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Discovering new living *Pinna nobilis* populations in the Sea of Marmara

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Abstract

Fan mussel (*Pinna nobilis*) is one of the largest bivalve species in the Mediterranean Sea. The situation of the species is critical as it faces widespread mass mortality attributed to pathogens in various parts of the Mediterranean Sea. The Sea of Marmara (SoM) offers a unique environment for fan mussel populations, with some areas hosting alive populations. This study aims to explore and describe new *P. nobilis* populations in the SoM that are known to be not affected by mortality. An area of 28,200 m² at 47 stations along the 105 km coastline in the southern part of the SoM was explored using underwater visual transects. A total of 544 alive fan mussels were recorded during the underwater surveys, ranging in total shell height from 11.8 to 31.4 cm. The mean density was estimated as 5.3 ind 100 m⁻² although maxima of 18.8 ind 100 m⁻² were recorded in some stations. These density hotspots were distributed from the shoreline to a 10 m depth range and 100 m distance from the shoreline in sandy and seagrass meadow habitats. The presence of juveniles provided evidence of successful recruitment. The distribution pattern and recorded mortalities were attributed to hydrodynamic factors and intense human activities. Potential environmental factors (low salinity and temperature) in the SoM may control or delay the possible spread of the lethal pathogens. Favorable conditions result in mussels' resilience and survival mechanisms. The SoM offer a promising larval reservoir for the recolonization of affected areas, such as those found in the Aegean Sea, through larval exportation.

Keywords Population density \cdot Demographic structure \cdot Spatial distribution \cdot Fan mussel \cdot Critically endangered species \cdot Sea of Marmara

Introduction

Fan mussel (*Pinna nobilis* Linnaeus, 1758), belongs to the family Pinnidae and is one of the largest bivalve species in the Mediterranean Sea (Vicente 1990). This species native to the Mediterranean Sea has a wide distribution from Spain to the Sea of Marmara (SoM) (Butler et al. 1993; Vázquez-Luis et al. 2017a). Although it grows quite fast during its first years of life (Kersting and García-March 2017), its growth

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slows down in older ages. Individuals can reach an impressive shell height of 120 cm and live up to 50 years (Zavodnik et al. 1991; Rouanet et al. 2015). Fan mussels exhibit a diverse distribution, thriving in various habitats such as seagrass meadows, sandy, gravelly, and rocky environments, as well as in rhodolith beds (Katsanevakis 2006; Basso et al. 2015; Kersting and García-March 2017; Karadurmuş and Sarı 2022a). Abundant along the coastal zone, fan mussels are found at depths of up to 60 m (Kersting and García-March 2017). They prefer areas with good environmental factors and gentle currents (Vázquez-Luis et al. 2017b; Prado et al. 2021). An essential ecological role of the fan mussel is its filtering capacity, with each individual capable of processing estimated 6 liters of seawater per hour (Trigos et al. 2014), effectively removing suspended particles and contributing to the clarification of the surrounding environment (Basso et al. 2015). Additionally, the fan mussel provides an attachment surface for sessile organisms, promoting biodiversity (Rabaoui et al. 2009). The diet mainly consists

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of detritus, phytoplankton, zooplankton, and pollen dust (Davenport et al. 2011; Alomar et al. 2015).

A destructive and geographically widespread mass mortality event in the Mediterranean basin has affected P. nobilis populations since 2016 (Vázquez-Luis et al. 2017a; Kersting et al. 2019; Katsanevakis et al. 2021). The first mortalities were caused by the protozoan pathogen Haplosporidium pinnae (Catanese et al. 2018; Vázquez-Luis et al. 2017a), with further studies revealing the impact of bacteria such as Mycobacterium sp. or Vibrio sp. on mortality (Catanese et al. 2018; Künili et al. 2021; Grau et al. 2022; Carella et al. 2023). The infection-associated deaths rapidly spread from west to east, starting in Spain and reaching the eastern Mediterranean Sea within a span of fewer than three years (Vázquez-Luis et al. 2017a; Kersting et al. 2019; Katsanevakis et al. 2021). As far as current knowledge goes, the infection extent has reached the Dardanelles Strait (Özalp and Kersting 2020; Künili et al. 2021), representing a significant geographic spread. Mortality rates have reached 80-100% across the Mediterranean Sea in most places, including Dardanelles Strait (Kersting et al. 2019; Özalp and Kersting 2020; Katsanevakis et al. 2021; Scarpa et al. 2021). While some populations in the Mediterranean Sea have managed to remain pathogen-free, they are situated in geographically isolated and environmentally specific regions (Cabanellas-Reboredo et al. 2019; Kersting et al. 2019; Katsanevakis et al. 2021). It is important to note that although extremely high salinity levels in these regions act as a deterrent to the spread and survival of the pathogen, they are not immune to anthropogenic pressure, including factors such as eutrophication and anoxia (Cabanellas-Reboredo et al. 2019; Giménez-Casalduero et al. 2020). Local-scale threats such as habitat loss, tourism, anchoring, illegal fishing, coastal constructions, and several factors like poor environmental conditions, eutrophication, and anoxia primarily contribute to deaths in certain areas (Basso et al. 2015; Giménez-Casalduero et al. 2020; Öndes et al. 2020a; Karadurmuş and Sarı 2022a). The World Union for Conservation of Nature (IUCN) has categorized the *P. nobilis* as Critically Endangered (Kersting et al. 2019) due to the severe mass mortality and significant reduction in stock sizes. As recommended by the IUCN, high-density regions should be protected, and containment measures include any impact that could cause incidental deaths (Kersting et al. 2019). Due to the high risk of extinction and the sharp decline in its global population, critical populations are drastically protected within various international and national directives (Kersting et al. 2019). Türkiye first provided protection to this species in 1997, and it remains on the list of forbidden species in the country (Karadurmuş and Sarı 2022a).

The existence of high-density alive *P. nobilis* populations in the SoM (Öndes et al. 2020a; Çınar et al. 2021a, b; Acarlı et al. 2022; Karadurmuş and Sarı 2022a) amidst mass mortalities elsewhere in the Mediterranean Sea, is of paramount importance for the natural recovery and survival of the species. These healthy populations serve as crucial larval exporting areas, offering a potential source of larvae for recolonization in mortality-affected sites, such as those in the Aegean Sea (Kersting et al. 2020; Papadakis et al. 2023). This study aims to describe new fan mussel populations that are not affected by widespread mass extinctions in the SoM and to examine their ecological aspects with a holistic approach (spatial and bathymetric distribution, abundance, mortality, population status). This study sheds light on the factors contributing to their resilience and survival of *P. nobilis* in the face of mass mortality events, providing valuable insights into potential conservation strategies for other affected areas.

Materials and methods

The study area covers Bandırma Bay and a large part of Kapıdağ Peninsula located in the south of the SoM (Fig. 1). In the study area, diving based on underwater observation was made at randomly determined stations at every 5 km intervals along the 105 km coastline. Additional dives were made in areas where critical density changes were observed. Underwater visual census includes techniques based on the assessment of habitat uses of target species and the determination of the abundance of target species within the transect area (Colvocoresses and Acosta 2007; Caldwell et al. 2016). These methods allow to calculate densities per unit area. Fan mussels were counted using the strip transect method in this study. This method gives more precise results since the boundaries are determined with precise strips and is frequently used in the estimation of fan mussel studies (García-March et al. 2002; Vafidis et al. 2014; Tsatiris et al. 2018; Karadurmuş and Sarı 2022a). Lead rope with a diameter of 3 mm and a length of 150 m, marked at intervals of five meters, was used as the transect material. The line was laid bottom vertically to the shore, starting from the shoreline. Two divers on either side of the strip used a 2 m wide lead made of plastic pipe with a center mark in their hands. The divers counted along the remaining strip in the transect area. In this way, two divers at each station scanned an area of $600 \text{ m}^2 (150 \text{ m} \times 4 \text{ m})$ in total. Divers recorded the number of individuals, the specimen status (dead or alive), depth, distance from shore, size of the individuals and the relative habitat type between the two marks by means of an underwater board for each segment.

During the census, the divers analyzed the living status of the individuals and recorded the alive and dead mussels counted in each subsection. Alive mussels respond by quickly closing their shell on approach or touch, while dead mussels contain no tissue and are unresponsive, their



Fig. 1 Map of study area (Southern part of the Sea of Marmara) showing the surveyed stations selected for the assessment of the status of *Pinna* nobilis populations

shells open and not closing. In addition, buried mussels with cracked or broken shells were defined as dead. Individuals completely plucked from sediment were not included in the count as they may have been carried indiscriminately by the flows. Recordings were taken throughout the entire field, including areas absent of fan mussels. The habitat type in each segment was classified and recorded as sandy, seagrass meadow, gravelly, rocky, shellfish, and muddy (IUCN 2023). Biometric measurements were made with a long-jaw multicaliper and depth caliper with 0.1 cm division in randomly selected individuals along the transect. Measurements were made by taking the maximum width (W), minimum width (w), and maximum length (UL) measured from the unburied part of the shell. Based on these measurements; maximum shell length (Ht) was estimated (García-March et al. 2002). The Ht was calculated using the equation Ht = UL + h, and the regression (h) 1.79w + 0.5 was used for the buried shell length.

The mean fan mussel density at the stations was estimated as the individual number in per hundred square meters (ind 100 m⁻²). Density statuses were classified as low (<5 ind 100 m⁻²), medium (5–10 ind 100 m⁻²), high (10–15 ind 100 m⁻²), and very high (>15 ind 100 m⁻²) (Rabaoui et al. 2008; Karadurmuş and Sarı 2022a). All analyzes were performed separately for alive and dead individuals. The accumulated percentage of mortality (%) was calculated from the ratio of the number of dead individuals to the total number of individuals at each station. The variation of all density data according to depth, distance from the shore and habitat type was examined in the examined segments. Individuals with a shell length less than 20 cm were defined as juveniles, and those larger than 20 cm as adults (Richardson et al. 1999).

The data sets to be used in the study were examined in terms of suitability before the analysis. The assumptions of normality of variables and homogeneity of variances were evaluated with the Shapiro–Wilk test (Shapiro and Wilk 1965) and Levene's test (Levene 1960). In the comparison of the quantitative data between two independent groups, parametric tests were used if the assumptions were significant, and non-parametric tests were used if the assumptions were significant. The significance level for the test statistics of the hypotheses in both the assumption controls and parameter estimations was determined as $\alpha = 0.05$ (Sokal and Rohlf 1969). Descriptive statistics (mean, standard error, quartiles, percentiles) were obtained according to the

distribution structure for all data obtained. Data analysis and modeling was performed with SPSS v0.26.

Results

A total area of 28,200 m² was scanned across 47 stations, revealing the presence of fan mussel individuals in only 17 of these stations. In total, 789 individuals were recorded, providing valuable insights into the distribution and abundance of fan mussels within the surveyed area. Among the recorded individuals, 544 (68.9%) were found to be alive, while 245 (31.1%) were observed in dead state. The mean alive mussel density was calculated as 5.3 ind 100 m^{-2} in the study area and five stations (S37, S44, S45, S46, and S47) were represented with above-average density. S45, located on the western side of the peninsula, exhibited the maximum mussel density with 18.8 ind 100 m⁻². The total area covered by areas with fan mussels accounted for only 11.3% of the scanned area. At two specific stations (S37 and S45), very high-density clustering of fan mussels was recorded, while most of the remaining stations were classified as having low densities. Stations (S9, S12, S14, and S16) located in the eastern portion of the study area were represented by a very limited number of individuals (one or two individuals in each station) (Table 1).

This study revealed distinct patterns of spatial distribution and habitat preferences for *P. nobilis*. Distribution patterns indicate that the fan mussel can be found in various habitats along the scanned area, encompassing both shallow and deeper waters (down to 30.1 m). The 5–10 m depth contours had the highest alive density with 8.3 ind 100 m⁻², followed by the 0–5 m depth contours with 6.1 ind 100 m⁻². Dense clustering was recorded in populations reaching 9.9 ind 100 m⁻² at a distance range of 50–75 m from the shore. Alive mussels most preferred sandy habitats with an average density of 13.6 ind 100 m⁻², followed by seagrass habitats with an average density of 7.3 ind 100 m⁻². The density of fan mussel was generally low in rocky, gravelly and shellfish habitats in this study. In contrast, *P. nobilis* was completely absent in muddy areas (Table 2).

The mean H_t was estimated as 20.1 ± 0.35 cm (11.8–31.4 cm) in alive and 23.2 ± 0.42 cm (11.8–32.5 cm) in dead individuals (Fig. 2). The ratio of juveniles to adults in the alive population was found to be evenly distributed (Fig. 3), with both comprising 50% of the population. A statistically significant difference in mean H_t was observed between alive and dead individuals ($t_{(357)} = -5.737$, p < 0.05). Accumulated mortality rates were recorded as 25.9% and 50.9% in juvenile and adult individuals, respectively.

Discussion

The study revealed the presence of thriving P. nobilis populations in the SoM, with evidence of successful recruitment and resilience to lethal pathogens. Despite the widespread mortality events in the Mediterranean Sea, the SoM continues to host alive individuals. Studies encompassing diverse regions, with a particular emphasis on the southern area of the SoM, revealed notable population densities: from 4.9 to 47.0 ind 100 m⁻² in Acarlı et al. (2022), up to 71.2 ind 100 m⁻² [alive] in Karadurmuş and Sarı (2022a), from 10 to 112 ind 100 m⁻² [both alive and dead] in Acarlı et al. (2021a), from 6 to 240 ind 100 m⁻² [alive] in Çınar et al. (2021a), and from 0.3 to 12 ind 100 m^{-2} in Cinar et al. (2021b), with a mean 24 ind 100 m^{-2} [alive] in Öndes et al. (2020b). Özalp and Kersting (2020) revealed the occurrence of mussel populations at high densities of up to 9 ind m^{-2} in shallow waters in the Dardanelles Strait, which connects the SoM with the Mediterranean Sea in the south. The patchy distribution pattern of the fan mussel in the study area implies localized concentrations rather than a homogenous distribution across the study area. The salinity difference between the Mediterranean Sea and the Black Sea causes water column stratification in the Dardanelles and Istanbul straits and the SoM. This stratification shapes the current system, with fresh Black Sea waters flowing on the surface towards the SoM and the Aegean Sea, while more saline Mediterranean Sea waters current towards the Black Sea at the bottom. The SoM exhibits lower temperature and salinity levels when contrasted with the Mediterranean Sea (Meric et al. 2018). These environmental factors may play a role in regulating or impeding the potential spread of the pathogen. creating a natural sanctuary (Prado et al. 2021; Papadakis et al. 2023). Additionally, meteorological factors like rainfall, wind speed, and direction can influence the currents, potentially causing local deviations in the main current systems in the SoM (Beşiktepe et al. 1994; Meriç et al. 2018). The water current characteristic of the northern shores of the peninsula (from S18 to S36) and the eastern side of the bay (from S1 to S3) are clearly different from the other stations in the study area. This area is greatly affected by bottom currents from the Mediterranean and surface currents from the Black Sea (Fig. 1). So, the deprivation of populations at these stations can be associated with strong water currents. Because high currents in certain areas may create turbulent conditions and strong water flows that make it challenging for the mussels to maintain their attachment to the substrate. Hydrodynamic processes, including coastal currents, upwelling, and eddies, play a crucial role in the longdistance dispersal and connectivity between different populations of fan mussels (Kersting and García-March 2017; Kersting et al. 2020). These currents may also facilitate

Table 1 Details on the location of surveyed stations and the status of Pinna nobilis populations in the study area

Station	Coordinate		Count (<i>n</i>)		Density (ind·100 m ⁻²)		Density status [*]	Mortal-
	Latitude	Longitude	Alive	Dead	Alive	Dead		ity rate (%)
S1	40.38686	28.12038	0	0	_	_	n/a	_
S2	40.37582	28.05474	0	0	_	_	n/a	_
S 3	40.37425	28.05044	0	0	_	_	n/a	_
S4	40.36149	27.98236	0	0	_	_	n/a	_
S5	40.35705	27.96824	0	0	_	_	n/a	_
S6	40.36185	27.93083	0	0	_	_	n/a	_
S 7	40.37523	27.91672	0	0	_	_	n/a	_
S 8	40.38542	27.89952	0	0	_	_	n/a	_
S9	40.39847	27.90824	2	0	0.3	_	low	0
S10	40.40741	27.92021	0	0	-	_	n/a	_
S11	40.40865	27.92348	0	0	_	_	n/a	_
S12	40.41181	27.93150	1	2	0.2	0.3	low	66.7
S13	40.42306	27.95409	0	0	_	_	n/a	_
S14	40.42433	27.96836	2	1	0.3	0.2	Low	33.3
S15	40.43972	27.99798	9	1	1.5	0.2	Low	10.0
S16	40.45417	28.02427	2	1	0.3	0.2	Low	33.3
S17	40.45926	28.02315	0	0	_	_	n/a	_
S18	40.46991	28.02912	15	4	2.5	0.7	Low	21.1
S19	40.47558	28.03070	0	0	_	_	n/a	_
S20	40.47920	28.02766	0	0	_	_	n/a	_
S21	40.48652	28.01089	0	0	_	_	n/a	_
S22	40.48691	27.99423	0	0	_	_	n/a	_
s23	40.48864	27,99000	0	0	_	_	n/a	_
S24	40.49167	27.97581	0	0	_	_	n/a	_
S25	40.49251	27.96547	0	0	_	_	n/a	_
S26	40.49494	27.96095	0	0	_	_	n/a	_
S27	40,49693	27.95068	0	0	_	_	n/a	_
S28	40.50659	27.90369	0	0	_	_	n/a	_
S29	40.50707	27.90100	0	0	_	_	n/a	_
S30	40 51256	27.87456	0	0	_	_	n/a	_
S31	40.51250	27.87430	0	0	_	_	n/a	_
\$32	40.51023	27.83627	0	0	_	_	n/a	_
S33	40.51337	27.80522	0	0	_	_	n/a	_
\$34	40.51822	27.80322	0	0	_	_	n/a	_
S35	40.52132	27.80380	0	0	_	_	n/a	
S36	40.51508	27.30237	0	0	_	_	n/a	
\$37	40.51508	27.79612	0	7	15.2	1.2	Wery high	- 7 1
539	40.50001	27.78014	91 12	5	2.0	0.8	Low	20.4
\$30	40.51751	27.74318	22	10	5.3	0.8	Average	29.4
S39 S40	40.52004	27.73400	32	19	3.5	3.2 2.2	Average	28.0
540 641	40.51941	27.72379	22	14	3.7	2.3	Low	20.4
541 642	40.51739	27.71219	20	15	3.3	2.2	Low	39.4 40.0
542 542	40.31/04	21.10/10	15	У 7	2.2	1.5	Low	40.9
343 844	40.50965	21.09803	15	/	2.5	1.2	Low	31.8 47.2
544 545	40.50093	27.08438	30 112	5U 02	9.3	ð.3	Average	47.2
545	40.49071	27.08/19	113	92	18.8	15.3	very nigh	44.9
S46	40.45622	28.07348	52	8	8.7	1.3	Average	13.3
S47	40.45225	28.08595	87	12	14.5	2.0	Hıgh	12.1

^an/a: none, low: <5 ind 100 m⁻², average: 5–10 ind 100 m⁻², high: 10–15 ind 100 m⁻², very high: \geq 15 ind 100 m⁻². Note: A total area of 600 m² was surveyed at each station

or prevent the dispersal and settlement of larvae (Kersting et al. 2020), thereby affecting the distribution of the species. Various researchers (Hendriks et al. 2011; Prado et al. 2021) have reported that fan mussel populations are vulnerable to severe weather events. As a defense mechanism, the mussels tend to avoid such areas and seek refuge in more sheltered habitats (Hendriks et al. 2011). The absence of individuals in areas with high currents is likely a result of the combined factors mentioned above.

The interior of Bandırma Bay is subjected to significant coastal uses and multiple industrial activities. The Bay is home to two heavy industry facilities, an international multipurpose port with large berths, and a fishing port. The bay is also exposed to domestic pollution of the highly populated district through the deep discharge system (Özen et al. 2023). Heavy industry activities can introduce pollutants such as chemicals, heavy metals, and toxins into the water, which can have detrimental effects on the fan mussels' health and survival. Most sites that still holding mussel populations are in highly anthropized areas, and a similar problem occurs at sites such as the Mar Menor in Spain (Giménez-Casalduero et al. 2020) and Gulf of Erdek in the SoM (Karadurmuş and Sarı 2022a). Additionally, domestic pollution can lead to eutrophication, oxygen depletion, and the accumulation of pollutants in the water, making it unsuitable for the fan mussels (Basso et al. 2015; Giménez-Casalduero et al. 2020). In addition, secondary bottom current and local surface current, which have a cyclical effect in the inner parts of the bay, prevent the pollution from leaving the bay and contribute to the continuity of the pollution pressure (Fig. 1). The bay also hosts the high-speed ferry route and busy anchorage area. Marine operations often involve negative impacts, which generate significant disturbances in the water and along the seabed (Öndes et al. 2020a). The absence or limited number of fan mussels at stations in this region, from S4 to S8, and in the bay is thought to be a result of the above-mentioned multi-faceted intense activities.

The western portion of the study area, spanning from S37 to S45, along with Fener Island (S46) and Hali Ada (S47) on the eastern side, both of which are isolated from the peninsula, exhibited fan mussel populations characterized by exceptionally high densities. Contrary to Acarlı et al. (2022) new discoveries of one or two individuals were obtained at the stations scanned in the area between S9 and S18 in Bandırma Bay. Previous research (Acarlı et al. 2022; Karadurmuş and Sarı 2022a) has also documented the presence of thriving populations in Erdek Bay, situated to the west of the peninsula, which further corroborates the region's ability to support survival and settlement populations of live fan mussels. These areas boast more stable environmental conditions and exhibit lower water velocities, factors that are conducive to the successful survival and settlement of fan mussels. We hypothesize that salinity

and temperature are potential factors behind the survival of populations in the SoM, as opposed to the Mediterranean Sea. The salinity of the SoM varies between 17.8 and 38.5 psu throughout the year and is characterized by an annual average of 22.5 psu. Cabanellas-Reboredo et al. (2019) reported that the onset of pathogen has been associated with temperatures above 13.5 °C and a narrow salinity range (36.5–39.7 psu). The fact that the SoM has a relatively lower temperature and salinity compared to the Mediterranean Sea may control or delay the possible spread of the pathogen. Notably, these regions remain entirely free from anthropogenic influences, preserving a sheltered and protected environment that further supports the proliferation of fan mussels. Understanding the complex relationships between environmental features and the distribution of fan mussels is crucial for effective conservation and management strategies. Further research in the study area should focus on understanding the multiple factors (environmental, physical, chemical, hydrodynamic, anthropogenic) that drive distribution patterns and explore additional potential areas that may contribute to the conservation of this species.

The distribution pattern of populations in the study area was mainly connected to depth and habitat type. Individuals were mostly concentrated in the shallow zone between the shoreline and 10 m, and the density decreased as depth increased. Similarly, areas of sandy bottom and seagrass meadow were also represented by a significant density of alive mussels. Habitats with high P. nobilis concentrations are very diverse and can be found in almost all habitat types down to depths of 40-50 m (Kersting and García-March 2017). Normally, seagrass areas are the primary habitat for fan mussel (Hendriks et al. 2011; Vázquez-Luis et al. 2014; Karadurmuş and Sarı 2022a). However, light transmittance decreases due to the high suspended solid load in the SoM, and, therefore, seagrass meadows can spread up to a maximum depth of 6-7 m (Cirik et al. 2006). Therefore, sandy habitats are the primary habitat for P. nobilis in the study area. Such depth (Siletić and Peharda 2003; García-March et al. 2007a; Karadurmuş and Sarı 2022a) and habitat (Katsanevakis and Thessalou-Legaki 2009; Hendriks et al. 2011; Deudero et al. 2015; Acarlı et al. 2022; Karadurmuş and Sarı 2022a) related diversities have an important impact on fan mussels' survival and recruitment. Key environmental factors such as hydrodynamics, light intensity, temperature, and productivity in shallow waters strongly support the existence of the species (Prado et al. 2014; Russo 2017; Tsatiris et al. 2018). Results indicating that sandy grounds and seagrass beds are the primary habitats where the species thrives. Sandy bottoms provide the most convenient environment for the implantation of byssus filaments (García-March et al. 2007b), while seagrass meadows ensure shelter against hydrodynamic factors (García-March and Kersting 2006; Hendriks et al. 2011). These habitats offer protection, food

	Variables	Count (<i>n</i>)		Density (ind• 100 m ⁻²)	
		Alive	Dead	Alive	Dead
Depth range (m)	0–5	246	98	6.1	2.4
	5-10	212	96	8.3	3.8
	10-15	64	26	3.3	1.3
	15-20	12	20	1.2	1.9
	20-25	8	2	2.5	0.6
	25-30	1	1	0.7	0.7
	30–35	1	2	1.0	2.0
	35-40	0	0	0	0
Distance from	0–25	25	43	1.8	3.2
shore (m)	25-50	105	46	6.2	2.7
	50-75	169	67	9.9	3.9
	75-100	128	34	7.5	2.0
	100-125	89	27	5.2	1.6
	125-150	26	26	1.5	1.5
	150-175	2	2	0.6	0.6
Habitat type	Sandy	144	120	13.6	11.3
	Seagrass	364	92	7.3	1.8
	Gravelly	15	12	1.5	1.2
	Rocky	13	16	1.6	2.0
	Shellfish	8	5	0.4	0.3
	Muddy	0	0	0	0
Total		544	245	5.3	2.4

Table 2 *Pinna nobilis* densities (ind 100 m^{-2}) in different depth, distance from shore and habitats



Fig. 2 Boxplot of the size $(H_t, \text{ cm})$ for the alive and dead *Pinna nobilis* in the study area

resources, and suitable substrate for attachment, making them important habitats for the species (Vázquez-Luis et al. 2014). Muddy substrates seem to be unsuitable for attachment and may not provide the necessary conditions for the survival and growth of the species (Katsanevakis 2006; Tsatiris et al. 2018).

Most of the current *P. nobilis* population were characterized by a predominance of juveniles across all areas studied. This situation often implies successful recruitment and reproduction and can generally indicate a stable population, good environmental conditions, and the absence of predation (Kersting and García-March 2017; Vázquez-Luis et al. 2017b). Kersting and García-March (2017) reported a scarcity of individuals below 45 cm in H_{t} in the population of the Columbretes Islands (NW Mediterranean). They emphasized that such a distribution model indicates that it can increase the survival rate and that this size may be a refuge size. On the other hand, the recorded mortalities are alarming, and as stated in previous studies (Öndes et al. 2020b; Cınar et al. 2021a, b; Acarlı et al. 2021a, 2022; Karadurmuş and Sarı 2022a), mortalities in *P. nobilis* have been recorded in the SoM due to multiple factors such as fisheries, anchoring, tourism, and diving activities. Although it is well known that environmental factors and human effects cause mortalities, it is still unknown whether pathogens, parasites, or other disease-causing agents are responsible for deaths in the SoM (Karadurmuş and Sarı 2022a; Papadakis et al. 2023). The size distribution revealed that larger size older individuals were represented in fewer numbers in the population (Fig. 3). Adults in the large size group were more affected by mortalities, thus leading to slow population dynamics and to the species' vulnerability to unexpected events. Juveniles may indicate inclusion in the productive stock, and rapid growth in the first years of life may help shorten the period of vulnerability of the stock (Kersting and García-March 2017). Juveniles may exhibit a high survival rate due to their superior responsiveness to pathogens (Sarić et al. 2020). The presence of pathogens that cause mass deaths in the Mediterranean is not yet known in the study area. As far as we know best, the last spread of the disease is limited to the Dardanelles Strait (Ozalp and Kersting 2020; Künili et al. 2021). In the most recent occurrence, a significant marine mucilage outbreak took place in the SoM, spanning from November 2020 to August 2021. Consequently, numerous benthic species within marine communities experienced extensive mass mortality (Karadurmuş and Sari 2022b; Karakulak et al. 2023). This phenomenon proved fatal to pelagic fish and crustaceans, with individuals being severely affected by suffocation due to the presence of mucilage or anoxia (Karadurmuş and Sarı 2022b). This devastating event may cover the benthic habitats where bivalves like P. nobilis reside, potentially smothering them and reducing their access to food and oxygen. The phenomena can alter the availability and distribution of planktonic food sources for bivalves, which might affect their feeding and distribution patterns. Acarlı et al. (2021b) claimed that the mucilage event in the SoM may have a negative impact on fan mussel populations either directly or through habitat loss (especially seagrass meadows). There is no conclusive evidence that mucilage is responsible for fan mussel deaths, although various potential effects have been predicted.

Fig. 3 Histogram graph of the size (H_t , cm) distribution of alive and dead *Pinna nobilis* in the study area. Individuals less than 20 cm H_t represent juveniles, individuals larger than 20 cm H_t represent adults



Conclusion

This study demonstrates the discovery of new P. nobilis populations in addition to previous studies in the south of the SoM. Findings present a unique opportunity to investigate the factors contributing to the mussel's resilience and survival capacity in the face of widespread extinction events. Despite the concerning situation observed in many areas, the SoM continues to serve as a vital refuge for fan mussel populations, with certain regions supporting thriving and high-density communities. The SoM can play a vital role in facilitating natural recoveries in sites affected by mortality events through larval export (Kersting et al. 2020; Papadakis et al. 2023). The presence of juveniles indicates successful recruitment and settlement, but the higher mortality rates among adults raise significant concerns. The fact that the SoM is characterized by lower temperature and salinity than the Mediterranean appears to be the main factors preventing the entry of the pathogen, but further research is still needed. The lessons learned from the survival of these populations could provide valuable insights into guide conservation strategies aimed at rejuvenating the species in its former Mediterranean Sea habitats. To ensure the conservation and survival of P. nobilis, continued monitoring and in-depth studies on the species are imperative, including investigations into potential threats from pathogens and anthropogenic activities. While this study sheds light on the southern part of the SoM, it is crucial to conduct additional research covering the entire of the SoM to gain a comprehensive understanding of fan mussel populations in the region. Immediate investigation into the presence of pathogens causing mortality is strongly recommended to address potential threats to the species' survival. By undertaking such efforts, we can enhance our ability to safeguard this ecologically significant species and contribute to the preservation of biodiversity in the Mediterranean Sea.

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Data availability Data will be made available on request.

Declarations

Conflict of interest The authors have no relevant financial or non-financial interests to disclose.

Ethical approval This is an observational study. No ethical approval is required.

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