RESEARCH PAPER

Determination of priority phenolic pollutants exploiting an in-syringe dispersive liquid—liquid microextraction—multisyringe chromatography system

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Abstract An automatic phenolic compounds analyzer is presented. The system performs online magnetic-stirring-assisted dispersive liquid-liquid microextraction before multisyringe chromatography (MSC) using a monolithic Chromolith Flash RP-18e column. The extraction behavior of the following phenolic pollutants: phenol, 2-nitrophenol, 4-nitrophenol, 2-chlorophenol, 2,4-diclorophenol, and 2,4,6trichlorophenol, has been studied. A critical comparison of extractants (tributyl phosphate, acetonitrile, hexane, and 1-chlorobutane) and disperser solvents (acetone, acetonitrile, ethanol, methanol, 1-propanol, and 2-propanol) was made. Tributyl phosphate and acetonitrile were chosen as the extractant and the disperser solvent, respectively, since these showed the best performance. Phenols were online back-extracted into NaOH and neutralized before multi-isocratic chromatographic separation. The proposed analyzer can be applied for wide linear working ranges, i.e., between 40 and 20,000 μ g L⁻¹. The precision of the developed system has been proved, with maximum values for the intraday and interday precision of 4.4 % and 5.2 %, respectively, expressed as relative standard deviation, and high preconcentration factors (9.3-10.5) for most of the compounds studied. The method developed was successfully applied to natural water samples.

Keywords Magnetic-stirring-assisted dispersive liquid–liquid microextraction · Multisyringe chromatography · Phenolic compounds · Environmental samples · Automation

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Introduction

Phenolic compounds can be found in the environment directly from anthropogenic sources and from the transformation of synthetic or natural chemicals [1, 2]. Phenol is widely used as a disinfectant for medical and industrial applications [3]. Its principal sources are discharges from different industries. It is also used in the production of explosives, pesticides, and dves [2]. The presence of chlorophenols in the environment is mainly related to the use of organic compounds such as pesticides and growth regulators, whose biodegradation leads to the formation of phenol, 2-chlorophenol (2-CP), and 2,4dichlorophenol (2,4-DCP) [2]. 2-CP and 2,4-DCP are also used as herbicides [4]. Nitrophenols, particularly 2nitrophenol (2-NP) and 4-nitrophenol (4-NP), can be formed in the reaction of phenol with nitrite ions in water under UV (sunlight) irradiation [5] and also during the production of pesticides, polymers, and drugs [6] in addition to other industrial activities. Given their toxicological [7-9] and organoleptic [10] effects, the United States Environmental Protection Agency has listed some phenolic compounds as priority pollutants [11]. Thus, the monitoring of phenolic pollutants is essential to notice and control illegal discharges from industry or contamination by pesticides.

Phenolic compounds have been determined using different strategies, mainly exploiting separation techniques, e.g., gas chromatography and liquid chromatography [12, 13]. Liquid chromatography is one of the most used techniques since derivatization of analytes is avoided [13, 14]. Moreover, the advent of monolithic columns made possible the development of low-pressure chromatographic techniques such as sequential injection chromatography [15] and multisyringe chromatography (MSC) [16] as a result of the implementation of monolithic columns in sequential injection analysis (SIA) and multisyringe flow injection analysis (MSFIA) [17] systems, respectively. These techniques are low-cost alternatives

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to high-performance liquid chromatography (HPLC) providing a reduction of sample and organic solvent consumption, which leads to reduction of waste production [18]. Moreover, multi-isocratic chromatography can be easily implemented using MSC. Furthermore, the use of a multisyringe burette allows precolumn and postcolumn operations permitting the performance of the whole analytical protocol in a fully automated way within a single instrumental assembly.

Usually sample pretreatment is required before chromatographic separation, especially when dealing with environmental or biological samples, in order to attain analyte enrichment and matrix removal. Solid-phase extraction is the technique most commonly used for this purpose [14, 19]. However, recently there have been interesting developments in liquidliquid extraction (LLE), such as the advent of dispersive liquid-liquid microextraction (DLLME) [20, 21] and the development of lab-in-syringe systems [22], which allow automatic LLE inside the syringe, saving time, reagents, and solvents, and consequently reducing the generation of residues. DLLME is based on the use of a ternary component (disperser solvent) miscible in both water and the extractant. The disperser is mixed with the immiscible extraction solvent and is injected into the aqueous sample, producing a cloudy solution of small droplets of the extractant, enhancing the contact surface area between the phases. DLLME has been automated by exploiting flow techniques. For example, Anthemidis and Ioannou [23] developed an SIA-DLLME system for the determination of lead and copper by exploiting flame atomic absorption spectrometry. DLLME has also been automated using MSFIA, e.g., for the determination of benzo[a]pyrene [24], chromium [25], and total phenol index [26]. The automation of DLLME improves its reproducibility and accuracy, since flow techniques allow the precise handling of small volumes. Furthermore, both DLLME and flow techniques reduce sample and reagent consumption, attaining improved environmentally friendly systems.

In this work a DLLME–MSC system for determination of six phenolic pollutants is presented. Several extractants and disperser solvents are studied. Multi-isocratic chromatography was implemented, and the potential of the proposed system as a powerful tool for screening of phenolic pollutants in environmental samples is studied.

Materials and methods

Reagents, samples and solutions

All solutions were prepared with distilled water (resistivity greater than $1.8 \times 10^5 \Omega$ cm) from a Milli-Q system (Millipore, Bedford, MA, USA). HPLC-grade acetonitrile (ACN; Scharlau, Barcelona, Spain) was used to prepare the mobile phases. Acetic acid (Sigma-Aldrich, Madrid, Spain), sodium

chloride, sodium dihydrogen phosphate, and trifluoroacetic acid (Scharlau) were of analytical reagent grade. The other reagents used were tributyl phosphate (TBP), hexane, acetone, ethanol, 1-propanol, and 2-propanol from Scharlau and 1chlorobutane and ammonium sulfate from Sigma-Aldrich.

The phenolic compounds analyzed, i.e., phenol, 2-NP, 4-NP, 2-CP, 2,4-DCP, and 2,4,6-trichlorophenol (2,4,6-TCP), were purchased from Sigma-Aldrich. Stock solutions of each phenolic compound were prepared by accurately weighing the appropriate mass of the phenolic compound and by dissolving it in HPLC-grade methanol (Scharlau) in order to obtain a final concentration of 1,000 mg L⁻¹. Working standard solutions containing either mixtures or individual compounds were prepared by dilution of the stock solutions in water.

The water samples analyzed were tap water from the laboratory, commercial mineral water, well water from a well in Son Reus, Palma (Spain, geographic coordinates, X473247, Y4388551), and a leachate sample from a demolition waste treatment plant in Santa Margalida (Spain, geographic coordinates, X 512401, Y 4398015). The pH was adjusted to 2 using 0.1 M HCl.

Magnetic-stirring-assisted DLLME-MSC system and software

The basic element of an MSFIA system is a multisyringe burette allowing the simultaneous movement of four syringes. Each syringe has a three-way solenoid valve (N-Research, Caldwell, NJ, USA) placed at the head which allows the injection of the reagents into the system only when required and their return to their reservoirs when they are not required. The proposed magnetic-stirring-assisted (MSA) DLLME-MSC system (Fig. 1c) comprises a multisyringe burette module (CRISON, Alella, Spain) placed upside down and equipped with two 5-mL glass syringes (S1 and S4; Hamilton, Bonaduz, Switzerland). This way the organic droplet of the extractant, which is less dense than water, stays at the top and the aqueous phase (Fig. 1a) can be discarded before back-extraction with NaOH. To enhance the extraction efficiency, an MSA system (Sciware Systems, Bunyola, Spain) was mounted on S1 (Fig. 1b), where the extraction and backextraction occur. The MSA system allows homogeneous and rapid mixing of phases. It consists of a small magnetic bar (10mm length, 3-mm diameter) placed inside the syringe, an external stirring support placed around the syringe's body, a motor connected to the external agitation support by a rubber band, which forces the rotation of the external agitation support, and a circuit that controls both the starting/stopping and the motor revolutions through one of the multisyringe outputs. The external stirring support is a ring (14-mm inner diameter, 30-mm outer diameter) with two small magnets facing each other creating a rotating magnetic field around the body of the syringe. When the motor is switched on, the ring starts to

Fig. 1 a Extractant drop after phase separation. b The agitation system. c Manifold. ACN acetonitrile, D detector, Db debubbler, HC holding coil, M magnetic-stirring-assisted system, MC monolithic column, MP mobile phase, S syringe, SV selection valve, TBP tributyl phosphate, V solenoid valve, W waste



rotate and with it the magnetic bar inside the syringe, mixing the phases. A schematic depiction of the MSA system is shown in Fig. 1.

The system has two switching selection valves (SV1 and SV2), two external solenoid valves (V1 and V2), a monolithic column, a debubbler (Sciware Systems), and a spectrophotometric detector. S1 is connected to an external solenoid valve (V1; MTV-3-N 1/4 UKG; Takasago, Nagoya, Japan) that drives the flow to waste ("on" position) or to the eight-port selection valve (SV1; CRISON) ("off" position), which has the extraction and back-extraction reagents located on its peripheral ports, as follows: water (port 1), cleaning solution (10:90 v/v acetone:H₂O; port 2), 0.1 M NaOH (port 3), autosampler (port 4), extractant-disperser solution (port 5), and waste (port 8). The second syringe (S4) is used to perform the multi-isocratic separation by dispensing two different mobile phases through the monolithic column. As the valves of the multisyringe module can handle up to 2 bar, the valve of S4 was replaced by a Kel-F one-way connector (Sciware Systems) and an external solenoid valve (V2) (Takasago) which can withstand up to 6 bar. The "on" position of V2 connects S4 to the mobile phase 1 reservoir, and the "off" position connects it to a holding coil connected to a second, six-port selection valve (SV2; VICI Valco, Schenkon, Switzerland) whose port configuration is as follows: a neutralization tank (port 5), a Chromolith Flash RP-18e monolithic column (25 mm×4.6-mm inner diameter) protected with a Chromolith RP-18e guard column (5 mm×4.6-mm inner diameter) (Merck, Darmstadt, Germany) (port 3), 0.5 M acetic acid (port 2), mobile phase 2 (port 4), and waste (port 6).

The detection system was a USB2000+ spectrophotometer equipped with a 1-cm flow cell (Ocean Optics, Dunedin, FL, USA). Measurements were recorded simultaneously at 270, 290, 300, and 315 nm. The manifold was constructed with polytetrafluoroethylene tubing (0.8-mm inner diameter).

Instrumental control, data acquisition, and data processing were done with AutoAnalysis 5.0 (Sciware Systems). The distinctive feature of this software based on dynamic link libraries at 32 bits is the possibility of using a single and versatile application without further modification for whatever instrumentation and detection system needed. It involves a basic protocol which allows the implementation of specific and individual dynamic link libraries, addressing the configuration of the assembled flow analyzer.

MSA-DLLME-MSC procedure and calculations

The analytical procedure is detailed in Table 1. DLLME is performed in S1 using a mixture of TBP and ACN (10:90 v/v) as the extractant and the disperser solvent, respectively. S1 is

	Step description	Multisyringe burette movement	SV1 (port)	SV2 (port)	S 1	S4 (V2)	Motor	V1
1	Start loop							
2	Fill HCs with MP1	Dispense 1 mL	8	6	Off	Off	Off	Off
3	Precondition the column	Dispense 2 mL	8	3	Off	Off	Off	Off
4	Empty S1 and S4	Dispense 2 mL	8	6	Off	Off	Off	Off
5	Load sample	Pick up 4.006 mL	4	6	Off	On	Off	Off
6	Load organic phase (with agitation)	Pick up 0.862 mL	5	6	Off	On	On	Off
7	Agitation	-	5	6	Off	Off	On	Off
8	Wait 20 s	-	5	6	Off	Off	On	Off
9	Stop agitation	-	5	6	Off	Off	Off	Off
10	Discard aqueous phase	Dispense 4.86 mL	8	6	Off	On	Off	On
11	Load NaOH with agitation	Pick up 0.35 mL	3	6	Off	On	On	Off
12	Agitation	-	5	6	Off	Off	On	Off
13	Wait 40 s	-	5	6	Off	Off	On	Off
14	Stop agitation	-	5	6	Off	Off	Off	Off
15	Deliver NaOH into HC2	Dispense 0.2 mL	8	6	On	On	Off	Off
16	Dispense NaOH into the neutralization tank	Dispense 0.1 mL	8	5	On	Off	Off	Off
17	Dispense the rest of NaOH from HC2 into the neutralization tank with MP1	Dispense 0.2 mL	8	5	Off	Off	Off	Off
18	Load CH ₃ COOH	Pick up 0.1 mL	8	2	Off	Off	Off	Off
19	Neutralization	Dispense 0.1 mL	8	5	Off	Off	Off	Off
20	Load neutralized extract	Pick up 0.385 mL	8	5	Off	Off	Off	Off
21	Inject extract and MP1	Dispense 6.9 mL	8	3	Off	Off	Off	Off
22	Rinse S4 with MP2 and S1 with cleaning solution	Priming pick up, dispense, pick up	2	4	Off	On	Off	Off
23	Second elution with MP2	Dispense 6.2 mL	8	3	Off	Off	Off	Off
24	Syringe cleaning procedure							
25	Repeat loop (×3)							
26	Sample change procedure							

 Table 1
 Detailed analytical procedure of the magnetic-stirring-assisted dispersive liquid–liquid microextraction–multisyringe chromatography (MSA-DLLME–MSC) system for determination of phenolic compounds

HC holding coil, MP mobile phase, S syringe, SV selection valve, V three-way solenoid valve

coupled to an external solenoid valve (V1). Its "off" position connects S1 to the central port of SV1. First, the sample is loaded into S1, followed by the mixture of TBP and ACN at a high flow rate (15 mL min⁻¹) and with the MSA system activated. As a result, a cloudy solution is formed inside S1. Agitation is maintained for 20 s, and then it is stopped. After phase separation has been achieved, the aqueous phase is discarded to waste with S1 in the "off" position and V1 in the "on" position in order to avoid cross-contamination. To combine the DLLME with the chromatographic separation, a back-extraction with NaOH is performed. NaOH is aspirated from port 3 of SV1 with the MSA system activated. Agitation is maintained for 40 s, and then NaOH is dispensed via S1 in the "on" position to holding coil 2 and S4 pushes it to the neutralization tank in port 5 of SV2. There, NaOH is neutralized with acetic acid. After neutralization, the resulting solution is injected onto the monolithic column using S4. Thus,

the sample is injected into the chromatographic column, followed by 6 mL of mobile phase 1. Then S4, holding coil 1, and holding coil 2 are filled with mobile phase 2 to dispense 6.2 mL of mobile phase 2 to elute the phenolic compounds still retained on the column. All solutions pass through the debubbler before reaching the detector. Finally, a cleaning procedure is performed. It consists in filling S1 with cleaning solution (10:90 v/v acetone:H₂O) before rinsing it twice with Milli-Q water, while S4 is simultaneously rinsed with mobile phase 1.

The limit of detection (LOD) was calculated according to IUPAC [27, 28] as three times the standard deviation of ten blanks divided by the slope of each calibration curve, the limit of quantification was calculated as ten times the standard deviation of ten blanks divided by the slope of each calibration curve, and relative bias in percent were calculated by dividing the difference between the mean of the results and the reference value by the reference value [29].

Optimization of experimental conditions

To optimize the operational conditions of the proposed system, a series of experiments were conducted using a multivariate optimization technique. Multivariate designs provide relevant knowledge of the effect of variables within the entire experimental domain selected, and the variance of the estimate of the response at every point of the domain is better than that obtained by univariate methods [30]. First, a preliminary study with wide ranges was performed to find out which variables had a significant effect on the response and to select the appropriate ranges to study. Then, a screening of the independent variables and their possible interactions was performed by a two-level full factorial design (2^k) . The evaluation of the screening results allows the identification of the variables with a significant influence on the analytical response and allows to discard those with negligible effects [31]. Then, a facecentered central composite design was applied in order to find the critical values of the significant variables. Three center points were included to identify any irregularities, such as the loss of linearity in the center of the interval in both studies. The statistical software program Statistica 7.0 (StatSoft, Tulsa, OK, USA) was used for the entire multivariate analysis.

Results and discussion

Study of extractants and disperser solvents

The selection of the extractant and the disperser solvent was performed off-line with aqueous solutions containing each phenolic compound at 10 mg L⁻¹ (pH 2–3). The extractants studied were: ACN adding 10 g of ammonium sulfate to 50 mL of sample [32], 1-chlorobutane [33, 34], hexane [35], and TBP [21]. For ACN with ammonium sulfate, the absorbance of the extraction blank was too high, so this option was discarded. The results obtained with the other extractants are shown in Fig. 2a. Results are expressed as relative intensity (%), with 100 % for the solvent providing the highest absorbance. TBP showed the best performance for most of the phenolic compounds studied but for 2-NP, for which hexane was slightly better. Taking into account these results, we selected TBP as the extractant.

Then, mixtures of TBP with different disperser solvents were tested, i.e., methanol [36], ethanol [36], ACN [26, 36], acetone [36, 37], 1-propanol [26], and 2-propanol. The TBP– acetone mixture was discarded because of the high absorbance of the extraction blank. As can be seen from Fig. 2b, TBP– ACN showed the best performance in general, and especially for those phenols giving a lower absorbance signal; therefore, it was selected for further assays.

Fig. 2 Results of the study of extractants (**a**) and disperser solvents (**b**) expressed as relative intensity with error bars. *2-CP* 2-chlorophenol, *2,4-DCP* 2,4-dichlorophenol, *2-NP* 2-nitrophenol, *4-NP* 4-nitrophenol, *P* phenol, *2,4,6-TCP* 2,4,6-trichlorophenol



Multivariate optimization of the DLLME and back-extraction

A multivariate optimization of the parameters affecting the DLLME and the back-extraction was performed. First, a preliminary study with wide ranges was conducted to find out which variables had a significant effect on the response and to select the appropriate ranges to study. The volume of NaOH had a significant positive effect. However, because of the limitations of the system (chromatographic separation) it was fixed at 0.3 mL. Thus, the percentage of TBP, the volume of the organic phase, and the concentration of NaOH were selected as variables for the multivariate study.

A full factorial screening (2^k) , including three center points, with adjusted ranges was performed. The curvature and the three variables were significant; therefore, according to the results obtained (ANOVA table, Pareto chart), the studied ranges were readjusted to the following to apply a response surface face-centered central composite design: TBP percentage, 10–40 %; volume of the organic phase, 0.5–1 mL; NaOH concentration, 0.1–1 mol L⁻¹. The desirability graph was studied to obtain the optimum values. The critical values were set to the lowest percentage of TBP studied (10 %), 0.832 mL of the organic phase, and 0.1 M NaOH.

Effect of the ionic strength

The effect of the ionic strength was studied by adding different amounts of NaCl (from 0 to 1 mol L⁻¹) to the aqueous samples. An increase of the extraction efficiency was obtained when the NaCl concentration increased from 0 to 0.5 mol L⁻¹, but no significant differences were observed when concentrations up to 1 mol L⁻¹ were added. Thus, 0.5 M NaCl was selected for further assays.

Effect of agitation time

The influence of agitation time was examined. No signal improvement was observed when the DLLME agitation time was increased from 20 to 40 s. Thus, the extraction agitation time was fixed at 20 s. For back-extraction, a higher response was obtained when the agitation time was increased from 20 to 40 s, but no signal improvement was observed when it was increased to 60 s. Therefore, the back-extraction time was set at 40 s.

Chromatographic separation

Multi-isocratic chromatographic separation was performed at a flow rate of 0.6 mL min⁻¹, the maximum flow rate showing no back-pressure problems. Different proportions of ACN in water were tested. A 16:84 (v/v) ACN:H₂O mixture was selected as mobile phase 1 to elute phenol, 2-NP, 2-CP, and 4-NP. Then, a second mobile phase, 35:65 (v/v) ACN:H₂O,

was used to elute 2,4-DCP and 2,4,6-TCP. Two mobile phases were required to perform the separation of the phenolic compounds studied in a reasonable time.

The effect of adding trifluoroacetic acid was studied, and with the addition of 0.1 % trifluoroacetic acid better resolution was achieved. A chromatogram obtained with the optimum conditions detailed above is shown in Fig. 3.

Quantification of phenolic compounds

As can be seen in Fig. 3, 2-CP and 2-NP were coeluted. We did not succeed in separating these two phenols with the column that our system permitted given its short length. However, both could be quantified by using one of the AutoAnalysis tools which allows to simultaneously measure up to four different wavelengths. This also allowed us to measure each phenol at the wavelength where better sensitivity is obtained. The quantification wavelengths are summarized in Table 2.

Thus, to quantify 2-CP and 2-NP, we used two wavelengths, i.e., 270 and 300 nm. At 270 nm both phenolic compounds absorb, but at 300 nm only 2-NP absorbs. The mathematical procedure followed was as follows: On the one hand, if both phenolic compounds are present in the sample, 2-NP is quantified at 300 nm. Then its concentration is interpolated in the curve obtained for 2-NP at 270 nm to reveal the area corresponding to this compound. Finally, the difference between the total area and the area associated with 2-NP at 270 nm corresponds to 2-CP. This way, both compounds are quantified. On the other hand, if no peak is observed at 300 nm, all the peak area measured at 270 nm corresponds to 2-CP, and its concentration can be calculated directly.



Fig. 3 Chromatogram obtained for an aqueous sample with 0.5 M NaCl, pH 2, and spiked with the six phenolic compounds studied at 2.5 mg L⁻¹. The chromatographic separation conditions were as follows: MP1, ACN:H₂O (16:84 v/v) with 0.1 % trifluoroacetic acid (TFA); MP2, ACN:H₂O (35:65 v/v) with 0.1 % TFA; flow rate, 0.6 mL min⁻¹. *I*) P, 2) 4-NP, *3*) 2-NP plus 2-CP, *4*) 2,4-DCP, *5*) 2,4,6-TCP, *AU* absorbance units

	Ρ	2-NP		4-NP		2-CP	2,4-DCP	2,4,6-TCP
Quantification wavelength (nm)	270	270	300	315 (without 2-NP)	315 (with 2-NP)	270	290	290
LOD (mg L ⁻¹)	0.10	0.04	0.08	0.01	0.01	0.07	0.09	0.14
LOQ (mg L ⁻¹)	0.33	0.12	0.27	0.04	0.04	0.22	0.30	0.46
Linear range (mg L ⁻¹)	0.33 - 20	0.12 - 20	0.27-20	0.04-20	0.04-20	0.22 - 20	0.30-20	0.46-20
Calibration curve	y = 1.0049x - 0.1187	y = 2.7981x - 0.2645	y = 1.0678x + 0.0307	y = 6.8747x + 0.7961	y = 6.1933x + 0.2789	y=1.4691x+0.1093	y = 1.1056x - 0.0658	y = 0.712x
Sensitivity (AU·L mg ⁻¹)	1.0049	2.7981	1.0678	6.8747	6.1933	1.4691	1.1056	0.712
Determination coefficient (R^2)	0.9998	0.9995	0.9926	0.9997	0.9995	0.9999	0.9930	0.9948
Intraday precision (%) $(n=7)$	2.2	3.5	4.0	3.3	3.2	4.4	3.5	3.1
Interday precision (%) $(n=5)$	4.8	4.4	5.2	3.6	3.7	5.13	4.1	5.2
Injection frequency (h ⁻¹)	1.8	1.8	1.8	1.8	1.8	1.8	1.8	1.8
Extraction efficiency (%)	82		82		86	92 ^a	82	49
Experimental preconcentration factor	9.4		9.3		9.8	10.5	9.4	5.6
<i>AU</i> absorbance units, 2-CP 2-cl trichlorophenol	ilorophenol, 2,4-DCP	2,4-dichlorophenol, LO	D limit of detection,	LOQ limit of quantif	ication, 2-NP 2-nitrop	henol, 4-NP 4-nitroph	nenol, P phenol, 2,4,6-	TCP 2,4,6-

 Table 2
 Figures of merit of the proposed MSA-DLLME-MSC method

^a Calculated using the area corresponding to 2-CP at 270 nm.

Figures of merit of the proposed MSA-DLLME-MSC system

To evaluate the performance of the proposed MSA-DLLME–MSC method, its figures of merit were studied, and they are listed in Table 2.

Intraday precision is expressed as the relative standard deviation (RSD) obtained when analyzing a 5 mg L^{-1} standard containing the six phenolic compounds studied in seven consecutive determinations on the same day. As can be seen, RSDs of 4.4 % or less were obtained. Interday precision corresponds to the RSD of a 5 mg L^{-1} standard containing the six phenolic compounds studied analyzed on five different days. For all the analytes, an RSD of 5.2 % or less was obtained.

For the preconcentration factor, first the theoretical preconcentration factor was calculated, assuming 100 % extraction and back-extraction efficiency, by the relation of the initial sample volume and the final NaOH volume where all the phenolic compounds are back-extracted. Thus, the theoretical preconcentration factor was 11.43. Then, 1 mL of a solution of NaOH containing each phenol at 11.43 mg L⁻¹ was neutralized and injected into the chromatographic part of the system. Its response was compared with the signal of a solution containing each phenol at 1 mg L⁻¹ analyzed with the proposed system. The relation of these two absorbance values gives the efficiency of extraction and back-extraction of each phenolic compound shown in Table 2 as the extraction efficiency. Then, the experimental preconcentration factor was calculated by multiplying the extraction efficiency of each phenolic compound and the theoretical preconcentration factor. As can be seen from Table 2, high preconcentration factors between 9.3 and 10.5 were achieved for most of the phenolic compounds studied, except for 2,4,6-TCP (5.6). Compared with previously reported manual methods dealing with liquidliquid microextraction of phenolic compounds, better extraction efficiencies are achieved for most of the common phenolic compounds studied, except for 2,4,6-TCP. Fattahi et al. [38] reported extraction efficiencies of 28.7, 62, and 75.15 % for 2-CP, 2,4-DCP, and 2,4,6-TCP, respectively, Faraji et al. [39] reported extraction efficiencies of 21, 34, 63, and 86 % for phenol, 2-CP, 2,4-DCP, and 2,4,6-TCP, respectively, and Vera-Ávila et al. [36] reported extraction efficiencies of 13.1, 38.2, 76.9, and 86.2 % for phenol, 2-CP, 2,4-DCP, and 2,4,6-TCP, respectively. Furthermore, higher precision is achieved with the present method, i.e., RSDs between 6.4 and 10.4 % in comparison with RSDs of 5.2 % or less obtained with the proposed system. These methods report better LODs because of greater sample consumption and the use of more sensitive but more expensive detectors. In addition, with the proposed method, the consumption of organic solvents is reduced between three and ten times in comparison with manual methods [36, 37].

Regarding other automatic methods determining 2-NP and 4-NP [33, 34], better extraction efficiencies are achieved with the present method with a similar consumption of organic solvents: i.e., approximately 50 % in comparison with approximately 80 % with the present method.

The linear working range has been established from the limit of quantification to the highest concentration of the standard analyzed, being possible to quantify simultaneously concentrations within the range $0.33-20 \text{ mg L}^{-1}$ for phenol,

Sample	Phenolic compound added (mg L ⁻¹)	Recovery											
		Р		2-NP		4-NP		2-CP		2,4-DCP		2,4,6-TCP	
		mg L ⁻¹	%	mg L ⁻¹	%								
Tap water	0	0	-	0	-	0	-	0	-	0	-	0	-
	0.5	$0.54{\pm}0.03$	108	$0.55 {\pm} 0.05$	110	$0.51 {\pm} 0.08$	102	$0.54{\pm}0.05$	108	$0.54{\pm}0.06$	108	$0.49 {\pm} 0.05$	98
	1	$0.89{\pm}0.04$	89	$0.97 {\pm} 0.04$	97	$0.910 {\pm} 0.007$	91	$0.83{\pm}0.08$	83	$0.94{\pm}0.03$	94	$0.98{\pm}0.02$	98
Mineral water	0	0	-	0	-	0	-	0	-	0	-	0	-
	0.5	$0.46 {\pm} 0.07$	92	$0.50 {\pm} 0.04$	100	$0.445 {\pm} 0.006$	89	$0.51 {\pm} 0.04$	102	$0.51 {\pm} 0.03$	102	$0.49{\pm}0.02$	98
	1	$1.08 {\pm} 0.06$	108	$0.96 {\pm} 0.08$	96	$0.94{\pm}0.03$	94	$0.94{\pm}0.08$	94	$0.91 {\pm} 0.03$	91	$0.91 {\pm} 0.04$	91
Well water	0	0	-	0	-	0	-	0	-	0	-	0	-
	0.5	$0.486 {\pm} 0.011$	97	$0.46 {\pm} 0.04$	92	$0.471 {\pm} 0.013$	94	$0.54{\pm}0.05$	108	$0.490 {\pm} 0.011$	98	$0.45 {\pm} 0.02$	90
	1	$1.06 {\pm} 0.05$	106	$0.89 {\pm} 0.06$	89	$0.915 {\pm} 0.018$	92	$0.91 {\pm} 0.08$	91	$0.95 {\pm} 0.04$	95	$1.00 {\pm} 0.05$	100
Leachate	0	0	-	0	-	0	-	0	-	0	-	0	-
	0.5	$0.526 {\pm} 0.019$	105	$0.480 {\pm} 0.009$	96	$0.466 {\pm} 0.005$	93	$0.480 {\pm} 0.003$	96	$0.51 {\pm} 0.02$	102	$0.48 {\pm} 0.04$	96
	1	$1.060 {\pm} 0.008$	106	$1.03{\pm}0.08$	103	$1.008 {\pm} 0.015$	101	$0.92{\pm}0.09$	92	$0.96 {\pm} 0.09$	96	$1.07 {\pm} 0.04$	107

Table 3 Application of the MSA-DLLME-MSC method to environmental samples and addition-recovery test at two spiking levels

Sample conditions, pH 2 (acidified with 0.1 M HCl)

0.12–20 mg L⁻¹ for 2-NP, 0.04–20 mg L⁻¹ for 4-NP, 0.22– 20 mg L⁻¹ for 2-CP, 0.30–20 mg L⁻¹ for 2,4-DCP, and 0.46– 20 mg L⁻¹ for 2,4,6-TCP. The LODs achieved are similar to those obtained with previously reported automated methods. For example, Cladera et al. developed an SIA system [33] and an MSFIA system [34] for determination of nitrophenols based on LLE and spectrophotometric detection with similar LODs for 2-NP and 4-NP but lower extraction efficiencies as detailed above. Another MSFIA system exploiting nitrophenols solid-phase extraction and a detector similar to the one used in this work (USB2000, Ocean Optics) [19] obtained LODs four times higher, i.e., 0.17 and 0.04 mg L⁻¹ for 2-NP and 4-NP, respectively. Moreover, higher accuracy is accomplished with the present method.

The injection frequency is 1.8 samples per hour. This frequency is high since it has to be taken into account that it includes the time necessary for performing the extraction and back-extraction, the neutralization, and the chromatographic separation, in addition to system cleaning and reconditioning of the monolithic column. All these steps will take several working days if the analysis is performed in a classic manual approach.

Application to water samples

To investigate the applicability of the proposed method to environmental water samples, it was applied to water samples from different sources. Addition of NaCl to adjust the ionic strength was not necessary since the salinity of the samples was sufficient. Thus, samples were acidified to pH 2 with 0.1 M HCl. The samples analyzed did not contain the phenolic compounds studied, i.e., no signal was observed when analyzing them either with the proposed system or with a commercial HPLC system (PerkinElmer) equipped with an XTerra RP18 column (150 mm×3.9-mm inner diameter and 5-µm particle size). Samples were spiked with different amounts of phenolic compounds. The results are shown in Table 3 together with the recoveries obtained. The range of mean recoveries was 83-110 % for tap water, 89-108 % for mineral water, 89-108 % for well water, and 92-107 % for the analyzed leachate. Accuracy is normally studied as two components: trueness and precision [29]. Thus, the accuracy of the proposed method is proved given the good intraday and interday precision obtained and the low relative bias obtained (11 % or less in most samples) in the spiking-recovery test.

Moreover, to evaluate the possible matrix effect, the slopes of the calibration curves obtained with the standards and the natural water samples were statistically compared with a *t* test. The results showed that there were no significant differences between the slopes since all experimental *t* values obtained were lower than the critical *t* value at the 95 % confidence level. Thus, the applicability of the method to these kinds of matrices is validated.

Conclusions

In conclusion, a fast and simple method to determine six phenolic pollutants in environmental samples has been described. The low cost of the MSC system compared with an HPLC system makes this system a useful screening tool to determine phenolic compounds. In addition, the coupling of DLLME and MSC helped reduce reagent consumption and thus waste generation and the environmental impact per analysis. The use of MSC allowed the integration of the whole protocol in a single analyzer, permitting background operations and the easy implementation of multi-isocratic chromatographic separation. After a careful study of extractants and disperser solvents, TBP and ACN were selected, and we obtained high preconcentration factors and extraction efficiencies for most of the phenolic compounds studied. The coupling of DLLME and the chromatographic separation was achieved by stripping the phenolic compounds with NaOH and its neutralization. The method developed has proved to be accurate and reproducible, given the low RSDs obtained and the satisfactory recoveries achieved when applying it to environmental water samples and a leachate sample.

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