Determination of Polycyclic Aromatic Hydrocarbons, Azaarenes, and Thiaarenes Emitted from Coal-Fired Residential Furnaces by Gas Chromatography]Mass Spectrometry

G. Grimmer, J. Jacob, G. Dettbarn, and K.-W. Naujack

Biochemisches Institut fiir Umweltcarcinogene, Sieker LandstraBe 19, D-2070 Ahrensburg, Federal Republic of Germany

Bestimmung von polycyclischen aromafischen Kohlenwasserstoffen, Azaarenen und Thiaarenen in der Emission von Kohle-beheizten Haushaltöfen **Bestandsanfnahme durch GC]MS**

Zusammenfassung. Da die krebserzeugende Wirkung von Emissionen aus Steinkohlebrikett-beheizten Zimmeröfen nahezu vollständig durch die im Kondensat enthaltenene polycyclischen aromatischen Verbindungen (PAC) mit mehr als 3 Ringen verursacht wird, wurde die Zusammensetzung dieser PAC-Fraktion untersucht.

Mehr als 170 neutrale [111 polycyclische aromatische Kohlenwasserstoffe (PAH), 57 Thiaarene, 6 Oxaarene] sowie 70 basische Verbindungen (Azaarene) konnten mit der GC/MS-Kombination als PAC charakterisiert werden (Nachweisgrenze 0,1 mg/kg Brikett). Etwa 80 Verbindungen wurden durch den Vergleich mit Referenzsubstanzen identifiziert. Um in einigen zweifelhaften Fällen die Identität der isolierten mit der synthetischen Verbindung zu bestätigen, wurden bei 11 PAH mit einem Molgewicht von 302 zusätzlich auch UV- und Fluorescenzspektren verglichen [Naphtho(l,2-k)fluoranthen, Dibenz(e,k)acephenanthrylen, Naphth(2,3-e)acephenanthrylen, Naphtho(2,3-k)fluoranthen, Dibenzo(de,qr)naphthacen, Coronen, Dibenzo(fg,op) naphthacen, Naphtho(1,2,3,4-def)chrysen, Benzo(b)perylen, Dibenzo(def, j)chrysen, Benzo(rst)pentaphen]. Diese Verbindungen sind von besonderem Interesse, da PAC mit 6 und mehr Ringen im Tierversuch etwa 50% der carcinogenen Wirkung des Steinkohlenbrikett-Emissionskondensates verursachen.

Summary. Since the carcinogenic effect of emissions from hard coal briquet-fired furnaces is almost entirely caused by polycyclic aromatic compounds (PAC) with more than three rings, the composition of this fraction has been analyzed in detail

More than 170 neutral [111 polycyclic aromatic hydrocarbons (PAH), 57 thiaarenes, 6 oxaarenes] and 70 basic (azaarenes) compounds were characterized as PAC by means of GC/MS (limit of detection 0.1 mg/kg briquet). About 80 compounds were identified by comparison with reference standards. In order to obtain unequivocal evidence for the identification of 11 PAH with a molecular weight of 302, the UV- and fluorescence spectra were compared with those of the synthesized reference compounds [naphtho(1,2-k) fluoranthene, dibenz(e,k)acephenanthrylene, naphth-(2,3 e)acephenanthrylene, naphtho(2,3-k)fluoranthene, dibenzo-

Offprint requests to: Prof. Dr. G. Grimmer

(de,qr)naphthacene, coronene, dibenzo(fg,op)naphthacene, naphtho(l,2,3,4-def)chrysene, benzo(b)perylene, dibenzo- (-def,j)chrysene, benzo(rst)pentaphene]. In case of the emission condensate from hard coal briquets, PAC containing 6 and more rings cause about 50% of the carcinogenic effect in animal experiments.

Introduction

The condensate emitted from hard coal-fired residential furnaces provoke local tumors, exhibiting a clear cut doseresponse relationship after repeated application to the dorsal skin of mice [1]. To identify the substances chiefly responsible for the carcinogenic effect of this emission, the condensate was separated into a fraction free of polycyclic aromatic compounds (PAC) and into various PAC-containing fractions. The carcinogenic potency of all these fractions have been compared with that of the unseparated sample of the emission condensate by testing them in different doses onto the dorsal skin of mice. It should be noted that the fraction of PAC consisting of more than 3 rings account for almost the total carcinogenicity [1]. Subfractions containing polycyclic aromatic hydrocarbons (PAH), thiaarenes (S-PAC) and oxaarenes (O-PAC) with 4 and 5 rings as well as 6 and more rings were almost equally active, each contributing by about 50% to the total carcinogenicity [2], whereas the subfraction containing azaarenes (N-PAC) and nitroarenes (NO₂-PAH) accounted for only $4-7%$ of the total carcinomas. In contrast to these carcinogenic fractions, the PAC-free fraction and that containing PAC with 2 and 3 rings (together about 77% by wt) were almost inactive.

Consequently, the objective of this investigation was to identify or to characterize PAH and their nitrogen-, sulfur-, and oxygen-containing heterocyclic analogues (N-PAC; S-PAC; O-PAC) occurring in hard coal briquets emission condensate which was used for the animal experiments [1, 21.

In Germany as well as in other countries, most of the air suspended PAC originate from coal-fired residential furnaces, coke plants or gasoline- and diesel-driven vehicles, whereas oil- and gas-fired hearings as well as coal-fired power plants contribute only very small amounts to the air pollution by PAC. A comparison of the benzo(a)pyrene emission from 5 sources (in total 18 t during 1981) shows that about 56% of the benzo(a)pyrene results from coal-fired residential furnaces and minor consumers, whereas coke production and passenger cars contribute by about 31% and 13% to the total amount. In contrast to that, coal-fired power plants and residential oil hearings contribute by only 0.006 and 0.1%, respectively [3]. This confirms a previous investigation [4, 5] in which the PAC-mass concentrations where monitored during 50 weeks in various areas of a city (Essen, FRG). The areas characterized predominantly by domestic coal heatings, coke oven, or automobile traffic (tunnel) showed by far the highest PAC-concentrations, in contrast to the low concentration of the oil-heating area. The PACprofile analysis from the oil-heating area exhibited a better correlation with coal emission sources than with oil, which indicates that less than one third of the PAC-concentration in this ara originates from oil-heatings [4, 5].

The PAH-emissions from hard coal- or hard coal briquet-fired residential furnaces were investigated by various authors $[6-15]$. The PAH-mass concentration in the flue gas varies widely, depending on the nature of coal e.g. anthracite nuts, bitumen- or pitch-pound hard coal briquets, broken coke and other special coals. The mass emission of benzo(a)pyrene and other PAH varies under comparable operation conditions and furnaces, the average values were found for anthracite nuts (0.04 mg/kg), for hard coal briquets $(2.8-4.5 \text{ mg/kg})$, and for brown coal briquets $(3.0-4.2 \text{ mg/kg})$ [3, 16]. It furthermore depends on the type of residential furnace, such as slow combustion stove (Durchbrandofen) and Universal slow combustion stove (Universaldauerbrenner) [3] as well as on the operating conditions (full, medium or low load) regulated by the air input [17].

Moreover, the kind of binder for the hard coal briquets significantly influences the PAC-mass emission. Using the most popular type of slow burning residential furnace in Germany (Durchbrandofen), the pitch-bound briquets produce about 200 mg benzo(a)pyrene (BaP) per kg in contrast to bitumen-bound briquets manufactured from the same hard coal, emitting only about 5 mg BaP/kg burned under the same operation conditions (DIN I8890) [15]. This confirms the results of Beine [6] who reported an emission of 380 mg BaP/kg for pitch-bound hard coal briquets. Since the federal regulation for prevention of emissions (Verordnung fiber Feuerungsanlagen, I. BImSchV) became effective in 1974, the production of pitch-bound briquets was stopped.

In contrast to brown coal-fired residential furnaces [17], no data on the total profile of the flue gas particle-bound PAC emitted by hard coal-fired residential stoves are available.

In the flue gas, the equilibrium between particle-bound and gaseous PAC depends on the temperature. At outdoor temperatures the PAH with more than 3 rings are predominantly bound to particles. Therefore, PAC with boiling points below 400°C (fluorene, phenanthrene, anthracene, fluoranthene, pyrene and their methyl derivatives) are incompletely collected on particulate filters such as siliconebound glass fibre filters. On the other hand, these low boiling compounds re-evaporate readily from the filter, when trapping particulates by high volume air samplers operating with high-flow air stream velocities. Furthermore, the reactivity of various PAC, such as benz(a)anthracene, benzo(a)pyrene, anthanthrene, and cyclopenta(cd)pyrene already trapped on the filter, has to be taken into account, e.g. the destruction by UV-light, oxidation (SO_3, O_3) and

nitration (NO_x) . To perceive both effects, the chemical destruction of various PAC on the particle filter and the reevaporation of already collected compounds, two criteria can be used:

- Reproducibility of the sampling process: with identical mass streams per hour or simultaneous sampling using two different filter systems at the same location, the collection of PAC must result in identical profiles and in the same masses.

 $-$ Invariability of the PAC profiles during various sampling periods: variability of the PAC profile is a criterion for destruction and/or blow-off of the PAC already trapped on the filter due to instability or volatility of certain individual compounds [16].

The collection arrangement applied for this investigation meets these requirements [15, 17].

Experimental Section

Collection System

The condensate of the emission during combustion of hard coal briquets (RA-briquets, 24 g, Sophia-Jacoba, in total 3.66 kg) was obtained from Bergbau-Forschung GmbH, Essen, FRG. The nominal operation output of the slow combustion furnace (Franck'sche Eisenwerke AG, Dillenburg, FRG; Durchbrandofen type Oranier 4312 DB) was 20.93 MJ/h , 5.8 kW , respectively. The furnace is connected with a smoke stack, 4.2 m in length and terminates in a change-over flap. The total testbench comprising furnace and the collection arrangement has been published previously [15]. The collecting system consisting of a glass cooler and a glass fibre filter (area ca. 1 m^2 , Drägerwerk AG, Lübeck, FRG; type WB 50) separates all PAC with boiling points above 380°C (collection efficiency for particles of $0.3 - 0.5$ µm was 99.998%). The repeatability of the trapping procedure was checked in separate experiments [15, 17]. The experimental conditions were in accordance with the regulation of DIN 18 890. First, the basic glowing fire was prepared and continued for 2.5 h under full load operation conditions to standardize the state of the test-bench. During this time, the flue gas was directly released to the atmosphere.

For the main experiment the glowing fire was stoked, then the coal filled in (3.66 kg), and finally the flue gas led onto the collecting system by switching the change-over flap. The flue gas passed the glass cooler (25°C) and subsequently two parallel arranged filters (Micronfilter WB 50) followed by a control filter of the same type. The volume of the flue gas re-heated to 45° C was measured and the concentrations of O_2 , CO_2 , CO , NO_x , SO_2 and C_nC_m were determined. Furthermore, sub-atmospheric pressure was obtained by means of an automatic blower. Details on the concentration of the above flue gases have been reported elsewhere [15].

The condensate consisted of three parts: (a) particulate filter on the micron filter, (b) acetone glass cooler rinsing, and (c) condensed water originating from the combustion, which contains only very small amounts of PAC.

For the animal experiments [1, 2] the condensates of three operations 3×3.66 kg fuel) were used.

Methods

A small part of the condensate and the fractions (about 1%) used for the bioassay or alternatively, one thirtieth of the glass fibre filter from a separate experiment were required for the analytical procedure.

Extraction. 1/30 of the filter (6 ribs out of 180) was extracted twice with boiling toluene (300 ml) for 30 min (with or without internal standard) and the extract separated from particles by filtration. The particles were extracted again with methanol and aqueous ammonia (pH 10) and then filtered again. The combined solutions were evaporated by a rotary evaporator under reduced pressure (bath temperature 35° C).

Liquid-Liquid Partition. The residue dissolved in cyclohexane was extracted by methanol and aqueous ammonia (0.1 N) $(1 + 1)$ and washed with water. The cyclohexane phase was evaporated to about 10 ml, isopropanol (30 ml) added and evaporated again to about 5 ml.

Chromatography on Sephadex LH 20. The solution (5 ml) of the liquid-liquid partition was transferred to a Sephadex LH 20 column (50 g) and the non-aromatic compounds together with the PAC with 2 rings were eluted with isopropanol (175 ml). This fraction was discarded. PAC with more than two rings were eluted with acetone (500 ml). The solvent was evaporated to few milliliters and then a mixture of methanol and benzene $(5 + 1, 20$ ml) was added. Finally this solution was evaporated to 5 ml.

Separation by SP-Sephadex C 25. SP-Sephadex C 25 (8 g) was activated with aqueous hydrochloric acid (0.05 N, 400 ml) and subsequently washed with methanol/water $(1 + 1, 500 \text{ ml})$ and methanol (500 ml).

The methanol/benzene concentrate from the chromatography on Sephadex LH 20 was transferred to the column and the neutral PAC (PAH, S-PAC, O-PAC) were eluted with methanol (400 ml). Basic compounds (azaarenes) were eluted with a solution of ammonium chloride (5 N), aqueous ammonia (2.5 N), and methanol $(3 + 2 + 5, 300 \text{ ml})$ and this solution after dilution with water (300 ml) was extracted with cyclohexane (600 ml).

Chromatography on Sifiea Gel. The methanol residue from the separation by SP-Sephadex C 25 containing the neutral PAC was dissolved in cyclohexane (3 ml) and chromatographed on silica gel (10 g, 10% water content) by elution with cyclohexane (200 ml).

Fractionation of PAC with 3 and more Rings on Sephadex LH 20. To facilitate the interpretation of the capillary gas chromatographic and mass spectrometric results, the neutral PAC (PAH, S-PAC, O-PAC) were separated by Sephadex LH 20 (50 g) as described above, using isopropanol (190– 385 ml for PAC with $3-5$ rings; $385-1,000$ ml for PAC with more than 5 rings) and acetone (500 ml, for high condensed PAC).

Isolation of S-PAC. To facilitate the interpretation of the gas chromatographic and mass spectrometric data of the S-PAC, one part of the neutral PAC fraction from the separation by SP-Sephadex C 25 was oxidized. The resulting thiaarenesulfones were then separated from the non-polar PAH and O-PAC. $-$ To this end the neutral fraction was dissolved in benzene, pure acetic acid and H_2O_2 (30%) $(5 + 5 + 1)$, refluxed for 5 h, and subsequently diluted with water $(1 + 1)$. The benzene phase was collected, the aqueous phase extracted twice with dichloromethane and both phases washed with water before evaporation.

Separation of the S-PAC from the PAC-Mixture. The combined residues of both phases, dissolved in cyclohexane were transferred to a silica gel column (5 g, 10% water content), the PAH were eluted with cyclohexane (80 ml) and the thiaarenesulfones with a methanol-benzene mixture $(1 + 1, 80 \text{ ml})$. To the residue of the methanol-benzene mixture dissolved in ether-benzene $(4 + 1)$, LiAlH₄ (500 mg) was added and the mixture refluxed for 3 h, avoiding contact to air. Thiaarenes were then extracted with benzene, evaporated, the residue dissolved in cyclohexane, and transferred to a silca gel column (5 g, 10% water content) and the S-PAC eluted with cyclohexane (80 ml).

Further purification of the N-PAC as well as the S-PAC for capillary gas chromatography was carried out by chromatography on Sephadex LH 20 (10 g) with isopropanol $(0-38 \text{ ml discarded}; 38-200 \text{ ml collected})$.

Conditions for Gas Chromatography. Apparatus: Perkin-Elmer, model Sigma 2B. Columns: fused silica column $25 \text{ m} \times 0.32 \text{ mm}$ coated with methyl(94%)-phenyl(5%)vinyl(l%)-polysiloxane (SE 54, Oribond). Carrier gas: Helium 2 ml/min. Splitless injected at 115° C, split opened after 5 min; 6 min at 115°C followed by a $115\textdegree C - 180\textdegree C$ with 30° C/min, subsequently 180° C -270° C with 3° C/min temperature program (for S-PAC), then isothermically; 115° C -160° C with 30° C/min, subsequently 160° C 270° C with 3° C/min (for N-PAC). Integrator: SES GmbH, model Trilab 2002.

Conditions of Mass Spectrometry. Apparatus (PAH, S-PAC, O-PAC): Finnigan MAT GC/MS, model 5100 with fused

Fig. 1. Scheme of enrichment of PAH, S-PAC, O-PAC, and N-PAC

Fig. 2. Fraction of neutral PAC (PAH, S-PAC, O-PAC) of the emission from briquet-fired furnaces recorded as total ion current. Numbers correspond to those of Table 1. GC/MS conditions see "Methods"

Table 1 (continued)

a Detected also in automobile exhaust [18]

As can be seen from the table, in various cases peaks listed under one number consist of more than one compound

silica column $30 \text{ m} \times 0.25 \text{ mm}$ methyl(95%)-phenyl(5%)polysiloxane (SE 54). The capillary column was connected directly with the electron-impact ion source (70 eV, 200° C). Apparatus (N-PAC): Varian MAT, model 112S. The gas chromatograph of this model is replaced by one from Perkin-Elmer, model F 22 with fused silica column 60 m \times 0.32 mm coated with polydimethylsiloxane (SE 30).

Figure 1 demonstrates the scheme of enrichment for PAH, S-PAC, O-PAC, and N-PAC.

Results and Discussion

On the basis of a detection limit of 0.1 mg/kg burned hard coal briquets, more than 170 neutral PAC were detected by GC/MS. Figure 2 shows the gas chromatogram of this fraction which consists of PAH, S-PAC, and O-PAC recorded as total ion current.

The identification and quantitative composition of this PAC-mixture is presented in Table 1.74 Compounds named in Table I were identified by comparing their GC-retention times and their mass spectrometric fragmentation with those of reference standards.

To confirm the identification in some cases eleven out of these compounds with molecular weights of 300/302 were characterized also by UV and fluorescence spectrometry [19]. To isolate these compounds from the condensate, the fraction containing PAC with more than 3 rings was separated by vacuum sublimation (short distance distillation) under reduced pressure (10^{-8} bar) at 105° C during 6 h into volatile PAC of less than 5 rings (mass < 300), and into a non-volatile residue (mass \ge 300). The latter, containing PAC with 6 and more rings, was separated by repeated highperformance TLC on RP-18 plates using various solvents. This procedure resulted in 20 compounds. Subsequent chromatography by Sephadex LH 20 using isopropanol for elution, in some cases led to PAC in almost pure state as

Fig. 3. A Naphtho $(1,2-k)$ fluoranthene (No. 109, Table 1); B Dibenz(e,k)acephenanthrylene (naphtho(2,3-b)fluoranthene) (No. 111); C Naphth(2,3-e)acephenanthrylene (dibenzo(b,k)fluoranthene) (No. 111); D Naphtho(2,3-k)fluoranthene (No. 113); E Dibenzo(de,qr)naphthacene (naphtho(2,3-e)pyrene) (No. 116); F Coronene (No. 117); G Dibenzo(fg,op)naphthacene (dibenzo(e,l) pyrene) (No. 118); H Naphtho $(1,2,3,4$ -def)chrysene (dibenzo (a,e) pyrene) (No. 118); I Benzo(b)perylene (dibenzo(fg,qr)naphthacene) (No. 119); J Dibenzo(def, j) chrysene (naphtho $(2,1-a)$ pyrene) (No. 119); K Benzo(rst)pentaphene (dibenzo(a,i)pyrene) (No. 119)

indicated by GC, HPLC and UV analysis. Based on the GC retention times, mass-, UV-, and fluorescence spectra [19] in total eleven of these isolated PAH were found to be identical with reference PAH synthesized for this purpose. Their structures are presented in Fig. 3.

Five compounds with the mass of 300/302, although characterized by their UV- and fluorescence spectra, could not be positively identified due to the lack of reference materials.

To facilitate the interpretation of the gas chromatographic data, a part of the neutral PAC-eontaining fraction was oxidized and the resulting thiaarenesulfones were separated from the non-polar PAH and O-PAC. Subsequent

Fig. 4. Fraction of thiaarenes (S-PAC) separated from the neutral PAC (see Fig. 2). Numbers correspond to those of Table 1. GCconditions see "Methods"

Fig. 5. Fraction of azaarenes (N-PAC) recorded by FID. Numbers correspond to those of Table 2. GC-conditions see "Methods"

Table 2. Identification and characterization of basic nitrogen containing polycyclic aromatic compounds from coal emission

| No. | RT -min | Mass | Compound |
|----------------|----------------|------------|---------------------------------------|
| 1 | 15.23 | 179 | 4-Azaphenanthrene (Benzo(h)quinoline) |
| \overline{c} | 15.47 | 179 | Acridine |
| 3 | 16.10 | 179 | 1-Azaphenanthrene (Benzo(f)quinoline) |
| | 16.10 | 179 | 9-Azaphenanthrene (Phenanthridine) |
| 4 | 16.70 | 193 | Methyl-179 |
| 5 | 16.87 | 179 | Aza-178 |
| 6 | 17.23 | 193 | Methyl-179 |
| | 17.23 | 199 | n.i. |
| 7 | 17.50 | 199 | n.i. |
| 8 | 17.93 | 193 | Methyl-179 |
| 9 | 18.17 | 193 | Methyl-179 |
| 10 | 18.33 | 193 | Methyl-179 |
| 11 | 18.63 | 193 | Methyl-179 |
| 12 | 18.83 | 193 | Methyl-179 |
| 13 | 19.07 | 193 | Methyl-179 |
| 14 | 19.20 | 193 | Methyl-179 |
| 15 | 19.37 | 213 | Methyl-199 |
| 16 | 20.77 | 213 | Methyl-199 |
| 17 | 20.93 | 205 | Aza-204 |
| | 20.93 | 207 | Dimethyl-179 |
| 18 | 22.83 | 203 | Aza-202 |
| | | | 1-Azafluoranthene |
| 19 | 23.37 | 203 203 | |
| 20 | 23.57 | | Aza-202 |
| 21 | 24.03 | 219 | Methyl-205 |
| 22 | 24.30 | 219 | Methyl-205 |
| 23 | 24.93 | 203 | 1-Azapyrene |
| | 24.93 | 219 | Methyl-205 |
| 24 | 25.27 | 203 | Aza-202 |
| | 25.27 | 217 | Methyl-203 |
| 24 A | 25.73 | 203 | 4-Azapyrene |
| 25 | 26.70 | 217 | Methyl-203 |
| 26 | 27.20 | 217 | Methyl-203 |
| 27 28 | 27.77 | 217 | Methyl-203 |
| 29 | 28.20 | 217 217 | Methyl-203 |
| 30 | 29.10 29.30 | 217 | Methyl-203 Methyl-203 |
| 31 | 29.57 | 221 | Dimethyl-203 |
| 32 | 31.13 | 235 | n.i. |
| | 31.13 | 221 | Dimethyl-203 |
| 33 | 31.60 | 229 | Benz(c)acridine |
| 34 | 32.57 | 229 | Azabenzo(c)phenanthrene |
| 35 | 32.77 | 229 | Benz(a)acridine |
| 36 | 33.10 | 229 | Aza-228 |
| 37 | .33.27 | 229 | Aza-228 |
| 38 | 33.47 | 229 | Aza-228 |
| 39 | 33.93 | 227 | Aza-226 |
| 40 | 34.13 | 229 | Aza-228 |
| 41 | 34.40 | 229 | Aza-228 |
| 42 | 34.60 | 243 | Methyl-229 |
| | 34.60 | 249 | n.i. |
| | 34.60 | 255 | n.i. |
| 43 | 34.97 | 243 | Methyl-229 |
| 44 | 36.15 | 243 | Methyl-229 |
| 45 | 39.80 | 253 | Aza-252 |
| 46 | 40.45 | 253 | Aza-252 |
| 47 | 40.80 | 269 | Methyl-255 |
| 48 | 41.00 | 253 | Aza-252 |
| 49 | 42.05 | 269 | |
| 50 | 42.30 | 253 | Methyl-255 Aza-252 |
| 51 | 42.60 | 253 | Aza-252 |
| 52 | 42.85 | 253 | 10-Azabenzo(a)pyrene |
| 53 | 43.35 | 267 | Methyl-253 |
| 54 | 43.70 | 253 | Aza-252 |
| 55 | 48.20 | 281 | Dimethyl-253 |
| 56 | 48.65 | 285 | n.i. |
| 57 | 49.00 | 281 | Dimethyl-253 |
| | | | |

AS can be seen from the table, in various cases peaks listed under one number consist of more than one compound

reduction resulted in the original S-PAC as presented in Fig. 4 (the numbering corresponds to that of Table 1).

As mentioned above, the subfraction containing PAH, S-PAC and O-PAC with 4 and 5 rings and that containing 6 and more rings were almost equally active, each contributing by about 50% to the total carcinogenicity of the emission condensate from coal briquet-fired furnaces [2]. All 4 ring-containing PAC as the majority of 5 ring compounds listed in Table 1, are volatile under the sublimation conditions (10⁻⁸ bar, 105 \degree C, 6 h). Only 7% of the benzo(a)pyrene and 15% of the benzo(e)pyrene were found in the non-volatile residue. Overlapping appears in the range of indeno $(1,2,3$ -cd)pyrene (No. 93, Table 1) and benzo(ghi)perylene (No. 99), which remain to 50% and 80% in the non-volatile part, respectively. The higher boiling PAC, such as dibenzofluoranthenes, dibenzopyrenes, coronene and their thiaarene-analogues remain entirely in this non-volatile fraction, and caused about half of the carcinogenicity of the total condensate when applied to the skin of mice.

The basic N-PAC (azaarenes) account for about $4-7\%$ of the total carcinogenic activity of the emission condensate [1]. After separation by ion-exchange chromatography they were further analyzed by gas chromatography (Fig. 5) and mass spectrometry (Table 2).

The main constituents were structural analogues of PAH as found in the neutral fractions, namely azaphenanthrenes, azafluoranthenes, azapyrenes, benz(a)acridine, dibenz(a,h) acridine and others.

Under the sublimation conditions used most of the azaarenes were found in the volatile fraction and only very small amounts of them occur in the sublimation residue. Concentration of azaarenes consisting of more than 5 rings were below the detection limit.

Mass Spectrometry

The mass spectrometric identification (recorded at 70 eV) of the compounds listed in Table 1 is based on the following key fragments for PAH and hetero atoms containing PAC. A general discussion and a tabular presentation of the keyions for characterization of unsubstitued and monomethylsubstituted PAH, thiophene- or furan-containing compounds has been published previously [20].

PAH and Methyl Derivatives. Both PAH and their methyl derivatives exhibit intense molecular ions $(M⁺)$ and the corresponding double-charged ions (M^{2+}) . The parent PAH show intense $(M-2)^+$ - and $(M-26)^+$ -ions, whereas in monomethyl derivatives intense $(M-1)^+$, moderately intense

Original Papers

 $(M-3)^+$, but small $(M-2)^+$ -fragments are observed. The $(M-26)^+$ -ion is also shifted to $(M-27)^+$ in methyl-PAH. Accordingly, the same intensity ratios are found for the double-charged ions with intense $(M-1)^{2+}$ and $(M-27)^{2+}$ in methyl-PAH. Under mass-spectrometric conditions compounds possessing 'fjord'-structures such as naphth(1,2-a)anthracene (No. 87a, Table 1) undergo further aromatisation giving rise to $(M-2)^+$ which may exceed M^+ . Methylenebridged structures such as fluorene (No. 6, Table 1), cyclopenta(def)phenanthrene (4,5-methylenephenanthrene, No. 21, Table 1), and the compounds deriving from base structures with masses 228, 252, and 276 (Nos. 66, 82, 84, and 105, Table 1) may be recognized from the ratio $M^+/$ $(M-1)^+ \approx 1$ and the corresponding intense $(M-27)^+$ fragment.

S-PAC. According to the natural isotope ratio ³²S/³⁴S which is about 25:1, especially those with more than one S-atom (found among the higher mass S-PAC such as 282 (No. 97), 290 (Nos. 84a, 85, 86, 87, Table 1), 312 (No. 105a, Table 1), 314 (Nos. 107, 108, l12a, Table 1), and 332 (No. 105b, Table 1) may be recognized from the comparatively intense $(M + 2)^+$ ion. Highly indicative are the $(M-32)$ - and $(M-$ 45)-fragments resulting from sulfur- and CHS-elimination. Double-charged ions \mathbf{M}^{2+} and $(\mathbf{M-45})^{2+}$ are also observed. As in the spectra of PAH, shifts by less one masses are found in the spectra of methyl derivatives of $S-PAC$: $(M-1)^+$ becomes larger than $(M-2)^+$, $(M-33)^+$, and $(M-46)^+$ are observed instead of $(M-32)^+$ and $(M-45)^+$.

Dimethyl-substitution results in an intense (M-15)-fragment the intensity of which increases with the number of methyl groups present in the molecule.

O-PAC. In Table 1 also some O-PAC are listed (Nos. 4, 7, 10, 15a, 35, 36) which were identified as dibenzofuran, benzonaphthofurans as well as their methyl derivatives on the basis of the $(M-16)$ -, $(M-29)$ - and $(M-55)$ -fragments formed by elimination of oxygen, a CHO- and a \overline{C} H = $\overline{CH-O-CH}$ -group, respectively. Methyl substitution may be recognized from the ratio $(M-1)^+ > (M-1)$ 2 ⁺ and the occurrence of (M-17)- and (M-30)-fragments.

N-PAC. Odd-numbered parent peaks indicate the presence of nitrogen in the molecule. Elimination of nitrogen (M-14) is rarely observed; however, the elimination of $CH = N$ or CH = NH resulting in intense $(M-27)^+$ and $(M-28)^+$ ions are common in N-PAC, often accompanied by a (M-29) fragment. In addition, a (M-53)-fragment which result from the elimination of $\overline{-CH = N-CH = CH-}$ is observed and which is analogous to the (M-55)-fragment in case of O-PAC. In methyl derivatives of O-PAC again the ratio $(M-1)^+$ > $(M-2)^+$ is found and $(M-27)$ -, $(M-29)$ - and $(M-53)$ -fragments are shifted to $(M-28)$, $(M-30)$ and $(M-54)$, respectively. A more detailed discussion of the mass spectra of azaarenes has been published previously [21].

Apart from 70eV activation, for a preliminary characterization, mass spectra were also recorded at 15 eV in order to obtain the molecular ions exclusively.

In summary, it may be concluded that PAH, S-PAC, O-PAC, and N-PAC consisting of 4 and 5 rings as well as those of 6 and more rings, listed in Tables 1 and 2 each contribute to about half of the total carcinogenicity found for the emission condensate of hard coal briquet-fired residential furnaces.

Acknowledgements. The present studies were carried out in accordance with the environmental plan of the Federal Environment Agency by the order of the Federal Ministry of the Interior. The authors are greatly indebted to Dr. Ing. E. Ahland and E.-A. Ratajczak, Bergbau-Forschung Essen, for supplying them with hard-coal briquet emission condensate. They gratefully acknowledge the donation of various benzaceanthrylenes and benzophenanthrylenes by Dr. Avram Gold.

References

- 1. Grimmer G, Brune H, Deutsch-Wenzel R, Dettbarn G, Misfeld J, Abel U, Timm J (1984) Cancer Lett 23:167-176
- 2. Grimmer G, Brune H, Deutsch-Wenzel R, Dettbarn G, Misfeld J, Abel U, Timm J (1985) Cancer Lett (in press)
- 3. Ahland E, Borneff J, Brune H, Grimmer G, Habs M, Heinrich U, Hermann P, Misfeld J, Mohr U, Pott F, Schmähl D, Thyssen J, Timm J (1985) Mfinch Med Wchschr 127 : 218 - 221
- 4. Grimmer G, Buck M, Ixfeld H (1981) Immissionsmessungen yon polycyclischen aromatischen Kohlenwasserstoffen, Minister ffir Arbeit, Gesundheit und Soziales des Landes Nordrhein-Westfalen
- 5. Grimmer G, Neujack KW, Schneider D (1981) Int J Env Anal Chem $10:265 - 276$
- 6. Beine H (1970) Staub Reinhalt Luft 30:334--336
- 7. Brockhaus A, Tomingas R (1976) Staub Reinhalt Luft 36:96-- 102
- 8. Deangelis DG, Reznik RB (1979) Source assessment: Residential combustion of coal, EPA-600/2-79-019 a, PB 295649A; Research Triangle Park, NC, USA, Industrial Environmental Research Laboratory, pp 145
- Herlan A, Meyer J (1980) Gas-Wasserfach, Gas/Erdgas 121:301-307
- 10. Ahland E, Mertens H (1980) Luftverunreinigung dutch polycyclische aromatische Kohlenwasserstoffe - Erfassung und Bewertung. VDI-Verlag, Düsseldorf, pp 107-111
- 11. Highes TW, Deangelis DG (1982) Residential solid fuels: Environmental impacts and solutions. In: Cooper JA, Malek D (eds), Proceedings of 1981 International Conference, Portland, OR, USA, pp 333--348
- 12. Truesdale RS, Cleland JG (1982) Residential wood and coal combustion, Speciality Conference Louisville, KY, USA (APCA, eds), pp 115 - 128
- 13. Herlan A, Mayer J (1983) Gas-Wasserfach, Gas/Erdgas 124: 263- 267
- 14. Truesdale RS (1984) Research Triangle Institute, Research Triangle Park, NC, USA, private communication
- 15. Ratajczak EA, Ahland E, Grimmer G, Dettbarn G (1984) Staub Reinhalt Luft 44: 505-- 509
- 16. IARC Monograph Volume 32 (1983), p 32
- 17. Grimmer G, Jacob J, Naujack KW, Dettbarn G (1983) Anal Chem 55: 892- 900
- 18. Grimmer G, B6hnke H, Glaser A (1977) Zbl Bakt Hyg I. Abt Orig B 164:218--234
- 19. Schmidt W, Grimmer G, Jacob J, Dettbarn G, Naujack KW (1985) Anal Chem, submitted for publication
- 20. Grimmer G, Jacob J, Naujack KW (1981) Fresenius Z Anal Chem 306:347 - 355
- 21. Grimmer G, Jacob J, Naujack KW (1983) Anal Chem 55:2398- 2404

Received June 28, 1985