

Effects of *Microcystis* on Hypothalamic-Pituitary-Gonadal-Liver Axis in Nile Tilapia (*Oreochromis niloticus*)

Jiazhang Chen^{1,2} · Shunlong Meng^{1,2} · Hai Xu³  · Zhen Zhang³ · Xiangyang Wu³

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Abstract In the present study, Nile tilapia (*Oreochromis niloticus*) were used to assess the endocrine disruption potential of *Microcystis aeruginosa*. Male Nile tilapia were exposed to lyophilized *M. aeruginosa* or purified microcystin-LR (8.3 µg/L) for 28 days. The levels of serum hormones (17β-estradiol and testosterone) and transcripts of selected genes in the hypothalamus-pituitary-gonadal-liver axis were analyzed. The results showed that serum hormones were significantly up-regulated, and transcripts of 13 genes (GHRH, PACAP, GH, GHR1, GHR2, IGF1, IGF2, CYP19a, CYP19b, 3β-HSD1, 20β-HSD, 17β-HSD1 and 17β-HSD8) were significantly altered after *Microcystis* exposure. These results indicate that fish reproduction can be altered in a *Microcystis* bloom-contaminated aquatic environment.

Keywords *Microcystis* · MC-LR · Brain · Hypothalamus-pituitary-gonadal-liver axis · Nile tilapia

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Jiazhang Chen and Shunlong Meng have contributed equally to this work.

✉ Hai Xu
xuhai@ujs.edu.cn

- ¹ Freshwater Fisheries Research Center, Chinese Academy of Fishery Sciences, Wuxi 214081, China
- ² Key Laboratory of Fishery Eco-environment Assessment and Resource Conservation in Middle and Lower Reaches of the Yangtze River, CAFS, Wuxi 214081, China
- ³ School of Environment and Safety Engineering, Jiangsu University, Zhenjiang 212013, China

Toxic cyanobacterial blooms occur in worldwide freshwaters and represent a human health and ecological concern. There are many species of toxin-producing cyanobacteria which are harmful to aquatic animals (Malbrouck and Kestemont 2006). Microcystins (MCs) belong to a family of cyanotoxins that are mainly produced by the genus *Microcystis*. Up to now, more than 100 variants of MCs have been isolated from *Microcystis* and cultures, with microcystin-LR (MC-LR), microcystin-RR (MC-RR) and microcystin-YR (MC-YR) being the most common (Puddick et al. 2014).

In recent decades, the toxicity of MCs has been well documented in aquatic organisms. For example, a number of studies have demonstrated that MCs cause a range of effects including oxidative stress, developmental toxicity, neurotoxicity, immunotoxicity, as well as hepatotoxic, carcinogenic, mutagenic and cytotoxic effects (Malbrouck and Kestemont 2006; Chen et al. 2016). However, microcystins are only one kind of compound isolated from cyanobacteria. There are also other biologically active compounds in cyanobacteria. Some previous studies have shown that crude extracts from cyanobacterial biomass or *Microcystis* cells may have greater effects than those of purified cyanotoxins (Palíková et al. 2007; Rogers et al. 2011). In fact, aquatic organisms, such as fish, are not only exposed to MCs but rather to *Microcystis* cells and lysates that contain other bioactive substances in *Microcystis* blooms in the real aquatic environment. For example, *Microcystis* also can produce multitudinous peptides classified as aeruginosins (Ishida et al. 1999), micropeptins (Yamaki et al. 2005), and microoviridins (Rohrlack et al. 2003) that have diverse types of biological function (Smith et al. 2008). However, the toxic effects of the complex mixtures of chemicals in *Microcystis* blooms have not been studied in detail. Recently, several studies have revealed insight into the potential for cyanobacterial extracts to exhibit estrogenic

effects (Sychrová et al. 2012), and for *Microcystis* to represent a natural source of environmental estrogens that can up-regulate vitellogenin genes (*vtg*) of embryonic zebrafish (Rogers et al. 2011). Therefore, it is important to consider the toxicity of *Microcystis* in fish.

The Nile tilapia (*Oreochromis niloticus*) is an important economic species and a widely farmed freshwater food fish in southern China. In recent years, this species has been considered as a bioindicator for aquatic environmental contaminants, because of its significant tolerance to environmental stress as well as its potential to be used as an intensive aquaculture species. Due to the tropical growth environment, Nile tilapia may be exposed to *Microcystis* blooms which frequently occur in aquaculture ponds. Although there are several studies demonstrating the estrogenic activities of *Microcystis* in fish, the effects on the hypothalamic-pituitary-gonadal-liver (HPGL) axis remain to be elucidated. In the present study, we investigated serum hormones and several important steroidogenic enzymes of the HPGL axis in Nile tilapia following exposure to *Microcystis* cells and purified MC-LR. Our results suggest that *Microcystis* and MC-LR have the potential to affect the endocrine system in male Nile tilapia.

Materials and Methods

Male Nile tilapia weighing between 40 and 50 g at the beginning of the experiment were obtained from the fish farm of Freshwater Fisheries Research Center, Chinese Academy of Fishery Science (Wuxi, CN). Fish were acclimated under laboratory culture conditions for two weeks and maintained at $28 \pm 0.5^\circ\text{C}$ in a 12 h light: 12 h dark cycle in aquaria with recirculation and dechlorinated tap water. The adults were fed with commercial floating pellets at 10% of their body weight twice daily.

The *Microcystis* treatment was prepared from lyophilized cells of *Microcystis aeruginosa*, which was obtained from the Freshwater Algae Culture Collection at the Institute of Hydrobiology (Wuhan, CN). The total dry weight mass of algal cells was 10 g. For exposure of Nile tilapia, lyophilized *Microcystis* was reconstituted back to its original nominal concentration of 100 mg of lyophilized cells/L. The solution for the MC-LR treatment was prepared by dissolving 100 µg of purified microcystin-LR (CAS#: 101043-37-2, Beijing Lianlixin BioTech, Beijing, CN) and dilution to a series of concentrations in culture water.

Six Nile tilapia were randomly distributed into an aquarium and exposed to lyophilized *Microcystis*, purified MC-LR and control (culture water) for 28 days, in triplicate, for each treatment group. The exposure solutions were renewed daily. Water samples for quality measurements and microcystin analysis were taken during the experiment.

After 28 days of exposure, all fish were euthanized in ice water. Body weight was measured at the start and end of the 28-day exposure period, and the specific growth rate was determined. Blood was collected from the caudal vein using a heparinized syringe, and then centrifuged at $1200 \times g$ for 15 min. The separated serum samples were stored at -80°C for analysis of sex hormones. The brain, gonads and liver were removed from each fish per treatment group, rinsed with cold phosphate buffered saline (PBS, pH 7.0), and snap frozen in liquid nitrogen for later gene expression analysis.

Water quality parameters were measured daily, with dissolved oxygen = 7.3 ± 0.3 mg/L, pH 7.4 ± 0.1 , and total hardness = 28.3 ± 0.4 mg CaCO_3/L . The concentrations of MC-LR were measured using the Beacon Microcystin Tube Kit (Beacon Analytical Systems Inc., Saco, ME, USA) following the manufacturer's instructions. In the lyophilized *Microcystis* treatment, measured MC-LR concentration was 8.3 ± 0.5 µg of MC-LR equiv/L. Therefore, as a positive control, Nile tilapia were also exposed to a nominal concentration of 8.3 µg of purified MC-LR/L under the same conditions. The actual concentration of MC-LR was 9.6 ± 1.2 µg/L in the purified MC-LR-treated groups. There were no fish deaths in the treatment groups or control group during the exposure period.

Serum hormone 17β -estradiol (E_2) and testosterone (T) were measured using a commercially available enzyme-linked immunosorbent assay kit (R&D Systems, Minneapolis, MN, USA) following the manufacturer's protocol.

For gene transcription analysis, six of the same tissues from one aquarium were set as one sample ($n=1$), and there were triplicate samples from each treatment group. Total RNA was extracted from tissue samples using TRIzol reagent (Sangon, Shanghai, CN) following the manufacturer's protocol. The quality and quantity of RNA were determined by UV spectrophotometry and by 1% agarose gel electrophoresis. First-strand cDNA synthesis was performed using the AMV First Strand cDNA Synthesis Kit (Sangon, Shanghai, CN) according to the manufacturer's instructions. Real-time PCR with SYBR green detection was performed on the ABI StepOnePlus™ Real-Time PCR Systems (ABI, Foster, CA, USA) according to protocols established by the manufacturer (ABI SYBR Green PCR Master Mix, ABI). A total of 11 functionally relevant genes associated with the pathways of the HPGL axis of Nile tilapia were selected for the present study based on previous literature (Table S1, Supplementary Information). In addition, reference genes, actin and RPL8 were selected as internal quantitative controls.

Experimental data were checked for normality and homogeneity of variance using the Kolmogorov–Smirnov one-sample test and Levene's test. Intergroup differences were assessed using one-way analysis of variance

(ANOVA) followed by Duncan's test, using SPSS Statistics 18 (SPSS Inc., Chicago, IL, USA). The level for statistical significance was set at $p < 0.05$. All data are shown as mean \pm standard error (S.E.M.).

Results and Discussion

We investigated the potential of *Microcystis* and purified MC-LR to directly disrupt the levels of serum hormones and influence the expression of genes in HPGL axis of male Nile tilapia. Serum E_2 and T levels increased in both treatments ($p < 0.01$) (Fig. 1). No significant differences of the E_2/T ratio were observed. Sex steroid hormones play crucial roles at all stages of the reproductive cycle in vertebrates. Measurement of sex steroid hormones could be suggested as one of the most integrative and functional endpoints for understanding the effects of a chemical on reproduction in fish (Ma et al. 2012; Jo et al. 2014). In the present study, serum E_2 and T levels increased in *Microcystis* and MC-LR exposed Nile tilapia. These results are consistent with a previous study which demonstrated increases in those hormones in male zebrafish upon exposure to purified MC-LR (Liu et al. 2016). Therefore, our results suggest that both *Microcystis* and microcystin have the potential to alter sex hormone balance.

The transcription of genes in the growth hormone (GH)-insulin-like growth factor (IGF) pathway of the HPLG axis was affected in Nile tilapia after *Microcystis* and MC-LR exposure (Fig. 2; Table S2, Supplementary Information). In brain, the levels of GH mRNA were down-regulated and no significant differences were observed for growth hormone-releasing hormone (GHRH) and pituitary adenylate cyclase-activating polypeptide (PACAP) expression in both *Microcystis* and MC-LR group. Expressions of growth hormone receptor 1 and 2 (GHR1 and GHR2) in liver were not significantly different, while those of IGF1 and IGF2 were down-regulated in the *Microcystis* group. However, transcripts of GHR1, GHR2, IGF1 and IGF2 in liver did not

exhibit significant changes in MC-LR treatment relative to the control. In gonad, the transcripts of GHR1 and IGF1 were down-regulated in the *Microcystis*-treated group, while only IGF1 was down-regulated in the MC-LR-treated group. Our observations showed differential expressions of these genes between the *Microcystis* and purified MC-LR treatments. Rogers et al. (2011) evaluated gene expression in zebrafish larvae after exposure to *Microcystis* and purified MC-LR treatment, and their results suggested that there were biological effects beyond just the toxin MC-LR.

The GH-IGF axis plays a crucial role in the regulation of growth. In previous studies, the growth rate of fish had been reported to be associated with the expression levels of GH-IGF axis related genes (Reinecke et al. 2005). In aquaculture, cyanobacteria, including *Microcystis*, were suggested as feasible fish diets due to their nutritive protein (Dong et al. 2012; Ziková et al. 2010). However, most fish growth was inhibited after ingesting cyanobacteria-containing food (Kamjunke et al. 2002; Liang et al. 2015). Furthermore, microcystins from *Microcystis* were accumulated in fish tissue (Zhao et al. 2006; Palikova et al. 2011). In the present study, although GH-IGF axis related gene expressions exhibited different patterns between *Microcystis* and purified MC-LR treatments, the specific growth rate of Nile tilapia was slightly inhibited in the treatment groups compared with the control group (Table S3, Supplementary Information). These results indicated that *Microcystis* and the low concentration of MC-LR may have influenced growth in the present study.

The transcription of cytochrome P450 aromatase gene (CYP19a), 3β -hydroxysteroid dehydrogenase type 1 (3β -HSD1), 20β -hydroxysteroid dehydrogenase (20β -HSD), and 17β -hydroxysteroid dehydrogenase type 1 (17β -HSD1) exhibited similar expressed patterns in Nile tilapia between *Microcystis*- and MC-LR-treated groups (Fig. 2; Table S2, Supplementary Information). In detail, the mRNA levels CYP19a and 3β -HSD1 decreased in both exposed groups, while expression of 17β -HSD1 was significantly increased by exposure to *Microcystis*

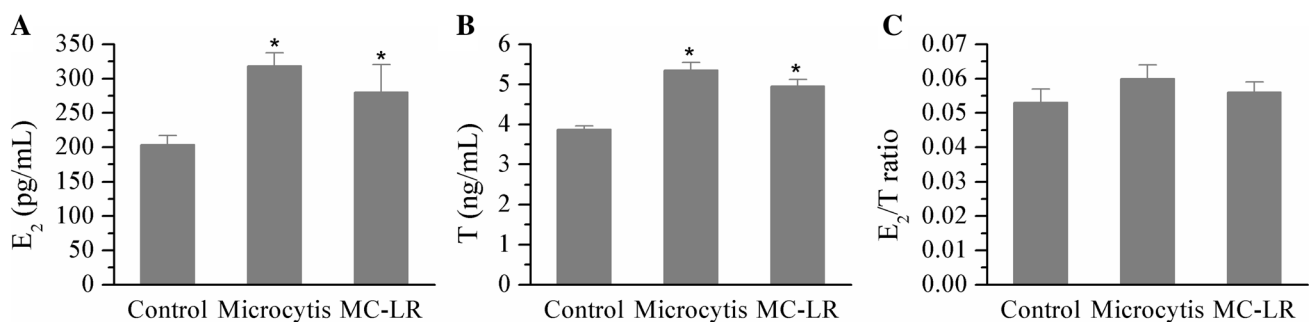
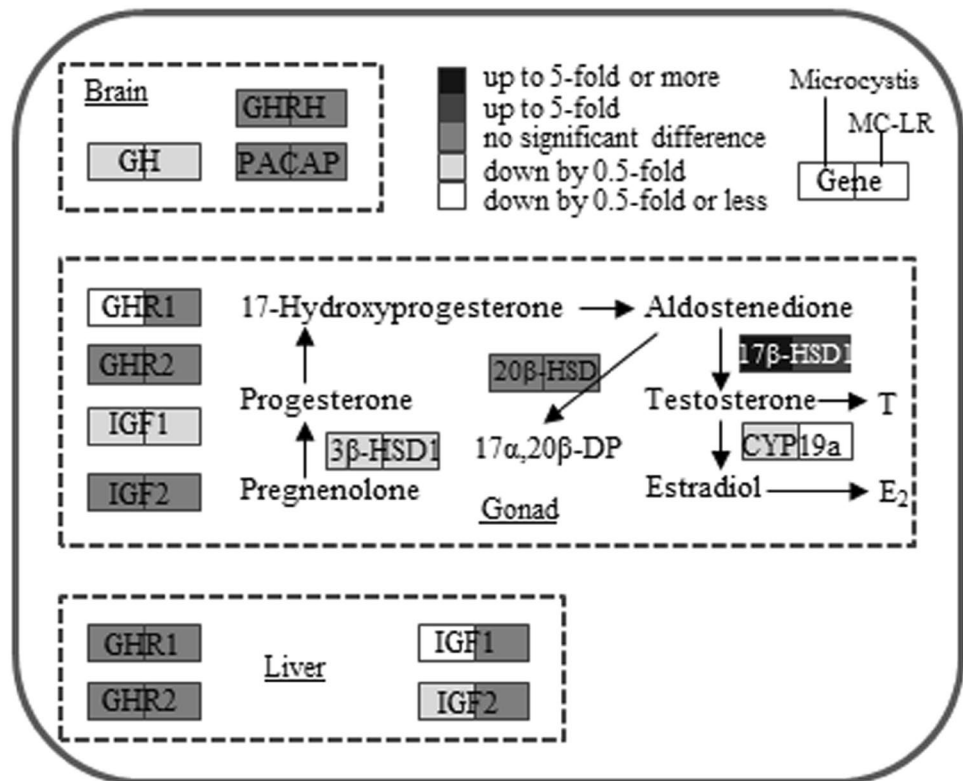


Fig. 1 The levels of serum hormones E_2 (a) and T (b), and the E_2/T ratio (c) in male Nile tilapia after exposure to *Microcystis* and purified MC-LR. The results are means \pm S.E.M. of six samples. * indicated significant difference at $p < 0.05$

Fig. 2 Effects of *Microcystis* and purified MC-LR on gene expression in the HPGL axis in male Nile tilapia. The shades designate different fold thresholds



and MC-LR. In teleosts, CYP19a is one isoform of aromatase involved in the synthesis of estrogens from androgens (Chang et al. 2005). 17β-HSD1 catalyzes the conversion of androstenedione (A) into T (Zhou et al. 2005). Liu et al. (2016) reported that the increase in 17β-HSD mRNAs and decrease in CYP19a mRNA agreed with the greater levels of T and 11-kekotestosterone (11-KT) in male zebrafish exposed to MC-LR. In the present study, we observed similar results in Nile tilapia exposed to *Microcystis* and MC-LR. The gonadal 3β-HSD is linked to the conversion of pregnenolone to progesterone and of dehydroepiandrosterone (DHEA) to A (Senthilkumaran et al. 2009). In the present study, 3β-HSD was down-regulated after exposure to *Microcystis* and MC-LR for 28 days. Adult male zebrafish showed similar decreases in 3β-HSD after 30 days of exposure to purified MC-LR (Liu et al. 2016). Generally, 3β-HSD has the potential to decrease testosterone in males. However, the reduction of 3βHSD level detected in both treatments did not decrease the serum levels of T in male Nile tilapia. We speculated that 17β-HSD1 induced the levels of T in male Nile tilapia. Fish 20β-HSD, expressed in various tissues, including testis and ovary, is known to be involved in the production of 17α, 20β-dihydroxy-4-pregnen-3-one (17α,20β-DP), the maturation inducing hormone (Senthilkumaran et al. 2002). In females, the alternation of this gene potentially suggests reproductive dysfunction

or disruption. However, the levels of 20β-HSD were not changed significantly after exposure to *Microcystis* and MC-LR in the present study.

In conclusion, our findings showed that exposure to *Microcystis* may interfere with the expression of genes involved in the HPGL axis and balance of sex hormones of male Nile tilapia. Therefore, the possibility that *Microcystis* may release endocrine disrupting substances is of considerable environmental interest. The results also indicated that *Microcystis* may pose a potential threat to fish growth and reproduction in a *Microcystis* bloom-contaminated aquatic environment.

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