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Application of respiration and FDA hydrolysis measurements for estimating microbial activity during composting processes

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Abstract Olive-tree leaves (OL) were mixed with olive press cake (OPC) and extracted olive press cake (EPC) at 1:1 dw/dw ratios to prepare two composting mixtures (OL+ OPC and OL+EPC). Both $CO₂-C$ evolution and fluorescein diacetate (FDA) hydrolysis, determined as estimates of the microbial activity during composting, were related to temperature fluctuations in the compost piles, showing greater values at the temperature peaks, compared to the end, of each thermophilic phase. This, however, was only shown after handling and incubating samples at the temperatures of the compost mixtures at the sampling times and not at a low standard temperature. Incubating samples from thermophilic phases at low standard temperatures resulted in underestimation of the microbial activity occurring during composting. The effect of incubation temperature was less dramatic for FDA hydrolysis compared to $CO₂$ –C evolution measurements, probably reflecting the reduced dependence of enzymes involved in FDA hydrolysis on the respective temperatures. However, FDA hydrolysis was a less sensitive indicator of microbial activity, probably due to extracellular cleavage of fluorescein by persistent esterases, at lowered microbial activity phases. Total microbial biomass, estimated by the fumigation–extraction method, was not consistently related to temperature fluctuations during composting and showed a clear increase at the end of composting, probably resulting from a large slow-growing mycelial community colonising the end products. Since high temperatures did not induce sig-

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nificant non-microbial $CO₂-C$ release and FDA degradation, we propose the performance of microbial activity measurements during thermophilic composting phases at the actual temperatures evolving in the composts.

Keywords Microbial activity \cdot Compost \cdot CO₂ evolution \cdot FDA . Thermophilic

Introduction

Composting is a widely used environment-friendly method for treating organic wastes and by-products before recycling them into soils. It consists of an aerobic self-heating biodegradation process of organic materials performed at high rates by a variety of mesophilic/thermotolerant and thermophilic micro-organisms, predominantly of soil origin (Finstein and Morris [1975;](#page-6-0) de Bertoldi et al. [1983](#page-6-0)). Therefore, methodologies for monitoring both microbial diversity and activity have been used to understand and control composting processes.

Respiration measurements have been used for a long time as standard methods for estimating microbial activity in composts measuring either $CO₂$ evolution or $O₂$ depletion (Shulze [1962](#page-7-0); Suler and Finstein [1977](#page-7-0); Manios and Balis [1983;](#page-7-0) Nakasaki et al. [1985](#page-7-0)). Their advantage is that they are broad evaluators of the spectrum of metabolic processes, integrated under the term "microbial activity", and they are directly related to decomposition. Low specificity enzyme assays have also been used for estimating microbial activity, detecting omnipresent classes of enzymes as dehydrogenases (Forster et al. [1993](#page-6-0); Benito et al. [2003](#page-6-0)), phosphatases (Vuorinen [1999\)](#page-7-0) and esterases (Craft and Nelson [1996\)](#page-6-0). They are sensitive and easy to perform assays, but they are more difficult to interpret in complex environmental samples, as they monitor enzymatic processes indirectly related to overall energy release and decomposition and do not distinguish intracellular enzyme activities from extracellular enzyme activities from potentially stabilized enzymes not reflecting current microbial activity (Insam [2001](#page-6-0); Nannipieri et al. [2002\)](#page-7-0). Moreover,

their results may depend on the taxonomic structure of the microbial community and the interference of other physicochemical processes in complex environmental media, affecting measurement precision. Among enzyme assays, the fluorescein diacetate (FDA) hydrolysis technique, evaluating the potential activity of ester-cleaving enzymes, frequently appears in recent studies of composting processes (Levanon and Pluda [2002](#page-6-0); Garcia-Gomez et al. [2003](#page-6-0); Ryckeboer et al. [2003;](#page-7-0) Smith and Hughes [2004\)](#page-7-0).

Both respiration and FDA hydrolysis methods were adapted from techniques developed for environment as soils (Birch and Friend [1956;](#page-6-0) Stotzky [1965](#page-7-0); Anderson [1982](#page-6-0); Schnürer and Rosswall [1982;](#page-7-0) Frankenberger and Dick [1983](#page-6-0)) or litter and container media (Swisher and Carroll [1980](#page-7-0); Schnürer and Rosswall [1982;](#page-7-0) Inbar et al. [1991](#page-6-0)) and were also applied in aquatic systems (Fontvieille et al. [1992\)](#page-6-0) and sediments (Meyer-Reil and Köster [1992\)](#page-7-0). These environments, however, do not show the large temperature increases observed during a typical composting process, which induce not only a dramatic microbial activity rise, up to temperatures around 50–60°C (McKinley and Vestal [1984](#page-7-0); Derikx et al. [1990](#page-6-0); Weppen [2002\)](#page-7-0), but also physiological adaptation of the microbial community at higher temperature optima (McKinley and Vestal [1984](#page-7-0); Lasaridi et al. [1996\)](#page-6-0). Therefore, applying the methods above at pre-set low temperatures (usually around 20– 30°C) may not be appropriate for estimating the microbial activity during thermophilic composting phases.

Residues/by-products from the olive-mill agro-industry, i.e. olive-tree leaves (OL), olive press cake (OPC) and extracted olive press cake (EPC), are produced in large amounts in olive cultivation regions and can be successfully composted (Madejon et al. [1998](#page-7-0); Filippi et al. [2002](#page-6-0); Garcia-Gomez et al. [2003\)](#page-6-0). They are typical xenobioticfree organic materials of vegetative origin and were used to make the two compost mixtures (OL+OPC and OL+EPC) employed in this work.

The aim of the study was to investigate how sample incubation conditions affect the estimation of microbial activity occurring during composting and determined by respiration and FDA hydrolysis. Measurements at temperatures resembling those developing in situ during the composting process were compared to low and standard temperature measurements.

Materials and methods

Composting materials and processes

Agricultural residues/by-products from the olive oil agroindustry, specifically OL, OPC and EPC, were mixed at 1:1 dw/dw ratios to prepare two compost mixtures, OL+ OPC and OL+EPC. The characteristics of the materials are shown in Table 1.

The composting process was performed in static perimetrically insulated containers (0.16-m^3) volume) for a period of 3 months. Aeration was passive from bottom to top, and a 2-cm-thick layer of polyurethane was used to cover the sides of the containers to minimize temperature losses. The moisture content was kept between 40 and 60% of the water-holding capacity throughout the composting process. Turnings were performed and water was added at the end of each thermophilic phase.

Microbiological analysis

FDA hydrolysis was performed according to Adam and Duncan ([2001](#page-6-0)), following the addition of 6 g of compost sample to 60 ml of 60 mM potassium phosphate buffer, pH 7.6. The enzymatic hydrolysis started by adding 0.4 ml FDA solution $(1 \text{ mg } \text{m}^{-1})$. The sample was shaken for 10 min in an orbital incubator, 15 ml sub-samples were obtained at 0 and 10 min and the hydrolysis reaction was terminated by adding 15 ml CHCl₃/CH₃OH (2:1 vol vol⁻¹). Following centrifugation $(700 \times g)$ and filtration of the aqueous phase, the absorbance of filtrates was measured at 490 nm. Blanks, without addition of FDA, were also included to correct for background absorbance, and the concentration of fluorescein release was determined against a calibration curve, produced by 0-, 1-, 2-, 3-, 4- and 5 - μ g ml⁻¹ fluorescein standards.

Carbon dioxide evolution was measured after $CO₂$ entrapment in alkali (Anderson [1982\)](#page-6-0). A dish with 7 g of compost mixture and a beaker containing 15 ml 0.5 N NaOH were placed in a 2-l screw-sealed bottle. The sample was incubated for 21 h, and subsequently, the alkali trap was removed and titrated with dilute HCl in the presence of phenolphthalein, following the addition of 6 ml 3 N BaCl₂. Blanks containing no compost were also included.

Table 1 Chemical characteristics of the composts (\pm values are SE of means, $n=3$)

	Density $(g \text{ ml}^{-1})$	C/N ratio	Ash content $\frac{6}{20}$ wt wt ⁻¹)	Total N $(\% \text{ wt w}^{-1})$	Total P $(g \text{ kg}^{-1}$ dry wt)	Total K $(g \text{ kg}^{-1}$ dry wt)	$NH4-N$ $(mg kg^{-1}$ dry wt	$NO3-N$ $(g \text{ kg}^{-1}$ dry wt
Start of composting								
$OL + EPC$	0.36 ± 0.01	32.5	5.9 ± 0.43	1.7 ± 0.10	1.2 ± 0.09	10.0 ± 0.13	13.3 ± 3.5	1.93±0.13
$OL + OPC$	0.32 ± 0.02	35.9	5.3 ± 0.30	1.5 ± 0.14	1.1 ± 0.08	10.6 ± 0.14	14.9 ± 4.5	1.86 ± 0.13
End of composting								
$OL + EPC$	0.54 ± 0.01	20.7	12.0 ± 0.31	2.5 ± 0.03	1.9 ± 0.06	14.5 ± 0.24	69.6 ± 3.7	1.3 ± 0.06
$OL + OPC$	0.53 ± 0.02	22.9	8.9 ± 0.18	2.3 ± 0.03	1.4 ± 0.06	11.8 ± 0.27	49.6 ± 2.9	1.3 ± 0.16

OL Olive-tree leaves, OPC olive press cake, EPC extracted olive press cake

At the peaks of the thermophilic phases, sample treatment and incubation for FDA degradation and $CO₂$ evolution analysis were performed at two temperatures: one equal to the temperature of the compost at the sampling time and the other at room temperature (25°C). At the end of each thermophilic phase, FDA degradation and $CO₂$ evolution analysis were performed at 25°C only, which matched the temperatures of the composts at the sampling time $(\pm 3^{\circ}C)$.

FDA hydrolysis determinations were also performed in the absence of compost samples following shaking for 10 min in the orbital incubator at a range of temperatures $(25, 30, 40, 50, 40, 60^{\circ}$ to check for thermal nonenzymatic FDA degradation, potentially induced by high temperatures. Fluorescein absorbance in the presence of compost and soil samples was measured at 490 nm from 6 g samples suspended in 60 ml of 60 mM potassium phosphate buffer, pH 7.6, containing 0.32 ml fluorescein solution (1 mg ml^{-1}) , following shaking for 10 min.

The occurrence of any $CO₂$ release induced by chemical oxidation of organic compounds at high temperatures (nonmicrobial CO_2 release) was determined by measuring CO_2 evolution as described above after incubating sterilized compost samples (121°C for 20 min) from both composts under sterile conditions at 25, 40, 50 and 60°C. The effect of sample pre-incubation (conditioning) on respiration was examined by pre-incubating samples obtained from a thermophilic phase (50 \degree C in situ) from OL+OPC for 48 h at 25° C before performing the $CO₂$ release determinations at the same temperature. Determination of microbial biomass was performed by the fumigation–extraction technique (Vance et al. [1987](#page-7-0)). Briefly, compost samples (10 g fresh wt) were extracted with 0.5 M K₂SO₄, shaken for 30 min and filtered, either directly or after fumigation (incubation for 5 days at 25°C in a vacuum desiccator, containing 30 ml alcohol-free chloroform). The extracts were further analysed for total C by a wet oxidation– titration procedure (Anderson and Ingram [1996](#page-6-0)).

Statistical analysis

Standard errors of means (±SE) were calculated, and analysis of variance and linear regression analysis were performed using SPSS software (version 10 for Windows OS).

Results

The composting of the OL+OPC and OL+EPC mixtures was carried out by performing successive turnings at the end of each of the respective three and four thermophilic phases (Fig. 1), respectively. The C/N ratios in the original materials decreased during the composting process, whereas the ash content increased (Table [1\)](#page-1-0), indicating significant C losses. The total N, P and K contents, the

concentration of NH_4^+ -N and the bulk density of the final compost products also increased during the process.

The $CO₂$ evolution and the degradation of FDA during composting of OL+OPC and OL+EPC were related to temperature fluctuations in the compost piles, showing greater values of fluorescein and $CO₂$ release when the composts were at their peak temperatures as compared to the values observed at the end of each thermophilic phase. This, however, was only shown when sample incubations were performed directly after sampling at the temperatures occurring in the compost piles (Figs. [2](#page-3-0) and [3\)](#page-3-0). The $CO₂$ evolution and the degradation of FDA from the OL+OPC and OL+EPC mixtures were not related to temperature fluctuations in the compost piles when the samples were incubated at room temperature (Figs. [2](#page-3-0) and [3](#page-3-0)). The differences between the two sample treatment processes were greater for the $CO₂$ compared to FDA determinations.

No $CO₂$ evolution was observed at either low or high incubation temperatures from any of the sterilized compost samples, where the microbial biomass had been killed, indicating lack of chemical oxidation at temperatures up to 60°C. Pre-incubation of samples obtained during a thermophilic phase (50 \degree C in situ) from OL+OPC at 25 \degree C resulted in only a marginal increase in the following respiration measurements, compared to direct respiration measurements at 25°C. In contrast, respiration measurements performed at 50°C showed dramatically greater values (Fig. [4](#page-4-0)).

A limited FDA hydrolysis was observed at temperatures up to 40° C (Fig. [5\)](#page-4-0), which was doubled at temperatures between 50 and 60°C. The compost materials used in this study showed either no direct fluorescein absorbance (OL+ EPC) or 10% fluorescein absorbance (OL+OPC) during incubations containing fluorescein instead of FDA. A correction was therefore applied on the FDA hydrolysis values obtained for OL+OPC.

The microbial biomass pattern was not consistent; however, at the end of the composting process when microbial

Fig. 2 CO_2 –C release during composting of OL+OPC (a) and OL+ EPC (b). Incubations were performed at the compost temperatures occurring at sampling (whole bars) and at 25°C (filled part of bars).

activity values were low, high microbial biomass was found in both composts (Fig. [6\)](#page-5-0). This indicates changes in the metabolic quotient $(C_{\text{respired}}/C_{\text{microbial}})$ probably related to differences in C availability and/or the composition of the microbial community.

Bars represent \pm SE of means, $n=3$ (*OL* Olive-tree leaves, *OPC* olive press cake, EPC extracted olive press cake)

Discussion

Higher fluorescein and $CO₂$ release at the peak compared to the end of the thermophilic phases are expected, since the temperature increases during composting result from

Fig. 3 FDA hydrolysis during composting of OL+OPC (a) and OL+EPC (b). Incubations were performed at the compost temperatures occurring at sampling (whole bars) and at 25°C (filled part of

bars). Bars represent \pm SE of means, n=3 (OL Olive-tree leaves, OPC olive press cake, EPC extracted olive press cake)

Fig. 4 Effect of incubation temperature on $CO₂$ –C release from compost material sampled at the peak of a thermophilic phase at 50°C and incubated directly at 50 and 25°C, or at 25°C following pre-incubation for 48 h at 25°C. Bars represent \pm SE of means, n=3

Fig. 5 Effect of temperature on FDA degradation in the absence of compost, during incubations at 25, 30, 40, 50 and 60°C. Bars represent \pm SE of means, $n=3$

extensive aerobic microbial activity (Finstein and Morris [1975](#page-6-0)). Micro-organisms show their greatest bio-oxidative potential for a wide range of materials between 45 and 60°C (Wiley [1957](#page-7-0); Jeris and Regan [1973](#page-6-0); Suler and Finstein [1977;](#page-7-0) MacGregor et al. [1981;](#page-6-0) Derikx et al. [1990](#page-6-0); Lopez-Real and Vere [1992;](#page-6-0) Stentiford [1996\)](#page-7-0). This pattern, however, was only obtained in our microbial activity measurements when sample incubations were carried out directly after sampling at temperatures similar to those occurring at the respective composting phases (Figs. [2](#page-3-0) and [3\)](#page-3-0). Incubating compost samples obtained during the thermophilic phases at low standard temperatures led to significant underestimation of the microbial activity, probably as a result of (1) physiological adaptation shock and/ or (2) poor metabolic activity at low temperature. Short pre-incubations to allow for physiological adaptation of the microbial community to the low-temperature incubations showed only a marginal effect on respiration activity,

indicating that reduced metabolic rates at lower temperatures were the main reason for the observed reduction in microbial activity. Incubation temperatures at 37°C have been suggested as a compromise (Iannotti-Frost et al. [1992](#page-6-0); Iannotti et al. [1993\)](#page-6-0). They could be acceptable for certain materials; however, $CO₂$ release was still underestimated as indicated by the measurements from the last thermophilic phases in our study (Fig. [2](#page-3-0)). Similar observations were also made by Mari et al. ([2003\)](#page-7-0) for incubations at 35°C.

In a study of vegetable and garden waste composting, Ryckeboer et al. ([2003\)](#page-7-0) observed a positive correlation between compost temperature and oxygen depletion but a weak negative correlation between temperature- and the log-transformed FDA hydrolysis data. This seems contradictory as both variables estimate microbial activity; it is, however, less surprising according to our findings since oxygen concentration measurements were made at the outlet gasses of the composter (produced at the actual composting temperatures), whereas FDA hydrolysis measurements were performed in compost samples incubated at 25°C. Relatively small respiration increases were reported during thermophilic composting phases by other researchers, applying respiration measurements at low temperatures (Tiquia et al. [1996](#page-7-0); Levanon and Pluda [2002](#page-6-0); Paredes et al. [2002](#page-7-0)), and small FDA hydrolysis values were shown (Levanon and Pluda [2002\)](#page-6-0). Moreover, in recent studies where respiration was monitored throughout the composting process by $CO₂$ evolution measurements at standard low-temperature incubations, only a part of the reported C losses seems to be accounted for by $CO₂-C$ evolution (Benito et al. [2003;](#page-6-0) Garcia-Gomez et al. [2003](#page-6-0)). Assuming adequate oxygenation levels necessary for obtaining the characteristic $CO₂$ flush at the initial composting stages (Michel and Reddy [1998](#page-7-0)), an underestimation of microbial activity during thermophilic composting phases may have contributed to the results above. On the contrary, large increases are shown at thermophilic composting phases when measurements are performed directly in the gas outlet of composters (Shulze [1962;](#page-7-0) Jeris and Regan [1973](#page-6-0); Nakasaki et al. [1985;](#page-7-0) Komilis and Ham [2000\)](#page-6-0). This respiration flush is particularly evident at the initial stages of thermophilic phases (Michel et al. [1993](#page-7-0); Atkinson et al. [1996](#page-6-0); Hellmann et al. [1997;](#page-6-0) Wong and Fang [2000\)](#page-7-0) when the microbial activity flush results in rapid self-heating of the compost mass.

To perform $CO₂$ evolution and FDA hydrolysis measurements at high incubation temperatures, however, significant chemo-oxidation of C substrates and thermal degradation of FDA, respectively, must be excluded. Indeed, the killing of the microbial biomass resulted in the lack of any $CO₂$ evolution in all sterilized samples incubated at temperatures up to 60°C, indicating no chemooxidation phenomena. Therefore, the higher $CO₂$ evolution values obtained at higher temperature incubations should be attributed to higher microbial activity alone. Our data also indicate that chemical hydrolysis of FDA at high temperatures was limited, reaching less than 3% of the FDA hydrolysis observed in compost incubations at temperatures up to 40°C and less than 6% for temperatures

Fig. 6 Microbial biomass size during composting of OL+OPC (a) and OL + EPC (b). Bars represent \pm SE of means, $n=3$ (OL Olive-tree leaves, OPC olive press cake, EPC extracted olive press cake)

up to 60° C (Figs. [3](#page-3-0) and [5](#page-4-0)). This is in line with Adam and Duncan [\(2001](#page-6-0)), who showed no spontaneous hydrolysis of FDA in the range from 20 to 40^oC. At higher temperatures, however, considerable spontaneous hydrolysis of FDA could occur, making comparisons with lower temperature incubations difficult, except if a reference FDA hydrolysis blank is used. Significant absorption of fluorescein may occur in humic materials. Their presence in solution would therefore reduce the fluorescein absorbance readings by the FDA hydrolysis assay. Our preliminary tests showed that the presence of different compost materials and soils in control fluorescein solution standards resulted in 0 to 29% lower fluorescein absorbance for a range of five compost materials and 4 to 13% for a range of seven soils (data not shown). Similarly, variable effects observed by Adam and Duncan [\(2001](#page-6-0)) indicate the need for preliminary tests before applying the FDA hydrolysis method in new environmental samples. Correction for lower fluorescein absorbance values in the presence of the compost materials was indeed needed for OL+OPC, but not for OL+EPC, which did not affect fluorescein absorbance values.

Much greater differences in microbial activity between the peak and the end of each thermophilic phase were shown for the $CO₂$ evolution compared to the FDA degradation. It appears that total respiration measurements are better evaluators of differences in microbial activity at each composting stage, compared to the fluorescein release, since microbial activity may be reduced towards the end of each thermophilic phase (limited availability of labile-C substrates which directly affects bio-oxidative processes and respiration). Indeed, enzymes, such as extracellular thermotolerant esterases, reactive to FDA may still maintain their structure and activity when microbial activity is

low (Insam [2001;](#page-6-0) Nannipieri et al. [2003](#page-7-0)). Remarkably stable and persistent extracellular esterases have been isolated from complex environmental media like soils (Satyanarayana and Getzin [1973\)](#page-7-0). This may lead to relatively high fluorescein values following the thermophilic composting phases. FDA is degraded by a broad range of enzymes capable of hydrolysing the ester bond, and therefore, any significant preservation of enzyme structure and activity could lead to methodological overestimation of the active bio-oxidative community in periods of lowered metabolic activity. The long lag phase in the drop in FDA hydrolysis, shown by Ryckeboer et al. ([2003\)](#page-7-0) following the thermophilic composting phase, is in accordance with this hypothesis. The smaller effect of the incubation temperature on FDA hydrolysis compared to $CO₂$ evolution in our samples obtained from thermophilic phases may be due to similar reasons.

The fluctuations in microbial activity and heat production were not necessarily accompanied by similar fluctuations in the microbial biomass, indicating changes in the metabolic quotient ($qCO₂-C$) during the composting process. At the end of the composting process, the high microbial biomass and low $CO₂$ evolution (reduced metabolic quotient) could be related to the depletion of labile-C and/ or changes in the composition of the microbial community: Increased microbial biomass is observed at the end of the thermophilic composting stages (Hellmann et al. [1997](#page-6-0); Tiquia et al. [1996](#page-7-0)), and fungi and actinomycetes dominate (Levanon and Pluda [2002;](#page-6-0) Tiquia et al. [2002;](#page-7-0) Ryckeboer et al. [2003\)](#page-7-0). Indeed, changes in the microbial community structure may affect, among other factors, the metabolic quotient (Wardle and Ghani [1995](#page-7-0); Insam et al. [1996\)](#page-6-0), and the increase in the fungal-to-bacterial biomass ratio has been directly related to $qCO₂$ decline (Sakamoto and Oba [1994](#page-7-0); Blagodatskaya and Anderson 1998). A reduced metabolic quotient has been proposed as a sensitive indicator of compost stability (Kostov et al. 1994); as fungi release less $CO₂$ per unit cell mass compared to bacteria (Alexander 1977), the qCO_2 reduction may derive from a shift towards filamentous micro-organisms rather than directly from the depletion of labile-C that occurs at the end of composting.

In conclusion, since (1) respiration is greatly underestimated when samples from thermophilic composting phases are treated and incubated at lower temperatures, (2) increased temperature incubations for respiration measurements are free of non-biological $CO₂$ release and (3) respiration measurements are direct indicators of microbial oxidative activity, we propose the incubation and respiration measurements at the temperatures occurring during the composting process as the best evaluators of the microbial activity occurring in situ. The FDA hydrolysis method, even when applied at the relevant temperatures, is probably a less sensitive evaluator of the oxidative microbial activity, especially when the latter is rapidly declining during the final stages of the process.

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