

# Hydrography and distribution dynamics of larval and juvenile fishes in the coastal waters of the Tanshui River estuary, Taiwan, with reference to estuarine larval transport

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**Abstract.** Distribution dynamics of fish larvae and juveniles in the coastal waters of the Tanshui River, Taiwan was studied fortnightly using surface horizontal tows with a larval net in daytime during the period from early April through early June 1991. Environmental factors, including water temperature, salinity, dissolved oxygen, pH, transparency and depth at sampling stations, were also monitored. A total of 10 737 fish eggs and 1387 individuals, representing 43 families and 93 species, was collected during five cruises from 12 stations in the coastal waters. Most fish were estuarine-dependent marine species. *Liza macrolepis*, *Ambassis gymnocephalus*, *Terapon jarbua*, Mullidae and Gobiidae were the most dominant, making up 64.7% of the total catch. Early life stages, including egg, preflexion, flexion and postflexion larvae were abundant in surface samples. However, yolk-sac larvae were absent in the surface water, probably due to an ontogenetic behavioral shift as a consequence of a change in specific weight during early development. The species composition of fish larvae and juveniles was related to the microhabitats found in the coastal waters. The physico-chemical conditions, along with ontogenetic behavior, played an important role in larval fish distribution in the coastal waters.

## Introduction

The coastal waters adjacent to the Tanshui River estuary make up a commercial fishing ground important for harvesting the juveniles of engraulids and clupeids for local consumers, as well as anguillid elvers for cultivation. There are rare mangroves, *Kandelia candel* (Rhizophoraceae), in the lower estuary of the river (Chou et al. 1987), which provide plentiful organic detritus and a link to the food web of the estuarine-dependent marine fish population during their early life stages (McErlean et al. 1973, Haedrich and Haedrich 1974, Bell et al. 1984, Robertson and Duke 1987, Blaber and Milton 1990). Consequently, the study area functions as a nursery and feeding ground for the onshore-offshore migratory fish.

Due to rapid economic and population growth, the river has been severely polluted, receiving domestic sewage from Taipei city and industrial wastewater from factories in the immediate area (Chou et al. 1987). Thus, the production of larval fishes in the Tanshui River estuary seems to be decreasing and the fishing grounds shifting seaward. In order to begin restoration of the river, a sewage and wastewater processing plant is now under construction at Bali. Treated wastewater will be discharged into the nearshore waters off the Tanshui River, a nursery ground for many commercially important fishes. In order to evaluate the effect of this discharge on distribution and abundance of fish larvae and juveniles in the ecosystem, baseline information, including hydrography, fauna and flora of the estuary is essential. Accordingly, a multidisciplinary team of scientists was formed, supported by the National Science Council, Republic of China, to study the mangrove estuarine ecosystem in the Tanshui River (Chou and Bi 1990).

The temporal and spatial use of estuaries and nearshore waters by estuarine, marine, freshwater and estuarine-dependent marine species has been schematically described by Deegan and Thompson (1985). A conceptual model of the transport of fish larvae and juveniles between nearshore and estuarine nursery areas was proposed on the basis of physical processes, activity and behavior of the fish, and environmental cues (Boehlert and Mundy 1988). In addition, the movement and vertical distribution of the fish larvae and juveniles from offshore to estuaries in relation to spawning mode and development of the fish were also suggested (Tanaka 1976).

The occurrence, abundance and species composition of fish larvae and juveniles in the estuary and nearshore waters varies with the spawning season of the fish and their seasonal onshore-offshore migrations, as well as the physico-chemical conditions during estuarine transportation (Blaber and Whitfield 1977 a, b, Weinstein 1979, Bell 1980, Blaber and Blaber 1980, Quinn, 1980, Yáñez-Arancibia et al. 1980, Bell et al. 1984, Loneragan et al. 1986, Mukai 1987, Powell et al. 1989, Blaber and Milton 1990, Robertson and Duke 1990, Sebatés 1990, Drake and Ari-

as 1991 b). The estuarine orientation of the fish is dependent on swimming ability and tolerance of the fish to the extremes of the environmental variables, which differs considerably among species (Kinne 1964, Whitfield et al. 1981), as well as between size and developmental stages of each species (Kinne 1964, Holliday 1971).

The species composition, structure and seasonal dynamics of the larval and juvenile fish community in the mangrove estuary of the Tanshui River has previously been studied (Wang et al. 1991, Tzeng and Wang 1992). The present paper aims to clarify the distribution and abundance of fish larvae and juveniles in relation to environmental factors in the coastal waters adjacent to the Tanshui River estuary.

## Materials and methods

### Study area

The Tanshui River, approximately 159 km long, is the largest river in northern Taiwan. The river flows through the Taipei basin and meets the sea at the town of Tanshui. The river mouth faces the shallow continental shelf in the northern part of the Taiwan Strait. The estuary belongs to a coastal plain estuary, its hydrography being greatly influenced by tidal currents. The direction of the tidal current is southward during flood and northward during ebb, with a tidal range in the river inlet from approximately 3.0 m at spring tide to 1.5 m at neap tide (Lee and Chu 1965). Twelve stations in the nearshore waters of the Tanshui River were selected for larval and juvenile fish sampling (Fig. 1). These stations were chosen because they will probably be affected by the treated wastewater discharged from the Bali wastewater processing plant.

The topography is deeper in the northern, offshore areas and shallower in the river inlet and in the southern part of the studied habitat. The depth of the sampling stations ranged from 10 to 50 m (Fig. 1).

### Sampling design

Fish larvae and juveniles were collected fortnightly from the 12 stations in the nearshore waters of the Tanshui River, during the period from early April to early June 1991. The investigation period is consistent with the main fishing season of the larval and juvenile fish. Sampling was conducted using surface horizontal tows with a larval net during the daytime flood tide. The duration of sampling was pre-set at 5 min for each station at a speed of ca. 2 knots. The larval net was a modified Maruchi-D larval net, net mouth diameter 1.3 m, length 4.5 m and mesh size 0.5 × 0.5 mm (Nakai 1962). A flowmeter mounted in the net mouth recorded filtered water volume. The fish collected were fixed immediately in 10% formalin seawater solution. Fish larvae and juveniles were identified to the lowest possible taxon, the developmental stages of the fish determined according to Kendall et al. (1984).

The environmental factors were monitored during sampling. Temperature was measured with a mercury thermometer, salinity with a salinometer (WTW: microprocessor conductivity meter, Model: LF 196), DO (dissolved oxygen) with Winkler's method, pH with a pH meter (Coring: pH meter, M107), transparency with a secchi disc, and the depth at station was measured with an echo sounder.

Patchy distribution, gear selection, net avoidance, and diel vertical migration of the larvae in relation to ontogenetic behaviour of the fish, and their influence on the accuracy of the estimation of fish abundance, has frequently been debated by many researchers. It has been suggested that replicated and vertically stratified samplings

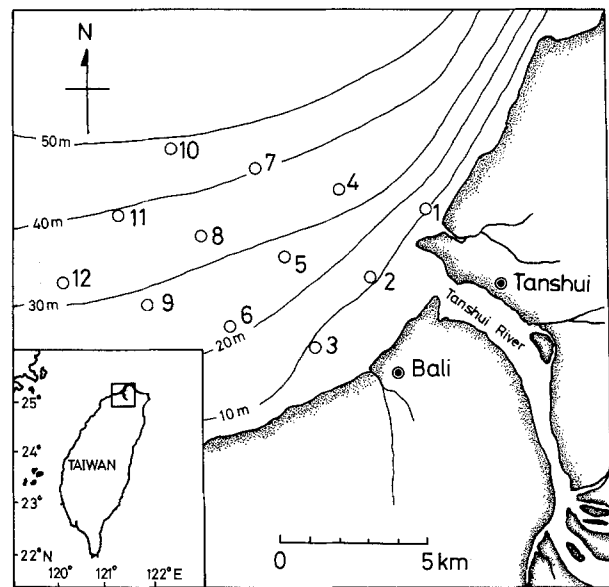


Fig. 1. Map showing sampling stations (1–12) and isodepth in the nearshore waters of the Tanshui River, Taiwan

can reduce the sampling bias (e.g. Clutter and Anraku 1968, Smith and Richardson 1977, Omori and Hamner 1982, Leis 1986, Brander and Thompson 1989). However, sampling in the present study was not replicated because the environmental conditions changed rapidly due to tidal currents. In addition, stratified sampling was difficult due to the shallow waters. The difference between night and day collections, gear selection and the vertical distribution of the larvae will be studied in a separate paper. The present study emphasized the horizontal distribution of the fish larvae in relation to environmental factors.

### Data analysis

The density of fish eggs and larvae was calculated from the flowmeter reading, then standardized according to the number of fish per 1000 m<sup>3</sup> seawater filtrated. The homogeneous and heterogeneous relationships of the environmental factors among stations were analyzed using multivariate analysis: principle coordinate analysis (PCA). To explain the zonation of the fish community according to habitat, the 12 stations were clustered and the dominant species ordinated based on species density data using PCA. For purposes of comparison, the similarities of species composition among stations was also calculated using Kimoto's (1976) overlap-degree index, and then clustered into different groups using the Mountford's (1962) method.

The community structure of fish larvae and juveniles was evaluated using Shannon-Weaver's species diversity ( $H'$ ) and Pielou's evenness ( $J'$ ) indices (Pielou 1966). The relationship between biotic and environmental factors was analyzed by canonical (Dillon and Goldstein 1984) and Spearman rank correlations (Siegel 1956).

## Results

### Physico-chemical conditions

The environmental factors in the estuary changed with tidal conditions and season. Hence, environmental conditions not only differed between the five surveys but also among the 12 stations themselves (Fig. 2).

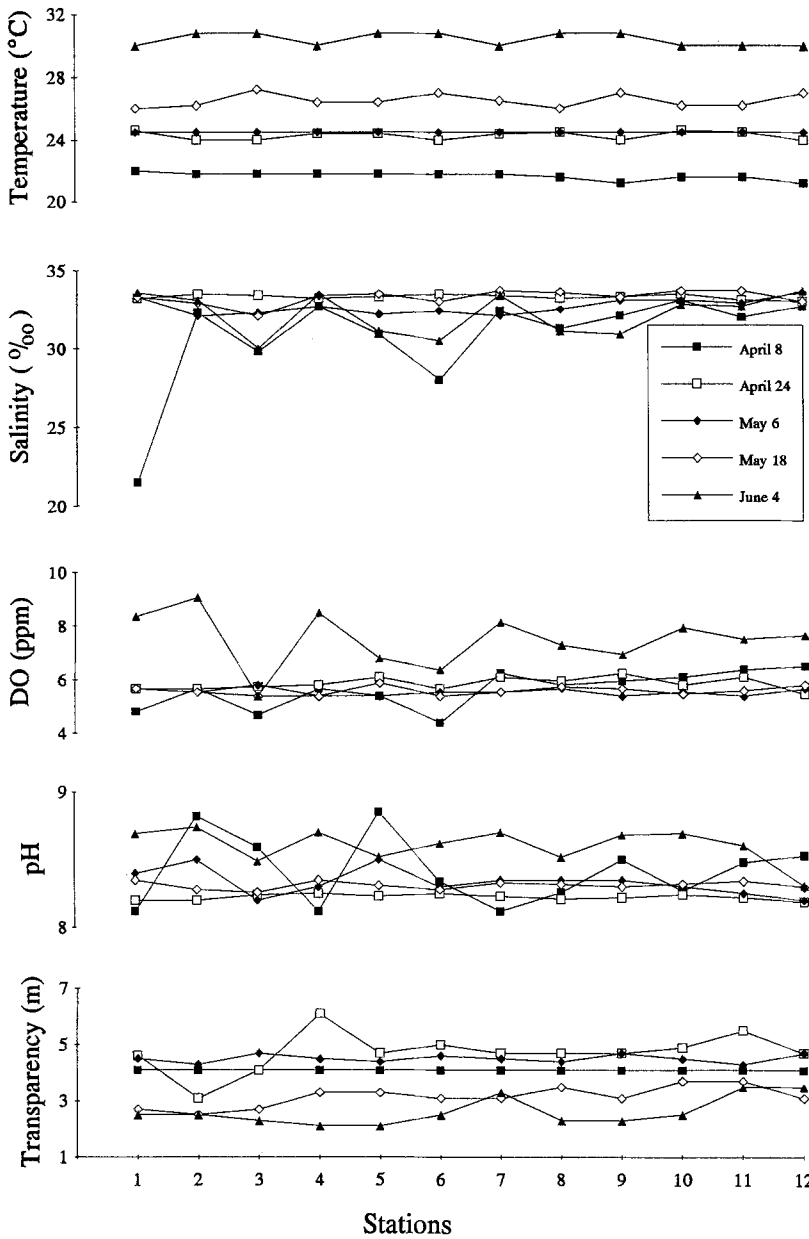


Fig. 2. Temporal and spatial variations of temperature, salinity, DO (dissolved oxygen), pH and transparency measured in the nearshore waters of the Tanshui River on 8 and 24 April, 6 and 18 May, and 4 June 1991

Water temperatures increased from 21 °C on 8 April to 31 °C on 4 June 1991. The temperatures in the southern stations were lower before and higher after 18 May (Stns 3, 6, 9, and 12) than in the northern and offshore stations of the studied habitat. This indicated that seawater temperatures were more variable in the shallower areas.

Salinity varied with sampling dates and fluctuated between the 12 stations (Fig. 2). Salinity was higher on 24 April and 18 May (greater than 33‰), moderate on 6 May (32 to 33‰) and less than 32‰ at some of the stations on 8 April and 4 June. The lower salinity was due to a strong river discharge, more greatly affecting the shallower, southern stations (Stns 3, 6, and 9). Therefore, the outflow of freshwater from the Tanshui River tended to be diverted to the southern part of the studied habitat during flood tides.

DO also varied with sampling dates and stations, and ranged from 4.5 to 9.0 ml l<sup>-1</sup> (Fig. 2). This range was above the minimal requirements of the fish. DO was

higher both during cold water mixing period (8 April) and during increased photosynthesis during the summer (6 June), but lower when affected by strong fluvial discharge, pH values were close to that of normal seawater (8.1 to 8.3), their fluctuations following those of DO. When DO increased, pH increased. DO and pH were related to biological activity and river discharge. Transparency was higher in offshore than in inshore waters but decreased with temperature and DO. This phenomenon indicated that transparency was correlated to plant biomass, as well as being influenced by the turbidity loading in the estuary.

Relationships among the five environmental factors at the 12 stations in the nearshore waters were analysed by PCA ordination. They indicated that physico-chemical conditions at these stations were very dynamic, differing on all five successive surveys. The 12 stations were classified into different groups according to the attributes of the five changing environmental factors (Fig. 3).

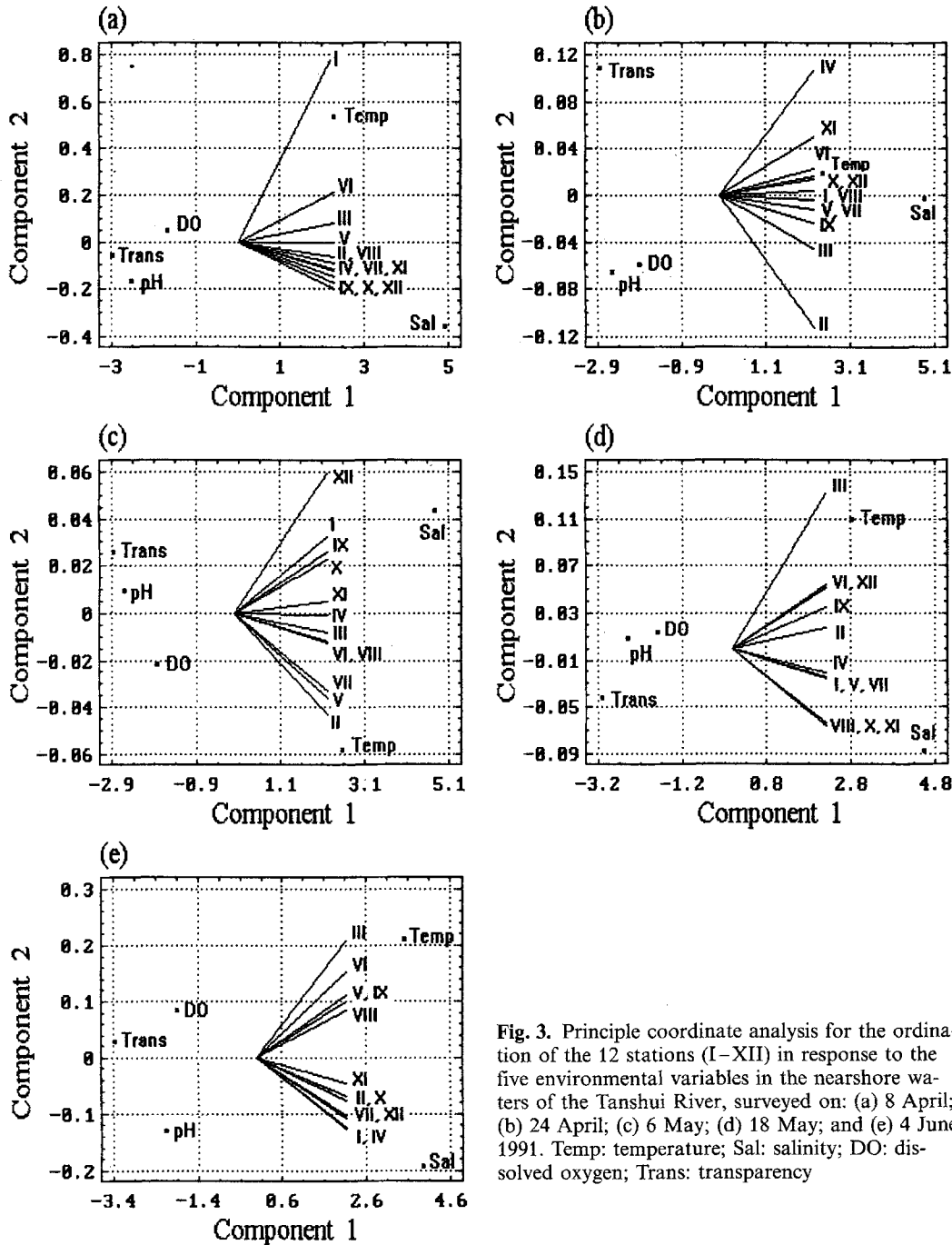


Fig. 3. Principle coordinate analysis for the ordination of the 12 stations (I–XII) in response to the five environmental variables in the nearshore waters of the Tanshui River, surveyed on: (a) 8 April; (b) 24 April; (c) 6 May; (d) 18 May; and (e) 4 June 1991. Temp: temperature; Sal: salinity; DO: dissolved oxygen; Trans: transparency

The ordination of the 12 stations as indicated from the first component of PCA in Fig. 3a, surveyed on 8 April 1991 was positively correlated to temperature and salinity, and negatively correlated to DO, transparency and pH. But the ordination of the 12 stations as indicated from the 2nd component was slightly divergent, the ordination of the shallower stations (Stns 1, 3, 6) being positively correlated to temperature and DO, and negatively correlated to salinity, pH and transparency; at the remaining nine stations, the relationships were reversed. This situation changed in each of the other four surveys (Fig. 3b–e).

#### Abundance and distribution of fish eggs and larvae

Fish eggs were more abundant at the deeper stations than at the shallower stations, except at Stn 1 on 8 April (Fig. 4). Fish eggs were probably spawned in the studied area or drifted with the current from offshore into the studied area. Fish larvae were more abundant in the deeper offshore stations (Stns 11 and 12) than in the shallower, more nearshore stations (Stns 1, 2, 3, and 6). This suggested that most fish eggs and larvae probably originated from the offshore-spawning marine species.

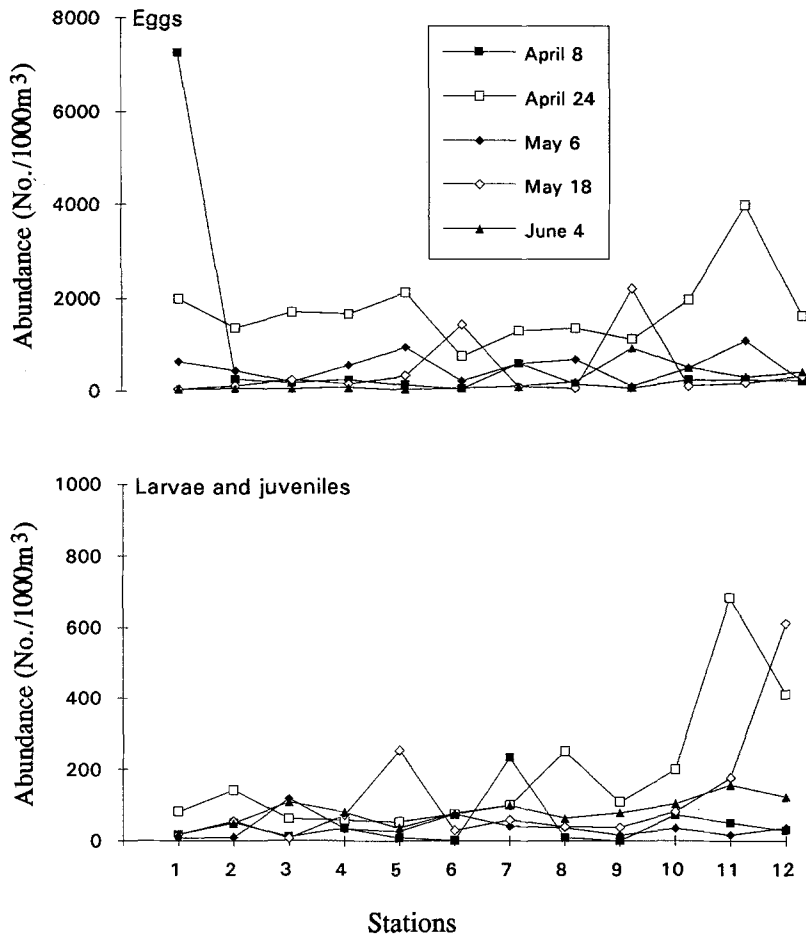


Fig. 4. Temporal and spatial variations in abundance of fish eggs and larvae collected in the nearshore waters of the Tanshui River on 8 and 24 April, 6 and 18 May, and 4 June 1991

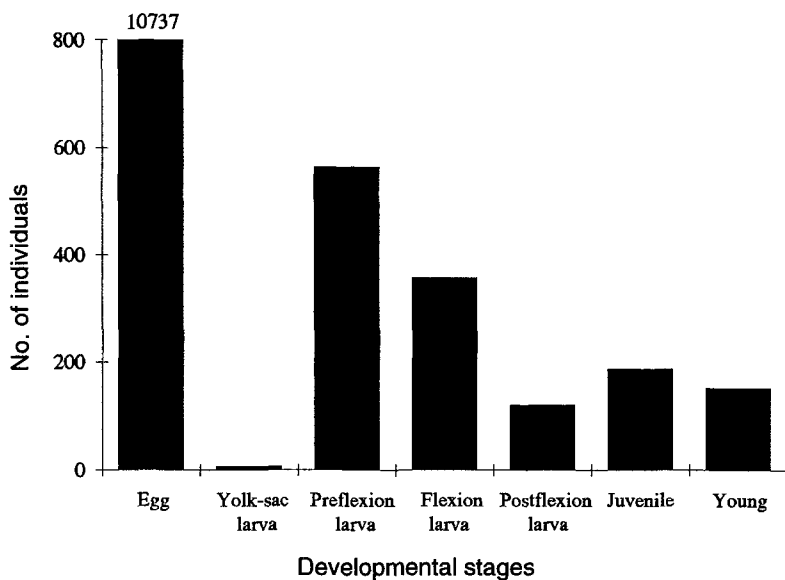


Fig. 5. Relative abundance at different developmental stages of the fish collected from the 12 stations in the nearshore waters of the Tanshui River on 8 and 24 April, 6 and 18 May, and 4 June 1991

Fish abundance associated with developmental stages

In general, due to mortality, the number of fish gradually decreased from egg to young. However, the number of yolk-sac larvae from surface sampling in the studied area was dramatically less than that of the next developmental stage (Fig. 5, Table 1). The number of yolk-sac larvae was

significantly lower than that of preflexion, flexion, and postflexion larvae. This indicates that the abundance of yolk-sac larvae does not fit the standard survival curve of fish in the early stages. This discontinuous phenomenon was probably due to ontogenetic behavioral changes. The specific weight of yolk-sac larvae has been reported to be greater than that of pelagic eggs (Tanaka 1990). Yolk-sac



Table 1 (continued)

Serial no.	Species and family	No. of fish at life stage						Total no. of fish	Frequency of occurrence (5 surveys at 12 stns)			Sum for all Stns	%		
		Ys	Pr	Fl	Po	Ju	Yg		Date						
44	sp. 1		1					1			D	1	1.67		
45	sp. 3			1				1		B		1	1.67		
46	sp. 6		3					3		B	D	2	3.33		
47	sp. 7		6					6	A			3	5.00		
48	sp. 8		2					2	A			2	3.33		
	Mullidae														
49	sp. 1		43	36	15	7	5	106		B	C D	16	26.67		
50	sp. 2		69	22				91		B	C D E	16	26.67		
51	Pempheridae sp. 1		3					3	A			1	1.67		
	Scatophagidae														
52	<i>Scatophagus argus</i>		1		1			2			D E	2	3.33		
	Pomacentridae														
53	sp. 1					4		4			D	2	3.33		
54	sp. 2						1	1			E	1	1.67		
	Mugilidae														
55	<i>Liza macrolepis</i>		31	83	16	60	143	333		B	C D E	28	46.67		
56	Sphyraenidae sp. 1		3	5				8		B	D E	8	13.33		
57	Scaridae sp. 1				2			2			D	1	1.67		
58	Mugiloididae sp. 2		1					1		B		1	1.67		
59	Percophididae sp. 1				1	1		2			C D	2	3.33		
	Blenniidae														
60	<i>Entomacrodus niuafoensis</i>					6		6			C D E	6	10.00		
61	sp. 2			1	1	4		6			C D E	6	10.00		
62	sp. 3		32	1				33	A	B	C D	17	28.33		
63	sp. 4						1	1			D	1	1.67		
64	sp. 5		7	1				8	A		C E	7	11.67		
65	sp. 7		1					1			E	1	1.67		
66	Tripterygiidae sp. 1		1					1			C	1	1.67		
	Gobiidae														
67	sp. 1		5			1		6	A			4	6.67		
68	sp. 8		5					5	A			3	5.00		
69	sp. 14		1					1	A			1	1.67		
70	sp. 15		1					1	A			1	1.67		
71	sp. 16		44	22	5	2		73		B	D E	19	31.67		
72	sp. 17					2		2			C D	2	3.33		
73	sp. 18		1	2	1			4			C E	3	5.00		
74	sp. 19					2		2			C	2	3.33		
75	sp. 20			1	2			3		B	D	3	5.00		
76	sp. 21					1		1			E	1	1.67		
77	sp. 22		1			1		2	A		E	2	3.33		
	Scombridae														
78	<i>Scomber japonicus</i>			1				1		B		1	1.67		
	Callionymidae														
79	sp. 1				2			2			C	2	3.33		
80	sp. 2		19	6	1	1		27	A	B	C D E	16	26.67		
	Bothidae														
81	sp. 1		3		1			4		B	C D	4	6.67		
82	sp. 2				1			1			D	1	1.67		
83	sp. 4					1		1			C	1	1.67		
84	Cynoglossidae sp. 1		2					2		B		2	3.33		
	Monacanthidae														
85	<i>Stephanolepis cirrhifer</i>					9		9		B		E	3	5.00	
86	sp. 2		1					1			D	1	1.67		
87	sp. 3					4		4			E	4	6.67		
	Tetradontidae														
88	<i>Fugu niphobles</i>					1		1			C	1	1.67		
89	sp. 2			1				1			E	1	1.67		
	Unidentified														
90	sp. 1	1						1		B		1	1.67		
91	sp. 2	1						1		B		1	1.67		
92	sp. 3		1					1			D	1	1.67		
93	sp. 4			2				2		B		1	1.67		
Total		7	564	357	120	187	152	1387	18	38	35	36	35	60	100.00

**Table 2.** Spearman rank correlation between fish abundance and abiotic factors for five surveys. A: 8 April; B: 24 April; C: 6 May; D: 18 May; E: 4 June. TEMP: temperature; SAL: salinity; DO: dissolved oxygen; DEP: depth of station; TRANS: transparency; DIST: distance from coast; H': Shannon-Weaver's species diversity index; J': Pielou's evenness index

Community variables	Abiotic factors																									
	TEMP			SAL			pH			DO			DEP			TRANS			DIST							
	A	B	C	A	B	C	A	B	C	A	B	C	A	B	C	A	B	C	A	B	C	A	B	C	D	E
No. of species	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
H'	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
J'	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Abundance (no./1000 m <sup>3</sup> )	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
<i>Liza macrolepis</i>	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
<i>Ambassis gymnocephalus</i>	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
<i>Terapon jarbua</i>	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Mullidae sp. 1	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Mullidae sp. 2	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Gobiidae sp. 16	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Nemipteridae sp. 2	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
<i>Decapterus maruadsi</i>	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
<i>Stephanolepis cirrhifer</i>	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Overall species	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±
Eggs	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±	±

± no significant correlation; + positive correlation; - negative correlation; \*\*  $p < 0.01$ ; \*  $0.01 < p < 0.05$ ; blank: no catch

larvae usually sink to the bottom at this stage. Therefore, yolk-sac larvae were not as accessible to surface sampling.

Species composition and retention

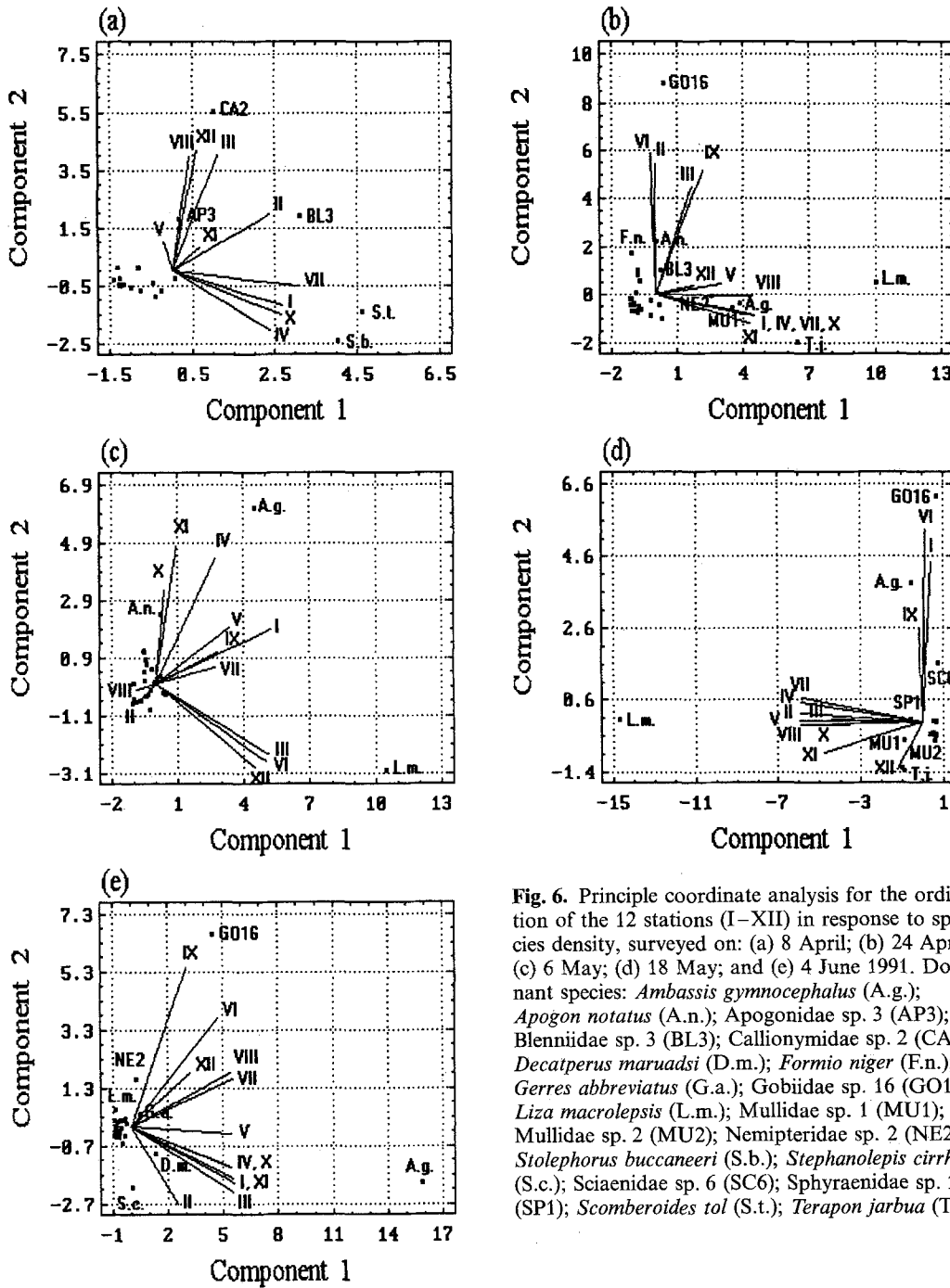
A total of 1387 individuals, representing 43 families and 93 species, was collected from 12 stations through five surveys, during the period from early April to early June 1991 (Table 1). The number of fish captured was consistent with their frequency of occurrence. The dominant species was widely distributed among the 12 stations and occurred with higher frequency in the five surveys. The top five dominant taxa – *Liza macrolepis*, *Ambassis gymnocephalus*, *Terapon jarbua*, Mullidae and Gobiidae – made up 64.7% of the total catch. Most species occurred in low numbers and at an early life stage, indicating that most species occurring in the area were short-term, temporary residents or rare species (Bell et al. 1984). In contrast, the dominant species, *Liza macrolepis* and *Ambassis gymnocephalus*, may be long-term, temporary residents, because of their higher frequency of occurrence in the nearshore waters and the larger number of life history stages in comparison with the other species. They may live in the nearshore waters for extended periods as juveniles before leaving.

Ordination of station by species density

The 12 stations clustered by PCA according to the species density data (Fig. 6), are similar to those clustered by Kimoto's (1976) index and Mountford's (1962) method (Fig. 7). The classification of the stations was not always consistent between surveys (Fig. 6).

According to species composition of fish larvae and juveniles, the 12 stations surveyed on 8 April can be classified into three groups: A (Stns I, II, IV, VII and X), B (Stns V and XI) and C (Stns III, VIII and XII). The dominant species in each group was different. Group A consisted mainly of *Scomberoides tol* and *Stolephorus buccaneeri*; Group B of Gobiidae sp. 1, Apogonidae sp. 3; and Group C of Callionimidae sp. 2 and Blenniidae sp. 3, respectively. Group A was mainly made up of pelagic fishes, while Groups B and C were mainly benthic fishes (Fig. 6a). Similarly, the same stations surveyed on 24 April were classified into only two groups: A (Stns I, IV, V, VII, VIII, X, XI and XII) and B (Stns II, III, VI and IX). The dominant species in the stations changed, Group A dominated by *Liza macrolepis* and *Terapon jarbua*; and Group B by Gobiidae sp. 16 and *Formio niger*, respectively (Fig. 6b). The stations of the other three surveys can also be classified into different groups, in which the dominant species again changed (Figs. 6c–e). These changes of station groupings and dominant species indicated that species composition of fish larvae and juveniles in the nearshore waters changed rapidly. This phenomenon was probably due to the seasonal succession of the species and/or environmental conditions.





**Fig. 6.** Principle coordinate analysis for the ordination of the 12 stations (I–XII) in response to species density, surveyed on: (a) 8 April; (b) 24 April; (c) 6 May; (d) 18 May; and (e) 4 June 1991. Dominant species: *Ambassis gymnocephalus* (A.g.); *Apogon notatus* (A.n.); Apogonidae sp. 3 (AP3); Blenniidae sp. 3 (BL3); Callionymidae sp. 2 (CA2); *Decapterus maruadsi* (D.m.); *Formio niger* (F.n.); *Gerres abbreviatus* (G.a.); Gobiidae sp. 16 (GO16); *Liza macrolepis* (L.m.); Mullidae sp. 1 (MU1); Mullidae sp. 2 (MU2); Nemipteridae sp. 2 (NE2); *Stolephorus buccaneeri* (S.b.); *Stephanolepis cirrhifer* (S.c.); Sciaenidae sp. 6 (SC6); Sphyraenidae sp. 1 (SP1); *Scomberoides tol* (S.t.); *Terapon jarbua* (T.j.)

**Relationships between biotic and abiotic factors**

A canonical correlation of the relationships between biotic and abiotic factors indicated that the abundance of fish eggs and larvae as well as species diversity indices were influenced by multiple abiotic factors (Fig. 8). The first two canonical correlations between six abiotic and five biotic factors were both significant ( $r_1 = 0.7298$ ,  $DF = 30$ ,  $\chi_1^2 = 74.917$   $P < 0.00001$ ;  $r_2 = 0.5959$ ,  $DF = 20$ ,  $\chi_2^2 = 36.88$ ,  $0.01 < P < 0.05$ ). In the first canonical variate, abundance of fish egg and larvae, and species diversity index ( $H'$ ) were positively correlated with temperature, DO, depth and transparency, but negatively correlated with salinity and pH, as well as number of species and

species evenness index ( $J'$ ). While in the second canonical variate, the relationship between biotic and abiotic factors was more complicated. Larval abundance was positively correlated with all the abiotic factors except pH, and negatively correlated with egg abundance and number of species (Fig. 8).

Furthermore, the community variables and abundance of dominant species in relation to each of the abiotic factors for each of the five surveys were conducted by Spearman rank correlation (Table 2). The community structure was more complicated with increasing pH, depth, transparency and the distance from the coast. Of the 18 selected species, half were significantly correlated with the abiotic factors. The abundance of most domi-

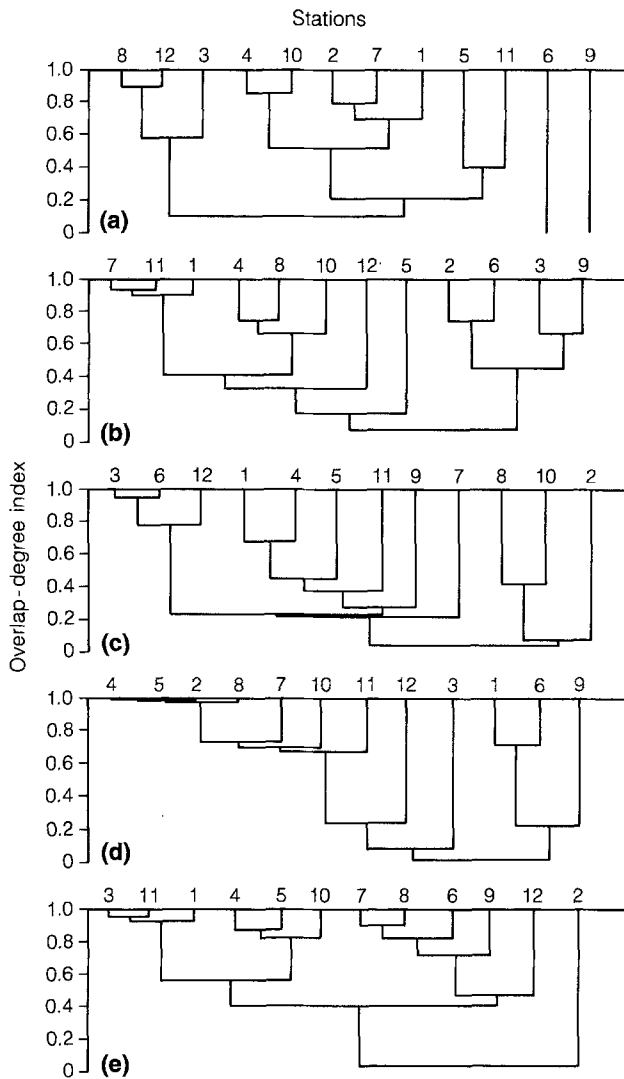


Fig. 7. Clustering of the 12 stations by Mountford's (1962) method based on Kimoto's (1976) overlap-degree index ( $C\pi$ ) of species composition between stations, surveyed on: (a) 8 April; (b) 24 April; (c) 6 May; (d) 18 May; and (e) 4 June 1991

nant species was positively correlated to depth and location of the station. *Ambassis gymnocephalus*, Nemipteridae, Mullidae and *Liza macrolepis* tended to live in low salinity water. Fish abundance was not significantly correlated with DO, probably due to the fact that DO was not lower than the required threshold of the fish. The positive correlation between fish egg and larval abundance, as well as the depth and location of the station, suggested that most fish larvae and juveniles in the nearshore waters of the Tanshui River originated from offshore.

## Discussion and conclusion

### Species composition and their origin

A total of 1387 individuals, representing 43 families and 93 species, was captured in the nearshore waters of the

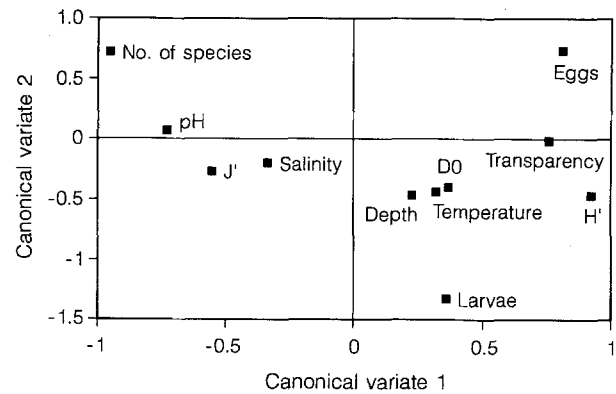


Fig. 8. Ordination of the biotic and abiotic factors by first two canonical variates. J': Pielou's evenness index; H': Shannon-Weaver's species diversity index

Tanshui River during the period from April to June 1991 (Table 1). The larval fish community is comparable with the fish fauna existing in other tropical mangrove estuarine ecosystems (Blaber et al. 1985, Robertson and Duke 1990). *Liza macrolepis*, *Ambassis gymnocephalus*, *Terapon jarbua*, Mullidae and Gobiidae are most dominant and frequently occur in the estuary. They are estuarine-dependent marine and estuarine species (Deegan 1989) and their retention in the estuary was longer than for the other species. The remaining marine species, e.g., Myctophidae, *Coryphaena hippurus*, and *Scomber japonicus*, occurred in low numbers and their larvae occasionally drifted with the tidal current into the estuary (Table 1). No freshwater species was found in the present study (Table 1), probably due to the salinity (Fig. 2). Freshwater eel, *Anguilla japonica*, eelers are abundant at nighttime in the study area during upstream migration in winter; however, they were not caught during the investigated period because the sampling was conducted during daylight.

Fish larvae and juveniles in the estuary were classified into permanent residents, long- and short-term temporary residents and rare species according to their number and duration in the estuary (Bell et al. 1984). Most of the fish captured seemed to be temporary residents or rare species, because the number of individuals of most species was fewer than four per sample (Table 1). Most fish larvae and juveniles in the nearshore waters of the Tanshui River were spawned in adjacent marine waters, their eggs and larvae passively drifting with the current into the study area, where they stayed through the early stages of their life history. On the other hand, Gobiidae are estuarine spawning and may remain in the estuary throughout their life cycle. There are three major migration patterns by which fish use estuarine systems for reproduction and juvenile feeding (Deegan and Thompson 1985), viz.: (1) saltwater spawning, followed by immigration of the larvae into an estuary; (2) estuarine spawning, in which the larvae remain, for the most part, within an estuary; (3) freshwater spawning, followed by the downstream drift or swimming of larvae and juvenile fish into an estuary.

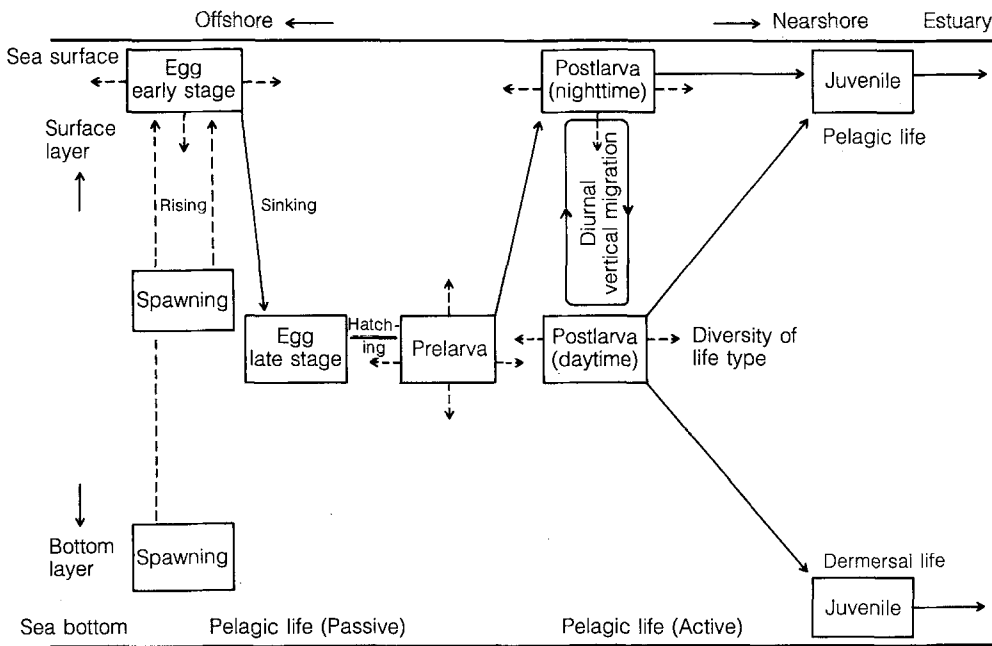


Fig. 9. Schematic diagram showing onshore movement of fish larvae and juveniles in relation to their ontogenetical behavioral changes and life-type (redrawn from Tanaka 1976)

To classify the species listed in Table 1 into the different migratory patterns, detailed information about the life history of the fish was necessary. This information was very limited except for some commercially important species, e.g. *Scomber japonicus*, *Anguilla japonica* and *Terapon jarbua* (Miu et al. 1990).

#### Larvae fish abundance in relation to estuarine transport and behavioral ontogeny

Transportation of fish larvae from offshore into the nearshore waters and further into estuarine nursery areas is influenced both by physical processes, as well as the fishes' activity and behavior (Tanaka 1976, 1985, 1990, Weinstein et al. 1980, Boehlert and Mundy 1988, Drake and Arias 1991 a, b). This process of transportation is important in determining the occurrence and abundance of fish eggs and larvae in the nearshore waters. Many studies indicate that the buoyancy of fish eggs changes during early developmental stages (Russell 1976, Coombs 1981, Tanaka 1981, Coombs et al. 1985, Tanaka 1990). Pelagic fish eggs passively drift with the current into the estuary. Upon reaching the yolk-sac stage, their specific weight increases and thus the larvae sink to subsurface layers, so that the yolk-sac larvae are not accessible to surface sampling. This behavior could prevent the larvae from being transported away from suitable habitats. When the larvae develop to the postflexion stage, their gas bladder is completely formed and they can perform diel vertical movement and active migration. Then, the larvae can adjust their buoyancy to allow selective tidal stream transport, using currents to move into an estuary (Kuwahara and Suzuki 1984, Boehlert and Mundy 1988). After the juvenile stage, the fish differentiated into pelagic and benthic forms. Benthic forms settle down and do not drift with the tidal current (Fig. 9).

Therefore, most benthic species beyond the juvenile stage were not caught by surface sampling, and the number of benthic species, e.g. *Terapon jarbua*, Mullidae, Apogonidae, Nemipteridae, Gobiidae, and Callionymidae, gradually decreased with increasing developmental stages (Table 1). In contrast, the pelagic forms, e.g. *Ambassis gymnocephalus* and *Liza macrolepis*, which are estuarine-dependent pelagic species, were abundant and stayed longer in the nearshore waters (Table 1). These facts indicated that ontogenetically behavioral change associated with developmental stage may play an important role in determining the abundance of the fish larvae in the nearshore waters (Fig. 5).

#### Distribution and abundance of fish larvae in relation to abiotic factors

The physico-chemical characteristics of the water in the studied area were significantly different in each of the five surveys (Figs. 2, 3). This indicates that the estuarine environment is very dynamic. The shallower area of the studied habitat was influenced by freshwater, lowering its salinity (Fig. 2). The abundance and species composition of fish larvae and juveniles also changed with the physico-chemical conditions (Table 2). At the shallow water stations, the fish community was dominated by benthic species, e.g. Callionymidae, Blenniidae, and Gobiidae. In contrast, the deep-water stations of the study area were dominated by nearshore pelagic species, e.g. *Ambassis gymnocephalus*, *Liza macrolepis*, *Coryphaena hippurus*, *Formio niger*, *Scomber japonicus*, *Stephanolepis cirrhifer*, Carangidae, and Engraulidae, as well as mesopelagic fish, Myctophidae (Fig. 6).

The extent of onshore migration of fish is dependent on their tolerance to extremes of environmental variables (Kinne 1964, De Vein 1978, Whitfield et al. 1981,

Claridge and Potter 1983, Tongiorgi et al. 1986, Boehlert and Mundy 1988, Tosi et al. 1988). The abundances of dominant species, *Liza macrolepis* and *Ambassis gymnocephalus*, were negatively correlated with salinity (Table 2), indicating that low salinity may act as a cue to guide these fish larvae and juveniles to inshore nursery grounds (Hughes 1969, Young and Carpenter 1977). However, this phenomenon may be different between species. Some species can not tolerate wide salinity ranges in the estuary and emigrate during later life stages. Thus, they appeared in low numbers and at very early stages of their life cycle (Table 1).

In conclusion, the community structure of fish larvae and juveniles in this estuarine area was very complex but consisted mainly of estuarine-dependent marine species. Their larvae drifted with the current into the estuary shortly after hatching. Not only physical factors, but also the development and ontogenetic behavior of the fish, played an important role in their transportation into the estuary.

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