RESEARCH PAPER



Deep and high-efficiency removal of sulfate through a coupling system with sulfate-reducing and sulfur-oxidizing capacity under haloalkaliphilic condition

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Abstract

Sulfide from anaerobic treatment of high-sulfate wastewater would always have some adverse effects on downstream processes. In this study, a coupling anaerobic/aerobic system was developed and operated under haloalkaliphilic condition to realize deep and high-efficiency removal of sulfate without production of sulfide. A haloalkaliphilic sulfur-oxidizing strain, *Thioalkalivibrio versutus SOB306*, was responsible for oxidation of sulfide. The anaerobic part was first operated at optimum condition based on a previous study. Then, its effluent with an average sulfide concentration of $674 \pm 33 \text{ mg} \cdot l^{-1}$ was further directly treated by a set of 1 l biofilter with SOB306 strain under aerobic condition. Finally, 100% removal rate of sulfide was achieved at aeration rate of $0.75 \, l \cdot l^{-1} \cdot min^{-1}$, ORP of -392 mV and HRT of 4 h. The average yield of elemental sulfur reached $79.1 \pm 1.3\%$ in the filter, and the CROS achieved a conversion rate of sulfate to sulfur beyond 54%. This study for the first time revealed the characteristics and performance of the haloalkaliphilic CROS in deep treatment of high-sulfate wastewater, which paved the way for the development and application of this method in the real world.

Keywords Sulfate-reducing · Sulfur-oxidizing · Haloalkaliphilic · Thioalkalivibrio versutus

Introduction

With the development of modern industrialization, a large amount of highly concentrated sulfate organic wastewater is discharged from some processes, such as chemical industry, monosodium glutamate production, pharmacy, leather, paper and so on. Even though sulfate does not result in distinct hazards in our surrounding environment, sulfate pollution can lead to several indirect environmental effects [1]. Wastewater with sulfate is normally treated with physicochemical

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and biological methods. It is well-known that, however, the physicochemical methods possess some underlying disadvantages which limit their applications, such as separation and appropriate disposal of solid phase, high cost and energy consumption [2].

As regards biological methods, organic sulfate-rich wastewater was generally treated in anaerobic processes in which sulfate-reducing bacteria (SRBs) could be responsible for sulfate reduction into sulfide [3]. As a matter of fact, a high concentration of sulfide could poison methane-producing archaea (MPA) which was another important member in the processes and cause a decline in methane production. In some conventional biological systems with about pH 6.0, sulfide existed in the form of H₂S which could penetrate into cells easily and generate a direct toxicity to MPAs, SRBs and other kinds of microorganisms, and inhibited the treatment effect of anaerobic processes [4, 5]. Moreover, treating wastewater using biological methods is economic, effective, and thorough, but the capacity of biosystems was easily subject to limitation under conditions of high salinity and pH [4, 6].

However, some haloalkaliphilic microorganisms displayed remarkable advantages in the treatment of such sulfate-rich wastewaters [7]. In this kind of haloalkaliphilic system, distinctly, the presence of hydrogen sulfide in solution was in the form of HS⁻ that cannot penetrate into cells easily [8, 9]. In our group, some interesting studies on performance of haloalkaliphilic bioreactors and bacterial communities were performed in recent years [4, 10, 11]. Generally, if there existed a high concentration of sulfide in effluent of an anaerobic reactor, the activity of some key microorganisms in some downstream processes was also suppressed. Sulfide is a kind of known inhibitor of nitrification and affects microbial communities in nitrifying treatment process, for instance, it can differentially inhibit ammonium-oxidizing and nitrite-oxidizing bacteria [12–14]. Therefore, removal of sulfide in solution using some methods would help minimize adverse effects on downstream processes as much as possible.

Biodesulfurization based on sulfur-oxidizing bacteria (SOBs) was a useful technology for removal of sulfide, and some relevant technologies have been presently developed. A specific group of haloalkaliphilic SOBs belonging to the genus Thioalkalivibrio isolated from soda lake sediments was successfully applied in some fed reactor systems for sulfide removal [15–17]. The members of *Thioalkalivibrio* were Gram-negative, halophilic, alkaliphilic, sulfur-oxidizing and chemolithotrophic bacteria. They could tolerate Na⁺ concentrations up to 5 M and their optimum pH for growth was between 9 and 10 [18]. They mainly gained energy by oxidizing reduced or partially reduced sulfur compounds and also fixed CO_2 from the atmosphere [19]. Under oxygenlimited conditions, dissolved sulfide was mainly oxidized into elemental sulfur (S^0) by these SOBs, whilst a part (typically less than 10%) was oxidized into sulfate (SO_4^{2-}) [20]. The use of these SOBs could circumvent many of the obstacles associated with conventional assays. As mentioned above, the effluent of anaerobic bioreactor using haloalkaliphilic microorganisms was also characteristic of high pH, high concentration of sodium and a certain concentration of sulfide, so it might be directly treated by haloalkaliphilic SOBs for sulfide removal.

In the present work, we offered a coupling bioprocess for treatment of wastewater with a high concentration of sulfate and sodium based on haloalkaliphilic and biological sulfate-reducing system and sulfur-oxidizing system. The performance of this process was investigated by treating modeling wastewater under optimum sulfate-reducing condition. Here we also come up with a significant demonstration that haloalkaliphilic sulfur-oxidizing biosystem could be directly applied in deep treatment of sulfate wastewater and high-efficiency removal of sulfide from some kind of wastewater.

Materials and methods

System description

A coupling sulfate-reducing and sulfur-oxidizing system (CROS) consisted of one bioreactor for removal of sulfate under anaerobic condition and the other for removal of sulfide under aerobic condition. The profile of CROS is displayed in Fig. 1. The anaerobic part with a working volume of 61 has been described detailedly and reported by our group [11]. Here, we attempted to construct a continuous and synchronized system for deep removal of sulfate by adding an aerobic biofilter as shown in Fig. 1 and Fig. S1A. The effluent of anaerobic process with highly concentrated sulfide was directly pumped into the biofilter at some constant flow rate, and sulfide was further oxidized into sulfur and a small amount of sulfate through keeping suitable aeration rate and redox potential (ORP). The biofilter was filled with hollow ceramic rings with the size of $\Phi 12 \text{ mm} \times 5 \text{ mm}$, and its working volume was 1 l. Two kind of sensors, pH (InPro3250i, Mettler-Toledo)





and ORP (Pt4805-DPA, Mettler-Toledo, Switzerland), were installed on the lid of biofilter for real-time monitoring of the running state.

Inoculum and nutrient media

The liquid culture of biofilter was directly taken from an 801 haloalkaliphilic sulfur-oxidizing bioreactor (HSOB) which

$$\begin{split} \left[SO_4^{2-} - S_{R-in} \right] &= \left[SO_4^{2-} - S_{R-out} \right] + \left[S_2 O_3^{2-} - S_{R-out} \right] + \left[S^{2-} - S_{R-out} \right] \\ &= \left[SO_4^{2-} - S_{O-in} \right] + \left[S_2 O_3^{2-} - S_{O-in} \right] + \left[S^{2-} - S_{O-in} \right] \\ &= \left[SO_4^{2-} - S_{O-out} \right] + \left[S_2 O_3^{2-} - S_{O-out} \right] + \left[S^{2-} - S_{O-out} \right] + \left[S_{O-out}^{0-} \right] \\ \end{split}$$

which was equipped with an electrical conductivity detector (Dionex Sunnyvale, CA). A Dionex IonPacTM AS14A analytical column (4×250 mm) was operated at 25 °C, the mobile phase was 8.0 mM Na₂CO₃/1.0 mM NaHCO₃ at a flow rate of 1.0 ml·min⁻¹ [24]. The injection volume was 10 µl. The concentration of elemental sulfur of effluent was calculated by the mass balance between total concentration of all dissolved sulfur products in the inlet and outlet. All experimental data were processed based on sulfur balance as the following equations:

was used for biodesulfurizing experiments in the long term in our group (Fig. S1B). *Thioalkalivibrio versutus* SOB306 was solely the functional strain responsible for oxidization of sulfide [21]. Medium FTD was used in contrast tests. The medium contained: Na₂CO₃ 46.0 g·l⁻¹, NaHCO₃ 23.0 g·l⁻¹, K₂HPO₄·₃H₂O 2 g·l⁻¹, KNO₃ g·l⁻¹, NH₄Cl 0.3 g·l⁻¹, MgCl₂·6H₂O 0.1 g·l⁻¹, Na₂S·9H₂O 5.3 g·l⁻¹. A trace elements solution was added (1 ml·l⁻¹) as described elsewhere [22]. The preparation of medium for anaerobic bioreactor was totally based on our previous study [11].

Operation of CROS

The anaerobic bioreactor was first operated according to a previous study [11]. After the performance of bioreactor reached the optimum condition stably, effluent was collected into the tank as shown in Fig. 1. Then, the effluent was pumped into biofilter with HSOB culture at a certain feed rate and aeration was regulated timely based on the value of ORP to guarantee CROS running continuously and stably. Through adjusting HRT and aeration rate, performance of bioreactor reached an optimal state with the greatest removal rate of sulfide. Samples were taken from A, C and D of CROS every 4 h.

Analytical methods

Before making measurement, all samples were centrifuged at 13,000 rpm and 10 min, and then the supernatant was taken and diluted based on the requirement. The concentration of dissolved sulfide was measured by colorimetry with a spectrophotometer (U-2910; Hitachi, Tokyo, Japan) [23]. Sulfate and thiosulfate were analyzed by ion chromatography (Dionex model ICS-900, Dionex, Sunnyvale, CA),

Removal percentage of sulfide (%) =
$$([S^{2-}-S_{O-in}] - [S^{2-}-S_{O-out}])/[S^{2-}-S_{O-in}] \times 100\%$$
 (2)

Herein, R and O represented anaerobic bioreactor and aerobic biofilter, respectively.

Results and discussion

Start-up of CROS

In this study, the anaerobic part of CROS was first operated under optimum condition totally based on the method reported by a previous study, with sulfate concentration of $3000 \text{ mg} \cdot 1^{-1}$ in influent, COD/SO₄²⁻ ratio of 4.0 and HRT of 24 h [11]. After stable operation, the average removal rate of sulfate reached 68.5%, and the average concentration of sulfide in effluent achieved $674 \pm 33 \text{ mg} \cdot 1^{-1}$, as shown in Fig. 2a. Its performance was line with that in the previous study, which demonstrated that the capacity of this anaerobic bioreactor was very stable [11]. 1 l culture was collected from previous HSOB. The values of pH and salinity of this kind culture were 9.5 and 1.0 M Na⁺, respectively. Figure 2b displayed the dose of some key sulfur compounds in the culture. Owing to long-term running, accumulation concentrations of sulfate and thiosulfate reached $9235 \pm 182 \text{ mg} \cdot l^{-1}$ and $5166 \pm 71 \text{ mg} \cdot l^{-1}$, respectively. However, the concentration of sulfide was only 32 ± 1 mg·l⁻¹. Then, the CROS was started up by coupling anaerobic bioreactor with biofilter through the action of peristaltic pump. The anaerobic bioreactor could produce around 61 effluent per day, so the pump rate was set to 4.15 ml·min⁻¹ to synchronously run biofilter. HRT of biofilter was about 4 h. The performance of CROS was regulated by adjusting aeration rate based on ORP.

Fig. 2 The average removal rate of sulfate and production of sulfide in anaerobic bioreactor (a). It was a display of treatment capacity of this anaerobic bioreactor. The concentration of some sulfur compounds in initial culture of biofilter (b)





Fig. 3 Relationship between the aeration rate and ORP and removal rate of sulfide

Aeration rate and ORP

The change laws of ORP with aeration rate are demonstrated in Fig. 3. The value of ORP gradually increased from -400 mV to -340 mV as aeration rate was improved under constant concentration of sulfide in the influent. ORP was one of the most important operating parameters for sulfide oxidizing selectively to elemental sulfur [25]. Under optimal ORP for sulfur formation, a maximum amount of sulfide would be converted into elemental sulfur [26]. When ORP was kept at -400 mV, the removal rate of sulfide was just $93.1 \pm 1.5\%$, but when the aeration rate was adjusted to $0.75 \text{ l}\cdot\text{l}^{-1}\cdot\text{min}^{-1}$ and even higher, sulfide was totally removed from liquid and transferred into other kinds of sulfur compounds without toxicity. And the value of ORP was -392 mV at this time. In the whole process, oxidizing reactions were dependent



Fig. 4 Effect of the rate of aeration on selectivity of S^0 and SO_4^{2-}

on some key enzyme systems in cells, like Fcc, Sox, Hdr, and so on [27, 28].

Moreover, selectivity of oxidization of sulfide to sulfur and sulfate was also obviously different at different aeration rates. As shown in Fig. 4, the change of sulfate was positively correlated with aeration rate. Even though sulfide had been oxidized completely, the selections for production of sulfate and elemental sulfur were 40% and 50% at the highest aeration rate, which was not an ideal state owing to lower production rate of sulfur and higher energy expenditure. When aeration rate was kept at $0.75 \ 1 \cdot 1^{-1} \cdot \text{min}^{-1}$, the production rate of elemental sulfur reached the highest level, namely 78%, and the rate of sulfate was decreased to 20%. As a matter of fact, the production rate, but the selection for production of elemental sulfur was dropped below 60% instead. In addition, when the value of ORP dropped, a lower level of oxygen led to a low oxidization level of sulfide and accumulation of thiosulfate [20]. Thiosulfate was formed during sulfide oxidation, which was likely related with some abiotic processes [29, 30]. As the aeration rate decreased, the oxidation capacity of the system also decreased, and sulfide could not be abundantly removed. That was mainly because the activity of SOB306 cells was repressed at a certain extent under lower oxygen input. Yet, the more elemental sulfur was generated, the better sulfur pollutants were removed thoroughly [31]. Taken together, the capacities of biofilter and CROS were optimum at aeration rate of $0.75 \, l \cdot l^{-1} \cdot min^{-1}$ and ORP of $-392 \, mV$ under continuous feeding.

Performance of CROS

Under aeration rate of $0.75 \text{ I}\cdot\text{I}^{-1}\cdot\text{min}^{-1}$ and ORP of -392 mV, the conversion of sulfide was further explored in this study. As shown in Fig. 5, before the first 20 h, the sulfur compounds in original culture from HSOB were oxidized by SOB306 strain and continuously discharged outside. Therefore, the concentrations of sulfate and thiosulfate were gradually reduced until it reached stable conditions after 20 h. The concentration of sulfate was kept below 1,400 mg·I⁻¹, the accumulation of thiosulfate was relatively low and the average yield of elemental sulfur reached 79.1 ± 1.3%. Eventually, the conversion rate of sulfate to sulfur achieved beyond 54% by the CROS.

Lastly, some contrast tests were performed by substituting effluent of anaerobic bioreactor with FTD medium. The concentration of sulfide was around 700 mg \cdot l⁻¹, which was similar to that in effluent of anaerobic part. Then, the performance of biofilter was tested at different flow rates, and



Fig. 5 Variation concentration of SO_4^{2-} and $S_2O_3^{2-}$ and production rate of S^0 with operation time

Fig. 6 The production rate of sulfur and accumulation rate of sulfate with variation of HRT in biofilter. These were some control experiments for displaying the performance of biofilter using FTD medium substituting for effluent of anaerobic bioreactor



aeration rate was still stayed at $0.75 \, l \cdot l^{-1} \cdot min^{-1}$. From Fig. 6, the performance of biofilter achieved optimum condition at HRT of 3 h, with $80.2 \pm 2.3\%$ of productive rate of sulfur and $19.6 \pm 0.5\%$ of accumulation rate of sulfate. It was speculated that the activity of SOB306 cells was suppressed partly by some kind of organic matters, metal ions and other microorganisms from effluent of anaerobic bioreactor. Certainly, S⁰ formation was accompanied by the growth of strain SOB306 which obtained energy by oxidizing sulfide into sulfur and sulfate under aerobic conditions. Compared with neutrophilic SOBs which just grew under some conditions with lower pH and salinity, SOB306 strain could be more adaptable to this kind of complex environment and the performance of biofilter and CROS could be further improved once cells covered onto packings by formation of biofilm over time [32]. However, it was still demonstrated that the effluent of anaerobic bioreactor could be directly treated by biofilter with SOB306 strain owing to its haloalkaliphilic and sulfur-oxidizing properties. Most of previously reported coupled anaerobic/aerobic treatments of high-sulfate systems were operated around neutral pH [32-34]. Therefore, this study revealed, for the first time, the characteristics and performance of haloalkaliphilic coupled anaerobic/aerobic system in deep removal of sulfate.

Conclusion

This work demonstrated successful operation of an integrated anaerobic/aerobic biosystem CROS for deep and high-efficiency treatment of high-sulfate model wastewater under high pH and salinity. The anaerobic part of CROS was operated at optimum condition with sulfate concentration of $3000 \text{ mg} \cdot 1^{-1}$ in influent, COD/SO₄²⁻ ratio of 4.0 and HRT of 24 h. The effluent with $674 \pm 33 \text{ mg} \cdot 1^{-1}$ of sulfide was further treated by a set of 11 biofilter under aerobic condition. To keep synchronous operation of two bioreactors, the flow rate of biofilter was set to 4.15 ml·min⁻¹. The capacity of biofilter got optimum condition with 100% removal of sulfide at aeration rate of $0.75 \, l \cdot l^{-1} \cdot min^{-1}$, ORP of $-392 \, mV$ and HRT of 4 h. The average yield of elemental sulfur in biofilter reached $79.1 \pm 1.3\%$, and the conversion rate of sulfate to sulfur achieved beyond 54% by the CROS. The main sulfur-oxidizing bacterium involved in this process was a haloalkaliphilic Thioalkalivibrio versutus SOB306, which could tolerate high pH and salinity. It turned out that the effluent of anaerobic bioreactor could be directly treated by biofilter with SOB306 strain owing to its haloalkaliphilic and sulfur-oxidizing properties. This study first revealed the characteristics and performance of haloalkaliphilic coupled with anaerobic/aerobic system in the advanced treatment of high-sulfate wastewater.

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Compliance with ethical standards

Conflicts of interest The authors declare no conflict of interest.

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