

J. A. Dijk · A. J. M. Stams · G. Schraa · H. Ballerstedt ·
J. A. M. de Bont · J. Gerritse

Anaerobic oxidation of 2-chloroethanol under denitrifying conditions by *Pseudomonas stutzeri* strain JJ

Received: 9 January 2003 / Revised: 1 April 2003 / Accepted: 4 April 2003 / Published online: 27 May 2003
© Springer-Verlag 2003

Abstract A bacterium that uses 2-chloroethanol as sole energy and carbon source coupled to denitrification was isolated from 1,2-dichloroethane-contaminated soil. Its 16 S rDNA sequence showed 98% similarity with the type strain of *Pseudomonas stutzeri* (DSM 5190) and the isolate was tentatively identified as *Pseudomonas stutzeri* strain JJ. Strain JJ oxidized 2-chloroethanol completely to CO₂ with NO₃⁻ or O₂ as electron acceptor, with a preference for O₂ if supplied in combination. Optimum growth on 2-chloroethanol with nitrate occurred at 30 °C with a μ_{\max} of 0.14 h⁻¹ and a yield of 4.4 g protein per mol 2-chloroethanol metabolized. Under aerobic conditions, the μ_{\max} was 0.31 h⁻¹. NO₂⁻ also served as electron acceptor, but reduction of Fe(OH)₃, MnO₂, SO₄²⁻, fumarate or ClO₃⁻ was not observed. Another chlorinated compound used as sole energy and carbon source under aerobic and denitrifying conditions was chloroacetate. Various different bacterial strains, including some closely related *Pseudomonas stutzeri* strains, were tested for their ability to grow on 2-chloroethanol as sole energy and carbon source under aerobic and denitrifying conditions, respectively. Only three strains, *Pseudomonas stutzeri* strain LMD 76.42, *Pseudomonas putida* US2 and *Xanthobacter autotrophicus* GJ10, grew aerobically on 2-chloroethanol. This is the first report of oxidation of 2-chloroethanol under denitrifying conditions by a pure bacterial culture.

Introduction

2-Chloroethanol (CE) is used in industry mainly for the synthesis of insecticides and as a solvent. Very little is known about the emission and fate of CE in the environment due to the fact that CE is not included in routine analyses for soil pollution. CE is metabolized both in vivo and in vitro by mammalian alcohol dehydrogenase to 2-chloroacetaldehyde, which is considered to be mutagenic (McCann et al. 1975).

Several bacteria have been described that can degrade CE aerobically, including *Pseudomonas* sp. strain CE1, *Pseudomonas putida* US2, *Pseudomonas* sp. strain DCA1, *Xanthobacter autotrophicus* GJ10, *Pseudomonas* sp. strain GJ1, *Ancylobacter aquaticus* strains AD20 and AD25, and *Mycobacterium* sp. strain GP1 (Hage and Hartmans 1999; Janssen et al. 1985, 1984; Poelarends et al. 1999; Strotmann et al. 1990; Stucki et al. 1981; van den Wijngaard et al. 1992). Usually, degradation proceeds via the intermediates 2-chloroacetaldehyde and chloroacetate to glycolic acid. Recently, Poelarends et al. (1999) showed that *Mycobacterium* sp. strain GP1 degraded CE via ethylene oxide, without production of the 2-chloroacetaldehyde intermediate.

Although the aerobic CE-metabolizing *Pseudomonas* sp. strain GJ1 and *Xanthobacter autotrophicus* GJ10 are able to reduce NO₃⁻, nothing is known about the degradation of CE under denitrifying conditions.

This report describes the isolation and characterization of *Pseudomonas stutzeri* strain JJ, the first microorganism known to grow anaerobically on CE with NO₃⁻ as electron acceptor. This is especially interesting, since many soils that are contaminated with chlorinated aliphatics are anoxic and NO₃⁻ is often present in the groundwater (Cox et al. 2000; Gerritse et al. 1999).

J. A. Dijk (✉) · A. J. M. Stams · G. Schraa
Laboratory of Microbiology, Wageningen University,
Hesselink van Suchtelenweg 4, 6703 CT Wageningen,
The Netherlands
e-mail: John.Dijk@wur.nl
Tel.: +31-317-483739
Fax: +31-317-483829

H. Ballerstedt · J. A. M. de Bont · J. Gerritse
TNO Environment, Energy and Process Innovation,
P.O. Box 342, 7300 AH Apeldoorn, The Netherlands

Materials and methods

Organisms

Pseudomonas stutzeri strain LMD 76.42 (Palleroni et al. 1970) and *Pseudomonas stutzeri* strain LMD 26.48 (Sijderius 1946) were purchased from The Netherlands Culture Collection of Bacteria (NCCB, Utrecht, The Netherlands). *Pseudomonas mendocina* (DSM 50017, type strain) (Skerman et al. 1980), *Pseudomonas stutzeri* (DSM 5190, type strain) (Skerman et al. 1980), (Middelhoven and Bakker 1982), *Xanthobacter autotrophicus* GJ10 (Janssen et al. 1984), *Pseudomonas stutzeri* strain KC (Criddle et al. 1990), *Pseudomonas putida* US2 (Strotmann et al. 1990) and *Alcaligenes xylosoxidans* subsp. *denitrificans* (DSM 6505) (Blake and Hegeman 1987) were obtained from the Deutsche Sammlung von Mikroorganismen und Zellkulturen (DSMZ, Braunschweig, Germany). *Pseudomonas* sp. strain P51 (van der Meer et al. 1987), *Pseudomonas putida* strain C 3024, *Pseudomonas stutzeri* strain 346B (DSM 6538) (Baggi et al. 1987) and *Pseudomonas aeruginosa* (ATCC 17933) (Çetin et al. 1965) were obtained from the culture collection of the Laboratory of Microbiology of Wageningen University. The new isolate, *Pseudomonas stutzeri* strain JJ, has been deposited in the DSM as strain DSM 15012.

Media and cultivation techniques

Batches were prepared under a N₂ atmosphere in 120-ml serum bottles that were crimp-sealed with butyl-rubber stoppers and contained 50 ml medium. Standard medium consisted of the following components: Na₂HPO₄/KH₂PO₄ buffer (20 mM, pH 7), 0.7 g (NH₄)₂SO₄/l, 0.1 g MgSO₄·7H₂O/l, 0.05 g CaCl₂·2H₂O/l, 1 ml trace elements/l (Gerritse et al. 1992) and 1 ml vitamin solution/l (Heijthuisen and Hansen 1986). Vitamins were filter sterilized (0.2-µm) and added to the media after autoclaving. Phosphate buffer was also added separately after autoclaving. Unless mentioned otherwise, CE (5 mM) was added as electron donor and NaNO₃ (10 mM) as electron acceptor from separately sterilized stock solutions. Batches were transferred weekly to fresh media to maintain actively growing cultures. Batch enrichments were initially incubated at 20 °C, but after isolation, strain JJ was incubated at 30 °C unless mentioned otherwise. Agar plates consisted of the above mentioned components plus 15 g Bacto-agar/l (Difco Laboratories, Detroit, Mich., USA). The plates were incubated in a nitrogen-flushed anaerobic jar or in air.

Enrichment, isolation and purity control

Strain JJ originates from an anaerobic enrichment culture growing on 1,2-dichloroethane as sole energy and carbon source in the presence of NO₃⁻ (Gerritse et al. 1999). One ml of the enrichment culture was transferred to standard medium, containing 1 mM CE and 5 mM NaNO₃. Upon growth, samples were plated on anoxic agar plates. Colonies were isolated from the plates and transferred to liquid medium. This process was repeated several times until a pure culture was obtained.

Purity was tested by phase-contrast microscopy and by aerobic growth on both solid and liquid standard media with glucose (5 mM) and yeast extract (5 g/l). Oxidase, catalase and Gram-type tests were done at The Netherlands Culture Collection of Bacteria (Utrecht, The Netherlands).

Identification

Physiological tests

The API 20NE and BIOLOG GN analyses were carried out at The Netherlands Culture Collection of Bacteria (Utrecht, The Netherlands).

DNA isolation, amplification, sequence determination, and analysis of 16S rDNA

Genomic DNA was isolated from whole cells by a combined method of SDS-treatment and bead beating in the presence of phenol, followed by subsequent phenol-chloroform-isoamylalcohol (25:24:1) extraction (Duarte et al. 1998; Ralebitso et al. 2001). The final DNA extracts were further purified, using the Wizard DNA Clean-Up System (Promega, Leiden, The Netherlands). PCR amplification of the 16S rRNA gene was done using 1 µl DNA extract as template, bacteria-specific primers FD1 and RP2 (Weisburg et al. 1991) (*Escherichia coli* numbering 8–27, and 1512–1492 (Brosius et al. 1978), AmpliTaq LD DNA-polymerase (Perkin Elmer Life Sciences, Zaventem, Belgium) and a thermal cycler temperature program as follows: denaturation at 95 °C for 5 min; touch down PCR step: 10 cycles of 94 °C for 30 s, 65 °C–55 °C for 30 s, 72 °C for 1 min; 27 similar cycles at 55 °C annealing temperature and with a 20-s extension of elongation phase per cycle. PCR product was purified via agarose gel separation stained with 1× Nuclistain (National Diagnostics, Manville, USA) and subsequent DNA extraction using a DNA extraction kit (MBI Fermentas, St. Leon-Rot, Germany). The purified 16S rDNA was ligated into a pGEM-T Easy Vector (Promega) and transformed into *E. coli* XL 1-Blue-MRF' cells (Stratagene, La Jolla, Calif., USA). The transformed cells were subjected to blue-white color screening, randomly picked white colonies were grown overnight in selective LB-medium as described elsewhere (Sambrook et al. 1989), and recombinant plasmids were subsequently isolated using QIAprep Spin Miniprep kit (Qiagen, Westburg b.v., Leusden, The Netherlands). Cloned inserts were reamplified using pGEM-T-Easy Vector binding primers PG1f and PG2r (Hantke 2000), and subsequently digested with restriction enzymes *MspI* and *RsaI* (Promega, Leiden, The Netherlands) overnight at 37 °C. The restriction patterns were monitored to confirm uniformity of the inserts using 4 % agarose gels in 1× TBE-buffer (Sambrook et al. 1989). Bidirectional sequence was determined by MWG Biotech AG (Ebersberg, Germany) using universal SP6 and T7 primers and primers 984f (Heuer et al. 1999) and 907r (Muyzer et al. 1995). The determined sequence was compared with 16S rDNA sequences obtained from the EMBL data library using the Fasta3 program (Pearson 1990).

Characterization of the CE-degrading strain JJ

To determine the optimum growth temperature, batch cultures of strain JJ with 5 mM CE and 10 mM NO₃⁻ were incubated at 4, 12, 20, 25, 30, 35 and 45 °C, respectively. Liquid samples (1, 5 ml) were taken at different time points and the optical density at 660 nm (OD₆₆₀) was determined. Specific growth rates were obtained by plotting the natural logarithm of the OD₆₆₀ against time. The determination of the optimum pH was done in the same way. Batch cultures of strain JJ were incubated at pH 6, 6.5, 7, 7.5, 8, 8.5, 9 and 9.5.

Electron donor utilization

Initially, utilization of chlorinated electron donor was screened in 10-ml vials containing 5 ml standard medium and crimp-sealed with butyl-rubber stoppers. Chlorinated electron donors were added from autoclaved stock solutions to a final concentration of 1 mM. Media were incubated anaerobically with 5 mM NO₃⁻, or under an air atmosphere. After 8 weeks of incubation, the Cl⁻ concentration was determined. Batches with a chloride production >0.2 mM were transferred to standard media in 120-ml serum bottles with 5 mM electron donor and incubated anaerobically with 10 mM NO₃⁻, or under air, respectively. After 3 weeks, growth was determined by visual detection of increased turbidity.

Growth on non-chlorinated electron donors was tested in 120-ml serum bottles with 50 ml standard medium, containing 10 mM electron donor with 10 mM NO₃⁻, or under air, respectively. After

3 weeks, growth was determined by visual detection of increased turbidity.

Electron acceptor utilization

Electron acceptors were added according to the theoretical amounts of reduction equivalents needed for the complete oxidation of 5 mM CE: NaNO₃ (10 mM), NaNO₂ (17 mM), MnO₂ (25 mM), Fe(OH)₃ (50 mM), Na₂SO₄ (6 mM), NaClO₃ (8 mM), O₂ (air atmosphere; 12 mmol O₂/l medium) and fumarate (25 mM). Amorphous Fe(OH)₃ was prepared by neutralizing a 0.4 M FeCl₃ solution with 1 M NaOH until a pH of 7 was achieved (Lovley and Phillips 1986). MnO₂ was prepared by mixing equal amounts of 0.4 M MnO₄ and 0.6 M MnCl₂ and adjusting the pH to 10 with 1 M NaOH (Lovley and Phillips 1988; Murray 1974). After preparation, the Fe(OH)₃ and MnO₂ suspensions were washed four times with milli-Q water by centrifugation. All stock solutions were flushed with N₂ before use. After 4 weeks of incubation, the CE concentration of the batches with the respective electron acceptors was determined. When more than 2 mM of CE had been consumed, the electron acceptor was considered to be utilized.

Growth on CE

Batch cultures of strain JJ were grown in 1-l screw-cap flasks with butyl-rubber stoppers, containing 500 ml medium with 5 mM CE and 10 mM NO₃⁻. The flasks were inoculated with 1–10 ml of an exponentially growing batch culture and incubated at 30 °C in a shaking waterbath (90 rpm). The optical densities and the CE, NO₃⁻, NO₂⁻, NH₄⁺ and Cl⁻ concentrations were monitored over time.

Experiments with NO₃⁻ and O₂ present at the same time were carried out in 500-ml screw-cap flasks with butyl-rubber stoppers. The flasks contained 250 ml medium with 10 mM CE, 1 mM O₂ and 2 mM NO₃⁻. The inoculum was pre-grown on CE and a combination of NO₃⁻ and O₂. The optical densities and the CE, NO₃⁻ and O₂ concentrations were monitored over time.

Comparison of the newly isolated strain JJ with other bacterial strains

Aerobic growth on CE of different bacterial strains was tested in standard medium containing 2.5 mM CE. After 3 weeks, the OD₆₆₀ was determined. For anaerobic growth experiments with CE, batch cultures were pre-grown on medium containing: 27 g sodium succinate/l, 1 g KNO₃/l, 1 g K₂HPO₄/l, 1 g (NH₄)₂SO₄/l, 0.1 gMgSO₄/l, 0.05 g CaCl₂ /l, and 1 ml trace elements/l (see Media and cultivation techniques) to induce denitrification. Subsequently, the batch cultures were transferred to anoxic standard medium (2% inoculum) containing 2.5 mM CE and 5 mM NO₃⁻. After 3 weeks, growth was determined by visual detection of increased turbidity.

Analytical methods

OD₆₆₀ was determined in a Perkin-Elmer 55 A UV-VIS spectrophotometer. Cl⁻, NO₃⁻ and NO₂⁻ were determined by suppressor-mediated ion chromatography (Dionex, Breda, The Netherlands) and conductivity detection. Eluent consisted of 1.8 mM Na₂CO₃ and 1.7 mM NaHCO₃ at a flow rate of 1 ml/min. The chromatograph was equipped with an IonPac AS9-SC column (Dionex, Breda, The Netherlands). NH₄⁺ was determined colorimetrically according to the Berthelot reaction (Richterich 1965). CE was measured by extracting 2-ml samples with 1 ml diethylether (containing 0.1 mM 1-butanol as internal standard) and injecting 2 μl into a Hewlett Packard HP 6890 gas chromatograph, equipped with a CPWax 52 CB column (length, 30 m; diameter, 0.32 mm; Chrompack, Middelburg, The Netherlands), using an FID

detector (300 °C). The oven temperature was 80 °C isothermal and the detection limit was 50 μM. CE concentrations are presented as nominal concentrations in the batch cultures, assuming that all CE was present in the liquid phase. O₂ and N₂ were measured on a Chrompack CP9000 gas chromatograph equipped with a CP Molsieve 5Å column (length, 30 m; diameter, 0.53 mm) and a TCD detector (150 °C). The oven temperature was kept at 50 °C. Protein was measured with the Lowry reagent with bovine serum albumin as a standard (Lowry et al. 1951).

Chemicals

All chemicals were of analytic grade with a purity of more than 97% and obtained from commercial sources.

Nucleotide sequence accession numbers

The nucleotide sequence was deposited in the GenBank database under accession number AF411219.

Results

Enrichment and isolation of strain JJ

A 1,2-dichloroethane (1,2-DCA)-degrading, denitrifying enrichment culture was previously obtained from 1,2-DCA-contaminated soil (Gerritse et al. 1999). Upon sub-culturing, the enrichment culture lost the ability to degrade 1,2-DCA. However, after supplying this culture with CE, a presumed intermediate of 1,2-DCA degradation, a well growing enrichment culture on CE (5 mM) and NO₃⁻ (10 mM) was obtained (Dijk et al. 2000). From this enrichment strain, JJ was isolated.

Identification of strain JJ

Strain JJ was identified by using the API 20NE, the ID 32 GN and BIOLOG GN systems and by sequencing the 16S rDNA. The API 20NE, ID 32 GN and BIOLOG GN tests identified strain JJ as *Pseudomonas stutzeri* (99.5% probability). The 16S rDNA sequence was compared with the BLAST database. Strain JJ showed 98% similarity with the type strain of *Pseudomonas stutzeri* (DSM 5190; U26262) (Bennasar et al. 1996). Therefore, the isolate was tentatively designated as *Pseudomonas stutzeri* strain JJ. The general characteristics of strain JJ are presented in Table 1.

Electron donor and acceptor utilization

No difference in the use of electron donor substrates was observed between aerobic and denitrifying conditions. Under both conditions, strain JJ grew on ethanol, propanol, butanol, acetate, pyruvate, glucose, chloroethanol and chloroacetate. The following substrates were not utilized by strain JJ: methanol, lactate, 1,1-dichloroethane, 1,2-dichloroethane, 1,1,1-trichloroethane, 1-

Table 1 Characteristics of *Pseudomonas stutzeri* strain JJ

Gram stain	–
Cell morphology	Rod-shaped
Cell size	1–2.5×0.6–1.3 μm
Colony morphology	Two types; wrinkled and smooth
Oxidase	+
Catalase	+
Motility	+
Spore formation	–
pH optimum	± 7.5
Temp. optimum	30 °C
Growth at 4 °C	–
Growth at 40 °C	+
Growth substrates	Glycogen, glucose, maltose, 3-methylglucose, α -methyl-D-glucoside, β -methyl-D-glucoside, α -methyl-D-mannoside, palatinose, stachyose, D-trehalose, turanose, D-sorbitol, xylitol, dextrin, α -cyclodextrin, inulin, proline, L-alanine, D-ribose, salicin, sedoheptulosan, D-sucrose, D-tagatose, D-xylose, inositol, mannitol, mannann, γ -hydroxybutyrate, γ -hydroxyphenylacetate, L-lactate, mono-methylsuccinate, succinate, <i>N</i> -acetyl L-glutamate, D-galacturonate, alaninamide, D-alanine, L-alanine, L-alanyl-glycine, L-asparagine, glycyl-L-glutamate, L-pyroglutamate, L-serine, putrescine, 2,3-butanediol, glycerol, adenosine, 2'-deoxy adenosine, thymidine, uridine, histidine, γ -cyclodextrin, adenosine-5'-monophosphate, uridine-5'-monophosphate, fructose-6-phosphate, glucose-1-phosphate, glucose-6-phosphate, D-L-glycerol phosphate, 2-ketogluconate, <i>N</i> -acetyl glucosamine, adipate, suberate, 3-hydroxybenzoate, 4-hydroxybenzoate and phenyl-acetate
No growth	Amygdalin, L-arabinose, D-arabitol, arbutin, cellobiose, D-fructose, D-galactose, gentiobiose, mannose, D-melibiose, rhamnase, D-ribose, salicin, sedoheptulosan, D-sucrose, D-tagatose, D-xylose, inositol, mannitol, mannann, γ -hydroxybutyrate, γ -hydroxyphenylacetate, L-lactate, mono-methylsuccinate, succinate, <i>N</i> -acetyl L-glutamate, D-galacturonate, alaninamide, D-alanine, L-alanine, L-alanyl-glycine, L-asparagine, glycyl-L-glutamate, L-pyroglutamate, L-serine, putrescine, 2,3-butanediol, glycerol, adenosine, 2'-deoxy adenosine, thymidine, uridine, histidine, γ -cyclodextrin, adenosine-5'-monophosphate, uridine-5'-monophosphate, fructose-6-phosphate, glucose-1-phosphate, glucose-6-phosphate, D-L-glycerol phosphate, 2-ketogluconate, <i>N</i> -acetyl glucosamine, adipate, suberate, 3-hydroxybenzoate, 4-hydroxybenzoate and phenyl-acetate

chloropropane, 2-chloropropane, 1,2-dichloropropane, 1,3-dichloropropane, 1,3-dichloro-2-propanol, 2-chloropropanol, vinyl chloride, 1,1-dichloroethene, *cis*-1,2-dichloroethene, *trans*-1,2-dichloroethene, trichloroethene and tetrachloroethene.

During growth on CE, strain JJ reduced NO_3^- , NO_2^- and O_2 . Not utilized as electron acceptor were SO_4^{2-} , $\text{Fe}(\text{OH})_3$, MnO_2 , ClO_3^- and fumarate.

Growth on CE

Growth on CE with NO_3^- started after a lag phase of about 12 hours and proceeded according to a maximum specific growth rate (μ_{max}) of 0.14 h^{-1} (Fig. 1A). After 45 days, 4.8 mM of Cl^- was produced and 4.4 mM CE had been consumed (Fig. 1B). Concurrently, 12.2 mM of NO_3^- was reduced, mainly to NO_2^- (6.2 mM) and presumably to N_2 (not measured), since NH_4^+ was not produced. The total amount of protein formed was 19.4 mg/l, corresponding to a yield of 4.4 g protein per mol CE metabolized. When grown on CE and NO_3^- , strain JJ displayed a temperature optimum at 30 °C. No growth was observed at 4 °C or at 45 °C.

When both O_2 and NO_3^- were simultaneously present in a batch culture, strain JJ had a preference for O_2 . The μ_{max} during the aerobic growth phase on CE was 0.31 h^{-1} . When the oxygen concentration dropped to below approximately 500 μM , strain JJ started to reduce NO_3^- to NO_2^- . NO_2^- was subsequently reduced to most likely N_2 gas, since no NH_4^+ formation was detected.

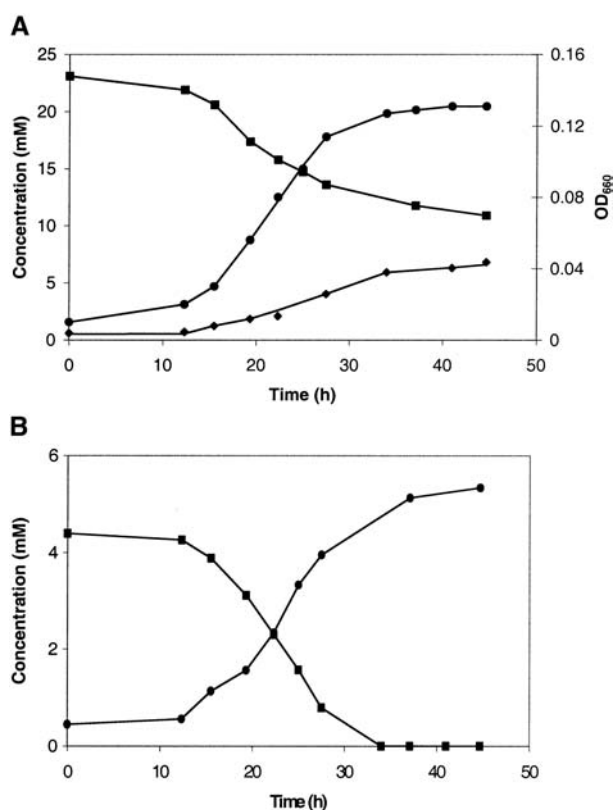


Fig. 1A, B Degradation of 2-chloroethanol (5 mM) with nitrate (10 mM) as electron acceptor by *Pseudomonas stutzeri* strain JJ. **A** Increase in optical density at 660 nm (●), the consumption of NO_3^- (■) and the formation of nitrite (◆). **B** Degradation of 2-chloroethanol (■) and the release of chloride (●)

Table 2 The ability of different *Pseudomonas* strains, *Xanthobacter autotrophicus* GJ10 and *Alcaligenes xylosoxidans* subsp. *denitrificans* to grow on 2-chloroethanol under denitrifying or oxic conditions

Species	Growth conditions		
	Succinate+NO ₃ ⁻	Chloroethanol+NO ₃ ⁻	Chloroethanol+O ₂
Strain JJ	+	+	+
<i>Pseudomonas stutzeri</i> type strain	+	-	-
<i>Pseudomonas stutzeri</i> strain LMD 76.42	+	-	+
<i>Pseudomonas stutzeri</i> strain LMD 26.48	+	-	-
<i>Pseudomonas stutzeri</i> strain KC	+	-	-
<i>Pseudomonas stutzeri</i> strain 346B	+	-	-
<i>Pseudomonas mendocina</i> type strain	+	-	-
<i>Pseudomonas</i> sp. strain P51	-	-	-
<i>Pseudomonas putida</i> strain C 3024	-	-	-
<i>Xanthobacter autotrophicus</i> GJ10	+	-	+
<i>Pseudomonas aeruginosa</i>	+	-	-
<i>Pseudomonas putida</i> US2	+	-	+
<i>Alcaligenes xylosoxidans</i> subsp. <i>denitrificans</i>	+	-	-

Comparison of newly isolated strain JJ with other bacterial strains

Eleven bacterial strains were tested for their ability to grow on CE under both aerobic and denitrifying conditions (Table 2). *Pseudomonas* sp. strain P51 and *Pseudomonas putida* strain P42 could not grow on succinate with NO₃⁻ as electron acceptor. *Pseudomonas stutzeri* strain LMD 76.42, *Pseudomonas putida* US2 and *Xanthobacter autotrophicus* GJ10 were able to oxidize CE under aerobic conditions. The other tested strains were able to denitrify, but did not oxidize CE. Only strain JJ was able to oxidize CE under both aerobic and denitrifying conditions.

Discussion

This report describes the isolation and characterization of a bacterium that is able to grow on CE under aerobic and denitrifying conditions. The isolated strain was designated as *Pseudomonas stutzeri* strain JJ and is the first pure bacterial culture described that is capable of anaerobic oxidation of CE with NO₃⁻ as electron acceptor.

Based on the present findings, the reaction equation for growth of strain JJ on CE plus NO₃⁻ is: CE+2HNO₃→2CO₂+HCl⁻+N₂+3H₂O and CE+5HNO₃→2CO₂+HCl⁻+5HNO₂+2H₂O. When NO₃⁻ was supplied in excess (23.1 mM NO₃⁻+4.4 mM CE), nitrite accumulated (up to 6.2 mM) in batches of strain JJ. It is known that nitrate can have an inhibitory effect on nitrite reduction by *Pseudomonas stutzeri* (Kodoma et al. 1969; Körner and Zumft 1989). Also, the nitrate concentration in the pre-culture is important, since nitrate starvation inactivates the existing capacity of NO₂⁻ reduction (Xu and Enfors 1996).

Various other bacterial strains that were tested, including some closely related *Pseudomonas stutzeri* strains, did not grow on CE with NO₃⁻, indicating that the ability to grow on CE with NO₃⁻ is an exceptional

feature of *Pseudomonas stutzeri* strain JJ. Possibly, strain JJ expresses a 2-chloroethanol dehydrogenase or a hydrolytic 2-chloroethanol dehalogenase, enabling it to grow on CE under denitrifying conditions. However, it is remarkable that *Xanthobacter autotrophicus* GJ10 cannot grow on CE with NO₃⁻. This strain is able to denitrify, to grow on CE aerobically, and it contains both dehydrogenases and dehalogenases, which do not require oxygen (Janssen et al. 1985, 1984). More insight into the regulation of the CE degradation pathway under denitrifying conditions may clarify this observation.

The ability of strain JJ to grow on CE and chloroacetate under denitrifying conditions may be important for survival in its natural habitat. The aquifer from which strain JJ was obtained was anoxic and contained small amounts of NO₃⁻ (<10 μM), chlorinated aliphatics, including 1,2-DCA (up to 3.5 mM), vinyl chloride and various chlorinated methanes and propanes (Gerritse et al. 1999). The ability to use low-chlorinated aliphatics as substrates under denitrifying conditions may add significantly to the competitive strength of strain JJ in the environment. In addition, if O₂ and NO₃⁻ are both present in the field, strain JJ can easily switch from aerobic growth to denitrification, when O₂ becomes limited, and vice versa.

It remains unclear which chlorinated compound(s) served as growth substrate for strain JJ in the field. CE and chloroacetate are degradation products from 1,2-DCA. Unfortunately, CE and chloroacetate are not covered in the standard analyses of the pollutants at this location. Another possibility would be that strain JJ could grow on 1,2-DCA in a mixed culture, since the enrichment from which strain JJ originates was able to degrade 1,2-DCA under denitrifying conditions (Gerritse et al. 1999). Finally, strain JJ may have lost the ability to degrade 1,2-DCA during isolation and cultivation on CE in the laboratory. The transfer and expression of a 1,2-DCA dehalogenase, e.g. from *Xanthobacter autotrophicus* GJ10, might enable strain JJ to grow on 1,2-DCA (Janssen et al. 1985).

The observation that denitrifying bacteria can use chlorinated aliphatics may have important implications for the natural attenuation and bioremediation of contaminated aquifers. The presence of NO_3^- in the groundwater may enable biodegradation of chlorinated aliphatics that entered the soil directly or have been formed in situ via partial dechlorination. Anaerobic oxidation pathways are of particular importance when chlorinated contaminated plumes spread from reduced (e.g. methanogenic, sulfate-reducing) to more oxidized (e.g. iron-reducing and denitrifying) redox zones. These groundwater situations frequently occur at polluted sites in Western Europe.

Nitrate-coupled bioremediation has several advantages because nitrate reducers grow rapidly, NO_3^- dissolves well in water and it does not form precipitates in anoxic groundwater, which frequently clog infiltration wells during aerobic bioremediation.

Oxidation of vinyl chloride and 1,2-DCA has also been demonstrated under denitrifying conditions (Cox et al. 2000; Dijk et al. 2000; Gerritse et al. 1999). Since no microorganisms that are able to oxidize vinyl chloride or 1,2-DCA with NO_3^- have been described thus far, the isolation of such bacteria would provide useful information for natural attenuation and bioremediation.

Acknowledgements We thank Harald Ruijsenaars (Hercules European Research Center) and Hans Visser (Wageningen University) for help with cloning the 16S rDNA. We also thank Dick B. Janssen, who kindly provided bacterial strains. The research described in this paper was supported by and carried out at the Research Centre on Soil, Sediment and Groundwater Management and Remediation WUR/TNO.

References

- Baggi G, Barbieri P, Galli E, Tollari S (1987) Isolation of a *Pseudomonas stutzeri* strain that degrades o-xylene. *Appl Environ Microbiol* 53:2129–2132
- Bennasar A, Rossello MR, Lalucat J, Moore ERB (1996) 16S rRNA gene sequence analysis relative to genomovars of *Pseudomonas stutzeri* and proposal of *Pseudomonas balearica* sp. nov. *Int J Syst Bacteriol* 46:200–205
- Blake CK, Hegeman GD (1987) Plasmid pCBI carries genes for anaerobic benzoate catabolism in *Alcaligenes xylooxidans* subsp. denitrificans PN-1. *J Bacteriol* 169:4878–4883
- Brosius J, Palmer JL, Kennedy JP, Noller HF (1978) Complete nucleotide sequence of a 16S ribosomal RNA gene from *Escherichia coli*. *Proc Natl Acad Sci USA* 75:4801–4805
- Çetin ET, Töreçi K, Ang Ö (1965) Encapsulated *Pseudomonas aeruginosa* (*Pseudomonas aeruginosa mucosus*) strains. *J Bacteriol* 89:1432–1433
- Cox EE, Major D, Edwards E (2000) Natural attenuation of 1,2-dichloroethane in groundwater at a chemical manufacturing facility. In: Wickramanayake GB, Gavaskar AR, Alleman BC, Magar VS (eds) *The Second International Conference on Remediation of Chlorinated and Recalcitrant Compounds*. Batelle, Monterey, California
- Criddle CS, Dewitt JT, D. G-G, McCarty PL (1990) Transformation of carbon tetrachloride by *Pseudomonas* sp. strain KC under denitrification conditions. *Appl Environ Microbiol* 56:3240–3246
- Dijk JA, de Bont JAM, Lu X, Becker PM, Bosma TNP, Rijnaarts HHM, Gerritse J (2000) Anaerobic oxidation of (chlorinated) hydrocarbons. In: Wickramanayake GB, Gavaskar AR, Alleman BC, Magar VS (eds) *The Second International Conference on Remediation of Chlorinated and Recalcitrant Compounds*. Batelle, Monterey, California
- Duarte GF, Soares Rosado A, Seldin L, Keijzer-Wolters AC, van Elsas JD (1998) Extraction of ribosomal RNA and genomic DNA from soil for studying the diversity of the indigenous bacterial community. *J Microbiol Meth* 32:21–29
- Gerritse J, van der Woude BJ, Gottschal JC (1992) Specific removal of chlorine from the ortho-position of halogenated benzoic acids by reductive dechlorination in anaerobic enrichment cultures. *FEMS Microbiol Lett* 79:273–280
- Gerritse J, Borger A, van Heiningen E, Rijnaarts HHM, Bosma TNP, Taat J, van Winden B, Dijk JA, de Bont JAM (1999) Assessment and monitoring of 1,2-dichloroethane dechlorination. In: Leeson A, Alleman BC (eds) *Engineered approaches for in situ bioremediation of chlorinated solvent contamination*. Batelle, Columbus, Ohio, pp 73–80
- Hage JC, Hartmans S (1999) Monooxygenase-mediated 1,2-dichloroethane degradation by *Pseudomonas* sp. strain DCA1. *Appl Environ Microbiol* 65:2466–2470
- Hantke J (2000) Molekularbiologische Analyse der mikrobiellen Zusammensetzung anaerober, Dioxine dechlorierender Mischkulturen. PhD thesis, Martin-Luther University Halle.
- Heijthuisen JHFG, Hansen TA (1986) Interspecies hydrogen transfer in co-cultures of methanol utilizing acidogens and sulfate-reducing or methanogenic bacteria. *FEMS Microbiol Ecol* 38:57–64
- Heuer H, Hartung K, Wieland G, Kramer I, Smalla K (1999) Polynucleotide probes that target a hypervariable region of 16S rRNA genes to identify bacterial isolates corresponding to bands of community fingerprints. *Appl Environ Microbiol* 65:1045–1049
- Janssen DB, Scheper A, Witholt B (1984) Biodegradation of 2-chloroethanol and 1,2-dichloroethane by pure bacterial cultures. In: Houwink EH, van der Meer RR (eds) *Innovations in biotechnology. Progress in industrial microbiology*, vol 20. Elsevier Biomedical, Amsterdam, pp 169–178
- Janssen DB, Scheper A, Dijkhuizen L, Witholt B (1985) Degradation of halogenated aliphatic compounds by *Xanthobacter autotrophicus* GJ10. *Appl Environ Microbiol* 49:673–677
- Kodoma T, Shimada K, Mori T (1969) Studies on anaerobic biphasic growth of a denitrifying bacterium, *Pseudomonas stutzeri*. *Plant Cell Physiol* 10:855–865
- Körner H, Zumft WG (1989) Expression of denitrification enzymes in response to the dissolved oxygen level and respiratory substrate in continuous culture of *Pseudomonas stutzeri*. *Appl Environ Microbiol* 55:1670–1676
- Lovley DR, Phillips EJP (1986) Organic matter mineralization with reduction of ferric iron in anaerobic sediments. *Appl Environ Microbiol* 51:683–689
- Lovley DR, Phillips EJP (1988) Novel mode of microbial energy metabolism: organic carbon oxidation coupled to dissimilatory reduction of iron or manganese. *Appl Environ Microbiol* 54:1472–1480
- Lowry OH, Rosebrough NJ, Farr AL, Randall RJ (1951) Protein measurements with the Folin phenol reagent. *J Biol Chem* 193:265–275
- McCann J, Simmon V, Streitwieser D, Ames BN (1975) Mutagenicity of chloroacetaldehyde, a possible metabolic product of 1,2-dichloroethane (ethylene chloride), chloroethanol (ethylene chlorohydrine), vinyl chloride and cyclophosphamide. *Proc Natl Acad Sci USA* 72:3190–3192
- Middelhoven WJ, Bakker CM (1982) Degradation of caffeine by immobilized cells of *Pseudomonas putida* strain C 3024. *European J Appl Microbiol Biotechnol* 15:214–217
- Murray JW (1974) The surface chemistry of hydrous manganese oxide. *J Colloid Interface Sci* 46:357–371
- Muyzer G, Teske A, Wirsén CO, Jannasch HW (1995) Phylogenetic relationship of *Thiomicrospira* species and their identification in deep-sea hydrothermal vent samples by denaturing gradient gel electrophoresis of 16S rDNA fragments. *Arch Microbiol* 164:165–172

- Palleroni NJ, Doudoroff M, Stanier RY, Solánes RE, Mandel M (1970) Taxonomy of the aerobic pseudomonads, the properties of the *Pseudomonas stutzeri* group. *J Gen Microbiol* 60:215–231
- Pearson WR (1990) Rapid and sensitive sequence comparison with FASTAP and FASTA. *Methods Enzymol* 183:63–98
- Poelarends GJ, van Hylckama Vlieg JET, Marchesi JR, Freitas Dos Santos LM, Janssen DB (1999) Degradation of 1,2-dibromoethane by *Mycobacterium* sp. strain GP1. *J Bacteriol* 181:2050–2058
- Ralebitso TK, Röling WFM, Braster M, Senior E, van Verselvel HW (2001) 16S rDNA-based characterisation of BTX-catabolizing microbial associations isolated from a South African sandy soil. *Biodegradation* 11:351–357
- Richterich R (1965) *Klinische Chemie*. Akademische Verlagsgesellschaft, Frankfurt
- Sambrook J, Fritsch EF, Maniatis T (1989) *Molecular cloning: a laboratory manual*, 2nd edn. Cold Spring Harbor Laboratory, Cold Spring Harbor, New York
- Sijderius R. (1946). PhD thesis, University of Amsterdam
- Skerman VBD, McGowan V, Sneath PHA (1980) Approved lists of bacterial names. *Int J Syst Bacteriol* 30:225–420
- Strotmann UJ, Pentenga M, Janssen DB (1990) Degradation of 2-chloroethanol by wild type and mutants of *Pseudomonas putida* US2. *Arch Microbiol* 154:294–300
- Stucki G, Brunner W, Staub D, Leisinger T (1981) Microbial degradation of chlorinated C1 and C2 hydrocarbons. In: Leisinger T, Cook AM, Hütter R, Nüesch J (eds) *Microbial degradation of xenobiotics and recalcitrant compounds*. Academic, London, pp 131–137
- Van den Wijngaard AJ, van der Kamp KWHJ, van der Ploeg J, Pries F, Kazemier B, Janssen DB (1992) Degradation of 1,2-dichloroethane by *Ancylobacter aquaticus* and other facultative methylotrophs. *Appl Environ Microbiol* 58:976–983
- Van der Meer JR, Roelofsen W, Schraa G, Zehnder JB (1987) Degradation of low concentrations of dichlorobenzenes and 1,2,4-trichlorobenzene by *Pseudomonas* sp. strain P51 in nonsterile soil columns. *FEMS Microbiol Ecol* 45:333–341
- Weisburg WG, Barns SM, Pelletier DA, Lane DJ (1991) 16S ribosomal DNA amplification for phylogenetic study. *J Bacteriol* 173:697–703
- Xu B, Enfors SO (1996) Influence of nitrate starvation on nitrite accumulation during denitrification by *Pseudomonas stutzeri*. *Appl Microbiol Biotechnol* 45:229–235