#### **REVIEW, CONCEPT, AND SYNTHESIS**



# **A review of a decade of lessons from one of the world's largest MPAs: conservation gains and key challenges**

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#### **Abstract**

Given the recent trend towards establishing very large marine protected areas (MPAs) and the high potential of these to contribute to global conservation targets, we review outcomes of the last decade of marine conservation research in the British Indian Ocean Territory (BIOT), one of the largest MPAs in the world. The BIOT MPA consists of the atolls of the Chagos Archipelago, interspersed with and surrounded by deep oceanic waters. Islands around the atoll rims serve as nesting grounds for sea birds. Extensive and diverse shallow and mesophotic reef habitats provide essential habitat and feeding grounds for all marine life, and the absence of local human impacts may improve recovery after coral bleaching events. Census data have shown recent increases in the abundance of sea turtles, high numbers of nesting seabirds and high fish abundance, at least some of which is linked to the lack of recent harvesting. For example, across the archipelago the annual number of green turtle clutches (*Chelonia mydas*) is  $\sim$  20,500 and increasing and the number of seabirds is  $\sim$  1 million. Animal tracking studies have shown that some taxa breed and/or forage consistently within the MPA (e.g. some reef fshes, elasmobranchs and seabirds), suggesting the MPA has the potential to provide long-term protection. In contrast, post-nesting green turtles travel up to 4000 km to distant foraging sites, so the protected beaches in the Chagos Archipelago provide a nesting sanctuary for individuals that forage across an ocean basin and several geopolitical borders. Surveys using divers and underwater video systems show high habitat diversity and abundant marine life on all trophic levels. For example, coral cover can be as high as 40–50%. Ecological studies are shedding light on how remote ecosystems function, connect to each other and respond to climate-driven stressors compared to other locations that are more locally impacted. However, important threats to this MPA have been identifed, particularly global heating events, and Illegal, Unreported and Unregulated (IUU) fshing activity, which considerably impact both reef and pelagic fshes.

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# **Introduction**

The growing recognition that marine ecosystems are threatened by biodiversity declines and habitat degradation (McCauley et al. [2015](#page-18-0)) has led to international calls

for protecting the world's ocean, including within Marine Protected Areas (MPAs) (Convention on Biological Diversity's Aichi Target 11 <https://www.cbd.int/sp/targets/>; Woodley et al. [2019](#page-20-0)). Negotiations at the United Nations are also ongoing to establish a new international treaty within which MPAs would be established in Areas Beyond National Jurisdiction (ABNJs) (O'Leary et al. [2020](#page-18-1)). A large body of research spanning over 50 years demonstrates that in general, MPAs lead to increases in biodiversity, abundance, size and biomass (e.g. Ballantine [2014](#page-15-0); Lester et al. [2009](#page-18-2)). Importantly, there is also clear evidence of fsheries benefts (Goñi et al. [2010](#page-17-0); Harrison et al. [2012\)](#page-17-1), well-being and social benefits (Ban et al. [2019](#page-15-1)), and resilience afforded by protection in the face of climate change (Mellin et al. [2016](#page-18-3); Roberts et al. [2017\)](#page-19-0). While there are recognised limitations (Devillers et al. [2015](#page-16-0); Edgar et al. [2014](#page-16-1); Giakoumi et al. [2018](#page-16-2)), impacts of protection are largely positive in coastal ecosystems.

Very Large Marine Protected Areas (VLMPAs), areas >  $100,000 \text{ km}^2$ , are fundamental to halting and reversing ocean health declines and to meeting global targets. The Aichi Target calls for a minimum of 10% of the world's ocean to be protected by 2020, a target that will not be met with currently only 2.5% of the ocean's surface in highly protected MPAs ([https://www.mpatlas.org/;](https://www.mpatlas.org/) Sala et al. [2018](#page-19-1)). Additionally, the  $30 \times 30$  initiative, supported by the analysis of O'Leary et al. [\(2016](#page-18-4)), suggests that a minimum of 30% of the ocean should be in highly protected MPAs. Positive conservation outcomes from large-scale protection are also expected to generate positive social, economic and equity outcomes with respect to food security and resource access (Sumaila et al. [2015\)](#page-19-2). However, the benefts of VLMPAs remain debated and empirical studies evaluating their efectiveness are essential. These studies have been limited due to the relatively young age of VLMPAs; the frst VLMPA to be established was the Pacifc Remote Islands National Marine Monument in 2009 (MPA Atlas, [https://mpatlas.org/mpa/](https://mpatlas.org/mpa/sites/7704395/) [sites/7704395/\)](https://mpatlas.org/mpa/sites/7704395/). Significant challenge also exists in delivering conservation research in remote regions and on large spatial scales that include offshore pelagic environments.

The British Indian Ocean Territory (BIOT) MPA was proclaimed by the UK Government in April 2010. It is classified as a VLMPA at  $640,000 \text{ km}^2$  and as an IUCN management category 1a strict nature reserve (Day et al. [2019](#page-16-3)), with efectively no permitted fshing. At the time of its designation, it was the largest contiguous highly protected MPA. The MPA includes a range of habitats with deep oceanic areas surrounding the shallow reef environments and reef islands of the Chagos Archipelago. Its recognition as an important site for conservation (reviewed previously by Sheppard et al. [2012](#page-19-3)) has helped drive a concerted programme of ongoing studies to understand the outcomes of the MPA's creation and its importance for the species and ecosystems it hosts.

At the same time, the legality of this MPA has been challenged (Appleby [2015](#page-15-2); United Nations [2019\)](#page-19-4). Given both the ongoing challenges to the BIOT MPA and the wealth of recent studies, here we assess the knowledge gains over the past decade regarding this MPA's conservation value. We also discuss the ongoing conservation challenges facing the BIOT MPA that continue to require new and innovative approaches and consider the implications of the lessons learnt for marine conservation planning and management more broadly across the globe.

#### **Materials and methods**

#### **Identifying case studies**

Marine research in BIOT extends back to the 1970s but has increased rapidly in the last 15 years. Recently, much of the research within the BIOT MPA has been coordinated through the Bertarelli Programme in Marine Science (BPMS). At the annual BPMS meeting in London (18–20 September 2019), programme-supported scientists were asked to describe their key recent fndings that highlight either the conservation value or the challenges facing the MPA. Experts who attended this meeting were also asked to identify other individuals from around the world who should be invited to participate in writing a review summarizing the last decade of research on the BIOT MPA. The assembled authors were able to provide comprehensive coverage of the breadth of recent work that has taken place concerning the BIOT MPA, including work on a range of habitats such as shallow coral reefs and pelagic realms as well as a range of taxa including fshes, seabirds and turtles. Case studies were identifed by taxonomic group, by habitat, or by ecological question and then experts in each area prepared text describing their recent discoveries, which are synthesised below.

#### **Background and overview of recent scientifc work**

Of the  $640,000 \text{ km}^2$  of the BIOT MPA, 19,120 km<sup>2</sup> is shallower than 100 m and the remainder is deep oceanic water with maximum depths of > 5000 m. The Chagos Archipelago consists of discrete atolls with around 58 associated islands, submerged banks, and an estimated 86 seamounts. The Great Chagos Bank is described as the world's largest atoll structure, covering an area of  $12,642 \text{ km}^2$  and water depths down to about 90 m (Fig. [1\)](#page-2-0). The land area of the islands within the archipelago totals only  $56 \text{ km}^2$ . These islands are surrounded by shallow fringing coral reefs and encompass lagoons with sheltered reefs, patch reefs, coral outcrops and seagrass meadows. The BIOT MPA covers the entire Economic Exclusion Zone (EEZ) with the exception of Diego Garcia atoll and a three-nautical mile bufer around



<span id="page-2-0"></span>**Fig. 1** The Chagos Archipelago. Inset shows the general location within the Indian Ocean and the MPA boundary (red). Main map shows the archipelago which lies at the heart of the MPA. The fve atolls with land are in bold, versus selected submerged reefs and atolls not in bold. Islands on the Great Chagos Bank include Danger Island, Eagle Island, Three Brothers islands and Nelsons Island. Blue shading indicates water shallower than approximately 100 m

it, noting that large parts of this atoll and waters receive separate protection under multiple legal and other regulatory controls (<https://biot.gov.io/>). From the eighteenth century until the 1970s, the archipelago was managed as a coconut oil plantation. When the fnal plantations closed, the archipelago was declared a military exclusion area, and the remaining population was relocated (Wenban-Smith and Carter [2017](#page-20-1)). Since then, commercial fshing—comprising licensed pelagic longline and purse seine fsheries and a relatively small-scale demersal fshery—was allowed up until 2010 at which point all legal commercial fshing ceased. Local human impacts on the reefs within the MPA have generally been minimal, but were signifcant on the islands when previously settled. Approximately half of Diego Garcia, which has the only current human settlement in the archipelago, has been extensively altered for the creation of a large military facility, with buildings and infrastructure, including coastal modifcation, ports and anchorages.



**Fig. 2** The breadth of recent studies in the BIOT MPA. Recent work in the BIOT MPA has used electronic tags to track the movements of sea turtles, seabirds and fsh. Pictured with tags attached **a** a green turtle (*Chelonia mydas*) with a Fastloc-GPS Argos tag on the carapace, **b** a red-footed booby (*Sula sula*) with a light-based geolocator tag on its leg, **c** a silvertip shark (*Carcharhinus albimarginatus*) prior to being ftted with a long-term, internal acoustic transmitter. **d** Habitat surveys using SCUBA and deployed instruments have shown long-term changes in reef environments and water temperature. **e** Counting tracks on beaches has revealed long-term increases in sea turtle nesting numbers. **f** Marine surveys have been extended using technology such as Baited Remote Underwater Video Systems (BRUVS) deployed in the open ocean or in shallow coastal areas. Pictured in (**f**) silvertip sharks. Images courtesy **a**, **e** Nicole Esteban and Graeme Hays, **b** Hannah Wood, **c** David Curnick, **d** Charles Sheppard, **f** Jessica Meeuwig

<span id="page-2-1"></span>The isolated and protected nature of the Chagos Archipelago means that many human infuences are minimal. This limited human presence and remote setting of the BIOT MPA provides a baseline to other systems more impacted by anthropogenic pressures. All else being equal, it might be expected that the MPA would result in positive species and habitat conservation outcomes. There have been considerable recent efforts, documented below, to quantify species abundances for comparison with other areas in the Indian Ocean, as well as assessing long-term changes within the archipelago. This work has shown the value of the MPA for sea turtles, pelagic and reef-associated fshes, seabirds, invertebrates and key habitats, such as coral reefs and seagrass beds (Fig. [2](#page-2-1)). To assess patterns of movement in relation to the MPA, a range of turtles, fshes and seabirds have been tracked using satellite (Argos and GPS), acoustic telemetry and archival biologging packages. Coral reef surveys have been conducted for four decades, thus informing research on how climate change impacts these ecosystems. Fish surveys on reefs and in pelagic areas with stereo Baited Remote Underwater Video Systems (BRUVS) have been used to describe species assemblages and relative abundance. More recently, detailed oceanographic studies have been undertaken to better understand the drivers behind the biotic patterns and behaviours observed, while remotely operated vehicles (ROVs) have been employed to study the health and diversity of mesophotic reefs and how they may act as refuges for shallow reefs. The temporal, spatial and bathymetric extent of data is thus now signifcant and increasing rapidly. In addition to these studies on abundance, trends and movements, the MPA has allowed a range of questions to be addressed on ecosystem functioning, movement ecology and animal behaviour in an environment relatively free of most human infuences. At the same time, patrols of the MPA provide indications of the extent of Illegal, Unreported and Unregulated (IUU) fshing activity.

#### **Review structure**

We begin by examining the importance of the BIOT MPA for coral reefs and coral-reef research. We then consider work with taxa that has included tracking individuals and/ or census surveys including coral-reef fsh, turtles, seabirds and pelagic fish. We then consider recent knowledge gains regarding invertebrate fauna and mesophotic reefs. We examine how the MPA has provided an environment for seminal work on natural behaviours and ecological relationships in the absence of anthropogenic infuences and we consider how the physical oceanography of the region may infuence its ecological value. Finally, we highlight the key threats the MPA faces, particularly climate warming impacts on coral reefs and IUU fshing impacts on fsh stocks.

#### **Results**

# **Importance of the BIOT MPA for coral reefs and coral‑reef research**

The BIOT MPA represents a valuable reference site for understanding coral community resilience in an ocean where most reefs have undergone signifcant and continuing declines in health. Although reefs in the Chagos Archipelago have not been spared from the effects of large climate-driven stressors (i.e. temperature driven coral bleaching), the MPA has afforded protection from many of the local threats that reefs face in other parts of the world such as destructive fshing practices, local pollution, or sedimentation and eutrophication from anthropogenic land-based sources.

Data collected following the major coral bleaching event of 1998 showed that despite its geographically isolated position, the Chagos Archipelago was not immune

from widespread coral mortality, which extended to depths of  $>40$  m in some locations (Sheppard et al. [2012\)](#page-19-3). However, most of the reefs recovered quickly and by 2012 coral cover on reefs in the BIOT MPA averaged 40–50% (Fig. [3](#page-4-0)a, d), with juvenile coral densities of 20–60 colonies  $m^{-2}$ (Fig. [3b](#page-4-0)) (Sheppard et al. [2017;](#page-19-5) Sheppard and Sheppard [2019\)](#page-19-6). Thus, the reefs had largely regained coral cover levels consistent with those documented prior to 1998 and coral recruitment was clearly prolifc. This high coral cover and return of dominant branching and tabular species on many fore reef sites supported high net positive carbonate budgets, an important metric infuencing reef growth potential and the maintenance of habitat complexity (Perry et al. [2015](#page-18-5)). Resultant estimates of average vertical reef accretion rates on *Acropora* dominated reefs  $(4.4 \pm 1.0 \text{ mm year}^{-1})$  were high in a global context, indicating that many of the reefs would have the capacity to track projected future sea level rise (Perry et al. [2018](#page-18-6)). For context it is important to note that not all reefs in the wider region recovered as well or as fast after the 1997–1998 bleaching event. For example, shallow reefs in the Maldives recovered to pre-bleaching states by 2013–2014, albeit comparatively slowly and displaying subtle changes in community composition (e.g. Morri et al. [2015\)](#page-18-7), whilst in the Seychelles reefs followed more divergent recovery trajectories. Some sites recovered well, while others regime-shifted to macroalgal or rubble dominated states with coral cover < 10% (e.g. Chong-Seng et al. [2014;](#page-16-4) Harris et al. [2014](#page-17-2); Graham et al. [2015](#page-17-3)). Regime-shifted sites had negative carbonate budgets and reef accretion rates (Perry et al. [2018](#page-18-6)).

It is clear that the absence of local impacts, provided by the remoteness of the Chagos Archipelago and the presence of the MPA, aided relatively rapid recovery of many reefs compared to other Indian Ocean sites (Sheppard and Sheppard [2019](#page-19-6)). In particular, water quality is emerging as an important factor shaping the response of corals and reefs to heat stress (Wooldridge and Done [2009](#page-20-2); D'Angelo and Wiedenmann [2014](#page-16-5); MacNeil et al. [2019;](#page-18-8) Lapointe et al. [2019](#page-18-9); Donovan et al. [2020](#page-16-6)). Specifcally, an increase in nitrogen (especially nitrate) coupled with phosphorous limitation, which are typical of land-based pollution, exacerbate the effects of heat stress and prolongs recovery time following bleaching events (Wiedenmann et al. [2013;](#page-20-3) Ezzat et al. [2016](#page-16-7); Burkepile et al. [2020](#page-15-3)). The absence of such stressors within the Chagos Archipelago is likely a key contributor to the rapid recovery observed on these reefs compared to other reefs within the region and within other MPAs (e.g., the Florida Keys National Marine Sanctuary and the Great Barrier Reef Marine Park) (MacNeil et al. [2019](#page-18-8); Lapointe et al. [2019](#page-18-9)).

However, it is also relevant to note that these reefs have not been immune from repeated disturbances over the last decade. Localised outbreaks of crown-of-thorns starfsh



<span id="page-4-0"></span>**Fig. 3** Metrics of reef health on ocean-facing coral reefs across the Chagos Archipelago. **a** Live coral cover (%) at diferent depths 1995–2019; **b** Juvenile coral densities (individuals m−2) at diferent depths 2012–2019; **c** Coral carbonate production rate (kg m−2 year−1) and rugosity at  $8-10$  m depth 2015–2020. All values are means $\pm$ SD.

(*Acanthaster planci*) were observed in 2013, causing high mortality of branching *Acropora* spp.. Further, White Syndrome disease was prevalent on many reefs in 2014 and 2015, causing widespread mortality of tabular *Acropora* colonies (Wright [2016;](#page-20-4) Sheppard et al. [2017](#page-19-5)). Most signifcantly, however, the reefs were again heavily impacted by the recent global heat stress event, which caused back-to-back coral bleaching and mortality in 2015 and 2016. Intensive research efforts in BIOT over the last five years are providing detailed insights into subsequent ecological changes across a wide range of depths and habitats.

As after the 1998 event, widespread coral mortality reduced average coral cover to around 10% in 2017, mainly afecting reefs to a depth of 15 m (Fig. [3](#page-4-0)a, e) (Sheppard et al. [2017;](#page-19-5) Head et al. [2019\)](#page-17-4). This decline in coral cover was driven primarily by a~90% decline in *Acropora* spp. cover in shallow and mid depths, shifting community composition from competitive to stress-tolerant taxa and leaving *Porites*

Shaded areas represent major coral bleaching events. Photographs show reef states in **d** 2015, **e** 2018 and **f** an example of young *Acropora* spp. growing on a dead table coral in 2019. Note that 2020 data in c are based on a subset of survey locations. Photographs: **d** Chris Perry, (**e**, **f**) Ines Lange

spp. as the dominant coral genus post-bleaching (Head et al. [2019](#page-17-4); Lange and Perry [2019](#page-18-10)). In deeper water (20 m+), the largest losses were of foliacious forms. No evidence of coral acclimation following 1998 can thus be inferred. Soft corals have also been lost, especially on shallow reefs and seaward facing exposed reefs, and now occupy less than 4% in the 15–25 m depth range. Sponges showed an initial increase in 2018, especially in deep waters, but have declined to about 12% cover in 2019 (Sannassy Pilly et al. unpubl. data). Despite the decrease in coral cover, feshy macroalgae are very rare, which may be attributed to absence of nutrient stress from fertilizer and sewage runoff that negatively affect reefs in many coastal areas (Fabricius [2005;](#page-16-8) Lapointe et al. [2019\)](#page-18-9). The only life form to show a mean increase across reefs are calcifying algae (especially *Halimeda* spp.), which have increased from negligible values to 12% in shallow waters and to 15–16% in deeper waters. Crustose coralline algae cover has increased from 8% to around 25% in shallow

water and to around 20% in deeper waters in 2019 (Benkwitt et al. [2019](#page-15-4); Sannassy Pilly et al. unpubl. data). From a geoecological perspective, the main consequence of the above community changes has been a major decline in carbonate production rates, which have dropped by an average of 77% (Fig. [3c](#page-4-0)). At the same time, mean reef rugosity declined by 16% (Fig. [3c](#page-4-0)) and rubble cover doubled between 2015 and 2018 (Lange and Perry [2019](#page-18-10)).

Critical questions at present are whether the reefs will follow the same recovery trajectories as after 1998, or whether more divergent trajectories will occur in diferent sites and locations (see section below on Key Ongoing Threats). The presence of the BIOT MPA guarantees that recovery trajectories will not be impeded by local stressors such as anthropogenically-derived nitrogen enrichment and altered nutrient ratios, which can exacerbate coral disease and bleaching and has led to reef degradation in other protected areas, e.g. the Florida Keys National Marine Sanctuary (Lapointe et al. [2019\)](#page-18-9). Still, recovery potential will ultimately depend on recurrence intervals and magnitudes of future heat stress events.

# **Coral reef fshes are much more abundant than in other Indian Ocean locations**

The first underwater visual surveys of fish biomass and community structure in the Chagos Archipelago were conducted on the outer reef slopes of the atolls in 2010, the year the MPA was established. The archipelago had also been a de facto MPA for reef fshes, with very limited reef fshing since the 1970s (Koldewey et al. [2010\)](#page-17-5). Fish biomass on these reefs was six times greater than even the best-protected smaller MPAs surveyed across eight other countries in the Western Indian Ocean (WIO) (Graham and McClanahan [2013\)](#page-17-6). Much of this biomass was made up of species targeted by fshing elsewhere in the region, higher trophic level species and larger body-sized fshes (Graham et al. [2013](#page-17-7)). These species often have large home ranges (Green et al. [2015](#page-17-8)), making them vulnerable to fshing pressures outside smaller MPAs. The trophic structure of fish communities across the Indian Ocean changes dramatically with fshing pressure (Barley et al. [2017,](#page-15-5) [2020](#page-15-6)) and in the Chagos Archipelago forms a concave shape, with biomass accumulating at the top and bottom of the trophic structure, allowing for efficient energy transfer through the food-web (Graham et al. [2017\)](#page-17-9). The semi-pristine fsh community allowed for baselines in a range of community-level life history and functional metrics, including maximum length, length at maturity and abundance of top predators and grazers, to be benchmarked across the region (McClanahan and Graham [2015](#page-18-11); McClanahan et al. [2015\)](#page-18-12), and regional-level management priorities to be set (McClanahan et al. [2016](#page-18-13)).

The high biomass values and relatively intact community structure have also been informative to global fish ecology and fsheries studies. Along with some remote locations in the Pacifc, fsh biomass and structure in the Chagos Archipelago enabled estimates of unfshed biomass for coral reefs globally (MacNeil et al. [2015](#page-18-14)) and the functional structure of semi-pristine fsh communities to be established (D'Agata et al. [2016\)](#page-16-9). Globally, the reef fsh biomass in the Chagos Archipelago stands out as a 'bright spot', being greater than would be expected based on the human and environmental conditions experienced alone (Cinner et al. [2016](#page-16-10)), with indications that deep-water refuges and the natural flow of nutrients may contribute to this high biomass (Graham et al. [2018](#page-17-10)). Further, the biomass and proportion of reefs with top predators helped identify the key role of distance to markets as a driver of resource condition inside and out of MPAs (Cinner et al. [2018](#page-16-11)), as has been also observed for pelagic species (Letessier et al. [2019](#page-18-15)). Reef fish otolith studies in the region have revealed the efects of fshing pressure on life spans and patterns of mortality of fshes in other locations across the Indo-Pacifc (Taylor et al. [2019\)](#page-19-7). Biochronological reconstructions of growth histories of fsh species have furthermore helped to refne ecological feedback loops between parrotfshes and habitat disturbance (Taylor et al. [2020a\)](#page-19-8) as well as decadal growth responses to oceanographic conditions (Taylor et al. [2020b\)](#page-19-9).

# **A climate resilient nesting sanctuary for turtles from across the Western Indian Ocean (WIO)**

Green (*Chelonia mydas*) and hawksbill (*Eretmochelys imbricata*) turtles nest in the Chagos Archipelago with both species heavily exploited for two centuries prior to protection being introduced in 1968–1970, with the creation of the MPA further reinforcing this protection (Mortimer et al. [2020\)](#page-18-16). Ongoing census data have highlighted both regionally important nesting populations as well as upwards trends in abundance. For example, estimates of the annual number of clutches across the archipelago for the period 2011–2018 are 6300 and 20,500 for hawksbill and green turtles respectively, increasing 2–5 times for hawksbills and 4–9 times for green turtles since 1996 (Mortimer et al. [2020](#page-18-16)). These upward trends in nesting for both species presumably refect, at least in part, the fact that there has been no known human exploitation of eggs or adults in the Chagos Archipelago for ~ 50 years. Regional estimates indicate that the Chagos Archipelago accounts for 39–51% of hawksbill and 14–20% of green turtle clutches laid across the entire south-western Indian Ocean (Mortimer et al. [2020\)](#page-18-16).

Satellite tracking of nesting green turtles in the Chagos Archipelago has shown that they disperse widely across the WIO at the end of their nesting season, which peaks during June to October (Fig. [4\)](#page-6-0) (Hays et al. [2020;](#page-17-11) Mortimer et al.



15 B Relative clutches per season 13  $11$ 9  $\overline{7}$ 5 3  $\overline{1}$ 1995 2000 2005 2010 2015

<span id="page-6-0"></span>**Fig. 4** The value of the Chagos Archipelago for sea turtles. **a** The archipelago provides a nesting sanctuary for green turtles that forage at distant sites throughout the Western Indian Ocean. Tracks of 35 adult female green turtles are shown, with individuals equipped with tags on nesting beaches on Diego Garcia and then dispersing widely at the end of the nesting season. The extent of the MPA is indicated by the blue hatched area. Stars denote the foraging locations of turtles, i.e. the end-point of migrations where turtles remained for many

months before tags failed (modifed from Hays et al. [2020](#page-17-11)). **b** The significant positive trend ( $p < 0.01$ ,  $r^2 = 0.88$ ) in the estimated number of green turtle clutches laid throughout the Chagos Archipelago. Numbers are scaled relative to those estimated in 1995, i.e. abundance in 1995 appears as one, to highlight the extent of the increase (modifed from Mortimer et al. [2020](#page-18-16)). Between 2011 and 2018, the estimated mean number of clutches per year throughout the archipelago was 20,500 (Mortimer et al. [2020](#page-18-16))

[2020\)](#page-18-16). While some individuals travel to foraging grounds around 80 km away on the Great Chagos Bank, others travel to foraging grounds 1000s of km away, for example, in the Seychelles, Maldives and mainland Africa. The Chagos Archipelago thus provides a key nesting sanctuary for adult green turtles foraging across much of an ocean basin. Ongoing work is assessing migration patterns in adult hawksbill turtles after their nesting season, which peaks during October–February (Mortimer et al. [2020\)](#page-18-16). These green and hawksbill turtle tracking data are being used to inform marine spatial planning broadly across the WIO, helping, for example, to determine boundaries of protected areas in the Seychelles. Investigation of foraging grounds within the MPA have led to discoveries of extensive, deep-water seagrass meadows across the south-east Great Chagos Bank (Esteban et al. [2018\)](#page-16-12). Little is known about these newly discovered habitats, but they appear to support abundant and diverse fsh communities (Esteban et al. [2018\)](#page-16-12). As marine mega-herbivores can act as indicators of the presence of seagrass meadows (Hays et al. [2018](#page-17-12)), future tracking of green turtles in BIOT may increase knowledge of the distribution of these important habitats broadly across the entire WIO. In addition, immature hawksbill and green turtles foraging at Diego Garcia are also being satellite tracked to assess their patterns of space use.

Sand temperature monitoring has shown that the nesting beaches at Diego Garcia are particularly climate resilient with regard to incubation temperatures (Esteban et al. [2016](#page-16-13)). The sex of sea turtle hatchlings is determined by the temperature in the nest in the middle third of incubation. Around the world there is concern that, with a warming climate, populations are becoming increasingly feminised, as females are produced at warmer temperatures. A lack of male hatchlings may ultimately lead to population extinction. At many sites globally, hatchling production is already heavily female skewed (Hays et al. [2014\)](#page-17-13). However, at Diego Garcia, the sand at nest depths is relatively cool, most likely because of a combination of heavy rainfall and shading provided by vegetation behind the nesting beaches. As a consequence of these cool incubation temperatures, it is estimated that hatchling sex ratios are currently balanced (Esteban et al. [2016\)](#page-16-13). Hence, in scenarios of climate warming, excessive feminisation of hatchlings will be much less likely to occur in the Chagos Archipelago than at most other nesting sites around the world. The Chagos Archipelago also supports immature foraging green and hawksbill turtles and ongoing work with drone surveys is estimating the size of these populations and their regional importance (Schofeld et al. [2019](#page-19-10)).

# **The BIOT MPA protects globally signifcant seabird populations**

Research in the Chagos Archipelago has reinforced the important role seabirds play in tropical marine ecosystems. The WIO has been estimated to support ~ 19 million seabirds of 30 species, with the Chagos Archipelago supporting~ 1 million (or 5% of the WIO total) individuals (Danckwerts et al. [2014](#page-16-14)). However, their status and distribution required updating, and until recently virtually nothing was known about their at-sea distribution. A recent synthesis of seabird

status and breeding distribution across the Chagos Archipelago based on visits to all 55 islands, estimated 281,596 breeding pairs of 18 species (Fig. [5](#page-7-0)a). Of these, 96% comprised three species, the sooty tern (*Onychoprion fuscatus* 70%), lesser noddy (*Anous tenuirostris* 18%) and red-footed booby (*Sula sula* 8%) (Carr et al. [2020](#page-16-15)). Assuming 50% breeding success, 281,596 breeding pairs (563,192 individuals) will produce 140,798 offspring, equating to  $\sim$  704,000 breeding adults and immatures, or  $\sim$  4% of the regional total (Danckwerts et al. [2014](#page-16-14)). Current estimates are considerably lower than those proposed by Danckwerts et al. [\(2014](#page-16-14)), and there is strong evidence from early visiting naturalists (Bourne [1886\)](#page-15-7) and guano mining records (Edis [2004](#page-16-16); Wenban-Smith and Carter [2017](#page-20-1)) to suggest this is a fraction of the historic breeding seabird populations. Yet, it is unclear whether trends observed in BIOT are representative of the WIO. Therefore, updated estimates from across the WIO are now needed to reassess the status of breeding seabirds for this region.

At-sea behaviour and distribution of one of the most widely distributed and abundant species in the archipelago, the red-footed booby, is being revealed through the deployment of GPS loggers on breeding adults. Tracking reveals adults commute long-distances over relatively straight paths to feed in deeper waters beyond the Great Chagos Bank (Fig. [5b](#page-7-0)) and suggests at-sea segregation as seen elsewhere with seabirds from diferent colonies (Wakefeld et al. [2013](#page-19-11)). As the vast majority of individuals remained within the MPA (Fig. [5b](#page-7-0)), the lack of commercial fshing within the MPA may help ensure high availability of forage fsh and reduce threats from fsheries bycatch. The restriction of suitable



breeding habitat due to the persistence of introduced rats and associated abandoned coconut plantations across 95% of the terrestrial landmass, remains a constraint to seabird recovery and the MPA delivering its full potential as a seabird sanctuary, although a feasibility study for eradicating rats across the archipelago has recently been completed.

#### **The large no‑take MPA encompasses important pelagic wildlife**

The relatively recent establishment of VLMPAs, combined with the logistical and methodological challenges of sampling remote, expansive regions means that empirical data on the efectiveness of these MPAs for pelagic species are currently limited and conclusions are sometimes conficting. Some studies suggest that MPAs are beneficial for mobile species, with the benefts of MPAs increasing with size, remoteness and age (Edgar et al. [2014](#page-16-1)). The BIOT MPA therefore represents an excellent reference site for such studies.

Since the establishment of the MPA, electronic tagging studies have reported, albeit with relatively low numbers and limited durations, higher than expected residency of pelagic fish species, such as silky sharks (*Carcharhinus falciformis*), sailfish (*Istiophorus platypterus*) and yellowfin tuna (*Thunnus albacares*) (Carlisle et al. [2019](#page-16-17)). The historical fishing record shows that large yellowfin tuna have also been reported to occur in the archipelago year-round (Curnick et al. [2020\)](#page-16-18). Further, activity spaces of all pelagic species tagged around the Chagos Archipelago were significantly smaller than the extent of the



<span id="page-7-0"></span>**Fig. 5** Seabird abundance and movements. **a** Seabird species richness and abundance varies across the Chagos Archipelago. Data are from breeding seabird counts on all 55 islands 2008–2018 (Carr et al. [2020](#page-16-15)). **b** Centrally placed red-footed boobies breeding on the Chagos Archipelago largely forage within the MPA and show evidence

of colony-specifc at-sea segregation. Data are from 192 individuals at three colonies (*DG* Diego Garcia, 2016–18, *n*=99; *DI*: Danger island, 2019 *n*=30; *NI* Nelson's Island, 2018–2019, *n*=63). Study colony locations are marked with triangles and the grey line delineates the MPA

MPA, suggesting it may be large enough to provide a refuge for extended periods of time (Carlisle et al. [2019](#page-16-17)).

Increased understanding of large pelagic species around the Chagos Archipelago has also been informed through the use of fisheries independent mid-water stereo-BRUVS (Fig. [2f](#page-2-1)). Assessments of pelagic richness and biomass using mid-water stereo-BRUVs (in 2012, 2015 and 2016) showed variation among pelagic habitats associated with atolls, seamounts and a deep-sea trench (Meeuwig unpubl. data). This is consistent with historical fisheries data that show high spatial heterogeneity in the distributions of species such as yellowfin tuna (Dunn and Curnick [2019\)](#page-16-19). Pelagic richness and biomass around the Chagos Archipelago are also relatively high compared to global averages (Letessier et al. [2019\)](#page-18-15).

The BIOT MPA was established for biodiversity conservation and not as a fisheries management tool. Studies elsewhere have shown benefits to adjacent tuna fisheries by VLMPA establishment (Boerder et al. [2017](#page-15-8)) and residency behaviour in yellowfin tuna to remote locations (Richardson et al. [2018\)](#page-19-12). Yet a recent study of commercial catch data found no direct evidence that indices of yellowfin tuna abundance have improved in the areas immediately surrounding the MPA (Curnick et al. [2020](#page-16-18)). However, since the MPA's establishment, mismanagement of the yellowfin tuna fishery and a failure to adhere to catch reduction measures (Andriamahefazafy et al. [2020\)](#page-15-9) has resulted in the stock being downgraded to "overfished and subject to overfishing" since 2015 (IOTC-SC21, [2018](#page-17-14)). It is therefore not surprising that a single MPA tiny in size compared to the fished region would be sufficient to turn around such declines, arguing the need for greater regional protection.

All pelagic shark species evaluated by the Indian Ocean Tuna Commission (IOTC)—with the exception of the blue shark (*Prionace glauca*) – have no or uncertain stock assessments (IOTC–SC[21](#page-17-14) [2018\)](#page-17-14). Tracking studies have shown that pelagic sharks may travel across the Indian Ocean to the BIOT MPA, providing further evidence that the MPA may provide an important sanctuary for this group (Queiroz et al. [2019\)](#page-18-17). So, while tracking data confirm sometimes protracted residence of pelagic species within the BIOT MPA (Carlisle et al. [2019\)](#page-16-17) and BRUVs data show high pelagic species richness (Letessier et al. [2019\)](#page-18-15), benefits may also be partly negated by overfishing in the surrounding region (IOTC–SC[21](#page-17-14) [2018](#page-17-14); Curnick et al. [2020\)](#page-16-18) and/or the ongoing IUU fishing activity (see below). Combined, these initial studies suggest that the BIOT MPA and its habitats could have considerable benefits for pelagic wildlife, particularly in the context of high fishing pressure in the region (Kroodsma et al. [2018](#page-17-15)).

# **The BIOT MPA hosts exceptionally high cryptofauna diversity**

First estimates of the decapods in the Chagos Archipelago, one of the most speciose cryptofauna groups on coral-reef microhabitats (Stella et al. [2011\)](#page-19-13), recorded 1868 individuals across 164 nominal species on 54 dead coral colony microhabitats (Head et al. [2018\)](#page-17-16). This number of species is exceptionally high relative to similar studies in other locations (e.g. Preston and Doherty [1990;](#page-18-18) Plaisance et al. [2009](#page-18-19); Enochs and Moanzello [2012;](#page-16-20) Head et al. [2018](#page-17-16)) and community structure is unusual due to a prevalence of obligate coral-dwelling decapods, such as Trapezia crabs (Head et al. [2015](#page-17-17)). Studies are now being undertaken across the archipelago to identify the most important environmental drivers of cryptofauna communities.

# **The BIOT MPA protects diverse mesophotic coral ecosystems**

Mesophotic coral ecosystems (MCEs) are typically found at depths of 30 to  $> 150$  m (Turner et al. [2017](#page-19-14)). Much of our knowledge of MCEs in BIOT is based on diver surveys from the 1970s (Sheppard [1980](#page-19-15)) and a small number of brief ROV surveys in 2016 (Andradi-Brown [2019](#page-15-10)). Building on these studies, in late 2019, high-resolution multibeam and a sophisticated ROV ftted with a HD camera were used to conduct extensive surveys of both upper and lower mesophotic communities from 30 to 150 m at seven sites around Egmont Atoll and Sandes Seamount. Preliminary analysis has revealed diverse and abundant MCEs at all locations surveyed, hosting communities of zooxanthellate scleractinian corals, soft corals, sea fans and sponges. A number of scleractinian coral specimens were also sampled at multiple sites and depths during the surveys. Using molecular techniques, work is ongoing to identify the species of corals sampled and to assess genetic connectivity among shallow and mesophotic reefs. Preliminary observations indicate that the MCEs of BIOT offer huge potential in the level of diversity they encompass and the extension of the shallow-water reefs into deeper waters, which is especially pertinent given recent bleaching events in the region (Head et al. [2019](#page-17-4)). Thus, the BIOT MPA has signifcant value in protecting extensive areas of diverse mesophotic coral ecosystems, which have the potential to support both local and regional shallow-water reefs in the face of climate change.

# **Long‑term protection preserves habitat connectivity, natural behaviours and ecological relationships**

Remote areas like the BIOT MPA can act as natural laboratories that deepen our ecological understanding of reef ecosystems. The BIOT MPA is home to numerous species of seabirds and mobile teleost and elasmobranch fshes that play an important role in connecting discrete habitats. Due to their proximity to deeper waters, the atoll ecosystems are spatially heterogeneous and temporally dynamic with resource availability continually shifting under the infuence of diel and seasonal cycles, as well as oceanographic processes. Quantifying connectivity across these seascapes is important for understanding the degree to which populations should be treated and managed as distinct units (Jacoby and Freeman [2016](#page-17-18)) and to uncover the functional role that mobile species play in nutrient transfer (Williams et al. [2018a](#page-20-5)), predation pressure (Heupel et al. [2014](#page-17-19)) or local measures of biodiversity (Benkwitt et al. [2020](#page-15-11)).

Seabirds in the Chagos Archipelago forage in the open ocean, far from the islands on which they roost and breed (Fig. [5](#page-7-0)). In doing so, they transfer large quantities of nutrients from pelagic food webs to terrestrial systems. This pathway of nutrient flow from seabird guano to coral reefs is illustrated by elevated nitrogen signatures in terrestrial soils and plants, benthic marine organisms, such as sponges and algae, and marine consumers, including herbivorous damselfsh (Graham et al. [2018\)](#page-17-10). These nutrient subsidies, in turn, bolster the growth rates of individual coral-reef fshes, and lead to enhanced biomass and ecosystem functioning (including secondary productivity, grazing and bioerosion rates) of entire fsh assemblages (Graham et al. [2018](#page-17-10); Benkwitt et al. [2020\)](#page-15-11). Contrary to anthropogenically-derived nutrient inputs, which negatively affects coral physiology and increase susceptibility to bleaching (Wooldridge and Done [2009;](#page-20-2) Wiedenmann et al. [2013;](#page-20-3) D'Angelo and

Wiedenmann [2014](#page-16-5); MacNeil et al. [2019](#page-18-8); Donovan et al. [2020](#page-16-6)), naturally-derived nutrients provide nitrogen and phosphorous in optimal ratios and can thus increase coral growth (Shantz and Burkepile [2014;](#page-19-16) Savage [2019](#page-19-17)) and may reduce susceptibility to heat stress (Ezzat et al. [2016](#page-16-7)). Indeed, nutrient inputs from seabirds can also alter the response of coral reefs to marine heatwaves, as demonstrated in part by the proliferation of calcifying algae (e.g., crustose coralline algae) around islands with abundant seabirds following the 2015/2016 mass coral bleaching event in the Chagos Archipelago (Benkwitt et al. [2019](#page-15-4)) (Fig. [6\)](#page-9-0).

Since 2013, a large network of acoustic receivers installed across the archipelago, and annual deployments of both acoustic and satellite tags, are beginning to reveal the extent to which large mobile fshes utilise and link diferent areas across atoll archipelagos (Carlisle et al. [2019;](#page-16-17) Jacoby et al. [2020\)](#page-17-20). Acoustic tracking of grey reef and silvertip sharks, both of which are a principal target of IUU fishing activity in the BIOT MPA, has revealed a few key locations where connectivity is unexpectedly high (Jacoby et al. [2020\)](#page-17-20). A closer look at the reef shark assemblage, using network analyses of the telemetry data, reveals how these species play diferent roles in connectivity across the MPA, with grey reef sharks exhibiting more residential/site-attached behaviour, while silvertip sharks have considerably more dynamic movements (Carlisle et al. [2019](#page-16-17); Jacoby et al. [2020\)](#page-17-20). Interestingly, the movement patterns, and thus connectivity of these sympatric species, vary both diurnally and seasonally suggesting both spatial and temporal segregation within the reef shark assemblage, corroborating patterns observed through stable isotope analyses in BIOT (Curnick et al. [2019\)](#page-16-21).



<span id="page-9-0"></span>**Fig. 6** Benefts of rat-free islands to coral reefs. On rat-free islands in the Chagos Archipelago, seabird guano supplies nutrients to the adjacent coral reefs. These nutrient subsidies, in turn, bolster the growth rates of individual coral-reef fshes, leading to enhanced biomass and ecosystem functioning. Additionally, these nutrient inputs from seabirds can also alter the response of coral reefs to marine heatwaves,

as demonstrated by responses to the 2015/2016 mass coral bleaching event. Even though seabird nutrients did not enhance communitywide resistance to bleaching, they may still promote recovery of these reefs through their positive infuence on **a** calcifying algae (e.g., crustose coralline algae) and **b** herbivorous fshes (modifed after Benkwitt et al. [2019](#page-15-4))

For large-bodied, wide-ranging planktivores like reef manta rays (*Mobula alfredi*), habitat selection is strongly infuenced by prey availability (Stewart et al. [2018](#page-19-18)). Telemetry and biologging approaches are beginning to show that the reef manta rays found in the BIOT MPA frequently utilise atoll ecosystems, sometimes with long-term site fdelity and aggregation sites, such as at Egmont and Salomon atolls (Carlisle et al. [2019](#page-16-17); Harris [2019;](#page-17-21) Andrzejaczek et al. [2020](#page-15-12)). Connectivity is greatly facilitated by dynamic reef manta movements over frequent short-distances  $(< 10 \text{ km})$  and infrequent long-distance (>200 km) horizontal movements as well as dives recorded as deep as 500 m (Andrzejaczek et al. [2020\)](#page-15-12). Characterising the portion of the population that is highly mobile will enable us to better understand drivers of connectivity across the archipelago.

A range of unusual or rarely observed behaviours have been studied in the Chagos Archipelago, which are likely linked to its isolation. Examples include moray eels (*Gymnothorax pictus*) diurnally hunting shore crabs on land (Graham et al. [2009\)](#page-17-22), day octopus (*Octopus cyanea*) hunting cooperatively with fshes (Bayley and Rose [2020\)](#page-15-13) and coconut crabs (*Birgus latro*) predating on adult seabirds (Laidre [2017\)](#page-17-23). All such behaviours are rarely seen, if at all, in highly human-impacted systems elsewhere (Graham and McClana-han [2013](#page-17-6)). Furthermore, parrotfish and surgeonfish in the archipelago exhibit reduced 'fight' behaviour compared to fished areas, showing either an inherited or learned effect of wariness in response to fshing pressure (Januchowski-Hartley et al. [2015\)](#page-17-24). Protected or wilderness areas can therefore provide a valuable window into the natural ecological interactions and behaviours, which have otherwise disappeared or been modifed.

In remote systems such as the Chagos Archipelago, characterised by high consumer biomass (Graham and McClanahan [2013](#page-17-6)), general ecological theories can be tested about relationships and behaviours. Such locations are ideal for investigating what mechanisms maintain trophic structure, drive variation in structure and complexity, and what the implications are for individual behaviours, species interactions, or food-web stability and productivity (McCauley et al. [2012](#page-18-20), [2018;](#page-18-21) Woodson et al. [2018](#page-20-6)). Current work in the Chagos Archipelago has just begun to test such broader ecological theories, for example, the biodiversity-ecosystem function relationship (Benkwitt et al. [2020](#page-15-11)). Thus, not only can remote MPAs like the Chagos Archipelago inform conservation, but also contribute to broader basic ecology research.

# **Understanding the physical oceanography driving biodiversity across the archipelago**

Deep oceanic fushing of cold water into the atolls across the Chagos archipelago drives plankton distributions and ecosystem functioning within the sheltered lagoons (Sheehan et al. [2019\)](#page-19-19). Seamounts are also particularly important features within BIOT and include relatively shallow features such as the Sandes and Swartz seamounts west of Diego Garcia. Their biological signifcance has been suggested from acoustic surveys during which backscatter indicated 100×higher biomass in close proximity to seamounts and a "halo" infuence of the seamount of approximately 1.8 km (Letessier et al. [2016\)](#page-18-22). Recognised as a hotspot for pelagic sharks (Tickler et al. [2017](#page-19-20)), studied seamounts exhibit internal lee waves that fush the summits with nutrient rich, cool water (Hosegood et al. [2019\)](#page-17-25). The steep and narrow seamounts found throughout the archipelago, however, prohibit the formation of Taylor Columns that are frequently cited as the mechanism causing the local retention of nutrients and the subsequent primary production over seamounts (Genin, [2004\)](#page-16-22). Instead, the local generation of turbulent and energetic currents associated with the lee waves are proposed to encourage schooling behaviour of lower trophic levels upon which sharks prey and thereby explain the corresponding acoustic signature in biomass over the drop-off where the internal wave impacts are most pronounced. Acoustic surveys during 2019 over the slopes surrounding Egmont Island, further confrmed that the intensifcation of biomass is not limited to seamounts but extends to the steep slopes surrounding islands and atolls throughout the archipelago (Fig. [7\)](#page-11-0).

#### **Key ongoing threats**

#### **Illegal fshing poses a major threat to vulnerable habitats and species in the BIOT MPA**

IUU fshing activity is a considerable challenge inside the BIOT MPA. Historically, IUU occurred alongside a licensed tuna fshery and it has persisted since the fshery closure in 2010 (Fig. [8](#page-12-0)). From 2002 to 2018, the majority (78%) of vessels have originated from Sri Lanka, although vessels from south-west India are also active (12% of sightings). The Sri Lankan vessels are medium-sized (10–15 m) operating both gill-net and long-line gears, often using illegal wire trace to target sharks (MRAG [2015\)](#page-18-23) (Fig. [8](#page-12-0)).

Enforcement occurs primarily through use of the BIOT Patrol Vessel, which is responsible for the detection and apprehension of IUU fshing vessels within the MPA. Ferretti et al. ([2018\)](#page-16-23) estimated that 20–120 boats enter the area annually. However, determining the actual level of IUU threat is complicated by temporal and spatial variation in patrolling effort. Although patrolling has occurred since 1996, patrol effort data have only been logged consistently since December 2013. That notwithstanding, trends in IUU vessel encounters suggest that the MPA's implementation has had little discernible impact on the IUU activity (Fig. [8](#page-12-0)).



<span id="page-11-0"></span>**Fig. 7** Use of sonar and cameras to reveal mid-water fauna. 38 kHz raw Sv echograms of submerged banks at **a** Sandes and **b** Egmont (lower). Dense dark red echogram returns show the seabed and second echo at Sandes, with aggregations of biomass (fsh and zooplank-

ton) in shallower water, confrmed opportunistically using camera drops. **c** and **d** cruise tracks showing seabed depth (with red showing echogram portion. **e** and **f** camera validation of targets (Hosegood, Williamson and Embling, unpublished data, [2019\)](#page-17-25)

Spatial and temporal analyses of all vessel encounters suggest that suspected IUU is focused on the shallow reefs and northern sectors (Fig. [8](#page-12-0)) with peaks in activity in the months of May–June and December (MRAG, unpublished data).

IUU fshing appears to have driven declines in some shark populations within the MPA (Ferretti et al. [2018;](#page-16-23) Tickler et al. [2019](#page-19-21)) and so may impair the MPA's function as a refuge for these species (Letessier et al. [2019\)](#page-18-15). From the catch data, Ferretti et al. ([2018](#page-16-23)) estimated that between 1,745 and 23,195 sharks were caught between 1996 and 2015 within the MPA. The number of sharks seen per scientifc dive in the archipelago reduced from  $\sim$  4 in the 1970s to  $\sim$  1 since the mid-1990s (Graham et al. [2010](#page-17-26)). Recent re-surveys (2018–2019) of the reef fsh community structure and biomass on the outer reef slopes at the same sites, using the same methods, and by the same observer, have indicated substantial declines in biomass (Graham et al. unpubl. data) that have also been linked to a reported increase in reef fsh within confscated catches (MRAG [2015](#page-18-23)).

Similar to the temporal surveys on the outer reef slopes, substantial declines in reef fsh and sharks were observed in BRUVS surveys within the atoll lagoons between 2012 and 2016 (Meeuwig unpubl. data). Important exploited families, such as serranids and lethrinids, decreased by 74% and 53%,

while coral feeding groups, such as chaetodontids, declined by 37% (Meeuwig unpubl. data). Among the shark species, whitetip reef sharks (*Triaenodon obesus*) declined by 81% and 60% in relative abundance and size, respectively. The grey reef shark declined by 76% in relative abundance and by 4% in size. The tawny nurse shark (*Nebrius ferrugineus*) reduced in relative abundance and size by 37% and 60% (Meeuwig unpubl. data). These declines in relative abundance and size were coincident with recorded poaching incidents (MRAG [2015\)](#page-18-23).

Currently, the BIOT Patrol Vessel has to balance patrol activities, border protection, scientifc research support, as well as service and maintenance outside the territory. As such, there have been recent efforts to improve enforcement capacity through the trialling of additional technologies within the MPA through the UK's Blue Belt Programme with a Technology Roadmap under development. Importantly, the continued threat from IUU fshing highlights the need to improve monitoring and understanding of the human dimensions (e.g. socio–economic drivers of illegal fshing) of large MPAs which, although remote, are interconnected within wider socio-ecological systems (Gruby et al. [2015](#page-17-27)). Concerns have also been raised about the adequacy and efectiveness of punitive measures, whereby risks of capture



<span id="page-12-0"></span>**Fig. 8** The threat of Illegal, Unreported and Unregulated fshing. **a** Heat-map of AIS activity from fshing feets operating in the British Indian Ocean Territory area of interest (BIOT AOI) between 1 January 2014 and 31 December 2019. Fishing vessel identities were confrmed and the activity shown is restricted to AIS transmissions associated with speeds between 0.5 and 5 knots, speeds typically associated with fshing operations and fshing activity at sea. The extension and level of fshing activity is represented by positional densities that vary from: black=no activity, transparent-green=lower activity (low positional densities) to red/higher activity (hotspots). Legal activity within 3 nautical miles of Diego Garcia (white cross) and slow transits to and from port are not shown. The activity in the northern MPA is produced by small-scale commercial fshing vessels (feet) transiting regularly at slow speed and shaping these lanes between the northeast and northwest boundaries. However, these vessels very frequently deploy fshing gears inside the MPA while on

combined with low costs associated with any arrest may still leave IUU fishing as a viable option for some fishers.

#### **Coral reefs in the Chagos Archipelago are not immune to bleaching events**

Reefs in the Chagos Archipelago have repeatedly been impacted by global coral bleaching events, and the current ecological condition of the reefs suggests they are presently at a critical recovery stage. While coral cover is starting to increase, structural complexity changes are likely to continue for several years, as the remaining reef continues to degrade due to intense external and internal bio-physical erosion. transit and need to be accounted for within the overall fshing activity. Overall, fshing activity is high and widespread through the adjacent high seas. The east and west boundaries of the MPA show high risk due to fshing activity encroaching and entering the marine protected area, with short and repetitive incursions. Additionally, low positional densities inside the south-west MPA are produced from infrequent longer incursions. **b** Vessels suspected of IUU activity that were either detained by authorities or escaped capture from 2002 to 2020. The dashed line indicates MPA implementation (2010). Flag of origin indicated in legend, other=Indonesia, Mauritius, Japan, Taiwan. Source: MRAG, unpublished data, 2020. **c** Location of detained or escaped vessels suspected of IUU from 2002 to 2020. Numbers represent the number of vessels from that same site. The cross indicates the location of Diego Garcia. Source: MRAG, unpublished data, 2020. **d** An example of a confscated catch in the BIOT MPA (photo Tom B Letessier)

Shallow reefs are increasingly covered by the bioeroding sponge *Cliona* spp., decreasing the area suitable for new coral settlement. Additionally, an outbreak of coralline fungal disease has been observed in 2018, potentially impacting coral recruitment further (Williams et al. [2018b\)](#page-20-7). Indeed, data from 2017 indicates that the density of newly settled coral recruits  $(< 1$  year-old) has reduced by approximately 90% since 2013 (Fig. [3b](#page-4-0)). Larger young corals  $(>1 \text{ year})$ are present in greater numbers, though most are located on unstable dead table corals or mobile rubble (Fig. [3f](#page-4-0)), and therefore are likely to experience high mortality rates (Sheppard et al. [2017\)](#page-19-5). Measured growth rates for several coral species were also comparatively low in 2018–2019,

suggesting prolonged efects of heat stress on coral physiology (Lange and Perry [2020\)](#page-18-24). Since the late 1970s, several coral species and key species assemblages in the Chagos Archipelago have gone regionally or functionally extinct. Although species diversity remains high at present, local extinctions may increase in the future, following a spiral of positive feedback through low recruitment and lack of suitable settlement substrate (Sheppard et al. [2020\)](#page-19-22).

Importantly, the remote and protected nature of the BIOT MPA has previously supported rapid coral community recovery following widespread mortality in 1997–1998, giving hope for future recovery (Sheppard et al. [2008\)](#page-19-23). However, it is unclear whether all reefs will restructure in the same way that they did after 1998, whether recovery will be as fast at all sites, or whether some sites may regime-shift to other states. The return of *Acropora* spp. dominated communities will be crucial to restore the key geo-ecological functions of habitat complexity and carbonate production that local reefs delivered pre-bleaching (Lange and Perry [2019\)](#page-18-10). Ultimately, the primary control on coral-reef recovery in the Chagos Archipelago will be the recurrence intervals and magnitudes of future heat stress events. Unfortunately, BIOT is predicted to see a large increase in the frequency of annual severe bleaching events in the coming decades, even under conservative emission scenarios (van Hooidonk et al. [2016\)](#page-19-24). Additionally, atmospheric nitrogen deposition is projected to increase in the future, negatively afecting even remote coral reefs (Chen et al. [2019](#page-16-24)).

# **Discussion**

#### **Future research directions for large MPA science**

Here, we have shown how recent research in the BIOT MPA has helped to identify not only its conservation benefts, such as increased abundance of various species, habitat diversity and resilience, but also the physical and ecological processes that drive these benefts. Fundamental to these fndings has been the multi-year monitoring that has identifed important conservation successes, such as the increase in nesting turtle numbers, the recovery of coral reefs following bleaching and mortality, or the preservation of natural processes such as seabird subsidies improving reef vigour. Global climate change remains a huge threat to coral reefs, both within the BIOT MPA and elsewhere (e.g. Bates et al. [2019\)](#page-15-14), with the frequency of temperature anomalies and extent of ocean acidifcation likely to play key roles in dictating the type of shallow reefs that survive into the future. Such monitoring needs to be continued and expanded. Long-term monitoring of mesophotic reefs will help identify if they are more resilient than shallow reefs to global heat waves and if these deep reefs help the recovery of bleached areas. It will also identify

if the encouraging trends of increased sea turtle nesting continue in the future as well as the impact of potential threats to sea turtle and seabird nesting posed by rising sea levels. Finally, long-term monitoring of pelagic species at BIOT will also demonstrate the degree to which the MPA generates conservation benefts for mobile exploited species that contribute to regional fsheries.

The BIOT MPA houses regionally significant fish assemblages that play an important role in the resilience of its coral reefs to climate threats but that continue to be impacted by IUU fshing. Future research should focus on improving the understanding of the scale and nature of IUU fshing in the MPA, as well as its drivers to assist with improved enforcement and compliance. Targeted research is also needed to develop efficient mechanisms to combat IUU fishing given the huge area of the BIOT MPA poses signifcant logistical challenges. Innovative methods to combat IUU fshing have started to be implemented, often with methods tailored to target the specifc IUU fshery (e.g. Tickler et al. [2019\)](#page-19-21) and need expanding.

It is important to assess the extent of animal movements in relation to MPAs so that threats to mobile species can be identifed and benefts of diferent sized protected areas can be objectively assessed (Dwyer et al. [2020](#page-16-25)). Given that many marine species may travel many thousands of km (Hays and Scott [2013](#page-17-28)), even the largest protected areas, such as the BIOT MPA, may sometimes not encompass the full extent of marine animal movements. While a number of species have been tracked (e.g. green turtles and red-footed boobies) important knowledge gaps remain. For seabirds, their movements outside the breeding season remain unknown. Initial studies suggest that the BIOT MPA and its habitats could have considerable benefts for pelagic fsh. Yet, a challenge remains to humanely capture and equip a large enough number of individuals to assess the overall patterns of movement for pelagic fsh species. Interestingly, some pelagic sharks equipped with tags 1000s of km away off southern Africa, have travelled across the Indian Ocean to the BIOT MPA (Queiroz et al. [2019](#page-18-17)). So, for some taxa, tagging studies conducted within the BIOT MPA might usefully be blended with studies being conducted elsewhere to assess patterns of space use across the Indian Ocean and more broadly (Barkley et al. [2019](#page-15-15)). The huge value of such data-sharing in animal tracking studies has recently been emphasised (Sequeira et al. [2019](#page-19-25)). In some areas, such as marine animal tracking, routes by which data can drive conservation outcomes have been identifed (Hays et al. [2019](#page-17-29)) and the tracks of turtles equipped in the Chagos Archipelago that migrate broadly are already being used to help direct marine spatial planning both in BIOT and the Seychelles.

Little is known about some important habitats in the BIOT MPA. While coral reefs have been a focal habitat for concerted research for some time, a depth limit of 25 m is placed on diving activities to minimise the risks in such a remote location. Yet most of the Great Chagos Bank, the world largest atoll structure, is between 25 and 100 m deep. Deeper areas are only starting to be explored with, for example, the use of drop-down cameras and ROVs (remotely operated vehicles). Furthermore, research in the BIOT MPA to date has also been focussed on returning to sites previously surveyed, to build a robust, long-term time-series. Yet this has resulted in the majority of the archipelago remaining unexplored and under-studied, such as the seagrass beds on the Great Chagos Bank. Here, there may be a very useful synergy between animal tracking studies and habitat surveys, with hot-spots of space use identifed in tracking studies, being used to direct in-situ habitat surveys, i.e. tracking animals helps identify areas of particular interest (Jacoby et al. [2020\)](#page-17-20). An example here is the use of green turtles to identify the location of seagrass beds on the Great Chagos Bank that were hitherto unknown (Esteban et al. [2018\)](#page-16-12).

#### **Lessons learned of relevance to other VLMPAs**

While the number of MPAs across the world is increasing, their benefts continue to be debated (Edgar et al. [2014](#page-16-1); Bruno et al. [2019\)](#page-15-16). Set against this backdrop, case studies showing the value of MPAs are important (Murray and Hee [2019](#page-18-25)). One feature that is evident from much of the recent research is the importance of long-term monitoring throughout the system. It is well established how the value of ecological time-series grows as the time-series lengthen (e.g. see Edwards et al. [2010](#page-16-26)), allowing the drivers of longterm changes and inter-annual variability to be more clearly identifed. It is therefore important for long-term monitoring to occur in VLMPAs and that it embraces new technology. Such monitoring allows assessment of the success of conservation actions and identifcation of emerging threats. For instance, in the Florida Keys National Marine Sanctuary, whilst highly protected zones have benefted fshes relative to partially protected zones, this high level of protection has had no impact on the rate of coral decline (Toth et al. [2014\)](#page-19-26) which is driven both by large scale factors such as poor water quality and climate-related storms and bleaching.

That the BIOT MPA, despite its extreme remoteness, remains subject to incursions of IUU fshing with a demonstrable impact on biodiversity demonstrates the need for more efficient mechanisms to combat IUU fishing. This may be a common issue with remote MPAs and necessitates the need for innovative methods to combat IUU fshing (Park et al. [2020](#page-18-26)). For example, in the territorial waters around French Islands in the Southern Ocean, radar detecting tags carried by albatrosses are being used to detect large ships operating illegally (Weimerskirch et al. [2019](#page-20-8)). Further, interactions between large static MPAs and mobile fshing gears, such as fish aggregation devices (FADS) (Bucaram et al.

[2018\)](#page-15-17) and industrial fshing feets around their perimeters (Kroodsma et al. [2018;](#page-17-15) Curnick et al. [2020\)](#page-16-18) need to be better understood. Given the huge fshing pressures in unregulated high seas fsheries outside protected areas, the importance of large MPAs for pelagic species protection has been stressed (Queiroz et al. [2019\)](#page-18-17). Yet, we emphasise that large protected areas, such as the BIOT MPA, should not be considered as a silver bullet, but rather in conjunction with wider sustainable and efective fshery management regulations to provide the urgent conservation and management benefts needed for pelagic predators. The recent developments to expand the UN Convention on the Law of the Sea (UNCLOS) to include a new legally binding instrument on the conservation and sustainable use of marine life in Areas Beyond National Jurisdiction (General Assembly resolution 72/249) are therefore encouraging.

In addition to studying a range of marine habitats within MPAs, another important research direction is to better quantify the connections between terrestrial and marine environments. Although this research will take diferent forms in the BIOT MPA and other remote VLMPAs compared to smaller MPAs located closer to human population centres, prioritizing research and encouraging management across land-sea boundaries applies to all MPAs. Specifcally, land-based nutrient pollution plays a large role in declining coral health, especially when coupled with increasing warming events (Wooldridge and Done [2009;](#page-20-2) Donovan et al. [2020](#page-16-6)). As a result, there have been recent calls to better regulate run-off from land adjacent to MPAs to mitigate continuing coral loss and enhance recovery following bleaching events (Lapointe et al. [2019;](#page-18-9) MacNeil et al. [2019](#page-18-8)). In contrast to these human-derived nutrients, natural nutrient subsidies, such as those provided by seabirds nesting on islands, may beneft coral reefs and enhance their resilience to global heat waves (Graham et al. [2018](#page-17-10); Benkwitt et al. [2019\)](#page-15-4). Thus, while one research and management priority within BIOT is the restoration of such natural nutrients (e.g., by eradicating invasive rats and restoring seabird populations), less remote MPAs will likely need to simultaneously reduce human-derived nutrient run-off to have similar benefts for coral reefs. Still, jointly managing terrestrial systems in conjunction with MPAs may be broadly applicable, and may increase the efectiveness of MPAs at conserving coral reefs and other nearshore habitats.

Cutting across all the marine science work in the BIOT MPA, an important goal is to maximise the translation of the accumulated data into positive conservation outcomes, a theme that pervades across MPAs more broadly (Lubchenco and Grorud-Colvert [2015\)](#page-18-27). The BIOT MPA was one of the early wave of no-take VLMPAs implemented from 2006 to 2010 (with Papahānaumokuākea Marine National Monument, USA and Phoenix Islands Protected Area, Kiribati) as countries worked to meet Aichi Target 11 of 10% ocean protection by 2020 under the United Nations' (UN) Convention on Biological Diversity (CBD), later endorsed under Sustainable Development Goal 14. Today, only 5.3% of the world's ocean is protected with 2.5% highly protected in no-take MPAs [\(https://mpatlas.org/,](https://mpatlas.org/) accessed 26 May 2020). However, the UK government is leading the 30-by-30 initiative, pushing for at least 30% of the global ocean to be protected by 2030 with the hope that this goal will be ratifed at the 2020 CBD Conference of the Parties, now rescheduled for 2021. Research from the BIOT MPA therefore provides important insights to inform policy commitments around ocean protection, including the need for greater regional protection, as part of the actions identifed to rebuild ocean life (Duarte et al. [2020](#page-16-27)). Mechanisms to effectively achieve this science to policy interface will be aided by the UN Decade of Ocean Science for Sustainable Development (2021–2030). The wealth of new information from ongoing work in the BIOT MPA promises to help drive marine conservation both within the MPA and more broadly, which is, perhaps the most important legacy this work can leave.

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#### **Compliance with ethical standards**

**Conflict of interest** The authors declare that they have no confict of interest.

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