154
1	1509	60	361	11	130	20	101	42	VadinHA17
1	988	437	204	58	266	292	137	251	Tetrasphaera
	538	570	376	55	484	56	58	177	Iamia
	719	242	158	106	33	93	251	112	Clade
1	825	182	788	198	51	60	204	406	Anaerolineaceae
	555	302	636	12	75	58	164	77	TK10
1	29	125	115	1751	386	344	402	43	Ferruginibacter
	316	496	449	1645	52	121	121	143	1-20
	116	387	216	1989	97	65	52	31	Mycobacterium
	0	238	168	482	86	348	412	1	Dokdonella
	230	379	85	300	447	363	220	300	Rhodobacteraceae
	183	344	310	10	559	360	298	65	Thauera
	77	395	336	585	4	288	64	211	Mle1-27
	292	390	652	585	411	256	138	102	JG30-KF-CM45
	281	256	183	557	248	204	267	196	Peptostreptococcaceae
	29	71	42	832	143	565	573	1486	Flavobacterium
	39	74	65	42	190	577	613	316	Simpli cispira
	767	450	571	97	50	285	1310	440	Sphingobacteriales
	380	507	310	332	627	1715	852	568	NS9 marine group
	348	447	369	280	293	845	948	336	Cytophagaceae
	789	1067	652	720	288	491	776	767	Comamonadaceae
	314	760	418	1357	951	492	785	263	Terrimonas
	192	227	191	180	1651	2798	3026	1206	Trichococcus
	2136	490	658	694	1454	1675	876	865	Saccharibacteria
	355	490	348	392	2449	1927	409	361	Omithinibacter
	848	3267	3085	4867	7439	8976	9876	9763	Candidatus Microthrix
	2126	2736	1476	264	314	1152	4089	2260	Saprospiraceae
	1401	1552	2657	96	293	637	1165	1956	Caldilineaceae
	STS	ANS	XGS	XNS	KNW	ANW	KGW	MTZ	

and even in the Polar Arctic Circle.[1, 9, 12, 17, 30]. These research studies further support the judgment that the core AS bacterial phylum community structure shows no latitudinal gradient and ecological coherence.

At the genus level, AS bacterial community profiles were displayed by a hierarchically clustered heatmap based on Bray-Curtis similarity index, see in Fig. 1b, which shows that all samples were divided into two clades in accordance with seasonal changes. Although the relative abundance of core genera among different AS samples changed dramatically, these genera has some overlap across the winter and summer samples, including members of Candidatus M. parvicella, Saprospiraceae, Caldilineaceae, Trichococcus, Saccharibacteria (TM7), Terrimonas, among which Candidatus Microthrix and *Trichococcus* in winter were significantly richer than that in summer samples. This change is consistent with a phenomenon that filamentous foaming frequently appeared in winter at the surface of the aeration tank. So it maybe concluded that these two genera be linked to the sludge foaming and bulking. At all, if the core community was identified by the abundance and occurrence frequency of OTUs, the core community in AS from the sub-plateau region is clearly distinguished from other geographical regions, suggesting that the sub-plateau AS selects for a unique core community and shapes the specific AS ecology.

# The changes in bacterial community diversity and abundance

The abundance-based coverage of clean reads was observed to be higher than 98% in all AS samples, indicating that the obtained sequence could well cover the bacterial communities. After clustering and alignment, a total of 3094 OTUs was identified at 97% similarity across all samples with a range of 988-1738 OTUs in individual AS samples. The OTU number in the summer samples  $(1421 \pm 304)$  was significantly higher than that in the winter samples  $(1202 \pm 159)$ (Table S1), while 46.1% OTUs (n = 605) was shared between them. Moreover, Chao richness index among different group samples indicates that the abundance of bacterial community in the summer samples is higher than that in the winter samples (Fig. 2a), excluding the XN-MSTPs samples because of its exclusive process. Shannon's diversity index values were  $4.49 \pm 0.20$  and  $5.61 \pm 0.43$  in the winter and summer samples, respectively. Bacterial diversity was positively linked with sewage temperature ( $R^2 = 0.9139$ , p = 0.005) (Fig. 2c,



Fig. 2 The difference of core bacterial community in diversity (c, d), relative abundance (a), and community structure in different MSTPs between summer and winter samples (b). Abbreviations as in Fig. 1 caption

d). The changes of the bacterial community were further assessed by principal co-ordinates analysis (PCoA), represented as PC1 and PC2 at 33.46% and 23.65% of the variance in all community structure, respectively. AS samples collected from the same MSTP were distinctly separated into two groups in accordance with winter and summer samples, and the bacterial community in Xining MSTP winter samples were distinctly different from other samples (Fig. 2b), probably due to the sewage temperature was lowest in Xining MSTP. These results indicate that temperature is a crucial environmental factor driving the dynamics of bacterial community diversity and abundance, these changes could be attributed to the variation in sensitivity and resistance of microorganisms under different temperatures [5, 22].

## The bacterial populations enriched in the summer and winter AS samples

To determine the classified bacterial taxa with significant abundance differences between the summer and winter AS samples, we performed biomarker analysis using the method of linear discriminant analysis (LDA) effect size (LEfSe) and T test (95% confidence intervals). The result showed that, of the 46 phyla, 22 bacterial phyla were significantly distinguished in relative abundance between summer and winter samples. As shown in Fig. 3a, Proteobacteria and Chloroflexi are richer in summer samples, Actinobacteria and Firmicutes are more abundant in winter samples. The difference of bacterial community richness between summer and winter samples is significant ( $p \leq 0.05$ ). In contrast, the other 24 bacterial phyla populations showed no significance differences ( $p \ge 0.05$ ), including a few frequent bacterial phyla, such as Bacteroidetes, Saccharibacteria, Verrucomicrobia. At genus level, the dominant bacterial community populations with significant abundance differences are shown in Fig. 3b, including Candidatus M. parvicella, Trichococcus, Ornithinibacter, Comamonadaceae, etc., Nevertheless, there was also about half of the community populations which showed no significant differences in relative abundance between summer and winter samples. These results indicate that different core bacterial community assembly was developed in summer and winter, respectively. Moreover, this assembly pattern is different from the other geographic MSTPs [15, 18, 29]. For example, *Haliangium, Roseiflexus, Smithella,* and *Lachnospiraceae* were detected in the plateau Wastewater Treatment Plants (WWTPs), but the *Dokdonella, Nitrospira, Terrimonas, Chitinophagaceae, Saprospiraceae,* and *Haliangium* were dominant in the plain WWTPs [9].

Furthermore, a significant difference was observed among the 21 bacterial clades present with an LDA threshold of 4.0, as shown in Fig. 4; it is unexpected that the more bacterial populations were significantly enriched in the winter samples, only six clades showed abundance advantage in summer samples. In detail, Proteobacteria (phylum), Chloroflexi (phylum), Deltaproteobacteria (class), Gammaproteobacteria (class), Myxococcales (order), and Betaproteobacteria (class) were enriched in summer samples. Actinobacteria (phylum and class), Acidimicrobiales incertae sedis (order), Candidatus M. parvicella (genus), Acidimicrobiales (family), and Firmicutes (phylum) were enriched in winter samples. Meanwhile, the efficiency of sewage purification in summer was clearly higher than that in winter (Table 2). Thus, we could assume that the AS bacterial community assembly enriched in the summer played more effective roles in removing organic pollutants and nutrients. This result agrees with the previous study that Proteobacteria could metabolize acetate, butyrate, glucose, and propionate by multiple metabolic mechanisms [11]. Chloroflexi populations play an important role in stabilizing flocs in AS and removing pollutants [9, 27]. In contrast, Actinobacteria members enriched in the winter AS, such as Candidatus M. parvicella, were associated with the foams or the AS bulking in the sub-plateau MSTPs [2, 24]. It could be deemed that the overgrowth of Actinobacteria inhibited other bacterial groups and results in



Fig. 4 LEfSe analysis of bacte-

enriched in winter and summer

rial community populations

AS samples, respectively



LDA Score (log10)

the decline of sewage purification at low temperature. This result could be attributed to the decreasing of bacterial community diversity, relative abundance and the changes in core bacterial community structure.

# The changes in bacterial community related to environmental and operational factors

To investigate the principal factors responsible for the changes of bacterial community in different seasons, RDA was performed to explore the relationship between the relative abundance of community and environmental/operational parameters. As shown in Fig. 5a, acute angles emerged among temperature (T) and DO, altitude (A), SRT, and HRT, respectively, indicating that these factors had a synergetic impact on the relative abundance of community. AS bacterial community was RDA1-affiliated, which corresponded more closely with temperature and DO. The altitude, SRT, and HRT had little influence on AS bacterial community. Proteobacteria and Chloroflexi were positively correlated with temperature and DO, but Actinobacteria, Firmicutes, and Bacteroidetes were negatively correlated with temperature and DO, and SRT and HRT were positively correlated with altitude (Fig. 5a). The most abundant bacterial genus Candidatus M. parvicella and Trichococcus were negatively correlated with temperature and DO, but positively correlated with SRT and HRT (Fig. 5b). These results were in agreement with the previous study that these two genera overgrew and caused sludge foaming under low temperature [24]. In addition, RDA also showed a weak correlation between the bacterial communities, and HRT and altitude.

Pprevious studies also have demonstrated that the distinctive AS community abundance and structure are shaped by different operational parameters and environmental conditions [4, 26]. Among them, temperature and DO are the most important variables that shaped the particular AS ecosystem in sub-plateau MSTPs. For example, temperature exerts a profound impact on the biological hydrolysis performance and the microbial community structures such as the main protein or carbohydrate fermenting bacteria genera changed with temperature [4]. DO concentration also significantly affects the microbial community structure as well as the efficiency of sewage purification [23]. Thus, to some extent, an optimized microbial community structure would be established in a given environmental/operational parameters, and the performance of MSTP tight links with it.

### Conclusions

There was significant difference of the core bacterial community between summer and winter AS sample in sub-plateau MSTPs. The OTU number, community diversity, and richness of bacterial community have decreased dramatically in winter. The different core bacterial community assembly was developed in summer and winter, respectively. As *Proteobacteria* and *Chloroflexi* are enriched in summer, *Actinobacteria* and *Firmicutes* are enriched in winter. These changes are closely associated with the performance of sewage purification. Additionally, the core bacterial community assembly in sub-plateau region is clearly distinguished from other geographical regions, **Fig. 5** RDA of the correlation between bacterial community and environmental/operational variables at phylum level (**a**) and genus level (**b**). Abbreviations as in Fig. 1 caption



suggesting that the sub-plateau AS selects for a unique community assembly pattern and shapes the particular AS ecosystem. Further analysis shows that temperature and DO are the principal factors that drive the dynamics of bacterial community diversity and relative abundance in sub-plateau MSTPs.

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#### **Compliance with ethical standards**

**Conflict of interest** The authors declare that they have no conflict of interest.

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