The Solubility of Metal Arsenates

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It has often been considered that certain metal arsenates because of their low solubility provide a convenient form for the disposal of arsenic residues. In this paper stability diagrams for FeAsO₄, Ca₃(AsO₄)₂ and Mg₃(AsO₄)₂ are derived which show comparatively high solubility of these compounds. The presence of $CO₂$ in the air has a pronounced effect in greatly increasing the solubility of $Ca_3(AsO₄)_2$ and significantly affecting that of $Mg_3(AsO_4)$.

THE residues from nonferrous mining and processing operations often contain appreciable quantities of arsenic compounds which are in many cases considered to be stable. Arsenic is often "stabilized" as ferric arsenate, calcium arsenate, magnesium arsenate and other metal arsenic (v) compounds and disposed of in slag, tailings and residues dumps.

The solubility of the metal arsenic (v) compounds is generally considered to be extremely low. Seidell¹ gives solubility data for a number of metal arsenates but the original source of the data is often very old and, for reasons given later, probably most unreliable.

In 1956 Chukhlansev^{2,3} published work on the solubility of a number of metal arsenates, the results of which have been used in several compilations of thermodynamic data both in the form of solubility products⁴ and free energies of formations.⁵ Chukhlansev's experiments were performed in aqueous solutions using NaOH, H_2SO_4 , HNO_3 or HCl to adust pH. The solubilities show a general linear relationship between log solubility and pH and this has been used by other workers for extrapolating to conditions of higher pH and as the basis of the opinion that metal arsenates under specific conditions of pH are extremely insoluble. Direct extrapolation of solubility products without considering the particular equilbria involved can also lead to large errors when assessing solubilities at various pH values.

Recently, Tozawa Umetsu and Nishimura⁶ and Nishimura and Tozawa⁷ reported a variation to Chukhlansev's linear relationship between log solubility and pH for ferric, calcium and magnesium arsenates which indicated that extrapolation of Chukhlansev's results to higher pH could lead to solubilities which are two low by up to two orders of magnitude. Nishimura and Tozawa explained their results as being due to the decomposition of the particular arsenate to the hydroxide at higher pH, liberating the appropriate arsenic (v) oxyanion.

The solubilities of ferric, calcium and magnesium arsenates at various pH values can be fairly accurately predicted using available free energy data for the various substances in each system.

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Manuscript submitted May 8, 1980.

THERMODYNAMIC STABILITY DIAGRAMS

A convenient approach to this problem of metal arsenate solubility is to consider all of the known equilibria in the particular system and to construct the appropriate stability diagram. A general equation for all hydrolitic equilibria can be written:

$$
aA + xH^+ = bB + cH_2O
$$

and the standard reaction isotherm

$$
\Delta G^{\circ} = -RT \ln K
$$

expanded by taking $T = 298.15$ K, pH = $-\log a_{\text{H+}}$ and $a_{\text{H}_2O} = 1$, to obtain the equation

$$
x \text{pH} = \frac{-\Delta G_{298.15 \text{ K}}^{o}(k \text{J})}{5.708} = -\log \frac{a_B^b}{a_A^a}
$$

This equation applied to each of the known or deduced equilibria in a particular system results in a straight line on a plot of log activity of the dissolved ion *vs* pH (approximately log solubility *vs* pH).

A number of metal arsenate-water systems are considered here in this way and appropriate "solubility"/pH diagrams drawn.

THE IRON (111) ARSENATE-WATER **SYSTEM**

Free energy data for the various substances considered here were taken mostly from NBS Technical Note 270: Parts 3 and 4 are given in Table I.

The equilibria which have been considered in this system are set out as follows together with the calculated free energy change and the pH-concentration equation.

$$
FeOH^{2+} + H^{+} = Fe^{3+} = H_2O
$$

\n
$$
\Delta G^{\circ} = -12.38 \text{ kJ}
$$

\n
$$
pH = 2.17 - \log^{Fe3+} / FeOH^{2+}
$$
 [1]

$$
Fe(OH)2+ + H+ = FeOH2+ + H2O
$$

$$
\Delta Go = -28.53 kJ
$$

$$
pH = 5.00 - \log FeOH^{2+}/Fe(OH)2
$$

Table I. Free Energy Data tot Species Considered In the Iron (111) Arsenate-Water System

Formula	$\Delta G^o_{f, 298.15 \text{ K}}$ kJ
$H2O$ (liq)	-237.18
Fe^{3+} $(m = 1)$	-4.60
$FeOH2+$ (<i>m</i> = 1)	-229.41
$Fe(OH),^+$ (<i>m</i> = 1)	-438.06
$Fe(OH)$, (c)	-696.64
H_3AsO_4 (<i>m</i> = 1)	-766.09
$H_2AsO_4^ (m = 1)$	-753.29
$HAsO42-$ (<i>m</i> = 1)	-714.71
$AsO43-$ (<i>m</i> = 1)	-648.52
$FeAsOa$ (c)	$-768.64*$
	$-776.42+$

* Calculated from Ref. 4 and used in calculating the stability diagrams.

t Calculated from Ref. 7.

$$
H_2AsO_4^- + H^+ = H_3AsO_4
$$

\n
$$
\Delta G^\circ = -12.80 \text{ kJ}
$$

\n
$$
pH = 2.24 - \log \frac{H_3AsO_4}{H_2AsO_4^-}
$$
 [3]

$$
HAsO42- + H+ = H2AsO4-
$$

\n
$$
\Delta Go = -38.58 \text{ kJ}
$$

\n
$$
pH = 6.76 - \log \frac{H_2AsO_4}{HAsO_4^{2-}}
$$
 [4]

As
$$
O_4^{3-}
$$
 + H⁺ = HAs O_4^{2-}
\n $\Delta G^{\circ} = -66.19 \text{ kJ}$
\n $\text{pH} = 11.60 - \log \frac{\text{HAsO}_4^{2-}}{\text{AsO}_4^{3-}} \tag{5}$

$$
Fe(OH)3 + 3H+ = Fe3+ + 3H2O
$$

\n
$$
\Delta G^{\circ} = -19.50 \text{ kJ}
$$

\n
$$
pH = 1.14 - 1/3 \log Fe^{3+}
$$
 [6]

$$
Fe(OH)3 + 2H+ = FeOH2+ + 2H2O
$$

\n
$$
\Delta G^{\circ} = -7.15 \text{ kJ}
$$

\n
$$
pH = 0.63 - 1/2 \log FeOH2+
$$
 [7]

$$
FeAsO4 + 3H+ = H3AsO4 + Fe3+
$$

\n
$$
\Delta Go = -2.05 \text{ kJ}
$$

\n
$$
pH = 0.12 - 1/3 \log H3AsO4
$$

\n
$$
-1/3 \log Fe3+
$$
 [8

$$
FeAsO4 + H2O + 2H+ = H3AsO4 + FeOH2+
$$

\n
$$
\Delta Go = 10.33 \text{ kJ}
$$

\n
$$
pH = -0.90 - 1/2 \log H3AsO4
$$

\n
$$
-1/2 \log FeOH2+
$$
 [9]

$$
FeAsO4 + H2O + H+ = H2AsO4- + FeOH2+
$$

\n
$$
\Delta Go = 23.14 \text{ kJ}
$$

\n
$$
pH = -4.05 - \log H2AsO4-
$$

\n
$$
- \log FeOH2+ [10]
$$

$$
FeAsO4 + 3H2O = Fe(OH)3 + H2AsO4- + H+
$$

$$
\Delta Go = 30.25 \text{ kJ}
$$

$$
pH = 5.30 + \log H2AsO4- [11]
$$

Taking the case of equal concentrations of the particular arsenic and iron ions which are obtained by the solution of $FeAsO₄$ Fig. 1 can be derived. It can readily be seen from this diagram that the minimum solubility of ferric arsenate lies at about $pH = 2.2$ and under the conditions of equal dissolved iron and arsenic this concentration is equal to approximately 10^{-3} m As (or 75 mg/l).

Also shown on Fig. 1 are the experimental solubility points of Tozawa and coworkers.^{6,7} Reasonably good agreement is revealed except for the points to the middle right of the diagram. In this region ferric arsenate has decomposed to ferric hydroxide which has probably not allowed equilibrium to be attained due to its coating of the original ferric arsenate particles. Tozawa's experiments allowed only 24 h for equilibration.

Fig.1—The ferric arsenate-water system showing also some of the [9] experimental analyses for both iron and arsenic from Ref. 6.

THE CALCIUM-ARSENATE-WATER **SYSTEM**

Free energy data were taken from NBS Technical Note 2705 Parts 3 and 6 and these data for calcium ions and compounds are given in Table II.

The equilibria which have been considered in this system are set out as follows together with the calculated free energy changes and the pH-concentration equations.

CaOH⁺ + H⁺ = Ca²⁺ + H₂O
\n
$$
\Delta G^{\circ} = -72.34 \text{ kJ}
$$
\n
$$
pH = 12.68 - \log \frac{Ca^{++}}{CaOH^+}
$$
\n[12]

$$
Ca(OH)2° + H+ = CaOH+ + H2O
$$

\n
$$
\Delta Go = -87.45 \text{ kJ}
$$

\n
$$
pH = 15.32 - \log \frac{CaOH+}{Ca(OH)2°}
$$
 [13]

$$
Ca(OH)2 + 2H+ = Ca2+ + 2H2O
$$

\n
$$
\Delta G^{\circ} = -129.37 \text{ kJ}
$$

\n
$$
pH = 11.33 - 1/2 \log Ca2+
$$
 [14]

$$
Ca(OH)2 + H+ = CaOH+ + H2O
$$

$$
\Delta Go = -57.02 kJ
$$

$$
pH = 10.0 - \log CaOH+
$$
[15]

$$
Ca(OH)2 = Ca(OH)2o
$$

\n
$$
\Delta Go = 30.42 kJ
$$

\n
$$
log Ca(OH)2o = -5.33
$$
 [16]

$$
Ca_3(AsO_4)_2 + 4H^+ = 2H_2AsO_4^- + 3Ca^{2+}
$$

\n
$$
\Delta G^\circ = -104.10 \text{ kJ}
$$

\n
$$
pH = 4.56 - 1/2 \log H_2AsO_4^-
$$

\n
$$
-3/4 \log Ca^{2+}
$$
 [17]

$$
Ca3(AsO4)2 + 2H+ = 2HAsO42- + 3Ca2+
$$

\n
$$
\Delta Go = -26.94 \text{ kJ}
$$

\n
$$
pH = 2.36 - \log HAsO42-
$$

\n
$$
- 3/2 \log Ca2+
$$
 [18]

$$
Ca_3(AsO_4)_2 = 2AsO_4^{3-} + 3Ca^{2+}
$$

\n
$$
\Delta G^\circ = 105.44 \text{ kJ}
$$

\n
$$
\log AsO_4^{3-} = -9.24 - 3/2 \log Ca^{2+}
$$
 [19]

$$
Ca_3(AsO_4)_2 + 3H_2O = 2AsO_4^{3-} + 3CaOH^+ + 3H^+
$$

\n
$$
\Delta G^\circ = 322.42 \text{ kJ}
$$

\n
$$
pH = 18.83 + 2/3 \log AsO_4^{3-}
$$

\n
$$
+ \log CaOH^+
$$
 [20]

$$
Ca_3(AsO_4)_2 + 6H_2O = 2AsO_4^{3-} + 3Ca(OH)_2 + 6H^+
$$

\n
$$
\Delta G^\circ = 493.46 \text{ kJ}
$$

\n
$$
pH = 14.41 + 1/3 \log AsO_4^{3-}
$$
 [21]

Figure 2 has been derived for equal concentrations of the appropriate calcium and arsenic ions (very little difference is obtained if the ratio of the calcium ion to the arsenic ion is taken as 3/2 which is the solution product ratio).

When Tozawa's results for the solubility of calcium arsenate are compared to this diagram the agreement deviates by the "splitting" of the experimental values of arsenic and calcium in solution with the ion equivalent of arsenic about sixteen times the ion equivalent of calcium. This is certainly due to the formation of another solid compound which is very likely to be a carbonate formed due to the presence of CO₂ in the atmosphere.

Fig.2--The calcium arsenate-water system. Experimental points are taken from Ref. 6.

THE CALCIUM ARSENATE-CARBONATE-WATER SYSTEM

Stability diagrams for this system were drawn up using the additional free energy data from NBS Technical Note⁵ Parts 3 and 6 and given in Table III and the equilibria set out below.

$$
CaCO3 + H+ = Ca2+ + HCO3-
$$

$$
\Delta Go = -12.80 \text{ kJ}
$$

 pH = 2.24 - log Ca²⁺ - log HCO₃⁻ [22]

$$
CaCO3 = Ca2+ + CO32-\n\Delta Go = 46.15 kJ\nlog Ca2+ = -8.08 - log CO32- [23]
$$

$$
CaCO_3 + H_2O = CaOH^+ + \log CO_3^{2-} + H^+
$$

\n
$$
\Delta G^o = 119.75 \text{ kJ}
$$

\n
$$
pH = 20.98 + \log CaOH^+ + \log CO_3^{2-}
$$

$$
[24]
$$

$$
Ca_3(Aso_4)_2 + 3HCO_3^- = 3CaCO_3 + 2HAsO_4^{2-} + H^+ \n\Delta G^\circ = 7.70 \text{ kJ} \n\text{pH} = 1.35 + 2 \log HAsO_4^{2-} \n- 3 \log HCO_3^-
$$
\n[25]

$$
Ca_3(AsO_4)_2 + 3CO_3^{2-} + 2H^+ = 3CaCO_3 + 2HAsO_4^{2-}
$$

\n
$$
\Delta G^\circ = -169.16 \text{ kJ}
$$

\n
$$
pH = 14.80 - \log HAsO_4^{2-}
$$

\n+ 3/2 log CO₃²⁻

$$
Ca_3(AsO_4)_2 + 3CO_3^{2-} = 3CaCO_3 + 2AsO_4^{3-}
$$

$$
\Delta G^\circ = -36.78 \text{ kJ}
$$

$$
\log AsO_4^{3-} = 3.22 + 3/2 \log CO_3^{2-} \quad [27]
$$

$$
Ca(OH)2 + CO32 - + 2H+ = CaCO3 + 2H2O
$$

$$
\Delta G^{\circ} = -176.77 \text{ kJ}
$$

$$
pH = 15.49 + 1/2 \log CO32 -
$$

[28]

[26]

Figure 3 was drawn up for a carbonate or bicarbonate activity of 10^{-4} molal which is an appropriate value for a closed system which has initially been in contact with the atmosphere at pH below 7. This would have been a likely condition for Tozawa's experiments which were carried out over a 24 h period and not equilibrated with the atmosphere where $p_{\text{CO}_2} = 10^{-3.52}$ atmospheres.

Figure 4 was calculated for the system in equilibrium with $CO₂$ in the atmosphere where:

(a)
$$
\log[H_2CO_3^\circ] = -4.97
$$

- (b) $\log[\text{HCO}_3^-] = pH 11.35$
- (c) $\log [CO_3^{2-}] = 2pH 21.68$

This diagram shows that under these conditions the minimum solubility of $Ca₃(AsO₄)$, is at a pH of 7.8 and is equivalent to an arsenic concentration of approximately 0.5 g/1.

THE MAGNESIUM ARSENATE-CARBONATE-WATER SYSTEM

Free energy data were taken from NBS Technical Note 2705 Parts 3 and 6 and those data for magnesium ions and compounds are given in Table IV.

MgOH⁺ + H⁺ = Mg²⁺ + H₂O
\n
$$
\Delta G^{\circ} = -65.23 \text{ kJ}
$$
\n
$$
pH = 11.43 - \log \frac{Mg^{2+}}{MgOH^{+}}
$$
\n[29]

$$
Mg(OH)2° + H+ = MgOH+ + H2O
$$

\n
$$
\Delta Go = -94.52 kJ
$$

\n
$$
pH = 16.56 - \log \frac{MgOH+}{Mg(OH)2} \qquad [30]
$$

$$
Mg(OH)2 + 2H+ = Mg2+ + 2H2O
$$

\n
$$
\Delta G^{\circ} = -95.60 \text{ kJ}
$$

\n
$$
pH = 8.37 - 1/2 \log Mg^{2+}
$$
 [31]

Table III. Free Energy Data for the Additional Species Considered in the Calcium Arsenate-Carbonate-Watar System

Formula	$\Delta G_{f, 298.15 \text{ K}}^{o}$ kJ
H, CO_3° (<i>m</i> = 1)	-623.16
HCO_3^- (<i>m</i> = 1)	-586.85
CO_3^{2-} (<i>m</i> = 1)	-527.90
$CaCO3$ (c) calcite	-1128.84

Table IV. Free Energy Data for the Species Considered In the Magnesium Arsenate-Water System

Formula	$\Delta G_{f, 298.15 \text{ K}}^{o}$ kJ
Mg^{2+} $(m = 1)$	-454.80
$MgOH^{+}$ $(m = 1)$	-626.76
$Mg(OH)_{2}^{o}$ $(m = 1)$	-769.44
$Mg(OH)$, (c)	-833.58
$Mg_3(AsO_4)$, (c)	$-2773.78*$
$MgCO$, (c)	-1012.11
$MgCO_3 \cdot 3H_2O$ (c)	-1726.32
$4MgCO_3 \cdot Mg(OH)_2 \cdot 4H_2O$ (c)	-5864.171
$Mg(OH)$, $MgCO$, $3H$, O	-2568.35
$MgHCO3$ (<i>m</i> = 1)	-1049.68 ⁺
$MgCO3$ (m = 1)	-1003.53 ⁺
*Calculated from Ref. 4.	
† Reference 8.	

 \ddagger Reference 9.

Fig.3--The calcium arsenate-carbonate-water system at a carbonate/bicarbonate activity of 10 -4 . Experimental points are taken from Ref. 6.

$$
Mg(OH)2 + H+ = MgOH+ + H2O
$$

\n
$$
\Delta Go = -30.38 kJ
$$

\n
$$
pH = 5.32 - \log MgOH+
$$
 [32]

$$
Mg(OH)2 = Mg(OH)2°
$$

\n
$$
\Delta G^{\circ} = 64.14 \text{ kJ}
$$

\n
$$
\log Mg(OH)2° = -11.24
$$

\n
$$
Mg(OH)2 + 4Ht = 3H A D = 1.2M2^{2}t
$$
 [33]

$$
Mg_3(AsO_4)_2 + 4H^+ = 2H_2AsO_4^- + 3Mg^{2+}
$$

\n
$$
\Delta G^\circ = -97.19 \text{ kJ}
$$

\n
$$
pH = 4.26 - 1/2 \log H_2AsO_4^- - 3/4 \log Mg^{2+}
$$
 [34]

$$
Mg_3(AsO_4)_2 + 2H^+ = 2HAsO_4^{2-} + 3Mg^{2+}
$$

\n
$$
\Delta G^\circ = -20.04 \text{ kJ}
$$

\n
$$
pH = 1.76 - \log HAsO_4^{2-}
$$

\n
$$
-3/2 \log Mg^{2+}
$$
 [35]

$$
Mg_3(AsO_4)_2 + 6H_2O = 3Mg(OH)_2
$$

+ 2H AsO₄²⁻ + 4H⁺

$$
\Delta G^{\circ} = 266.73 \text{ kJ}
$$

pH = 11.6 8 + 1/2 log HAsO₄²⁻
[36]

Fig.4--The calcium arsenate-carbonate-water system at a carbonate/bicarbonate activity in equilibrium with atmospheric carbon $\text{dioxide}(p_{\text{CO}_2} = 10^{-3.52} \text{ ats.})$ Experimental points are taken from **Ref. 6.**

$$
Mg(OH)_2 \cdot MgCO_3 \cdot 3H_2O + 3H^+
$$

= 2Mg²⁺ + HCO₃⁻ + 5H₂O

$$
\Delta G^{\circ} = -114.01 \text{ kJ}
$$

pH = 6.66 - 2/3 log Mg²⁺ - 1/3 log HCO₃⁻

$$
Mg(OH)2 · MgCO3 · 3H2O
$$

= 2 Mg(OH)₂ + CO₃²⁻ + H₂O + 2H⁺

$$
\Delta G^{\circ} = 136.11 \text{ kJ}
$$

$$
pH = 11.92 + 1/2 \log CO_3^{2-}
$$
 [38]

$$
2Mg_3(AsO_4)_2 + 3HCO_3^- + 15H_2O
$$

= 3[Mg(OH)₂ · MgCO₃ · 3H₂O]
+ 4HAsO₄²⁻ + 5H⁺

$$
\Delta G^{\circ} = 301.96 \text{ kJ}
$$

pH = 10.58 - 3/5 log HCO₃⁻ + 4/5 log HAsO₄²⁻

$$
4MgCO3 \cdot Mg(OH)2 \cdot 4H2O + 6H+
$$

= 5Mg²⁺ + 4HCO₃⁻ + 6H₂O

$$
\Delta Go = -180.33 \text{ kJ}
$$

$$
pH = 5.27 - 2/3 \log HCO3- - 5/6 \log Mg2+
$$

$$
5Mg_3(AsO_4)_2 + 12HCO_3^- + 18H_2O
$$

= 3[4MgCo₃ · Mg(OH)₂ · 4H₂O]
+ 10HAsO₄²⁻ + 8H⁺

$$
\Delta G^\circ = 440.74 \text{ kJ}
$$

$$
pH = 9.65 + 5/4 \log HAsO42- - 3/2 \log HCO3-
$$

[41]

[40]

$$
4MgCO_3 \cdot Mg(OH)_2 \cdot 4H_2O + 2H^+
$$

= 5Mg²⁺ + 4CO₃²⁻ + 6H₂O

$$
\Delta G^{\circ} = 55.48 \text{ kJ}
$$

pH = -4.86 - 2 log Co₃²⁻ - 5/2 log Mg²⁺ [42]

$$
5Mg_3(AsO_{4})_2 + 12CO_3^{2-} + 18H_2O + 4H^+
$$

= 3[4MgCo₃ \cdot Mg(OH)₂ \cdot 4H₂O]
+ 10HAsO₄²⁻

$$
\Delta G^{\circ} = -266.69 \text{ kJ}
$$

pH = 11.68 - 5/2 log HAsO₄²⁻ + 3 log CO₃²⁻ [43]

Figure 5 is a stability diagram for the magnesium arsenate-carbonate-water system at a carbonate (or bicarbonate) activity of $10⁻³$ molal. This activity of carbonate was chosen to approximate a closed system which had initially been in equilibrium with the $CO₂$ in air (10^{-3.52} atmospheres) at a pH less than neutral, so that Tozawa's results could be explained. Under these conditions the basic carbonate which is stable at pH greater than 9.3 is the compound $Mg(OH)_{2}$. MgCO₃. \cdot 3H₂O (antinite). Tozawa's results fit well into the stability region of this compound. A carbonate activity of 10^{-4} molal would result in a decrease in this stability region to the point of disappearance.

Figure 6 is a stability diagram for the same system but under conditions where equilbrium exists with the carbon dioxide in the atmosphere ($p_{CO_2} = 10^{-3.52}$ atmospheres). Under these conditions it can be shown that the basic carbonate which is stable is the compound $4MgCO₃ \cdot Mg(OH)₂ \cdot 4H₂O$ (hydromagnesite), over a very wide range of pH. Minimum solubility of $Mg_3(AsO₄)₂$ is at pH = 9.3 equivalent to an arsenic concentration of approximately 75 mg/l.

CONCLUSIONS

Stability diagrams for iron (111) arsenate, calcium arsenate, and magnesium arsenate, which have been calculated using reliable free energy data, reveal that the solubility of these compounds is significantly higher in the region of minimum solubility than has been

Fig. 5-The magnesium arsenate-carbonate-water system at a carbonate/bicarbonate activity of 10^{-3} . Experimental points are taken from Ref. 6.

Fig. 6-The magesium arsenate-carbonate-water system at a carbonate/bicarbonate activity in equilibrium with atmospheric carbon dioxide ($p_{CO_2} = 10^{-3.52}$ ats.) Experimental points are taken from Ref. 6.

widely assumed. The presence of carbonate at very low concentration has a large affect in greatly increasing the minimum solubility of calcium arsenate by approximately two orders of magnitude, and also significantly affecting the solubility of magnesium arsenate.

The experimental results which have been published by Tozawa and coworkers^{6,7} show fairly close agreement with the calculated diagrams but an accurate equilibrium study of the systems involved would be appropriate.

The significance of the higher than recognized solubilities of metal arsenates leads to important considerations in relation to the disposal of arsenic containing residues from metal extraction processes. The misuse of published solubility products has led to the widely accepted conclusion that metal arsenates (under specific conditions of pH) are extremely insoluble. In this work an attempt has been made to take into account all of the reactions which could influence the solubility of each of the three arsenates considered. Some of the equilibria which are presented in the equations above do not appear as lines on the stability diagrams but are important in attaining the end result. Other equilibria for which equations are not presented were also considered.

The equilibrium considerations reported here relate to a simple system consisting of pure components. In relation to the mobility of arsenic in ground waters associated with residues from metal extraction processes this chemical model is too simple by far and should be considered together with all of the possible heterogeneous interactions in the natural aquatic environment including adsorption and organic complexing. Holm *et al* and Gulens *et al* have discussed various aspects of the occurrence and mobility of arsenic in gound waters.¹⁰ Chemical modelling in such systems is a field which is evolving rapidly and one in which the hydrometallurgist must become more active.

ACKNOWLEDGMENT

This work was carried out at the U.S. Bureau of Mines Research Center, Salt Lake City, Utah, in September 1979, while the author was visiting that laboratory under a U.S./Australia Agreement for Scientific and Technical Co-operation.

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