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# Evaluation and Evolution of MAX-DOAS-observed Vertical NO<sub>2</sub> Profiles in Urban Beijing

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## **ABSTRACT**

Multiaxis differential absorption spectroscopy (MAX-DOAS) is a newly developed advanced vertical profile detection method, but the vertical nitrogen dioxide  $(NO<sub>2</sub>)$  profiles measured by MAX-DOAS have not yet been fully verified. In this study, we perform MAX-DOAS and tower gradient observations to simultaneously acquire tropospheric  $NO<sub>2</sub>$  observations in the Beijing urban area from 1 April to 31 May 2019. The average values of the tropospheric  $NO<sub>2</sub>$  vertical column densities measured by MAX-DOAS and the tropospheric monitoring instrument are  $15.8 \times 10^{15}$  and  $12.4 \times 10^{15}$  molecules cm−2, respectively, and the correlation coefficient *R* reaches 0.87. The MAX-DOAS measurements are highly consistent with the tower-based in situ measurements, and the correlation coefficients R from the ground to the upper air are 0.89  $(60 \text{ m})$ , 0.87 (160 m), and 0.76 (280 m). MAX-DOAS accurately measures the trend of NO<sub>2</sub> vertical profile changes, although a large underestimation occurs by a factor of two. By analyzing the  $NO<sub>2</sub>$  vertical profile, the  $NO<sub>2</sub>$  concentration reveals an exponential decrease with height. The  $NO<sub>2</sub>$  vertical profile also coincides with the evolution of the boundary layer height. The study shows that the  $NO<sub>2</sub>$  over Beijing mainly originates from local sources and occurs in the boundary layer, and its vertical evolution pattern has an important guiding significance to better understand nitrate production and ozone pollution.

Key words: MAX-DOAS, NO<sub>2</sub>, tower-based in situ observation, TROPOMI, validation, vertical

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## Article Highlights:

- MAX-DOAS measurements are highly consistent with TROPOMI and tower-gradient in situ measurements.
- MAX-DOAS accurately measures the trend of  $NO<sub>2</sub>$  vertical profile changes, although underestimates its concentration by a factor of 2.
- $\bullet$  The NO<sub>2</sub> vertical profile evolution coincides with the BLH evolution.

# 1. Introduction

Nitrogen dioxide  $(NO<sub>2</sub>)$  is an important trace gas in the troposphere. It plays a pivotal role in atmospheric chem-

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istry by participating in the formation of tropospheric ozone  $(O_3)$ and nitrate aerosols ([Varshney and Singh, 200](#page-8-0)3). In China, the 90th percentile of the maximum average  $O_3$  concentration has trended higher, with the mixing ratio increasing from 64 to 79 ppb from 2013 to 2017 [\(Wang et al.](#page-8-1), [2020b](#page-8-1)). Another study demonstrated that nitrate aerosols are more important than sulfate aerosols during severe pollution episodes in Beijing [\(Tan et al., 2018](#page-8-2)a; [Wang et al](#page-8-1).,  $2020b$ ). Therefore, studies on the variation in  $NO<sub>2</sub>$ , as a precursor of  $O_3$  and nitrate aerosols, are very important to better understand their formation and influencing factors ([Tang](#page-8-3) [et al., 2021a](#page-8-3), [b](#page-8-4)).

Since the implementation of the State Council's Air Pollution Prevention and Control Action Plan (the Clean Air Action), China's environmental air quality has notably improved ([Wang et al., 2020b](#page-8-1)). However, while the tropospheric  $NO<sub>2</sub>$  columns have greatly decreased ([Zhang et al.,](#page-8-5)  $2019$ ), the surface  $NO<sub>2</sub>$  concentrations have slowly [decreased or even d](#page-8-1)istinctly increased at certain locations ([Wang et al., 2020b\)](#page-8-1). To address this problem, we need comprehensive horizontal and vertical observations. Since regular ground monitoring can only provide information horizontally, it is necessary to observe  $NO<sub>2</sub>$  in the vertical direction to control the  $NO<sub>2</sub>$  concentration in all aspects.

There are several vertical observation methods, including tower-based observations, tethered balloon observations, aircraft aerial surveys, ground-based remote sensing, and satellite observations. Given the tower height, towerbased in [situ observations o](#page-8-6)[nly measure the n](#page-8-7)e[ar-surface atmo](#page-7-0)sphere ([Meng et al., 2008](#page-8-6); [Sun et al., 2010](#page-8-7); [Li et al., 2020](#page-7-0)). With onboard instruments, balloons and aircraft detect a wide range of he[ights but do not m](#page-7-1)eet the needs of continuous monitoring [\(Glaser et al., 2003](#page-7-1)). Satellite observations cover a lar[ge area, but the measu](#page-7-2)[red results are u](#page-7-3)s[ually under](#page-7-4)[estim](#page-7-4)ated ([Celarier et al., 2008](#page-7-2); [Ma et al., 2013](#page-7-3); [Chan et al.,](#page-7-4) [2019\)](#page-7-4). Thus, sate[llite observation](#page-7-3)s usually reveal a low observation accuracy ([Ma et al., 2013\)](#page-7-3). As a spectral remote measurement method, multiaxis differential absorption spectroscopy (MAX-DOAS) is a no-contact and high-accuracy detection method upon which long-term and real-time online monitoring can be performed, and this method does not require gas sampling or calibration. Therefore, it has been widely applied to acquire tropospheric column information on  $NO<sub>2</sub>$ and other pollutants in various regions and has become a new technique for the [measurement of atmo](#page-8-8)[spheric pollut](#page-8-9)[ant ga](#page-8-9)[ses and aerosols](#page-8-10) [\(Wittrock et al., 2004](#page-8-8); [Wagner et al.,](#page-8-9) [2011;](#page-8-9) [Tan et al., 2018b](#page-8-10)).

In recent years, much comparison and verification work has been executed concerning the inversion results of aerosols and their precursors determined by MAX-DOAS. Although the formaldehyde and  $NO<sub>2</sub>$  concentrations determined by MAX-DOAS suitably compare to those determined by the scanning imaging absorption spectrometer for atmospher[ic cartography and Ozo](#page-8-11)n[e Monitoring Inst](#page-7-4)rument satellites ([Vigouroux et al., 2009](#page-8-11); [Chan et al., 2019](#page-7-4)), the MAX-DOAS measurement results demonstrate a low sensitivity

above 2.5 km and exhibit low seasonal amplitudes ([Vigour](#page-8-11)[oux et al., 2009\)](#page-8-11). In addition, clean areas yield higher correlations than polluted areas([Chan et al., 201](#page-7-5)5). Recently, MAX-DOAS observation results have been comprehensively verified against aircraft aerial survey data, and it has been found that the  $NO<sub>2</sub>$  and sulfur dioxide results are notably better than the formaldehyde results, while the  $NO<sub>2</sub>$  results are better than the sulfur dioxide results in the presence of clouds([Wang et al., 20](#page-8-12)19). The trace gas profiles obtained via inversion are different when various a priori profiles are adopted as input.

Although many comparisons and verifications have been conducted, most of these verifications were based on satellite-observed column densities and near-ground observation results. Hence, it is urgent to verify the vertical profiles retrieved by MAX-DOAS against conventional vertical detection results. In this study, MAX-DOAS observation and comparison experiments were performed in Beijing in spring. By comparing in situ measurements and the tropospheric monitoring instrument (TROPOMI) satellite observations, we analyzed the applicability of MAX-DOAS as well as the diurnal and vertical evolution characteristics of  $NO<sub>2</sub>$  at the experimental site.

# 2. Methodology

#### 2.1. *Experimental site*

A MAX-DOAS instrument was set up on the roof of the building of the Chinese Academy of Meteorological Sciences, which has a height of approximately 40 m. The tower is 9 km away from Iron Tower Branch, Institute of Atmospheric Physics, Chinese Academy of Sciences (39.974°N, 116.372°E), which is surrounded by buildings with heights ranging from 30−60 m. At a distance of 200 m to the east of the tower, the Beijing-Tibet Expressway runs north-south, and 50 m to the north, the Beitucheng West Road runs eastwest. In addition, there are no factories around the experimental sites, and the important emission sources come from vehicles.

## 2.2. *MAX-DOAS observation*

#### 2.2.1. *Measurement method*

MAX-DOAS relies on a Skype-2D-200 system, which acquires observations parallel to the tower direction. It is mainly composed of a telescope, two spectrometers (AvaSpec-ULS2048 L-USB2; UV range: 296−408 nm; visible range: 420−565 nm), a spectrometer incubator, a two-dimensional rotating platform, and a computer for control and data acquisition purposes. Approximately 15 minutes are required to complete a scan, while the elevation angle of the telescope is set to  $6^\circ$ ,  $8^\circ$ ,  $10^\circ$ ,  $15^\circ$ ,  $30^\circ$ , and  $90^\circ$ , and below 6°, the angle is increased at steps of 1°. Detailed processing after observation is shown in Text S1 in the electronic supplementary material (ESM).

With oxygen dimer differential slant column densities

(dSCDs) retrieved through a specific algorithm (please refer to section 2.2.2) and the subsequently retrieved aerosol extinction,  $NO<sub>2</sub>$  vertical profiles are finally obtained combined with the  $NO<sub>2</sub>$  slant column densities (SCDs) observed at the different elevation angles. In this study, we considered data with a solar zenith angle smaller than 75° to avoid the strong influence of stratospheric absorption.

For the comparison of the  $NO<sub>2</sub>$  column density, we deleted the  $NO<sub>2</sub>$  vertical column density (VCD) measured by MAX-DOAS at the corresponding time under cloudy weather conditions from 1200–1500 LST (LST = UTC + 8) because the presence of clouds greatly complicates MAX-DOAS measurements concerning the accurate analysis of the tropospheric trace gas VCD. To obtain the  $NO<sub>2</sub>$  VCD, we need to convert dSCDs using differential air mass factors (dAMFs) by Eq. (1) [\(Wagner et al., 2010](#page-8-13)):

$$
\text{VCD} = \frac{\text{dSCDs}}{\text{dAMFs}} = \frac{\text{dSCD}_{\alpha \neq 90^\circ} - \text{dSCD}_{\alpha = 90^\circ}}{\text{AMF}_{\alpha \neq 90^\circ} - \text{AMF}_{\alpha = 90^\circ}} \,. \tag{1}
$$

a gas in the atmosphere.  $\alpha$  denotes the elevation angle of the The air mass factor calculation is solved using the radiative transfer model (RTM) to describe the absorption path of telescope. Combined with downloaded TROPOMI observation data, there are 24 data sets for comparison purposes.

In terms of the comparison of the  $NO<sub>2</sub>$  vertical profiles, any anomalous measured tower-based in situ values were eliminated, and the MAX-DOAS observation data were screened according to a relevant threshold. First, the retrieved  $NO<sub>2</sub>$  data were deleted when the chi-square value exceeded 200, the relative error of the  $NO<sub>2</sub>$  VCD exceeded 50% or there was less than one degree of freedom (DF), and any anomalous points were then deleted from the MAX-DOAS observations. Second, because clouds greatly influence the MAX-DOAS detection method, we deleted any MAX-DOAS data measured under cloud cover based on the aerosol optical depth (AOD) level 1.5 (cloud-shielded) data measured by the CIMEL automatic solar photometer obtained from the AERONET website. Since MAX-DOAS inverts the AOD at a wavelength of 360 nm, only the corresponding moment when the AOD is not missing at wavelengths of 340 and 380 nm on the AERONET website is regarded as a cloud-free moment, and the  $\rm NO_2$  data measured by MAX-DOAS at the corresponding moment are then retained. In addition, the all-day AOD not detected according to the AERONET website was regarded as indicating cloud cover, whereby all the associated MAX-DOASobserved  $NO<sub>2</sub>$  data were deleted. The method of deleting some data to avoid [the influence of clou](#page-8-14)ds was introduced in a previous study [\(Wagner et al., 2014](#page-8-14)), and we also performed an analysis in the case where clouds were not removed, as shown in Fig. S1 in the ESM.

#### 2.2.2. *Inversion algorithm*

Vertical profiles of the aerosols and  $NO<sub>2</sub>$  are obtained by the HEIPRO inversion algorithm (Heidelberg Profile, developed by the IUP of Heidelberg University, Germany)

sion approach, and the minimized value function  $\chi^2$  is adop-[\(Frieß et al., 2006](#page-7-6), [2011](#page-7-7)). This inversion algorithm is based on the optimal estimation method ([Rodgers, 2000](#page-8-15)), with the use of RTM SCIATRAN as the forward model of the inverted to determine the optimal estimated state quantity between the measurement and prior information, thereby calculating the state quantity most closely approaching the real atmosphere through inversion. The inversion expression is as follows:

$$
\chi^{2}(x) = [y - F(x, b)]^{T} S_{\varepsilon}^{-1} [y - F(x, b)] +
$$
  

$$
(x - x_{a})^{T} S_{a}^{-1} (x - x_{a}),
$$
 (2)

where  $F(x, b)$  is the forward model of the physical properties of the system, i.e., the RTM,  $x$  is the aerosol extinction crete height intervals,  $b$  represents the system parameters but impacting the measurement results, y represents the ent elevation angles,  $x_a$  is the a priori vector, and  $S_\epsilon$  and  $S_a$ are the covariance matrices of  $y$  and  $x_a$ , respectively. To express the sensitivity of the inversion state  $\hat{x}$  relative to the true state  $x$ , the OEM describes this important inversion propdefined as  $A = \frac{\partial \hat{x}}{\partial x}$ . In addition, to reflect the high resolucoefficient or the trace gas concentration in a series of dis-(pressure, temperature, etc.) not involved in the inversion dSCDs of the aerosols or trace gases measured at the differerty through the averaging kernel (AK) matrix. The AK *A* is tion of the inversion process, namely, the independence of each information layer, we add values to the diagonal of the mean nucleus as DFs to characterize this inversion property, and the more DFs there are, the higher the inversion accuracy is. More specific settings of this retrieval approach are provided in Text S2 in the ESM.

#### 2.3. *Tower in situ observation*

The atmospheric observation tower is 325 m high, containing 15 observation platforms. Each layer provides continuous observations of meteorological elements and characterizes the vertical evolution of the four key meteorological parameters: wind direction, wind speed, temperature, and relative humidity. We applied ground-based telemetry to obtain the height of the atmospheric boundary layer (BL). To determine the vertical distribution of  $NO<sub>2</sub>$  concentrations in the area, point instruments were placed on the 60-, 160- and 280-m platforms of the tower, which could achieve the simultaneous observation of these three elevations.  $NO<sub>2</sub>$  monitoring adopted a  $NO-NO_2-NO_x$  analyzer (model MMS-AC32M) produced by the French ESA company, using the chemiluminescence method to detect the  $NO<sub>x</sub>$  concentration. To provide a stable working environment for the measuring instruments, they were placed in a custom observation box, which was equipped with air conditioning to provide an appropriate operating temperature. A detailed description has been provided elsewhere [\(Tang et al., 2009](#page-8-16), [2012\)](#page-8-17). An enhanced single-lens ceilometer (CL-51, Vaisala, Finland) is installed near the tower to analyze the dilution capability based on strobe laser lidar (laser detection and range measurement) technology. The boundary layer height (BLH) was

determined by the gradient method([Steyn et al., 199](#page-8-18)9) based on the backscatter density [\(Tang et al., 2016](#page-8-19)).

#### 2.4. *TROPOMI satellite observations*

TROPOMI is the only payload carried on the Sentinel-5 precuS5P satellite launched in October 2017. The transit time is generally 1330 local time. As the next-generation atmospheric composition monitor, TROPOMI inherits the advances of the global ozone monitoring experiment, the scanning imaging absorption spectrometer for atmospheric cartography, and the ozone monitoring instrument with a wider wavelength range and higher spatial resolution. It has a maximum spatial resolution of  $7 \text{ km} \times 3.5 \text{ km}$ , and an observation period of daily global coverage. In addition, it has three spectral regions, seven bands, and can measure in ultraviolet-visible (270−500 nm), near-infrared (675−775 nm) and short-wave infrared (2305−2385 nm) wavelengths. In this paper, the TROPOMI Level 2 Offline  $NO<sub>2</sub>$  product of version 1.03.01 was used to extract the time series of  $NO<sub>2</sub>$  tropospheric column density. The product can be accessed via the Sentinel-5P Pre-Operations Data Hub ([https://s5phub.coperni](https://s5phub.copernicus.eu/dhus/#/home)[cus.eu/dhus/#/home](https://s5phub.copernicus.eu/dhus/#/home)). The observation quality of individual pixels depends on many factors, including clouds, surface albedo, spectral saturation, geometric elements, etc. TRO-POMI NO<sub>2</sub> products provide a "qa\_value" for each pixel of observation to provide users of the data with a simple way to filter out inaccurate results. The filtering criteria used in this paper is qa\_value > 0.75. Observations at a given location are defined as the mean of valid observations around  $0.1^{\circ}$  of the included pixels. To reduce the comparison error, this study selected the average MAX-DOAS  $NO<sub>2</sub>$  VCD from 1200–1500 LST to compare with the TROPOMI NO<sub>2</sub> VCD.

## 3. Results

## 3.1. *MAX-DOAS evaluation*

#### 3.1.1. *Column density*

The correlation analysis is shown in [Fig. 1](#page-3-0). The correlation coefficient R between the two data sets is 0.87, indicating that the  $NO<sub>2</sub> VCD$  results obtained by the two methods are consistent. The daytime mean  $NO<sub>2</sub> VCDs$  (in units of 1015 molecules cm−2) varied from 4.0 to 46.5 for MAX-DOAS and from 2.3 to 37.3 for TROPOMI. The average values of the  $NO<sub>2</sub> VCD$  measured by  $MAX-DOAS$  and  $TRO-$ POMI are 15.8 and 12.4, respectively. In conclusion, the TROPOMI observation results are generally close to the MAX-DOAS observation results with significantly lower average results. Please refer to section 4.1 for the error analysis of the MAX-DOAS and TROPOMI observation results.

## 3.1.2. *Vertical profile*

To examine the MAX-DOAS  $NO<sub>2</sub>$  vertical profile, we selected the first three layers (10−100 m, 100−200 m, and

<span id="page-3-0"></span>

Fig. 1. Correlation analysis of the MAX-DOAS and TROPOMI observations from 1 April to 31 May 2019.

200–300 m) of the MAX-DOAS  $NO<sub>2</sub>$  volume mixing ratio (VMR) inversion results and compared them to the in situ  $NO<sub>2</sub>$  results obtained at 60, 160, and 280 m with the meteorological observation tower in Beijing. Because the MAX-DOAS instrument can only operate during the day when there is sunlight, this study only analyzes the data measured by MAX-DOAS from 0800−1700 LST. A diagram of the  $MAX-DOAS$  and tower  $NO<sub>2</sub>$  concentration time series during the observation period after data screening is shown in [Fig. 2.](#page-4-0)

This paper studies the obtained data in two ways. On the one hand, the MAX-DOAS three-layer data are compared to the corresponding in situ three-layer, tower-based data, and the results are shown in [Figs. 3a](#page-4-1)−[c](#page-4-1). The threelayer correlation coefficients R from the ground to the upper air are 0.89 (60 m), 0.87 (160 m), and 0.76 (280 m), which decrease with increasing height. On the other hand, as the bottom concentration obtained by MAX-DOAS via inversion is the average  $NO<sub>2</sub>$  concentration from the ground to a certain height along the effective horizontal optical path, the MAX-DOAS result represents the average concentration of one layer, while the tower observation adopts the in situ measurement method, and the measured value represents the air quality close to the instrument, which represents a point concentration. Therefore, to enable the comparison, we averaged the concentrations of the three tower layers and compared the average concentration to that of the three MAX-DOAS layers. The correlation analysis result is shown in [Fig. 3d](#page-4-1). The correlation coefficient  $R$  of the average concentration of the three layers measured by the two methods reaches 0.88. Moreover, the analysis in the case where the cloud is not removed is shown in Fig. S1 in the ESM. The correlation coefficients are 0.76, 0.78, and 0.69 from the ground to the high level, respectively, while the *R* of the average concentration is 0.80, which is significantly lower. Overall, regardless of the comparison with different layers or with average results, these high R values indicate that the  $NO<sub>2</sub>$  concentration measured by MAX-DOAS is highly consistent with that measured in situ by the tower, which also demonstrates that MAX-DOAS suitably reflects the evolution trend of  $NO<sub>2</sub>$ 

<span id="page-4-0"></span>

Fig. 2. NO<sub>2</sub> time series of the MAX-DOAS and tower-based in situ observations from April–May 2019.

<span id="page-4-1"></span>

Fig. 3. Correlation analysis of the MAX-DOAS and tower-based in situ VMR results in the case where the cloud is removed.

near the ground. In addition, the points in the four correlation analysis graphs of [Fig. 3](#page-4-1) are essentially distributed below the 1:1 line, which reveals that some deviations occur in the magnitude of the  $NO<sub>2</sub>$  concentration measured by MAX-DOAS. Furthermore, we also compared the performance in the situations of high and low AOD, and the results are shown in Fig. S2 in the ESM. The inversion of  $NO<sub>2</sub>$ requires aerosol profiles as a priori information, which means that the influence of aerosols has been taken into account during the inversion. Consequently, the influence of aerosols is not significant. From Fig. S2, we can see that both *R* and slope are similar to one other, indicating that the influence of aerosols on the retrieval of  $NO<sub>2</sub>$  is limited.

## 3.2. *Evolution rule of NO<sup>2</sup>*

It should be noted that the below daily variation curve and profile analyzed by MAX-DOAS were obtained from the convolution of the true atmospheric  $NO<sub>2</sub>$  concentration and the AK profile. However, due to height limitations, the tower-based in situ method does not detect the complete AK profile, so the tower observation data cannot be smoothed. In particular, the  $NO<sub>2</sub>$  daily variation curve and vertical profile provided by the tower-based in situ method given below were obtained by analyzing the original atmospheric  $\mathrm{NO}_2$  concentration. Therefore, the diurnal variation and  $\rm NO_2$  mean vertical profiles measured by the two methods cannot be quantitatively analyzed, and only qualitative analysis can be conducted.

## 3.2.1. *Diurnal variation in NO<sup>2</sup>*

To examine the MAX-DOAS measurements at different times, diurnal variation analysis was conducted for the  $MAX-DOAS NO<sub>2</sub>$  and tower-based in situ  $NO<sub>2</sub>$  concentra[tions.](#page-5-0) The  $NO<sub>2</sub>$  during the day and night are displayed in [Fig. 4](#page-5-0), which shows that the  $NO<sub>2</sub>$  concentration exhibits a large range of variation throughout the day. Unfortunately,

MAX-DOAS cannot record the variations due to the lack of light at night. However, the relatively high  $NO<sub>2</sub>$  concentration in the early morning and relatively low  $NO<sub>2</sub>$  concentration at noon confirm the diurnal variation in  $NO<sub>2</sub>$  measured by MAX-DOAS. In the early morning, the surface  $NO<sub>2</sub>$  concentration was high due to accumulation at night and morning peak traffic emissions. During mid-morning, at approximately 0900 LST, the surface  $NO<sub>2</sub>$  concentration greatly decreases due to consumption processes, such as photolysis. In the afternoon (1400−1500 LST), the photolysis rate is relatively high, and the traffic load is low. Hence, the  $NO_2$  concentration reaches its lowest level of the day. Soon thereafter, photolysis weakens, the cumulative loss rate of  $NO<sub>2</sub>$  is low, and the  $NO<sub>2</sub>$  production volume is large. Therefore,  $NO<sub>2</sub>$ begins to accumulate again, and the concentration increases with a peak at approximately 1700 LST. Moreover, BL dynamics also play an important role in the diurnal variation in  $NO<sub>2</sub>$ . The diurnal variation in BLH (please refer to Fig. S3 in the ESM) is low in the morning and high at noon. From morning to evening, the evolving BL dilutes  $NO<sub>2</sub>$ , and the corresponding diurnal variation in  $NO<sub>2</sub>$  is the opposite of that in BLH, resulting in  $NO<sub>2</sub>$  accumulation in the morning and evening. Since the  $NO<sub>2</sub>$  concentration in the bottom layer (0−100 m) is more sensitive to near-surface emissions and diurnal variations in the BLH, a high  $NO<sub>2</sub>$  concentration in the bottom layer is more evident in the morning and evening, when emissions are the primary source of the  $NO<sub>2</sub>$ poll[ution detec](#page-5-0)ted at the experimental site.

[Figure 4b](#page-5-0) shows the changes in the  $NO<sub>2</sub>$  gradient measured in the aforementioned two ways. In the daytime, the differences between the three layers are small in both measurements because the troposphere shows high mixing rates during the day, and pollutants are evenly distributed in the mixed layer. However, the  $NO<sub>2</sub>$  concentration of the third layer of MAX-DOAS is significantly lower, which implies that the near-ground  $NO<sub>2</sub>$  at the experimental point mainly ori-

<span id="page-5-0"></span>

Fig. 4. Comparison of the  $NO<sub>2</sub>$  diurnal variations ((a) mean concentration of the three MAX-DOAS layers and in situ tower (b) concentration of a single layer).

ginates from the ground. At night, the difference between the three layers becomes larger from the gradient measured in situ by the tower, which is contrary to the daytime. Due to the difference between the MAX-DOAS and in situ measurements concerning air mass quality detection [please refer to section 4.2 (1) for details], the  $NO<sub>2</sub>$  measured by MAX-DOAS is more spatially representative, and its vertical evolution rule is detailed in section 3.2.2.

## 3.2.2. *Vertical evolution of NO<sup>2</sup>*

Because MAX-DOAS is not sensitive to pollutants above 3 km ([Wang et al., 2013](#page-8-20)), this article only analyzes the vertical evolution of  $NO<sub>2</sub>$  below 3 km. [Figure 5a](#page-6-0) shows the vertical evolution trend of  $\rm NO_2$  obtained by the two methods. The average  $NO<sub>2</sub>$  vertical profile decreases exponentially with increasing height, and most of the high  $NO<sub>2</sub>$  concentrations occur below 1 km. Simultaneously, based on Fig. S3, we find that the  $NO<sub>2</sub>$  concentration was distributed mainly within the BL. As a result, the  $NO<sub>2</sub>$  changes are consistently reflected in the BL changes, which notably verifies that the MAX-DOAS inversion method is correct.

In [Fig. 5b](#page-6-0), we divided the daytime into three parts, 0800−1000 LST, 1100−1300 LST, and 1400−1700 LST, to analyze the typical  $NO<sub>2</sub>$  profiles in the morning, at noon, and in the afternoon, respectively. Near the ground, it is obvious that the  $NO<sub>2</sub>$  concentration is morning > afternoon > noon. Since the  $NO<sub>2</sub>$  in the urban atmosphere is mainly from the emission of mobile sources such as cars, nearground  $NO<sub>2</sub>$  is hi[ghly rel](#page-6-0)ated to traffic emissions. In addi-tion, as shown in [Fig. 5a](#page-6-0), the growth rate of  $NO<sub>2</sub>$  changes little in the vertical. As a result, the measured  $NO<sub>2</sub>$  in the [study a](#page-6-0)rea generally depends on the local distribution. From [Fig. 5b](#page-6-0), we also find that the  $NO<sub>2</sub>$  concentration above 1.4 km in the afternoon is approximately 1 ppb higher than that in the morning and noon. This can also be seen in Fig. S3. In Fig. S3, the BLH starts to decrease at 1500 LST and drops significantly at 1600 LST at approximately 1.4 km, which leads to the  $NO<sub>2</sub>$  concentration at approximately 1.4 km increasing in turn. Moreover, pollutants in the residual

layer (RL) can impact pollution near the surface [\(Liu et al.,](#page-7-8)  $2021$ ). With strong turbulence in the daytime, the NO<sub>2</sub> in the BL is then transported into the free atmosphere above the BL, and when the BL begins to decrease, the  $NO<sub>2</sub>$  in the free atmosphere is stored in the RL until the next morning, leading to a high  $NO<sub>2</sub>$  concentration above the BL in the early morning of the next day. Therefore, it is clearly shown in Fig. S3 that within the BL, the  $NO<sub>2</sub>$  concentration is low in the daytime when the BL develops, while it is higher in the afternoon when the BL begins to decrease.

# 4. Discussion

# 4.1. *Analysis of the difference between the MAX-DOAS and TROPOMI results*

Both the average observation value and measurement range of TROPOMI are lower than that of MAX-DOAS, which are mainly caused by the following two factors:

(1) It is partly related to the a priori vertical distribution of  $NO<sub>2</sub>$  used in the AMF calculation of the satellite retrieval, whichi[s not able to ful](#page-7-9)ly resolve emission hot spots over cities ([Chan et al., 2020](#page-7-9)), noting that the uncertainty in the estimate of the tropospheric AMF is the largest source of uncertainty in  $NO<sub>2</sub>$  satellite retrieva[ls in situatio](#page-8-21)ns [with](#page-8-21) enhanced trace gas concentrations([Wang et a](#page-8-21)l., [2020a](#page-8-21)).

(2) It is due to the effect of aerosols and cl[ouds, both of](#page-8-21) [which](#page-8-21) can mask some of the tropospheric  $NO<sub>2</sub>$  [\(Wang et al.,](#page-8-21)  $2020a$ ). Because aerosols have a longer lifetime than  $NO<sub>2</sub>$  in the atmosphere, ne[ar-ground aeroso](#page-7-3)ls influence satellite measurements of  $NO<sub>2</sub>$  ([Ma et al., 2013](#page-7-3)). Therefore, the sensitivity of satellite instruments is limited near the ground, while MAX-DOAS exhibits the highest sensitivity near the ground, and its sensitivity is almost zero at heights above 3−4 km.

# 4.2. *Analysis of the difference between the MAX-DOAS and tower in situ observation results*

<span id="page-6-0"></span>

Fig. 5. Averaged NO<sub>2</sub> vertical profiles during the observation period (a) and at certain times of day (b).

Regarding the low MAX-DOAS measurement results, the following reasons are given:

(1) The difference in spatial representation between the measured values increases the observation bias over the entire measurement period. The gas concentration measured by MAX-DOAS is the average gas concentration along the optical path from several hundred meters to several kilometers, which ensures more representative detection results. The in situ measurement method is more sensitive to the air mass at the measuring port of the instrument, and it only measures the air pollution in the sampling space at the monitoring site. When a high air mass concentration occurs, the measured value will be high.

(2) It is attributed to the difference in sensitivity of different hights measured by MAX-DOAS. From the perspective of the error, the MAX-DOAS inversion error mainly includes a smoothing error, which is caused by a smoothing of the real profile [\(Hong, 2019](#page-7-10)), and the smoothing effect mainly depends on the AK. The AK results in the MAX-DOAS method exhibiting different vertical resolutions and inversion sensitivities at various heights. With increasing height, the sensitivity of MAX-DOAS to trace gas and aerosol profile inversion systematically decreases. Previous experiments have also demonstrated that the MAX-DOAS-measured results after smoothing are lower than the real results, and in particular, the concentration of pollutants above 1 km is conspicuously underestimated [\(Wang et al., 2017](#page-8-22), [2019\)](#page-8-12), while profile inversion is more sensitive to the near-ground height [\(Hong, 2019\)](#page-7-10).

# 5. Conclusions

In this paper, we applied MAX-DOAS to conduct vertical observations of the tropospheric  $NO<sub>2</sub>$  concentration in Beijing from 1 April to 31 May 2019. Moreover, we compared the MAX-DOAS  $NO_2$  results to the TROPOMI  $NO_2$ VCD and tower-based in situ  $NO<sub>2</sub>$  VMR results. Several conclusions are drawn as follows:

(1) The MAX-DOAS results are highly consistent with the satellite and tower-based in situ measurement results, and the approach accurately detects the trend of the tropospheric  $NO<sub>2</sub>$  concentration. Although MAX-DOAS does not directly provide the  $NO<sub>2</sub>$  concentration near the ground, the bottom concentration of the  $NO<sub>2</sub>$  profile yielded by MAX-DOAS reflects the near-surface  $NO<sub>2</sub>$  trend, which proves that the profile inversion results are highly reliable.

(2) The  $NO<sub>2</sub>$  vertical evolution trend measured by  $MAX-DOAS$  is correct, but the measured  $NO<sub>2</sub>$  concentrations differ in magnitude.

(3) The inversion results of the  $NO<sub>2</sub>$  profile during the experiment indicate that the  $NO<sub>2</sub>$  observed at the experimental site mainly occurs in the near-surface layer below 1 km, and the average  $NO<sub>2</sub>$  vertical profile decreases exponentially with increasing height. The  $NO<sub>2</sub>$  vertical profile evolution also coincides with the BLH evolution.

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#### **REFERENCES**

- <span id="page-7-2"></span>Celarier, E. A., and Coauthors, 2008: Validation of ozone monitoring instrument nitrogen dioxide columns. *J. Geophys. Res.*, 113, D15S15, [https://doi.org/10.1029/2007JD008908.](https://doi.org/10.1029/2007JD008908)
- <span id="page-7-5"></span>Chan, K. L., and Coauthors, 2015: Observations of tropospheric  $NO<sub>2</sub>$  using ground based MAX-DOAS and OMI measurements during the Shanghai World Expo 2010. *Atmos. Environ.*, 119, 45−58, [https://doi.org/10.1016/j.atmosenv.](https://doi.org/10.1016/j.atmosenv.2015.08.041) [2015.08.041.](https://doi.org/10.1016/j.atmosenv.2015.08.041)
- <span id="page-7-4"></span>Chan, K. L., and Coauthors, 2019: MAX-DOAS measurements of tropospheric  $NO<sub>2</sub>$  and HCHO in Nanjing and a comparison to ozone monitoring instrument observations. *Atmospheric Chemistry and Physics*, 19(15), 10 051−10 071, [https://doi.org/10.5194/acp-19-10051-2019.](https://doi.org/10.5194/acp-19-10051-2019)
- <span id="page-7-9"></span>Chan, K. L., M. Wiegner, J. van Geffen, I. de Smedt, C. Alberti, Z. B. Cheng, S. Ye, and M. Wenig, 2020: MAX-DOAS measurements of tropospheric  $NO<sub>2</sub>$  and HCHO in Munich and the comparison to OMI and TROPOMI satellite observations. *Atmospheric Measurement Techniques*, 13, 4499−4520, <https://doi.org/10.5194/amt-13-4499-2020>.
- <span id="page-7-6"></span>Frieß, U., P. S. Monks, J. J. Remedios, A. Rozanov, R. Sinreich, T. Wagner, and U. Platt, 2006: MAX-DOAS  $O<sub>4</sub>$  measurements: A new technique to derive information on atmospheric aerosols: 2. Modeling studies. *J. Geophys. Res.*, 111, D14203,<https://doi.org/10.1029/2005JD006618>.
- <span id="page-7-7"></span>Frieß, U., H. Sihler, R. Sander, D. Pöhler, S. Yilmaz, and U. Platt, 2011: The vertical distribution of BrO and aerosols in the Arctic: Measurements by active and passive differential optical absorption spectroscopy. *J. Geophys. Res.*, 116, D00R04, [https://doi.org/10.1029/2011JD015938.](https://doi.org/10.1029/2011JD015938)
- <span id="page-7-1"></span>Glaser, K., U. Vogt, G. Baumbach, A. Volz‐Thomas, and H. Geiss, 2003: Vertical profiles of  $O_3$ ,  $NO_2$ ,  $NO_x$ , VOC, and meteorological parameters during the Berlin Ozone Experiment (BERLIOZ) campaign. *J. Geophys. Res.*, 108(D4), 8253,<https://doi.org/10.1029/2002JD002475>.
- <span id="page-7-10"></span>Hong, Q. Q., 2019: Study on the Distribution and Regional Transport of Trace Gases and Aerosol over YRD Region Based on MAX-DOAS. PhD dissertation, University of Science and Technology of China. (in Chinese with English abstract)
- <span id="page-7-0"></span>Li, L., and Coauthors, 2020: Tower observed vertical distribution of  $PM_{2.5}$ ,  $O_3$  and  $NO_x$  in the Pearl River Delta. *Atmos. Environ.*, 220, 117083, [https://doi.org/10.1016/j.atmosenv.](https://doi.org/10.1016/j.atmosenv.2019.117083) [2019.117083](https://doi.org/10.1016/j.atmosenv.2019.117083).
- <span id="page-7-8"></span>Liu, Y. S., G. Q. Tang, M. Wang, B. X. Liu, B. Hu, Q. Chen, and Y. S. Wang, 2021: Impact of residual layer transport on air pollution in Beijing, China. *Environmental Pollution*, 271, 116325,<https://doi.org/10.1016/j.envpol.2020.116325>.
- <span id="page-7-3"></span>Ma, J. Z., S. Beirle, J. L. Jin, R. Shaiganfar, P. Yan, and T. Wagner, 2013: Tropospheric NO<sub>2</sub> vertical column densities over

Beijing: Results of the first three years of ground-based MAX-DOAS measurements (2008−2011) and satellite validation. *Atmospheric Chemistry and Physics*, 13, 1547−1567, [https://doi.org/10.5194/acp-13-1547-2013.](https://doi.org/10.5194/acp-13-1547-2013)

- <span id="page-8-6"></span>Meng, Z. Y., G. A. Ding, X. B. Xu, X. D. Xu, H. Q. Yu, and S. F. Wang, 2008: Vertical distributions of  $SO_2$  and  $NO_2$  in the lower atmosphere in Beijing Urban areas, China. *Science of the Total Environment*, 390, 456−465, [https://doi.org/10.](https://doi.org/10.1016/j.scitotenv.2007.10.012) [1016/j.scitotenv.2007.10.012](https://doi.org/10.1016/j.scitotenv.2007.10.012).
- <span id="page-8-18"></span><span id="page-8-15"></span>Rodgers, C. D., 2000: *Inverse Methods for Atmospheric Sounding: Theory and Practice*. World Scientific, 238 pp.
- Steyn, D. G., M. Baldi, and R. M. Hoff, 1999: The detection of mixed layer depth and entrainment zone thickness from lidar backscatter profiles. *J. Atmos. Oceanic Technol.*, 16, 953−959, [https://doi.org/10.1175/1520-0426\(1999\)016<09](https://doi.org/10.1175/1520-0426(1999)016%3C0953:TDOMLD%3E2.0.CO;2) [53:TDOMLD>2.0.CO;2.](https://doi.org/10.1175/1520-0426(1999)016%3C0953:TDOMLD%3E2.0.CO;2)
- <span id="page-8-7"></span>Sun, Y., Y. Wang, and C. Zhang, 2010: Vertical observations and analysis of  $PM_{2.5}$ ,  $O_3$ , and  $NO_x$  at Beijing and Tianjin from towers during summer and autumn 2006. *Advances in Atmospheric Sciences*, 27(1), 123−136, [https://doi.org/10.1007/](https://doi.org/10.1007/s00376-009-8154-z) [s00376-009-8154-z.](https://doi.org/10.1007/s00376-009-8154-z)
- <span id="page-8-2"></span>Tan, T. Y., and Coauthors, 2018a: New insight into  $PM_{2.5}$  pollution patterns in Beijing based on one-year measurement of chemical compositions. *Science of the Total Environment*, 621, 734−743, [https://doi.org/10.1016/j.scitotenv.2017.11.](https://doi.org/10.1016/j.scitotenv.2017.11.208) [208](https://doi.org/10.1016/j.scitotenv.2017.11.208).
- <span id="page-8-10"></span>Tan, W., and Coauthors, 2018b: Tropospheric  $NO_2$ ,  $SO_2$ , and HCHO over the east China Sea, using ship-based MAX-DOAS observations and comparison with OMI and OMPS satellite data. *Atmospheric Chemistry and Physics*, 18, 15 387−15 402,<https://doi.org/10.5194/acp-18-15387-2018>.
- <span id="page-8-16"></span>Tang, G., X. Li, Y. Wang, J. Xin, and X. Ren, 2009: Surface ozone trend details and interpretations in Beijing, 2001- 2006. *Atmospheric Chemistry and Physics*, 9, 8813−8823, <https://doi.org/10.5194/acp-9-8813-2009>.
- <span id="page-8-17"></span>Tang, G., Y. Wang, X. Li, D. Ji, S. Hsu, and X. Gao, 2012: Spatial-temporal variations in surface ozone in Northern China as observed during 2009-2010 and possible implications for future air quality control strategies. *Atmospheric Chemistry and Physics*, 12, 2757−2776, [https://doi.org/10.5194/acp-12-](https://doi.org/10.5194/acp-12-2757-2012) [2757-2012](https://doi.org/10.5194/acp-12-2757-2012).
- <span id="page-8-19"></span>Tang, G. Q., and Coauthors, 2016: Mixing layer height and its implications for air pollution over Beijing, China. *Atmospheric Chemistry and Physics*, 16, 2459−2475, [https://doi.org/10.5194/acp-16-2459-2016.](https://doi.org/10.5194/acp-16-2459-2016)
- <span id="page-8-3"></span>Tang, G. Q., and Coauthors, 2021a: Bypassing the NO*<sup>x</sup>* titration trap in ozone pollution control in Beijing. *Atmospheric Research*, 249, 105333, [https://doi.org/10.1016/j.atmosres.](https://doi.org/10.1016/j.atmosres.2020.105333) [2020.105333.](https://doi.org/10.1016/j.atmosres.2020.105333)
- <span id="page-8-4"></span>Tang, G. Q., and Coauthors, 2021b: Low particulate nitrate in the residual layer in autumn over the North China Plain. *Science of the Total Environment*, 782, 146845, [https://doi.org/](https://doi.org/10.1016/j.scitotenv.2021.146845) [10.1016/j.scitotenv.2021.146845.](https://doi.org/10.1016/j.scitotenv.2021.146845)
- <span id="page-8-0"></span>Varshney, C. K., and A. P. Singh, 2003: Passive samplers for NO<sup>x</sup> monitoring: A critical review. *Environmentalist*, 23, 127−136,<https://doi.org/10.1023/A:1024883620408>.
- <span id="page-8-11"></span>Vigouroux, C., and Coauthors, 2009: Ground-based FTIR and

MAX-DOAS observations of formaldehyde at réunion Island and comparisons with satellite and model data. *Atmo[spheric Chemistry and Physics](https://doi.org/10.5194/amt-3-129-2010)*, 9(24), 9523−9544, <https://doi.org/10.5194/acp-9-9523-2009>.

- <span id="page-8-13"></span>Wag[ner, T., O. Ibrahim, R. Shaiganfar, a](https://doi.org/10.5194/acp-9-9523-2009)nd U. Platt, 2010: Mobile MAX-DOAS observations of tropospheric trace gases. *Atmospheric Measurement Techniques*, 3, 129−140, [https://doi.org/10.5194/amt-3-129-2010.](https://doi.org/10.5194/amt-3-129-2010)
- <span id="page-8-9"></span>Wag[ner, T., and Coauthors, 2011: Inversion o](https://doi.org/10.5194/amt-4-2685-2011)f tropospheric profiles of aerosol extinction and HCHO and  $NO<sub>2</sub>$  mixing ratios from MAX-DOAS observations in Milano during the summer of 2003 and comparison with independent data sets. *Atmospheric Measurement Techniques*, 4[, 2685−2715,](https://doi.org/10.5194/amt-7-1289-2014) <https://doi.org/10.5194/amt-4-2685-2011>.
- <span id="page-8-14"></span>Wag[ner, T., A. Apituley, S. Beirle, S. Dörner](https://doi.org/10.5194/amt-4-2685-2011), U. Friess, J. Remmers, and R. Shaiganfar, 2014: Cloud detection and classification based on MAX-DOAS observations. *[Atmospheric Meas](https://doi.org/10.3390/atmos11060636)[urement Techniq](https://doi.org/10.3390/atmos11060636)ues*, 7, 1289−1320, [https://doi.org/10.5194/](https://doi.org/10.5194/amt-7-1289-2014) [amt-7-1289-2014](https://doi.org/10.5194/amt-7-1289-2014).
- <span id="page-8-21"></span>Wan[g, C., T. Wang, P](https://doi.org/10.5194/amt-7-1289-2014) Wang, and V Rakitin. 2020a: Comparison and validation of TROPOMI and OMI  $NO<sub>2</sub>$  observations over China. *Atmosphere*, 11(6), 636. [https://doi.org/10.3390/](https://doi.org/10.3390/atmos11060636) [atmos11060636.](https://doi.org/10.3390/atmos11060636)
- <span id="page-8-20"></span>Wan[g, Y., P. H. Xie, A](https://doi.org/10.3390/atmos11060636). Li, and H. Chen, 2013: Distribution of aerosol and NO<sub>2</sub> measured by MAX-DOAS. Proc. 7th Control *and Monitoring Technology Conf. on SO<sup>2</sup> , NO<sup>x</sup> , Mercury Pollutant Control Technology and Fine Particular Matter (PM2.5)*. Hangzhou: Ch[inese Society of Environmental Sci](https://doi.org/10.5194/acp-17-2189-2017)[ences](https://doi.org/10.5194/acp-17-2189-2017). (in Chinese)
- <span id="page-8-22"></span>Wang, Y., J. Lampel, P. H. Xie, S. Beirle, A. Li, D. X. Wu, and T. Wagner, 2017: Ground-based MAX-DOAS observations of tropospheric aerosols,  $NO_2$ ,  $SO_2$  and HCHO in Wuxi, China, from 2011 to 2014. *[Atmospheric Chemistry and Phys](https://doi.org/10.5194/acp-17-2189-2017)[ics](https://doi.org/10.5194/acp-17-2189-2017)*, 17(3), 2189−2215, [https://doi.org/10.5194/acp-17-2189-](https://doi.org/10.5194/acp-17-2189-2017) [2017](https://doi.org/10.5194/acp-17-2189-2017).
- <span id="page-8-12"></span>Wan[g, Y., and](https://doi.org/10.5194/acp-19-5417-2019) Coauthors, 2019: [Vertical profiles of NO](https://doi.org/10.5194/acp-19-5417-2019)<sub>2</sub>, SO<sub>2</sub>, HONO, HCHO, CHOCHO and aerosols derived from MAX-DOAS measurements at a rural site in the central western North China Plain and their relation to e[mission sources](https://doi.org/10.1093/nsr/nwaa032) [and effects of region](https://doi.org/10.1093/nsr/nwaa032)al transport. *[Atmospheric Chemistry](https://doi.org/10.5194/acp-19-5417-2019) [and Physic](https://doi.org/10.5194/acp-19-5417-2019)s*, 19, 5417−5449, [https://doi.org/10.5194/acp-19-](https://doi.org/10.5194/acp-19-5417-2019) [5417-2019](https://doi.org/10.5194/acp-19-5417-2019).
- <span id="page-8-1"></span>Wang, Y. H., and Coauthors, 2020b: Contrasting trends of PM<sub>2.5</sub> and surface-ozone concentrations in China [from 2013 to](https://doi.org/10.1093/nsr/nwaa032) 2017. *[National Scienc](https://doi.org/10.1093/nsr/nwaa032)e Review*, 7[, 1331−1339,](https://doi.org/10.5194/acp-4-955-2004) [https://doi.org/](https://doi.org/10.1093/nsr/nwaa032) [10.1093/nsr/nwaa032](https://doi.org/10.1093/nsr/nwaa032).
- <span id="page-8-8"></span>Wittrock, F., H. Oetjen, A. Richter, S. Fietkau, T. Medeke, A. Rozanov, and J. P. Burrows, 2004: MAX-DOAS measurements of atmospheric trace gases in Ny-Ålesund−Radiative transfer studies and their application. *[Atmospheric Chemistry](https://doi.org/10.1038/s41377-019-0210-6) [and](https://doi.org/10.5194/acp-4-955-2004) [Physi](https://doi.org/10.5194/acp-4-955-2004)cs*, 4, 955−966, [https://doi.org/10.5194/acp-4-955-](https://doi.org/10.5194/acp-4-955-2004) [2004](https://doi.org/10.5194/acp-4-955-2004).
- <span id="page-8-5"></span>Zhang, C. X., and Coauthors, 2019: Satellite UV-Vis spectroscopy: Implications for air quality trends and their driving forces in Chi[na during 2005−2017.](https://doi.org/10.1038/s41377-019-0210-6) *Light: Science & Applications*, 8, 100,<https://doi.org/10.1038/s41377-019-0210-6>.