POPULATION ECOLOGY – ORIGINAL RESEARCH

Fitness correlates of age at primiparity in a hunted moose population

StineS. Markussen¹[®] • Anne Loison² • Ivar Herfindal¹ • Erling J. Solberg³ • Hallvard Haanes⁴ • Knut H. Røed⁵ • **Morten Heim³ · Bernt‑Erik Sæther1**

Received: 30 January 2017 / Accepted: 22 November 2017 / Published online: 2 December 2017 © Springer-Verlag GmbH Germany, part of Springer Nature 2017

Abstract

Trade-ofs between ftness-related traits are predicted from the principle of resource allocation, where increased fecundity or parental investment leads to reduced future reproduction or survival. However, ftness traits can also be positively correlated due to individual diferences (e.g. body mass). Age at primiparity could potentially explain variation in individual ftness either because early primiparity is costly, or it may lead to higher lifetime reproductive success. Based on long-term monitoring and genetic parentage assignment of an island population of moose, we quantifed reproductive performance and survival, and examined whether early maturing females have higher total calf production than late maturing females. We explored if harvesting of calves afected the subsequent reproductive success of their mothers, i.e. also due to a postweaning cost of reproduction, and whether there are any intergenerational efects of female reproductive success. There was a positive relationship between current and future reproduction. The probability to reproduce was lower for females that were unsuccessful the year before, indicating a strong quality efect on productivity. Females that started to reproduce as 2-year olds had a slightly higher total calf production compared to those starting at age three or four. High-performing mothers were also correlated with daughters that performed well in terms of reproductive success. Our results suggest that the observed individual heterogeneity in ftness could be associated with diferences in age at primiparity. This heterogeneity was not affected by reproductive costs associated with tending for a calf post-weaning.

Keywords Age of maturity · Cost of reproduction · Individual heterogeneity · Life history · Ungulates

Communicated by Jean-Michel Gaillard.

Electronic supplementary material The online version of this article [\(https://doi.org/10.1007/s00442-017-4021-2\)](https://doi.org/10.1007/s00442-017-4021-2) contains supplementary material, which is available to authorized users.

 \boxtimes Stine S. Markussen stine.markussen@ntnu.no

- ¹ Department of Biology, Centre for Biodiversity Dynamics, Norwegian University of Science and Technology, Trondheim, Norway
- ² CNRS, UMR 5553, Laboratoire d'Écologie Alpine, Université de Savoie, Le Bourget du Lac, France
- ³ Norwegian Institute for Nature Research (NINA), Trondheim, Norway
- ⁴ The Norwegian Radiation Protection Authority (NRPA), Oslo, Norway
- ⁵ Department of Basic Sciences and Aquatic Medicine, Norwegian University of Life Sciences, Oslo, Norway

Introduction

Life history theory predicts that iteroparous species should allocate resources to maximize individual ftness, resulting in trade-ofs between ftness-related traits (Stearns [1992](#page-11-0)). Trade-ofs in life history evolution are often based on the principle of resource allocation (Williams [1966](#page-11-1)), where resources allocated to one ftness component will lead to a decrease in resources allocated to other ftness components. One example is the cost of reproduction, in which increased fecundity or parental investment causes reduced longevity of the parent or a reduced capacity for future reproduction (Reznick [1985](#page-10-0)). Such a negative covariation between competing traits is an element of most early models of life history evolution (e.g. Williams [1966;](#page-11-1) Schafer [1974;](#page-11-2) Michod [1979](#page-10-1)). However, the amount of resources available and the acquisition and handling abilities of individuals are rarely constant (van Noordwijk and de Jong [1986\)](#page-11-3). Accordingly, fitness-related traits can also positively covary because the costs may be higher in some individuals than others

(Partridge [1987](#page-10-2)) or due to individual variation in resource acquisition and/or utilization (e.g. Bell and Koufopanou [1986;](#page-9-0) Cam et al. [2002](#page-9-1); Hamel et al. [2009a](#page-9-2), [b](#page-9-3); Chambert et al. [2013](#page-9-4)). This individual variation could lead to diferences in individual quality where quality is defned as "an axis of among-individual heterogeneity that is positively correlated with ftness" (Wilson and Nussey [2010\)](#page-11-4), such as diferences in body condition (Bérubé et al. [1999\)](#page-9-5), experience or age (McNamara and Houston [1996\)](#page-10-3).

The age at frst reproduction (hereafter age at primiparity) is an important trait with substantial potential for infuencing lifetime reproductive success (Bell [1980\)](#page-9-6) and population dynamics (Reiter and LeBoeuf [1991\)](#page-10-4). Within species, earlymaturing individuals can have higher lifetime reproductive success than late-maturing individuals as they may obtain a higher number of successful reproductive events (Stearns [1992;](#page-11-0) Oli and Dobson [1999](#page-10-5); Herfndal et al. [2015](#page-10-6)). However, early start of reproduction may also involve a ftness cost such as reductions in growth, survival and future reproduction (Pyle et al. [1997;](#page-10-7) Proaktor et al. [2008](#page-10-8)).

If age at primiparity is related to individual quality, its trade-off with future fitness may be less apparent (Sæther and Heim [1993;](#page-11-5) Festa-Bianchet et al. [1995;](#page-9-7) Neuhaus et al. [2004](#page-10-9)). In capital breeders (Gaillard et al. [2000a\)](#page-9-8), reproduction and age at primiparity are often related to body condition (Jorgenson et al. [1993;](#page-10-10) Gaillard et al. [2000b](#page-9-9)), and large, high-quality females reproduce more successfully and with lower future ftness costs than small females—particularly in species with low or no variability in litter size (Green and Rothstein [1991;](#page-9-10) Festa-Bianchet et al. [1998](#page-9-11); Weladji et al. [2008](#page-11-6)). Moreover, variation in quality among females may generate size and quality differences in their offspring (Solberg et al. [2007](#page-11-7); Hamel et al. [2012](#page-9-12)) and in turn their ftness (Hamel et al. [2010\)](#page-9-13), leading to possible intergenerational efects of varying age at primiparity. Such intergenerational efects have, however, rarely been investigated in long-lived species (but see Moore et al. [2016](#page-10-11)).

Measuring the costs of reproduction is challenging for at least two reasons. First, reproduction involves several phases during which energy expenditure varies. In mammals, diferent expenditures during gestation, lactation and post-weaning care may lead to diferent costs (Clutton-Brock et al. [1983\)](#page-9-14). The pre-weaning care is related mainly to foetus growth and milk production, as opposed to post-weaning care, which is associated with fending for the calf from weaning to independence (Clutton-Brock [1991](#page-9-15)). The latter may involve reduced food consumption of the mother due to direct feeding competition, or because mothers trade of good feeding opportunities for better anti-predator conditions for the calf (Green et al. [1989](#page-9-16); White and Berger [2001](#page-11-8)). Substantially fewer studies have examined maternal costs associated with the post-weaning than the pre-weaning stage in ungulates (L'Heureux et al. [1995](#page-10-12); Andersen et al. [2000](#page-9-17); Froy et al. [2016](#page-9-18)). However, each of these stages may vary diferently with age or environmental factors, which means that teasing apart which stage is more likely to lead to future ftness consequences is essential for a better understanding of life history strategies. Still, this is difficult in the wild because of the need for detailed data on success or failure during the diferent steps of the reproductive period. The second challenge for evaluating the costs of reproduction is methodological. Methods used usually rely on phenotypic or genetic correlations, and experimental manipulations (Lande [1982;](#page-10-13) Reznick [1985;](#page-10-0) Bell and Koufopanou [1986](#page-9-0)). Experimental manipulations have been performed in the wild mostly in birds and lizards, e.g. by manipulating litter size or phenotypes (Lindèn and Møller [1989;](#page-10-14) Nilsson and Svensson [1996;](#page-10-15) Cox and Calsbeek [2010\)](#page-9-19), but have turned out to be difficult to perform in mammals (but see Tavecchia et al. [2005](#page-11-9); Gélin et al. [2015](#page-9-20)). Hence, in most studies of wild populations of large mammals, the cost of reproduction is determined by phenotypic correlations (e.g. Clutton-Brock et al. [1982;](#page-9-21) Bérubé et al. [1999\)](#page-9-5). This is considered a rather weak method, due to individual heterogeneity in available resources, and should if possible also include some kind of manipulation of ftness traits. Such a manipulation can be induced when an offspring dies from non-natural causes, such as harvesting. While such an "experimental" setting has led to some studies of the efect of post-weaning care on ofspring performances when a mother was harvested (e.g. Festa-Bianchet et al. [1994;](#page-9-22) Holand et al. [2012\)](#page-10-16), it has seldom been used in the context of measuring reproductive costs in females.

Here, we analysed the age-specifc costs of reproduction in a moose population that is harvested by recreational hunting. Based on long-term monitoring (28 years) and genetic parentage assignment of individually marked moose (Haanes et al. [2013\)](#page-9-23), we focused on the consequences of age at primiparity and post-weaning costs on the reproductive performance and survival of females. We frst examined if early maturing females have higher total calf production than late maturing females, or alternatively, if late maturing females may be able to compensate by increasing reproductive output later in life. All else being constant, early maturing females are expected to produce more calves throughout their lifetime simply because they will gain additional breeding attempts. Then, we tested if mothers experience a post-weaning cost of reproduction by analysing their reproductive success in relation to their loss of calves during the previous hunting season. If there is a post-weaning cost of reproduction, we expected a lower reproductive output in females that successfully retained a calf after the hunting season than those that did not. Finally, we assessed whether there were any intergenerational correlations in reproductive performance, i.e. if high performing females are correlated with high performing offspring.

Materials and methods

Study system and data collection

Vega (65°40′N, 11°55′E) is an island of approximately 119 km², 13 km off the coast of northern Norway. The moose population on the island was founded by two yearling females and one yearling male that swam to the island in 1985. Harvesting started in 1989 and has since 1992 been used to keep the breeding population at 20–40 moose, resulting in 15–26 calves born per year (Haanes et al. [2013\)](#page-9-23). Calves are generally weaned in mid-September (Bubenik [1998\)](#page-9-24).

Moose hunting on the island can occur from the 25th of September to the 31st of October, but in most years started in early October, i.e. after the start of the rut. Hunting permits are typically issued as a fxed number of calves, adult females and adult males (or unspecifed animals), but given that hunters have a harvesting plan approved by the local wildlife authorities, they are granted some fexibility with regard to how many of each category they can harvest within a year. Hunters can among themselves decide to focus their harvest on young or old adult males or females. In general, harvested females are either without calves that year, or the calves are shot together with the female at the same occasion.

During 1992 and 1993, all individuals on the island were radiocollared for the frst time. In January–March every year (except 2003 and 2008), this procedure was repeated to mark and measure all new calves (born in May–June) which survived the annual hunt. In addition, we collected sex, age, carcass mass, and tissue samples from almost all moose harvested on Vega. This included 4- to 5-month-old calves that were not yet collared. For more information about the moose population and study site, see Sæther et al. [\(2003](#page-11-10), [2004](#page-11-11), [2007](#page-11-12)), Solberg et al. ([2007](#page-11-7), [2008](#page-11-13)) and Kvalnes et al. [2016](#page-10-17).

By the use of genetic parentage assignment based on 22 microsatellite loci, we developed a 28-year-long pedigree for all > 4-month-old moose recorded alive (Haanes et al. [2013\)](#page-9-23) from 1984 to 2012. This includes individual histories of 222 females, of which 59 reproduced successfully at least once. 36 females started to reproduce at the age of 2 years while the remaining 23 started at the age of 3 or 4. The natural mortality rate, φ , is in general low $(0.05 \pm 0.02$ for calves and 0.02 ± 0.01 for females aged 3–9 years old), making hunting the main mortality factor on the island. 19 of the 59 reproducing females were still alive at the end of the study. To increase our sample size, we included the 19 females still alive in our life span variable (age at last observation) used in our analysis of total calf production and intergenerational correlations.

However, excluding those still alive did not change the best model and direction of the results. Based on previous studies (Sæther and Haagenrud [1983](#page-11-14); Garel et al. [2009](#page-9-25)), we divided females into fve age classes: calves, yearlings, 2-year-olds, 3- to 9-year-olds (prime-aged females) and \geq 10-year-olds (potentially senescent females; Sæther et al. [2007](#page-11-12)). Calf body mass was measured as live body mass during winter (approximately 8 months old) and adjusted relative to the date of weighing (Herfndal et al. [2006;](#page-9-26) Kvalnes et al. [2016](#page-10-17)). We had calf body mass for 37 of 59 females recorded to reproduce. Total calf production was defned as the total number of calves produced per female at age of death or age at last observation (year 2012) if alive.

Statistical analysis

Initially, we addressed whether total calf production was related to age at primiparity, accounting for life span (natural log transformed age at death/last observation) with a generalized linear model with Poisson error structure. As modelling of reproductive output can be challenging (Kendall and Wittmann [2010](#page-10-18)), the total calf production was also analysed with a Poisson error structure with square root link, a generalized Poisson distribution and linear model with Gaussian error structure. However, the choice of statistical model did not affect our results qualitatively. To assess how survival and reproduction were related to age, previous reproductive state, age at primiparity and their interactions, we used multistate capture-mark-recapture models (Lebreton et al. [2009\)](#page-10-19) in E-SURGE (version 1.9.0, Choquet et al. [2009a](#page-9-27)). This method allows for simultaneous estimation of survival probabilities (from hunting or natural causes) and probabilities of transitioning between reproductive states. Due to the thorough pedigree (Haanes et al. 2013), we assumed that the fate of all individuals was known (no error in determination of reproductive or survival state). As we also know which individuals were hunted or died naturally, the capture probability was fxed to 1. Before starting our analysis, we conducted a goodness-of-ft (GOF) assessment of the Jolly-Move (JMV) model (Brownie et al. [1993\)](#page-9-28) in U-CARE (version 2.3.2, Choquet et al. [2009b\)](#page-9-29). It satisfactorily ftted the data $(\chi^2 = 53.882, d.f. = 83, p = 0.995)$.

We divided the analysis into four steps: frst, we focused on estimating hunting mortality and reproduction probabilities for all individuals. The models included three types of parameters:

 η_a^r , the probability that an individual in state *r* at age *a* was not hunted until $a + 1$.

 φ_a^r , the probability that an individual in state *r* at age *a* survived until $a + 1$.

 ψ_a^{rs} , the probability that an individual in state *r* at age *a* was in state s at $a + 1$, given that the individual survived from age *a* to age $a + 1$.

The possible states were as follows: not reproduced (ψ^{NR}) , produced one calf (ψ^1) , and produced twins (ψ^2) . The general model included the effect of previous reproductive state (PR) and age class *a* on hunting mortality *η* and natural mortality φ . For reproduction ψ probabilities also the transition to current reproductive state (CR) was included:

$\eta_{a \times PR}, \varphi_{a \times PR}, \psi_{a \times PR \times CR}.$

Second, we assessed the effects of age, age at primiparity (AP) and previous reproductive success on the probabilities to reproduce (NR, 1C or 2C) the following year for those that reproduced, given the general model:

$\eta_{a\times PR}, \varphi_{a}, \psi_{a\times PR\times AP\times CR}$

Because few mature individuals died of natural causes, *φ* included only an efect of age. Age at primiparity as 3- and 4-year-olds were pooled for sample size reasons.

Third, we included the body mass of each female as calf as a covariate to the best model in step two. This was to investigate if body mass in early life could explain any variation in reproductive success that was not already accounted for by other covariates (e.g. age at primiparity). The general model included the same variables as in the second step.

In the last step, we investigated the post-weaning cost of reproduction by examining to what extent hunting of calves afected the subsequent reproductive success of mothers. For this, we introduced two potential states in the model: females had calf/calves that survived, and females had calves that were hunter-killed during the following hunting season. Due to small sample size, we also divided the females into two age classes: 2- and \geq 3-year-olds. All models were run with random initial values repeated fve times to protect against local minima. Including individual identity as random efect in the reproductive transitions in our best models led to models with equal or higher deviance and higher AIC_c -values and were left out (including individual identity: all individuals $\Delta AIC_c = 1.99$, reproducing individuals $\Delta AIC_c = 10.74$.

For more information about the models, see supplementary materials.

To search for intergenerational relationships, we examined whether age at primiparity of daughters was correlated with the age at primiparity of their mother with a χ^2 test. Fifty-four unique mother–daughter pairs included daughters that reproduced at least once $(N_{\text{mothers}} = 27)$. We used generalized linear mixed models with a Poisson error structure and maternal identity as random factor, to analyse the variation in total calf production of daughters (≥ 2 years of age). Explanatory variables included natural log transformed age at death or age at last observation of the daughters, the

mother total calf production divided by her age at death or age at last observation (mean age-specifc calf production), maternal age and her age at primiparity and their interactions as covariates. Finally, we analyzed the variation in daughter body mass in relation to mothers' age, age at primiparity, and total calf production, as well as birth date, twin status and their interactions as covariates. We used generalized linear mixed models with a Gaussian error structure with year and maternal identity as random factors.

Akaike's Information Criterion corrected for sample size $(AIC_{\alpha},$ Burnham and Anderson 2002) was used for model selection; the best models being the most parsimonious (lowest AIC_c value). We also computed Akaike weights (AIC_{c-w}) to compare the relative performance of different models (Burnham and Anderson [2002](#page-9-30)).

Results

Age at primiparity and calf production

During the study period, 54% of all 2-year-old females $(N = 67)$ reproduced successfully and 7% produced twins. At age 3, 82% of all females reproduced ($N = 57$, among which 35% were primiparous) and 37% had twins. 37% of mature females skipped reproduction in one or several years throughout their life, or they lost their calves too early to be detected. Total calf production was positively related to age at last observation in interaction with age at primiparity (AIC_{c-w} = 0.407, evidence ratio to the second best model = 1.109; $\beta \pm \text{SE} = 0.43 \pm 0.28$, 95% CI – 0.10; 0.99). The accumulated calf production of females at the age at last observation or death was on average higher for those starting to reproduce at the age of two compared to those that reproduced successfully as 3- or 4-year-olds (Fig. [1](#page-4-0)). For a typical life span of 5 years old, a female that started to reproduce as a 2-year-old produced 4.95 calves, while a female that started at the age of three or four produced on average 3.29 calves.

Individual heterogeneity in life history traits

The best model explaining the probability to reproduce in females included interactive effects of age class and the previous and current reproductive state (Table [1A](#page-5-0)). Prime-aged females (3–9 years old; $N = 63$) had higher probability to reproduce than females aged 10 years and older $(N = 20)$, independent of their reproductive success the previous year (for females that had not reproduced the year before: $\psi^{NR \to R} = 0.84 \pm 0.07$ vs $\psi^{NR \to R} = 0.70 \pm 0.08$; for females that had a singleton: $\psi^{1C\rightarrow R} = 0.86 \pm 0.04$ vs $\psi^{1C\rightarrow R} = 0.73 \pm 0.06$; for females that had twins: $\psi^{2C\rightarrow R} = 0.94 \pm 0.04$ vs $\psi^{2C\rightarrow R} = 0.88 \pm 0.05$; for 3–9 years

Fig. 1 Mean total calf production at age at last observation (age at death or at the end of study period if still alive) \pm SE for female moose according to their age at primiparity, with regression lines \pm SE. Black square, solid line Age at primiparity as 2-year-old $(n = 36)$; Black circle, dashed line age at primiparity as $3-4$ -year-old ($n = 23$). Age at primiparity 2: ln(Total calf production) = $(1.36 \pm 0.13 \times \ln(A)$ at last observation)) -0.59 ± 0.27 ; Age at primiparity 3–4: ln(Total calf production) = $(1.79 \pm 0.35 \times \ln(A)$ at last observation)) – 1.69 ± 0.61

 $vs \geq 10$ years old, respectively; Fig. [2;](#page-5-1) Table [2](#page-6-0)A). Moreover, the probability to produce twins was lower for females that were unsuccessful the year before than for females that produced singletons or twins, indicating a strong quality efect on calf production (Fig. [2](#page-5-1); Table [2A](#page-6-0)).

Age at primiparity also influenced reproductive success in a more complex way. For females with at least one previous reproductive event, the reproduction transitions were best explained by also accounting for the threeway interaction between age class, age at primiparity and current reproductive state (Table [1B](#page-5-0)). Early maturing females showed a slightly higher probability to reproduce (singletons and twins, age 3–9: $\psi^{NR \to R} = 0.88 \pm 0.08$, $\psi^{1C\rightarrow R} = 0.87 \pm 0.05$, $\psi^{2C\rightarrow R} = 0.94 \pm 0.05$) compared to those that started to reproduce at the age of 3 or 4 $(\text{age } 3-9: \psi^{\text{NR}\rightarrow \text{R}} = 0.83 \pm 0.08, \psi^{\text{IC}\rightarrow \text{R}} = 0.83 \pm 0.06,$ $\psi^{2C\rightarrow R}$ = 0.92 \pm 0.06; Fig. [3;](#page-7-0) Table [2](#page-6-0)B). This heterogeneity

was more pronounced when females were ≥ 10 years old (for early maturing females: $\psi^{NR\rightarrow R} = 0.76 \pm 0.08$, $\psi^{1}C\rightarrow R$ = 0.75 ± 0.06, $\psi^{2}C\rightarrow R$ = 0.88 ± 0.04; for late maturing females: $\psi^{NR \to R} = 0.69 \pm 0.08$, $\psi^{1C \to R} = 0.68 \pm 0.08$, $v^{2C\rightarrow R} = 0.84 + 0.06$.

For yearlings, the probability to be killed by hunting was 0.33 ± 0.04 (95% CI 0.25; 0.42), while for 2 years and older the hunting mortality depended on their reproductive status before the hunting season. Females that were accompanied by a calf had almost half the probability to be killed during hunting than females without calf, and the probability was lowest for those accompanied by twins [age 3–9: $\eta^{NR} = 0.18 \pm 0.06$ (95% CI 0.10; 0.33); $\eta^{1C} = 0.09 \pm 0.03$ $(95\% \text{ CI } 0.05; 0.17); \eta^{2} = 0.05 \pm 0.02 (95\% \text{ CI } 0.03; 0.10)$.

Adding females body mass as calf did not improve the model fit ($\Delta AIC_c = 16.74$; Table [1C](#page-5-0)), possibly because of its close covariation with age at primiparity. Indeed, females that started to reproduce at the age of 2 had nearly signifcant higher mean body mass as calf ($\beta \pm SE = 186.09 \pm 4.22$ kg, 95% CI 177.53; 194.65, *N* = 21) than late-maturing females $(\beta \pm \text{SE} = 173.44 \pm 6.42 \text{ kg}, 95\% \text{ CI} 151.86; 195.03,$ $N = 16$). Hence, body mass in early life seems to have no other efects on reproductive performance than the indirect effects on age at primiparity and current reproductive status.

Fitness cost of recruitment

Caring for a calf/calves during winter did not seem to increase the cost of reproduction as the probability to reproduce next year was similar for females whose calves survived or were lost during hunting (for early maturing females aged ≥ 3 , $N = 33$: $\psi^{\text{Calves recruit}\rightarrow R} = 0.90 \pm 0.04$, $N_{\text{females}} = N_{\text{transitions}} = 78$; $\psi^{\text{Calves lost}\rightarrow R} = 0.91 \pm 0.06$, $N_{\text{transitions}} = 43$; for late maturing females aged ≥ 3 , $N = 23$: $\psi^{\text{Calves} \text{ recruit}\rightarrow R} = 0.87 \pm 0.06$, $N_{\text{transitions}} = 49$; $\psi^{\text{Calves lost}\rightarrow R} = 0.85 \pm 0.06$, $N_{\text{transitions}} = 34$; Tables [1](#page-5-0)D, [2C](#page-6-0)). Females that successfully recruited calves to the adult population had a higher probability of giving birth to twins (for early maturing females aged ≥ 3 : $\psi^{\text{Calves recruitment}} = 0.59 \pm 0.05$, $N_{\text{transitions}} = 53$; for late maturing females aged ≥ 3 : $\psi^{\text{Calves} \, \text{recruit}\rightarrow 2C} = 0.61 \pm 0.06$, $N_{\text{transitions}}$ = 33) than singletons (for early maturing females aged ≥ 3 : $\psi^{\text{Calves recruitment}}$ = 0.31 \pm 0.04, $N_{\text{transitions}} = 27$; for late maturing females aged ≥ 3 : ψ^{Calves} recruit→1C = 0.26 ± 0.06 , $N_{\text{transitions}} = 17$; Fig. [4,](#page-7-1) Table [2C](#page-6-0)).

For females that started to reproduce as 2-year-olds and that gave birth to twins, the probability that both of the twin calves was hunter-killed was 0.36 ± 0.05 , while it was 0.28 ± 0.06 for those who started to reproduce later. If the mother had only a single calf, the probability that it was harvested was 0.25 ± 0.05 for early and 0.46 ± 0.08 for late maturing individuals.

Table 1 Results from model selection of multistate models for estimation of hunting mortality (*η*), natural mortality (*φ*) and future reproductive performance (*ψ*) for female moose on Vega, Norway, 1984–2012, for (A) all females, (B) females with at least one reproductive event, (C) including calf body mass and (D) reproductive females with hunting mortality of their calves

Only the three highest ranked models according to AIC_c value are shown. The effects considered were variation between age classes (Age), previous reproductive state (PR), current reproductive state (CR), age at primiparity (AP), calf body mass (BM). The symbol " \times " was used to denote the statistical interaction between the efects, while the symbol "+" was used for additive efects. For further information about models, see "[Materials and methods"](#page-2-0). *K* number of parameters in the model. AIC_c weights (AIC_{c-w}), defned as the relative likelihood of a model given the data and set of models. The best model in (A) had an AIC_c = 1076.208, (B) AIC_c = 830.632, (C) AIC_c = 490.345 and (D) AIC_c = 1167.493

Fig. 2 Probability to reproduce one (ψ^1) or two (ψ^2) calves \pm SE for female moose according to age class and their previous reproductive state. Black square Not reproduced (NR), black circle reproduced 1 calf, and black triangle reproduced two calves at previous state. Probabilities are from the best model in Table [1](#page-5-0)A

Intergenerational correlations

There was no signifcant association between the age at primiparity of mothers and their offspring $(\chi^2 = 1.32)$, $p = 0.251$). The variation in the daughter's calf production was best explained by their age at last observation $(\beta = 1.53 \pm 0.13, \text{ AIC}_{c-w} = 0.339, \text{ evidence ratio to the})$ second best model $= 1.215$) and the mother's mean agespecific calf production ($\beta = 0.63 \pm 0.30$). Mothers with above average reproductive success were more likely to produce daughters that also performed well in terms of reproductive success. The variation in daughter's body mass was best explained by the age of the mother ($AIC_{c-w} = 0.240$, evidence ratio to the second best model $= 1.437$). Older mothers gave birth to heavier calves irrespective of their age at primiparity (β = 2.91 \pm 0.89).

Discussion

Twenty-eight years of detailed reproductive histories of female moose revealed positive relationships between current and future reproduction. Early age at primiparity was associated with higher total calf production (Fig. [1](#page-4-0)), and successful breeders had a higher probability to reproduce at similar levels in subsequent years (Figs. [2](#page-5-1), [3](#page-7-0)). Moreover, the probability to reproduce twins was higher for those

Table 2 Parameter estimates \pm SE (95% CI) for the transition probabilities (from state \rightarrow to state) of future reproductive performance (ψ) for female moose on Vega, Norway, [1](#page-5-0)984–2012, from the highest ranked models according to AIC_c model selection in Table 1

Model	Age	Group	From state	Parameter estimates SE (95% CI) To state		
				(A)	$\sqrt{2}$	
$3 - 9$		NR	0.55 ± 0.07 (0.41; 0.69)		0.29 ± 0.07 (0.18; 0.43)	
		1 ^C	0.29 ± 0.04 (0.21; 0.39)		0.57 ± 0.05 (0.47; 0.66)	
		$2\mathrm{C}$	0.32 ± 0.04 (0.25; 0.41)		0.62 ± 0.04 (0.53; 0.70)	
$10+$		NR	0.46 ± 0.09 (0.29; 0.64)		0.24 ± 0.07 (0.13; 0.39)	
		1 ^C	0.25 ± 0.05 (0.17; 0.36)		0.48 ± 0.07 (0.34; 0.62)	
		$2\mathrm{C}$	0.30 ± 0.04 (0.23; 0.39)		0.58 ± 0.05 (0.48; 0.67)	
(B)	\overline{c}	AP ₂	NR	0.83 ± 0.06 (0.68; 0.92)	0.14 ± 0.06 (0.06; 0.29)	
	$3 - 9$		$\rm NR$	0.55 ± 0.09 (0.38; 0.71)	0.33 ± 0.08 (0.19; 0.50)	
			1 ^C	0.31 ± 0.05 (0.22; 0.41)	0.56 ± 0.05 (0.46; 0.67)	
			2C	0.32 ± 0.05 (0.24; 0.42)	0.62 ± 0.05 (0.52; 0.71)	
	$10+$		NR	0.48 ± 0.09 (0.29; 0.67)	0.29 ± 0.08 (0.16; 0.46)	
			1 ^C	0.27 ± 0.05 (0.32; 0.55)	0.49 ± 0.07 (0.35; 0.63)	
			2C	0.30 ± 0.05 (0.22; 0.40)	0.58 ± 0.05 (0.47; 0.68)	
	$3 - 9$	$AP3+$	NR	0.51 ± 0.08 (0.36; 0.65)	0.32 ± 0.07 (0.20; 0.47)	
			$1\mathrm{C}$	0.28 ± 0.06 (0.18; 0.40)	0.55 ± 0.07 (0.42; 0.67)	
			$2\mathrm{C}$	0.30 ± 0.05 (0.21; 0.42)	0.62 ± 0.06 (0.50; 0.72)	
	$10+$		$\rm NR$	0.42 ± 0.09 (0.29; 0.64)	0.27 ± 0.07 (0.15; 0.43)	
			1 ^C	0.23 ± 0.06 (0.13; 0.37)	0.45 ± 0.09 (0.29; 0.63)	
			2C	0.28 ± 0.05 (0.18; 0.42)	0.56 ± 0.07 (0.48; 0.67)	
(C)	2	AP ₂	Not recruit	0.83 ± 0.06 (0.68; 0.92)	0.14 ± 0.06 (0.06; 0.29)	
	$3+$		1 ^C			0.75 ± 0.05 (0.64; 0.83)
			2C			0.64 ± 0.05 (0.54; 0.73)
			Recruit	0.31 ± 0.04 (0.23; 0.40)	0.59 ± 0.05 (0.50; 0.68)	
			Not recruit	0.39 ± 0.07 (0.27; 0.53)	0.52 ± 0.07 (0.39; 0.64)	
	$3+$	$AP3+$	1 ^C			0.54 ± 0.08 (0.38; 0.69)
			2C			0.72 ± 0.06 (0.59; 0.83)
			Recruit	0.26 ± 0.06 (0.17; 0.39)	0.61 ± 0.06 (0.48; 0.73)	
			Not recruit	0.45 ± 0.06 (0.33; 0.58)	0.40 ± 0.06 (0.29; 0.53)	

(A) All females, (B) females with at least one reproductive event, and (C) reproductive females with hunting mortality of their calves. Age $=$ age classes, Group: $AP2$ = age at primiparity as 2 year olds, $AP3$ + = age at primiparity as 3 years old and older. States: NR = not reproduced, 1C = reproduced one calf, 2C = reproduced two calves, not recruit = no calf/calves survived to year *t* + 1, recruit = one or two calves survived to year $t + 1$

females that already had successfully recruited calves into the population (Fig. [4\)](#page-7-1). This suggests that the ftness contribution to the next generation from reproduction and survival is substantially higher in some females than in others and could potentially last over generations since high-performing females were correlated with high-performing calves in terms of total calf production.

The among-females diferences in calf production were partly related to individual diferences in age at primiparity; the most successful females matured at the age of 2 years. Such a positive relationship between age at maturity and other ftness components has also been found in other moose

populations (Sæther and Haagenrud [1983,](#page-11-14) [1985\)](#page-11-15), as well as in other ungulates (Moyes et al. [2006](#page-10-20); Weladji et al. [2008\)](#page-11-6) and long-lived mammals (Pistorius et al. [2001;](#page-10-21) Moore et al. [2016\)](#page-10-11). However, it is not found in some long-lived birds, where early maturing individuals tend to suffer higher mortality rates (e.g. Pyle et al. [1997;](#page-10-7) Barbraud and Weimerskirch [2005](#page-9-31)).

In many ungulate species, age at maturity is often inversely related to body mass: large females mature earlier than smaller ones (Jorgenson et al. [1993;](#page-10-10) Gaillard et al. [2000a](#page-9-8)). This also occurs at Vega where a favourable environment results in high body masses (Solberg et al. [2015\)](#page-11-16) and

Fig. 3 Probability to reproduce one (ψ^1) or two (ψ^2) calves \pm SE for female moose according to age at primiparity, age class and their previous reproductive state. Black square Not reproduced (NR), black circle reproduced 1 calf, and black triangle reproduced 2 calves at previous state. Probabilities are from the best model in Table [1](#page-5-0)B

high population growth rates (Sæther et al. [2007](#page-11-12)). Still, there is large inter-individual variation in offspring production among females. This shows that individual heterogeneity does not only occur in populations facing stressful environmental conditions (Lomnicki [1978](#page-10-22)) and suggests that there are consistent individual diferences among females in their ability to acquire resources (van Noordwijk and de Jong [1986](#page-11-3)). In moose, variation in body mass of young females may be related to availability of high-quality food plants during summer (Sæther [1985;](#page-10-23) Sæther and Heim [1993](#page-11-5)) or variation in snow conditions during late winter (Sæther and Gravem [1988\)](#page-11-17). This indicates that individual diferences in calf production may be related to spatial heterogeneity in quality of the females' home ranges (Pettorelli et al. [2001](#page-10-24)), a hypothesis that should be tested further as we currently lack studies showing how heterogeneity in habitat quality transfers to heterogeneity in individual demographic performances in large herbivores (Gaillard et al. [2010](#page-9-32)).

Individual diferences in female reproductive success may be related to the timing of conception, which can be afected by the sex- and age-structure of the population. On Vega, the sex and age structure was manipulated by selective

Fig. 4 Probability to reproduce one (ψ^1) or two (ψ^2) calves \pm SE for female moose dependent on whether her previous calf/calves survived the hunting season. Black square At least one calf survived the hunt, black circle calf/calves lost during the hunt. Probabilities are from the best model in Table [1](#page-5-0)D

harvesting during the period 1994–1999 (Sæther et al. [2003\)](#page-11-10) leading to a female-biased sex ratio and low mean male age. A mating preference for older males seems to make females inclined to delay conception when the sex ratio is very biased (Sæther et al. [2003](#page-11-10), [2004;](#page-11-11) Herfndal et al. [2014](#page-10-25)). This in turn results in relatively large variation in calving date, which can affect the demography. Such effects may provide a link between social organization, demography and population dynamics in moose.

The reproductive costs associated with postweaning maternal care have rarely been investigated in ungulates (Green et al. [1989](#page-9-16); Andersen et al. [2000;](#page-9-17) Holand et al. [2012](#page-10-16); Froy et al. [2016\)](#page-9-18). L'Heureux et al. ([1995](#page-10-12)) investigated postweaning mother–daughter associations in bighorn sheep and concluded that post-weaning associations did not appear costly to mothers. Costs to mothers of protracted associations with ofspring have, however, been found in red deer (*Cervus elaphus*; Clutton-Brock et al. [1982](#page-9-21); Froy et al. [2016\)](#page-9-18), red-necked wallaby (*Macropus rufogriseus*; Johnson [1986\)](#page-10-26) and bison (*Bison bison*; Green et al. [1989\)](#page-9-16). Loss of calves from hunting did not seem to increase the probability to reproduce for mothers, indicating that maternal post-weaning costs may be of minor importance in moose and that calf harvest has no or little effect on female reproduction. Instead, females that recruited calves to the adult population had higher probability to give birth to twins the next year, reinforcing the individual quality hypothesis.

The cost of reproduction in ungulates is more pronounced in young and old individuals (Clutton-Brock et al. [1983](#page-9-14); Tavecchia et al. [2005](#page-11-9)). There seems to be a threshold in body mass that females must reach to reproduce, and they may take reproductive pauses or adjust the effort later in the reproductive cycle (e.g. abortion; Hamel et al. [2009b;](#page-9-3) Milner et al. [2013\)](#page-10-27) according to their current body reserves. As females reach prime-age, the cost of reproduction is likely lowered by the completion of body growth and increased maternal experience (Weladji et al. [2006](#page-11-18)), giving a higher probability to reproduce and recruit ofspring into the population. At old age, senescence leads to reductions in survival and/or breeding success (e.g. Loison et al. [1999](#page-10-28); Lemaître et al. [2015\)](#page-10-29), which is observed as lower reproductive probability in the old females.

The cost of earlier age of frst reproduction depends on the relative position of a species along the fast–slow continuum (Gaillard et al. [1989](#page-9-33); Sæther and Bakke [2000](#page-10-30)). Moose have a rapid life cycle with short generation time and high reproductive output compared to other, similar-sized ungulates (Gaillard [2007;](#page-9-34) Gaillard et al. [2000a\)](#page-9-8). To reach their high reproductive output, moose have to trade quality of offspring (small newborns compared to newborns of other, similar-sized ungulates) for a higher quantity, as female moose produce twins as soon as environmental conditions allow. Probability to produce twins is, therefore, an additional indication of individual quality which infuence the extent of the inter-individual heterogeneity found in this study.

Being under high hunting pressure, female moose may try to reproduce as early as possible, and not be as conservative as, e.g. bighorn sheep (Bérubé et al. [1999\)](#page-9-5) or red deer (Moyes et al. [2006\)](#page-10-20). At Vega, this effect is enhanced by differential hunting mortality, as hunters kill non-reproducing females with a higher probability than females accompanied by one or two calves (see Ericsson [2001](#page-9-35) for similar results). As females are observed with similar probability independent of calf status (Rolandsen et al. [2003;](#page-10-31) Solberg et al. [2010](#page-11-19)), we believe that the varying hunting mortality rate is because hunters avoid killing calf-rearing females and not because non-producing females are more easily detected. Females can, therefore, increase their survival probability by reproducing early and may gain a ftness beneft even if the capacity for future reproduction is reduced (Nilsen and Solberg [2006\)](#page-10-32). However, so far we see no indications of such a trade-off between current and future reproductive performance in moose on Vega.

Contrary to the Vega population, several ungulate populations, throughout the world, has natural predation as an important source of mortality. Natural predators can afect ungulate demography both through predation-associated mortality (Krivan and Eisner [2003\)](#page-10-33) and indirectly through the costs of anti-predator behavioural and physiological responses (Creel and Christianson [2008\)](#page-9-36) leading to, e.g. changes in habitat use (Creel et al. [2005](#page-9-37)). Predation can change the distribution of life history traits over generations through prey selectivity (Reznick et al. [1990\)](#page-10-34). Dependent on their selectivity and the preys individual variation in predation risk (due to, e.g. diferences among the sexes, age, phenotypic attributes, habitat use), predation can lead to higher or lower between-individual variation in reproductive success of prey (Pettorelli et al. [2011](#page-10-35)).

Theoretical studies show that individual heterogeneity in demography may strongly afect the population dynamics by, e.g. afecting the demographic variance (e.g. Vindenes et al. [2008](#page-11-20); Plard et al. [2016\)](#page-10-36) and temporal variation in genetic composition (Lee et al. [2011\)](#page-10-37), and should be accounted for in population models applied to moose. Recent studies have incorporated individual heterogeneity using mixed or mixture models (e.g. Cam et al. [2002](#page-9-1); Péron et al. [2010](#page-10-38); Chambert et al. [2013\)](#page-9-4). In our case, we grouped the individuals a priori into groups according to their previous reproductive status and age at primiparity. The groupings of individuals may successfully account for the major axes of variation in the data, but could in theory bias the parameters.

To summarize, our study of individual variation in reproductive success among female moose showed individual heterogeneity in demography, associated with individual diferences in age at maturity. These diferences were not afected by variation in reproductive costs of tending for a calf after the end of lactation.

Acknowledgements We are grateful to the local moose hunters and managers for cooperation in collecting data. This study was supported by the European Research Council (ERC-2010-AdG 268562), the Research Council of Norway (SFF-III 223257/F50 and project number 244647) and the Norwegian Environment Agency.

Author contribution statement SSM, AL and BES developed the concept and idea. HH performed genetic analyses and constructed the pedigree in collaboration with KHR. Statistical analyses were performed by SSM with input from AL, IH, EJS and BES. MH was in charge of the feldwork, tissue sampling and the life history databases. SSM wrote the article with input from all co-authors.

Compliance with ethical standards

Conflict of interest The authors have no conficts of interest to declare.

Ethical statement All applicable national guidelines for the care and use of animals were followed.

References

- Andersen R, Gaillard J-M, Linnell JDC, Duncan P (2000) Factors affecting maternal care in an income breeder, the European roe deer. J Anim Ecol 69:672–682. [https://doi.](https://doi.org/10.1046/j.1365-2656.2000.00425.x) [org/10.1046/j.1365-2656.2000.00425.x](https://doi.org/10.1046/j.1365-2656.2000.00425.x)
- Barbraud C, Weimerskirch H (2005) Environmental conditions and breeding experience afect costs of reproduction in blue petrels. Ecology 86:682–692.<https://doi.org/10.1890/04-0075>
- Bell G (1980) The costs of reproduction and their consequences. Am Nat 116:45–76.<https://doi.org/10.1086/283611>
- Bell G, Koufopanou V (1986) The cost of reproduction. Oxf Surv Evol Biol 3:83–131
- Bérubé CH, Festa-Bianchet M, Jorgenson JT (1999) Individual differences, longevity and reproductive senescence in bighorn ewes. Ecology 80:2555–2565. <https://doi.org/10.2307/177240>
- Brownie C, Hines JE, Nichols JD, Pollock KH, Hestbeck JB (1993) Capture-recapture studies for multiple strata including non-Markovian transitions. Biometrics 49:1173–1187. [https://doi.](https://doi.org/10.2307/2532259) [org/10.2307/2532259](https://doi.org/10.2307/2532259)
- Bubenik AB (1998) Behavior. In: Franzmann AW, Schwartz CC (eds) Ecology and management of the North American moose. Smithsonian Inst Press, Washington, pp 173–222
- Burnham KP, Anderson DR (2002) Model selection and multimodel inference—a practical information-theoretic approach, 2nd edn. Springer, New York
- Cam E, Link WA, Cooch EG, Monnat JY, Danchin E (2002) Individual covariation in life-history traits: seeing the trees despite the forest. Am Nat 159:96–105.<https://doi.org/10.1086/324126>
- Chambert T, Rotella JJ, Higgs MD, Garrott RA (2013) Individual heterogeneity in reproductive rates and cost of reproduction in a long-lived vertebrate. Ecol Evol 3:2047–2060. [https://doi.](https://doi.org/10.1002/ece3.615) [org/10.1002/ece3.615](https://doi.org/10.1002/ece3.615)
- Choquet R, Rouan L, Pradel R (2009a) Program E-surge: a software for ftting multievent models. In: Thomson DL, Cooch EG, Conroy MJ (eds) Modelling demographic processes in marked populations. Springer, Berlin, pp 845–865
- Choquet R, Lebreton JD, Gimenez O, Reboulet AM, Pradel R (2009b) U-CARE: utilities for performing goodness of ft tests and manipulating CApture–REcapture data. Ecography 32:1071– 1074.<https://doi.org/10.1111/j.1600-0587.2009.05968.x>
- Clutton-Brock TH (1991) The evolution of parental care. Princeton University Press, Princeton
- Clutton-Brock TH, Guinness FE, Albon SD (1982) Red deer: behaviour and ecology of two sexes. University of Chicago Press, Chicago
- Clutton-Brock TH, Guinness FE, Albon SD (1983) The costs of reproduction to red deer hinds. J Anim Ecol 52:367–383. [https://](https://doi.org/10.2307/4560) doi.org/10.2307/4560
- Cox RM, Calsbeek R (2010) Severe costs of reproduction persist in Anolis lizards despite the evolution of a single-egg clutch. Evolution 64:1321–1330. [https://doi.](https://doi.org/10.1111/j.1558-5646.2009.00906.x) [org/10.1111/j.1558-5646.2009.00906.x](https://doi.org/10.1111/j.1558-5646.2009.00906.x)
- Creel S, Christianson D (2008) Relationships between direct predation and risk efects. Trends Ecol Evol 23:194–201. [https://doi.](https://doi.org/10.1016/j.tree.2007.12.004) [org/10.1016/j.tree.2007.12.004](https://doi.org/10.1016/j.tree.2007.12.004)
- Creel S, Winnie J Jr, Maxwell B, Hamlin K, Creel M (2005) Elk alter habitat selection as an antipredator response to wolves. Ecology 86:3387–3397. <https://doi.org/10.1890/05-0032>
- Ericsson G (2001) Reversed cost of reproduction in moose *Alces alces* through human harvest. Alces 37:61–69
- Festa-Bianchet M, Jorgenson JT, Lucherini M, Wishart WD (1994) Early weaning in bighorn sheep, *Ovis canadensis*, affects growth of males but not of females. Behav Ecol 6:21–27. [https://](https://doi.org/10.1093/beheco/5.1.21) doi.org/10.1093/beheco/5.1.21
- Festa-Bianchet M, Jorgenson JT, Lucherini M, Wishart WD (1995) Life history consequences of variation in age of primiparity in bighorn ewes. Ecology 76:871–881. [https://doi.](https://doi.org/10.2307/1939352) [org/10.2307/1939352](https://doi.org/10.2307/1939352)
- Festa-Bianchet M, Gaillard J-M, Jorgenson JT (1998) Mass- and density-dependent reproductive success and reproductive costs in a capital breeder. Am Nat 152:367–379. [https://doi.](https://doi.org/10.1086/286175) [org/10.1086/286175](https://doi.org/10.1086/286175)
- Froy H, Walling CA, Pemberton JM, Clutton-Brock TH, Kruuk LEB (2016) Relative costs of ofspring sex and ofspring survival in a polygynous mammal. Biol Lett 12:201604117. [https://doi.](https://doi.org/10.1098/rsbl.2016.0417) [org/10.1098/rsbl.2016.0417](https://doi.org/10.1098/rsbl.2016.0417)
- Gaillard J-M (2007) Are moose only a large deer?: some life history considerations. Alces 43:1–11
- Gaillard J-M, Pontier D, Allaine D, Lebreton JD, Trouvilliez J, Clobert J (1989) An analysis of demographic tactics in birds and mammals. Oikos 56:59–76. <https://doi.org/10.2307/3566088>
- Gaillard J-M, Festa-Bianchet M, Yoccoz NG, Loison A, Toigo C (2000a) Temporal variation in ftness components and population dynamics of large herbivores. Annu Rev Ecol Syst 31:367–393. <https://doi.org/10.1146/annurev.ecolsys.31.1.367>
- Gaillard J-M, Festa-Bianchet M, Delorme MD, Jorgenson JT (2000b) Body mass and individual ftness in female ungulates: bigger is not always better. Proc R Soc B Biol Sci 267:471–477. [https://doi.](https://doi.org/10.1098/rspb.2000.1024) [org/10.1098/rspb.2000.1024](https://doi.org/10.1098/rspb.2000.1024)
- Gaillard J-M, Hebblewhite M, Loison A, Fuller M, Powell R, Basille M, van Moorter B (2010) Habitat–performance relationships: fnding the right metric at a given spatial scale. Philos Trans R Soc Lond B Biol Sci 365:2255–2265. [https://doi.org/10.1098/](https://doi.org/10.1098/rstb.2010.0085) [rstb.2010.0085](https://doi.org/10.1098/rstb.2010.0085)
- Garel M, Solberg EJ, Sæther B-E, Grøtan V, Tufto J, Heim M (2009) Age, size, and spatiotemporal variation in ovulation patterns of a seasonal breeder, the Norwegian moose (*Alces alces*). Am Nat 173:89–104.<https://doi.org/10.1086/593359>
- Gélin U, Wilson ME, Coulson G, Festa-Bianchet M (2015) Experimental manipulation of female reproduction demonstrates its ftness costs in kangaroos. J Anim Ecol 84:239–248. [https://doi.](https://doi.org/10.1111/1365-2656.12266) [org/10.1111/1365-2656.12266](https://doi.org/10.1111/1365-2656.12266)
- Green WCH, Rothstein A (1991) Trade-offs between growth and reproduction in female bison. Oecologia 86:521–527. [https://doi.](https://doi.org/10.1007/BF00318318) [org/10.1007/BF00318318](https://doi.org/10.1007/BF00318318)
- Green WCH, Griswold JG, Rothstein A (1989) Post-weaning associations among bison mothers and daughters. Anim Behav 38:847– 858. [https://doi.org/10.1016/S0003-3472\(89\)80116-2](https://doi.org/10.1016/S0003-3472(89)80116-2)
- Haanes H, Markussen SS, Herfndal I, Røed KH, Solberg EJ, Heim M, Midthjell L, Sæther B-E (2013) Efects of inbreeding on ftnessrelated traits in a small isolated moose population. Ecol Evol 3:4230–4242. <https://doi.org/10.1002/ece3.819>
- Hamel S, Gaillard J-M, Festa-Bianchet M, Côté SD (2009a) Individual quality, early-life conditions, and reproductive success in contrasted populations of large herbivores. Ecology 90:1981–1995. <https://doi.org/10.1890/08-0596.1>
- Hamel S, Côté SD, Gaillard J-M, Festa-Bianchet M (2009b) Individual variation in reproductive costs of reproduction: high-quality females always do better. J Anim Ecol 78:143–151. [https://doi.](https://doi.org/10.1111/j.1365-2656.2008.01459.x) [org/10.1111/j.1365-2656.2008.01459.x](https://doi.org/10.1111/j.1365-2656.2008.01459.x)
- Hamel S, Gaillard J-M, Yoccoz NG, Loison A, Bonenfant C, Descamps S (2010) Fitness costs of reproduction depend on life speed: empirical evidence from mammalian populations. Ecol Lett 13:915–935.<https://doi.org/10.1111/j.1461-0248.2010.01478.x>
- Hamel S, Craine JM, Towne EG (2012) Maternal allocation in bison: co-occurrence of senescence, cost of reproduction, and individual quality. Ecol Appl 22:1628–1639. [https://doi.](https://doi.org/10.2307/41722878) [org/10.2307/41722878](https://doi.org/10.2307/41722878)
- Herfndal I, Solberg EJ, Sæther B-E, Hogda KA, Andersen R (2006) Environmental phenology and geographical gradients in moose

body mass. Oecologia 150:213–224. [https://doi.org/10.1007/](https://doi.org/10.1007/s00442-006-0519-8) [s00442-006-0519-8](https://doi.org/10.1007/s00442-006-0519-8)

- Herfndal I, Haanes H, Røed KH, Solberg EJ, Markussen SS, Heim M, Sæther B-E (2014) Population properties affect inbreeding avoidance in moose. Biol Lett 10:20140786. [https://doi.org/10.1098/](https://doi.org/10.1098/rsbl.2014.0786) [rsbl.2014.0786](https://doi.org/10.1098/rsbl.2014.0786)
- Herfndal I, van de Pol M, Nielsen JT, Møller AP, Sæther B-E (2015) Climatic conditions cause complex patterns of covariation between demographic traits in a long-lived raptor. J Anim Ecol 84:702–711.<https://doi.org/10.1111/1365-2656.12318>
- Holand Ø, Weladji RB, Mysterud A, Røed KH, Reimers E, Nieminen M (2012) Induced orphaning reveals post-weaning maternal care in reindeer. Eur J Wildl Res 58:589–596. [https://doi.org/10.1007/](https://doi.org/10.1007/s10344-012-0608-4) [s10344-012-0608-4](https://doi.org/10.1007/s10344-012-0608-4)
- Johnson CN (1986) Philopatry, reproductive success of females, and maternal investment in the red-necked wallaby. Behav Ecol Sociobiol 19:143–150. <https://doi.org/10.1007/BF00299949>
- Jorgenson JT, Festa-Bianchet M, Lucherini M, Wishart WD (1993) Efects of body size, population density, and maternal characteristics on age at frst reproduction in bighorn ewes. Can J Zool 71:2509–2517. <https://doi.org/10.1139/z93-344>
- Kendall BE, Wittmann ME (2010) A stochastic model for annual reproductive success. Am Nat 175:461–468. [https://doi.](https://doi.org/10.1086/650724) [org/10.1086/650724](https://doi.org/10.1086/650724)
- Krivan V, Eisner J (2003) Optimal foraging and predator–prey dynamics III. Theor Pop Biol 63:269–279. [https://doi.org/10.1016/](https://doi.org/10.1016/S0040-5809(03)00012-1) [S0040-5809\(03\)00012-1](https://doi.org/10.1016/S0040-5809(03)00012-1)
- Kvalnes T, Sæther B-E, Haanes H, Røed KH, Engen S, Solberg EJ (2016) Harvest-induced phenotypic selection in an island population of moose, *Alces alces*. Evolution 70:1486–1500
- L'Heureux N, Lucherini M, Festa-Bianchet M, Jorgenson JT (1995) Density-dependent mother-yearling association in bighorn sheep. Anim Behav 49:301–910.<https://doi.org/10.1006/anbe.1995.0122>
- Lande R (1982) A quantitative genetic theory of life-history evolution. Ecology 63:607–615.<https://doi.org/10.1111/evo.12952>
- Lebreton JD, Nichols JD, Barker RJ, Pradel R, Spendelow JA (2009) Modelling individual animal histories with multistate capture–recapture models. Adv Ecol Res 41:87–173. [https://doi.](https://doi.org/10.1016/S0065-2504(09)00403-6) [org/10.1016/S0065-2504\(09\)00403-6](https://doi.org/10.1016/S0065-2504(09)00403-6)
- Lee AM, Engen S, Sæther B-E (2011) The infuence of persistent individual differences and age at maturity on effective population size. Proc R Soc Lond B Biol Sci 278:3303–3312. [https://doi.](https://doi.org/10.1098/rspb.2011.0283) [org/10.1098/rspb.2011.0283](https://doi.org/10.1098/rspb.2011.0283)
- Lemaître J-F, Berger V, Bonenfant C, Douhard M, Gamelon M, Plard F, Gaillard J-M (2015) Early-late life trade-ofs and the evolution of ageing in the wild. Proc R Soc Lond B Biol Sci 282:20150209. <https://doi.org/10.1098/rspb.2015.0209>
- Lindèn M, Møller AP (1989) Cost of reproduction and covariation of life history traits in birds. Trends Ecol Evol 4:367–371. [https://](https://doi.org/10.1016/0169-5347(89)90101-8) [doi.org/10.1016/0169-5347\(89\)90101-8](https://doi.org/10.1016/0169-5347(89)90101-8)
- Loison A, Festa-Bianchet M, Gaillard J-M, Jorgenson JT, Jullien J (1999) Age-specifc survival in fve populations of ungulates: evidence of senescence. Ecology 80:2539–2554. [https://doi.](https://doi.org/10.2307/177239) [org/10.2307/177239](https://doi.org/10.2307/177239)
- Lomnicki A (1978) Individual diferences between animals and the natural regulation of their numbers. J Anim Ecol 47:461–475. <https://doi.org/10.2307/3794>
- McNamara JM, Houston AI (1996) State-dependent life histories. Nature 380:215–221. <https://doi.org/10.1038/380215a0>
- Michod RE (1979) Evolution of life histories in response to agespecific mortality factors. Am Nat 113:531–550. [https://doi.](https://doi.org/10.1086/283411) [org/10.1086/283411](https://doi.org/10.1086/283411)
- Milner JM, van Beest FM, Solberg EJ, Storaas T (2013) Reproductive success and failure: the role of winter body mass in reproductive allocation in Norwegian moose. Oecologia 172:995–1005. [https://](https://doi.org/10.1007/s00442-012-2547-x) doi.org/10.1007/s00442-012-2547-x
- Moore JF, Wells CP, van Vuren DH, Oli MK (2016) Who pays? Intra- versus inter-generational costs of reproduction. Ecosphere 7:e01236. <https://doi.org/10.1002/ecs2.1236>
- Moyes K, Coulson T, Morgan BJT, Donald A, Morris SJ, Clutton-Brock TH (2006) Cumulative reproduction and survival costs in female red deer. Oikos 115:241–252. [https://doi.](https://doi.org/10.1111/j.2006.0030-1299.15200.x) [org/10.1111/j.2006.0030-1299.15200.x](https://doi.org/10.1111/j.2006.0030-1299.15200.x)
- Neuhaus P, Broussard DR, Murie JO, Dobson FS (2004) Age of primiparity and implications of early reproduction on life history in female Columbian ground squirrels. J Anim Ecol 73:36–43. <https://doi.org/10.1111/j.1365-2656.2004.00793.x>
- Nilsen EB, Solberg EJ (2006) Patterns of hunting mortality in Norwegian moose populations. Eur J Wildl Res 52:153–163. [https://](https://doi.org/10.1007/s10344-005-0023-1) doi.org/10.1007/s10344-005-0023-1
- Nilsson JÅ, Svensson E (1996) The cost of reproduction: a new link between current reproductive effort and future reproductive success. Proc R Soc Lond B Biol Sci 263:711–714. [https://doi.](https://doi.org/10.1098/rspb.1996.0106) [org/10.1098/rspb.1996.0106](https://doi.org/10.1098/rspb.1996.0106)
- Oli MK, Dobson FS (1999) Population cycles in small mammals: the role of age at sexual maturity. Oikos 86:557–566. [https://](https://doi.org/10.2307/3546660) doi.org/10.2307/3546660
- Partridge L (1987) Is accelerated senescence a cost of reproduction? Funct Ecol 1:317–320.<https://doi.org/10.2307/2389786>
- Péron G, Crochet P-A, Choquet R, Pradel R, Lebreton J-D, Gimenez O (2010) Capture–recapture models with heterogeneity to study survival senescence in the wild. Oikos 119:524–532. [https://doi.](https://doi.org/10.1111/j.1600-1706.2009.17882.x) [org/10.1111/j.1600-1706.2009.17882.x](https://doi.org/10.1111/j.1600-1706.2009.17882.x)
- Pettorelli N, Gaillard J-M, Duncan P, Ouellet JP, van Laere G (2001) Population density and small-scale variation in habitat quality afect phenotypic quality in roe deer. Oecologia 128:400–405. <https://doi.org/10.1007/s004420100682>
- Pettorelli N, Coulson T, Durant SM, Gaillard J-M (2011) Predation, individual variability and vertebrate population dynamics. Oecologia 167:305–314. [https://doi.org/10.1007/](https://doi.org/10.1007/s00442-011-2069-y) [s00442-011-2069-y](https://doi.org/10.1007/s00442-011-2069-y)
- Pistorius PA, Bester MN, Kirkman SP, Taylor FE (2001) Temporal changes in fecundity and age at sexual maturity of southern elephant seals at Marion Island. Polar Biol 24:343–348. [https://doi.](https://doi.org/10.1007/s003000000217) [org/10.1007/s003000000217](https://doi.org/10.1007/s003000000217)
- Plard F, Gaillard J-M, Coulson T, Tuljapurkar S (2016) Des différences, pourquoi? Transmission, maintenance and effects of phenotypic variance. J Anim Ecol 85:356–370. [https://doi.](https://doi.org/10.1111/1365-2656.12477) [org/10.1111/1365-2656.12477](https://doi.org/10.1111/1365-2656.12477)
- Proaktor G, Coulson T, Milner-Gulland EJ (2008) The demographic consequences of the cost of reproduction in ungulates. Ecology 89:2604–2611.<https://doi.org/10.1890/07-0833.1>
- Pyle P, Nur N, Sydeman WJ, Emslie SD (1997) Cost of reproduction and the evolution of deferred breeding in the western gull. Behav Ecol 8:140–147. <https://doi.org/10.1093/beheco/8.2.140>
- Reiter J, Leboeuf BJ (1991) Life-history consequences of variation in age at primiparity in northern elephant seals. Behav Ecol Sociobiol 28:153–160. <https://doi.org/10.1007/BF00172166>
- Reznick D (1985) Costs of reproduction: an evaluation of the empirical evidence. Oikos 44:257–267.<https://doi.org/10.2307/3544698>
- Reznick DA, Bryga H, Endler JA (1990) Experimentally induced lifehistory evolution in a natural population. Nature 346:357–359. <https://doi.org/10.1038/346357a0>
- Rolandsen CR, Solberg EJ, Tufto J, Sæther B-E, Heim M (2003) Factors afecting detectability of moose *Alces alces* during the hunting season in northern Norway. Alces 39:79–88
- Sæther B-E (1985) Annual variation in carcass weight of Norwegian moose in relation to climate along a latitudinal gradient. J Wildl Manag 49:977–983.<https://doi.org/10.2307/3801382>
- Sæther B-E, Bakke Ø (2000) Avian life history variation and contribution of demographic traits to the population growth rate. Ecology 81:642–653.<https://doi.org/10.2307/177366>
- Sæther B-E, Gravem AJ (1988) Annual variation in winter body condition of Norwegian moose calves. J Wildl Manag 52:333–336. <https://doi.org/10.2307/3801245>
- Sæther B-E, Haagenrud H (1983) Life history of the moose (*Alces alces*): fecundity rates in relation to age and carcass weight. J Mammal 64:226–232.<https://doi.org/10.2307/1380552>
- Sæther B-E, Haagenrud H (1985) Life-history of the moose *Alces alces*: relationship between growth and reproduction. Ecography 8:100–106.<https://doi.org/10.1111/j.1600-0587.1985.tb01159.x>
- Sæther B-E, Heim M (1993) Ecological correlates of individual variation in age at maturity in female moose (*Alces alces*): the efects of environmental variability. J Anim Ecol 62:482–489. [https://](https://doi.org/10.2307/5197) doi.org/10.2307/5197
- Sæther B-E, Solberg EJ, Heim M (2003) Efects of altering sex ratio structure on the demography of an isolated moose population. J Wildl Manag 67:455–466. <https://doi.org/10.2307/3802703>
- Sæther B-E, Solberg EJ, Heim M, Stacy JE, Jakobsen KS, Olstad R (2004) Ofspring sex ratio in moose *Alces alces* in relation to paternal age: an experiment. Wildl Biol 10:51–57
- Sæther B-E, Engen S, Solberg EJ, Heim M (2007) Estimating the growth of a newly established moose population using reproductive value. Ecography 30:417–421. [https://doi.](https://doi.org/10.1111/j.0906-7590.2007.05006.x) [org/10.1111/j.0906-7590.2007.05006.x](https://doi.org/10.1111/j.0906-7590.2007.05006.x)
- Schaffer WM (1974) Selection for optimal life histories: the effects of age structure. Ecology 55:291–303. [https://doi.](https://doi.org/10.2307/1935217) [org/10.2307/1935217](https://doi.org/10.2307/1935217)
- Solberg EJ, Heim M, Grøtan V, Sæther B-E, Garel M (2007) Annual variation in maternal age and calving date generates cohort effects in moose (*Alces alces*) body mass. Oecologia 154:259–271. <https://doi.org/10.1007/s00442-008-1158-z>
- Solberg EJ, Garel M, Heim M, Grøtan V, Sæther B-E (2008) Lack of compensatory body growth in a high performance moose *Alces alces* population. Oecologia 158:485–498. [https://doi.](https://doi.org/10.1007/s00442-008-1158-z) [org/10.1007/s00442-008-1158-z](https://doi.org/10.1007/s00442-008-1158-z)
- Solberg EJ, Rolandsen CR, Heim M, Linnell JDC, Herfndal I, Sæther B-E (2010) Age and sex-specifc variation in detectability of moose (*Alces alces*) during the hunting season: implications for

population monitoring. Eur J Wildl Res 56:871–881. [https://doi.](https://doi.org/10.1007/s10344-010-0385-x) [org10.1007/s10344-010-0385-x](https://doi.org/10.1007/s10344-010-0385-x)

- Solberg EJ, Strand O, Veiberg V, Andersen R, Heim M, Rolandsen CR, Solem MI, Holmstrøm F, Jordhøy P, Nilsen EB, Granhus A, Eriksen R (2015) Moose, red deer and reindeer—results from the monitoring program for wild cervids, 2012–2014. NINA report 1177. Norwegian Institute for Nature Research
- Stearns S (1992) The evolution of life histories. Oxford University Press, Oxford
- Tavecchia G, Coulson T, Morgan BJT, Pemberton JM, Pilkington JC, Gulland FMD, Clutton-Brock TH (2005) Predictors of reproductive cost in female Soay sheep. J Anim Ecol 74:201–213. [https://](https://doi.org/10.1111/j.1365-2656.2005.00916.x) doi.org/10.1111/j.1365-2656.2005.00916.x
- van Noordwijk AJ, de Jong G (1986) Acquisition and allocation of resources: their infuence on variation in life history tactics. Am Nat 128:137–142.<https://doi.org/10.2307/2461293>
- Vindenes Y, Engen S, Sæther B-E (2008) Individual heterogeneity in vital parameters and demographic stochasticity. Am Nat 171:455– 467.<https://doi.org/10.1086/528965>
- Weladji RB, Gaillard J-M, Yoccoz NG, Holand Ø, Mysterud A, Loison A, Nieminen M, Stenseth NC (2006) Good reindeer mothers live longer and become better in raising offspring. Proc R Soc Lond B Biol Sci 273:1239–1244.<https://doi.org/10.1098/rspb.2005.3393>
- Weladji RB, Loison A, Gaillard J-M, Holand Ø, Mysterud A, Yoccoz NG, Nieminen M, Stenseth NC (2008) Heterogeneity in individual quality overrides costs of reproduction in female reindeer. Oecologica 156:237–247. <https://doi.org/10.1007/s00442-008-0961-x>
- White KS, Berger J (2001) Antipredator strategies of Alaskan moose: are maternal trade offs influenced by offspring activity? Can J Zool 79:2055–2062.<https://doi.org/10.1139/z01-170>
- Williams GC (1966) Natural selection, the costs of reproduction, and a refnement of Lack's principle. Am Nat 100:687–690. [https://](https://doi.org/10.1086/282461) doi.org/10.1086/282461
- Wilson AJ, Nussey DH (2010) What is individual quality? An evolutionary perspective. Trends Ecol Evol 25:207–214. [https://doi.](https://doi.org/10.1016/j.tree.2009.10.002) [org/10.1016/j.tree.2009.10.002](https://doi.org/10.1016/j.tree.2009.10.002)