# Changes in haemolymph ion concentrations of *Astacus astacus* L. and *Pacifastacus leniusculus* (Dana) after exposure to low pH and aluminium

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# Abstract

During exposure to soft water, acidified to pH 4.0, the haemolymph concentrations of Na<sup>+</sup>, K<sup>+</sup>, and Cldecreased whereas the Ca<sup>2+</sup> concentration fluctuated in *Astacus astacus*. The haemocyte content of K<sup>+</sup> decreased from 9% to 2% of the total haemolymph K<sup>+</sup> content after exposure to pH 3.7 for 3 days. Within 14 days, 250  $\mu$ g Al<sup>3+</sup> l<sup>-1</sup>, as Al<sub>2</sub>(SO<sub>4</sub>)<sub>3</sub> at pH 5.0, reduced the haemolymph Na<sup>+</sup> content in *Astacus astacus* and *Pacifastacus leniusculus*, however, the effects were less pronounced than earlier reported for fish. Disturbed ion regulation, mainly depending on low pH, is thought to contribute to the absence of these species in acid waters.

# Introduction

Effects of lake acidification on freshwater organisms have been thoroughly studied in both northern Europe and North America during the last ten years (e.g. Beamish & Harvey, 1972; Leivestad et al., 1976; Schofield, 1976; Almer et al., 1978). Low pH affects ion regulation in most freshwater organisms, as reported for fish by Packer & Dunson (1970), Leivestad et al. (1976), McWilliams (1980), McDonald & Wood (1981), Runn (1982) and several others. In decapods, mechanisms of ion regulation have been thoroughly studied (see Mantel & Farmer, 1982, for review), and effects of acid water on ion regulation of crayfish has been reported by Malley (1980) (Orconectes virilis), Morgan & McMahon (1982), McMahon & Morgan (1983) (Orconectes rusticus and Procambarus clarkii), and Appelberg (1984) (early stages of Astacus astacus). However, the effect of acidification on decapods is less well understood than it is on fish.

There is strong evidence that aluminium acts as a toxic agent in acidified natural waters, since it causes disturbances of ion regulation and blood

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oxygen tension in fish at sublethal pH levels (Leivestad *et al.*, 1976; Dickson, 1978; Baker & Schofield, 1980; Muniz & Leivestad, 1980a, b). It is mainly the inorganic aluminium fraction that is thought to be toxic to fish, and the effect is most pronounced in soft water at pH 5.0 to 5.5 (Muniz & Leivestad, 1980b; Dricsoll *et al.*, 1980). Although the drastic effects of aluminium on fish in acid water are well established, effects of aluminium on crayfish has not been reported.

In the present paper, the effect of exposure to acid water on the ion balance between haemolymph and the ambient water and on the ion balance between haemocyte and plasma of *Astacus astacus* has been studied. The effect of exposure to aluminium at pH 5.0 on the haemolymph ion concentration of *Astacus astacus* and *Pacifastacus leniusculus* is also reported.

# Material and methods

Three experiments were conducted. In experiment 1, eight A. astacus were exposed to water at pH 3.7, and sampling of haemocytes and plasma was made before exposure and after three days. In experiment 2, three groups of eight *A. astacus* were exposed to pH 6.8 (control), 5.0, and 4.0 for 11 days, and sampling of haemolymph was performed on days 0, 4, 8, and 11. In the third experiment the effect of aluminium at pH 5.0 was tested on both *A. astacus* and *P. leniusculus* at concentrations of 0, 250, 500, and 1000  $\mu$ g Al<sup>3+</sup> l<sup>-1</sup> (0, 9.3, 18.5, and 37.0 mmol l<sup>-1</sup>, respectively) for 14 days, using four individuals of each species in each group.

The A. astacus originated from the non-acidified lakes Tomtsjön (experiments 1 and 3) and Tisaren (experiment 2), and the P. leniusculus from Lake Erken, likewise non-acidified. All crayfish were caught in the autumn and kept in large plastic tanks at 10 °C, supplied with tap-water, during the winter. Only specimens that were regarded to be in intermoult stage (C4) were used in experiments. The time for acclimation to the experimental water and temperature was 6 days in experiments 1 and 3, and 8 days in experiment 2. The animals were not fed during the time of acclimation, nor during the experimental periods.

In experiment 1 the temperature, during the acclimation and experimental periods, was room temperature (20 °C), and in experiments 2 and 3 it was  $16.0 \pm 1.0$  °C.

To obtain water of constant quality, deionized water supplied with salts was used in all experiments. The major ions, as added, were Ca<sup>2+</sup> 0.090 mmol l<sup>-1</sup>, Mg<sup>2+</sup> 0.030 mmol l<sup>-1</sup>, Na<sup>+</sup> 0.090 mmol l<sup>-1</sup>, K<sup>+</sup> 0.015 mmol l<sup>-1</sup>, HCO<sub>3</sub>, Cl<sup>-</sup>, and NO<sub>3</sub> all 0.090 mmol l<sup>-1</sup> and SO<sub>4</sub><sup>2-</sup> 0.045 mmol l<sup>-1</sup>. The measured conductivity was 4.8 mS m<sup>-1</sup>, pH 6.8, and the calculated content of pCO<sub>2</sub>, after acidification of the water, was less than 2.5 mm Hg. Al<sub>2</sub>(SO<sub>4</sub>)<sub>3</sub> × 18 H<sub>2</sub>O (Mallininckrodt Chemical Works) was used to create different concentration of Al<sup>3-</sup> in the medium.

In experiment 1, a single 80 l aquarium was used. In experiments 2 and 3, an aquarium system composed of four separate units, consisting of four 13 l plastic beakers, connected to an 80 l aquarium, where replacements of water and pH adjustments were made. Two crayfishes were placed in each plastic beaker, and a circulating system between these and the aquarium was complemented with an extra aeration of the water. In experiments 1 and 2 the water was filtered, recirculated, and aerated, and half of the water volume was removed every third day. In experiment 3, flow-through water was used with a flow rate of  $0.5 \ h^{-1}$  through each 13 l beaker, and freshly mixed water was refilled every second day.

Samples of prebranchial haemolymph were collected from the ventral side of the abdomen, after washing with ethanol (70%). The soft part of the abdomen was pierced with a hypodermic needle  $(25 \times 0.6 \text{ mm})$ , and after the first drop of haemolymph was wiped away, 150  $\mu$ l samples were collected with a micropipette. Following sampling, the haemolymph of the test animals was allowed to clot by keeping the crayfishes in air for 15 min in order to minimize an unrecorded haemolymph loss. In experiment 1, plasma and haemocytes were separated by dropping 300  $\mu$ l haemolymph in a preweighed formaldehyde solution (3.3 mol l-1) to prevent disintegration of the haemocytes. Each sample was then reweighed and centrifugated at 700 g for 10 min (Beckman J-21C), and the haemocytes were washed three times with formaldehyde before being suspended in 0.1 M HCl. For measuring  $Ca^{2+}$ ,  $Na^{+}$ , and K<sup>+</sup>, amounts of 100  $\mu$ l haemolymph or 150  $\mu$ l plasma-formaldehyde solution were diluted in 10 ml0.1 M HCl. Measurements were made with an atomic absorption spectrophotometer (Unicam SP 1900). LaCl<sub>3</sub> was added to the  $Ca^{2+}$  samples to prevent binding to phosphate and organic material. For the  $K^+$  measurements, Na<sup>+</sup> was added to the standard solutions to reach a comparable concentration between these and the samples. 50  $\mu$ l haemolymph and 100  $\mu$ l plasma-formaldehyde solution were used for CI determinations, made within four hours with an automatic chloride titrator unit (Radiometer PHM 28).

In experiment 1, an unpaired Students t-test was used to determine whether differences in the concentrations of plasma and the haemocyte fraction before and after acid exposure were significant. The F-test was used to test the equality of variances (SAS Institute Inc., 1982). In experiments 2 and 3, differences between means of the control groups of treatment at day 0, were tested by an unpaired Students t-test. Analysis of variance, linear model (SAS Institute Inc., 1982), was used to test the significance of differences between changes in actual ion contents from day 0 and the following days of exposure. Least-square of means were calculated for each level of treatment.

## Results

Experiment 1 was performed to determine the ratio between the ion content in haemocytes and whole haemolymph of *Astacus astacus*, however, total haemocyte count was not determined. Haemocyte  $K^+$  was found to be an important component of the total  $K^+$  content of haemolymph, and ca. 9% of the total  $K^+$  content of haemolymph originated from the haemocytes in neutral water, while less than 1% of the total Na<sup>+</sup> content originated from the haemocytes (Table 1).

Acid exposure at pH 3.7 for three days reduced the haemocyte fraction of  $K^+$  significantly (p < 0.05), but the haemocyte Na<sup>+</sup> fraction was not affected. Contradictory, plasma Na<sup>+</sup> was reduced by acid treatment (p < 0.05), while plasma K<sup>+</sup> was not, and in the same experiment plasma  $Ca^{2+}$  rose and plasma Cl<sup>-</sup> decreased (p < 0.05 for both, Table 1). Total haemolymph concentrations of Na<sup>+</sup>, Ca<sup>2+</sup>, K<sup>+</sup>, and Cl<sup>-</sup> were significantly affected at pH 4.0 after 11 days (Fig. 1a-d). To test the significance of differences in concentrations between days and pH levels, analysis of variance was used. Na<sup>+</sup> and Cl<sup>-</sup> decreased at a constant rate during the experimental period, Na<sup>+</sup> slightly more rapidly than Cl<sup>-</sup>. The changes in  $K^+$  and  $Ca^{2+}$  concentrations were small; significant losses were recorded on day 11 at pH 4.0 (p < 0.05 and p < 0.01, respectively). A significant difference in Ca<sup>2+</sup> concentration between the control group and the other two groups at day 0 (p < 0.05) was also found. Although the observed

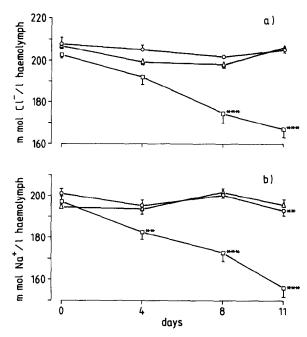


Fig. 1a and b. Haemolymph concentrations of Cl<sup>-</sup> and Na<sup>+</sup> in Astacus astacus exposed to acid water,  $\triangle - \triangle pH = 6.8$  (control group),  $\bigcirc - \bigcirc pH = 5.0$ ,  $\Box - \Box pH = 4.0$ . Means  $\pm$  S.E.M. N = 8. Asterisks indicte mean changes from day 0 significantly different from the control group by analysis of variance, least square-means. \* p < 0.05, \*\* p < 0.01, \*\*\* p < 0.001. When no vertical bar is present, the dot representing the mean is larger than S.E.M.

changes in haemolymph ion content were less than 30% at pH 4.0 after 11 days exposure, a significant loss of Na<sup>+</sup> and Ca<sup>2+</sup> were recorded at pH 5.0 (p < 0.01 for both), however, the losses were

Fraction	Before exposure (pH 6.8)		After exposure for 3 days to pH 3.7	
	μmol/g haemolymph	% of total haemolymph	μmol/g haemolymph	% of total haemolymph
Na <sup>+</sup>	······································		······································	
Plasma	191.1 ± 2.5	99.06	142.8 ± 5.7*	98.99
Haemocyte	$1.82\pm0.23$	0.94	$1.45 \pm 0.45$	1.01
K+				
Plasma	$3.45 \pm 0.13$	91.27	$3.38 \pm 0.23$	97.97
Haemocyte	$0.33 \pm 0.11$	8.73	$0.07 \pm 0.02*$	2.03
Ca <sup>2+</sup>				2.00
Plasma	$8.94 \pm 0.33$	-	$11.00 \pm 0.87*$	
Cl				
Plasma	$198.5 \pm 1.9$		186.8 ± 4.2*	

Table 1. Ion content in the plasma and haemocyte fractions of haemolymph of Astacus astacus before and after 3 days of exposure to pH 3.7. Means  $\pm$  S.E.M., N = 8.

\* Indicates means significantly different (p < 0.05) from corresponding pre-treatment value by unpaired t test.

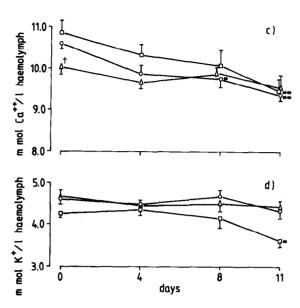


Fig. 1c and d. Haemolymph concentrations of Ca<sup>2+</sup> and K<sup>+</sup> in Astacus astacus exposed to acid water,  $\triangle - \triangle pH = 6.8$  (control group),  $\bigcirc - \bigcirc pH = 5.0$ ,  $\square - \square pH = 4.0$ . Means  $\pm$  S.E.M. N = 8. Asterisks indicate mean changes from day 0 significantly different from the control group by analysis of variance, leastsquare means. \* p < 0.05, \*\* p < 0.01, \*\*\* p < 0.001. + indicate that the mean of the control group is significantly different (p < 0.05) from the means of the other two groups on day 0 by unpaired *t* test. When no vertical bar is present, the dot representing the mean is larger than S.E.M.

smaller than at pH 4.0. The constant concentrations of the control group indicate that neither repeated sampling, nor starvation during the experimental period had any significant effect on the measured ion concentrations.

In order to study the effect of  $Al^{3+}$  on the haemolymph ion content of *A. astacus* and *P. leniusculus*, both species were exposed to four different concentrations of  $Al^{3+}(0, 250, 500, \text{ and } 1000 \ \mu g \ Al^{3+} l^{-1})$  at pH 5.0 for 14 days. Haemolymph was sampled at neutral pH before the addition of  $Al^{3+}$  (day 0), and at the 6th and 14th days (Fig. 2a-d). Changes in ion concentrations from day 0 to days 6 and 14 were compared between the control group and the experimental groups, and after 14 days significant reductions of haemolymph Na<sup>+</sup> were noted in all groups treated with aluminium (p < 0.05).

There was no significant difference on any day between the two species (Fig. 2a-d), nor was there any significant difference between the Na concentrations of the four different groups at day 0.

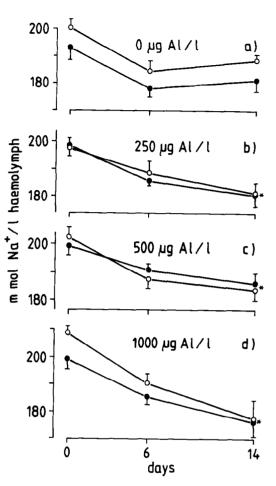


Fig. 2a-d. Haemolymph concentrations of Na<sup>+</sup> in Astacus astacus •---• and Pacifastacus leniusculus O---O during exposure to different concentrations of aluminium in water at pH 5.0. Means  $\pm$  S.E.M., N =4. Asterisks indicate mean changes of both species from day 0, significantly different (p < 0.05) from the control group (0  $\mu$ g Al<sup>3+</sup>/1) by analysis of variance, least-square means.

## Discussion

Studies by Höglund & Härdig (1969) and Neville (1979) show that a clear distinction between the effects of low pH and a raised level of free carbon dioxide, due to acidification, must be made. An increase of the carbon dioxide in the ambient water may induce a hypercapnic acidosis and affect haemolymph ion content (Dejours & Armand, 1982). In the experimental water, used in the present study, the calculated value for the maximal  $pCO_2$ was less than 2.5 mm Hg, which indicates that the changes in haemolymph ion content were not due to high  $pCO_2$  of the media. A comparison of ion concentrations in haemocytes and plasma shows that haemocyte  $K^+$  constitutes a considerable part of the total amount of haemolymph  $K^+$ . The total haemocyte count of haemolymph in crustaceans can be affected by several factors such as temperature, cuticle damages, and infections (Söderhäll, 1982; Smith & Söderhäll, 1983), and therefore care has to be taken when evaluating changes of  $K^+$  concentrations in samples of total haemolymph. The constant amount of Na<sup>+</sup> in the haemocyte fraction in experiment 1, indicates that the decrease of  $K^+$  in the haemocyte fraction after acid exposure is due to a loss of intracellular  $K^+$  and not a loss or lysis of haemocytes.

Acid exposure of fish has been reported to produce a loss of intracellular K<sup>+</sup> from heart tissue and muscle tissue to the plasma fraction (Fugelli & Vislie, 1980, and McDonald & Wood, 1981, respectively), which corresponds to the loss of haemocyte K<sup>+</sup> found here. It is suggested that this intracellular loss is a result of K<sup>+</sup> acting as a regulator of the cell volume, and caused by extracellular ion losses (Fugelli & Vislie, 1980). The non-linear decrease of haemolymph K<sup>+</sup>, recorded at pH 4.0, indicates a net loss of K<sup>+</sup> similar to that reported for *Salmo* gairdneri by McDonald & Wood (1981). A reduced net accumulation rate of K<sup>+</sup> is also found in developing eggs of A. astacus exposed to acid water at pH 4.9 (Appelberg, 1984).

A reduction of plasma concentrations of Na<sup>+</sup> and Cl- during acid stress has previously been reported for several fish species (Packer & Dunson, 1970; Leivestad et al., 1976; McWilliams, 1980; McDonald & Wood, 1981), and for crayfish by Morgan & McMahon (1982) and McMahon & Morgan (1983). This loss is probably due to a reduced uptake rate of Na<sup>+</sup> and Cl<sup>-</sup> over the gills, resulting largely from a disturbance of the Na<sup>+</sup>/H<sup>+</sup> and Cl<sup>-</sup>/basic anion exchange mechanisms, with a possible contribution from an increased passive efflux of Na<sup>+</sup> (Morgan & McMahon, 1982). The reduction of haemolymph Cl- may also be related to the actual decrease in haemolymph Na<sup>+</sup>, since the uptake rate of Cl- is dependent on haemolymph Na<sup>+</sup> concentration (Shaw, 1964).

Morgan & McMahon (1982), who recorded a drastic increase of haemolymph  $Ca^{2+}$  in *Procambarus clarkii* when exposed to water at pH 3.8, suggest that exoskeletal carbonate is used to buffer an internal acidosis. In the present study, plasma  $Ca^{2+}$ 

rose in experiment 1, when exposed to pH 3.7, while there was a slight decrease in total haemolymph  $Ca^{2+}$  in experiment 2, when exposed to pH 4.0 and 5.0. It is possible that the difference in pH levels may influence the suggested buffering by  $CaCO_3$  of the haemolymph, but the rate of  $Ca^{2+}$  loss from haemolymph may also be attributed to differences within the intermoult stage, with is found to influence the  $Ca^{2+}$  metabolism (Greenaway, 1972).

Despite a significant loss of Na<sup>+</sup> within 14 days, at an aluminium concentration of 250  $\mu$ g l<sup>-1</sup> at pH 5.0, the present effects on both A. astacus and P. leniusculus are small in comparison with those reported for fish. Driscoll et al. (1980), recorded a 50% survival time of 115 h of brook trout, Salvelinus fontenalis, at pH 5.2 and 0.42 mg Al<sup>3+</sup> l<sup>-1</sup>, and Muniz & Leivestad (1980b), recorded toxic effects of 190  $\mu$ g Al<sup>3+</sup> l<sup>-1</sup> at pH 5.0 to brown trout Salmo trutta. However, Havas & Hutchinson (1982) found that high concentrations of aluminium  $(20 \text{ mg} l^{-1})$  is highly toxic to smaller crustaceans, and it is possible that a higher aluminium concentration would have caused a more drastic response in the present study. The decrease in Na<sup>+</sup> content in all groups (Fig. 2) between days 0 and 6, seems to be a true effect of lowered pH at day 0 only.

Differences in ion composition of experimental waters influences comparisons between different studies, since it is known that the uptake rate of Na<sup>+</sup>, Cl<sup>-</sup>, and Ca<sup>2+</sup> in crayfish to some extent depends on the concentrations in the medium (Shaw, 1959, 1960; Greenaway, 1974). However, the changes in haemolymph ion concentration reported for *A. astacus* in the present study, in general agree with the results of Morgan & McMahon (1982) and McMahon & Morgan (1983) obtained for *Procambarus clarkii* and *Orconectes rusticus* during acid exposure.

The present study indicates that also slightly acid water (pH 5.0) may, in the long run, disturb the ion regulation in *A. astacus* and preliminary studies gave similar effects of acid stress on *P. leniusculus*. The absence of these species in acid waters (Svärdson, 1974), is possibly due to a complex of factors; changes of habitat, predators, inter-specific competition, and food organisms. As judged from the present findings, disturbance of the ion regulation, mainly depending on low pH, and to lesser extent possibly also aluminium, may contribute to this. More susceptible stages in the life history, such as reproduction and moulting, seem to be more critical during acid exposure compared to the intermoult, adult, crayfish used in the present study (Malley, 1980; Appelberg, 1984).

## Summary

- Haemolymph concentrations of Na<sup>+</sup>, K<sup>+</sup>, and Cl<sup>-</sup> decreased, whereas the concentration of Ca<sup>2+</sup> fluctuated in Astacus astacus L. during exposure to soft, acid water. Weak acid stress, pH 5.0, reduced the content of both Na<sup>+</sup> and Ca<sup>2+</sup> in the haemolymph within 11 days.
- 2. After exposure to pH 3.7 for 3 days the content of the haemocyte K<sup>+</sup> was reduced.
- Exposure to 250 μg Al<sup>3+</sup>l<sup>-1</sup> in water at pH 5.0 for 14 days, reduced the haemolymph content of Na<sup>+</sup> in Astacus astacus and Pacifastacus leniusculus.

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#### References

- Almer, B., W. Dickson, C. Ekström & E. Hörnström, 1978. Sulfur pollution and the aquatic ecosystem. In Sulfur in the environment; part II. In J. O. Nriagu (ed.), Ecological impacts. Publ. John Wiley & Sons Inc.: 273-286.
- Appelberg, M., 1984. Early development of Astacus astacus L. in acid water. Rep. Inst. Freshw. Res., Drottningholm, (in press).
- Baker, J. P. & C. L. Schofield, 1980. Aluminium toxicity to fish as related to acid precipitation and Adirondack surface water quality. In D. Drablös & A. Tollan (eds.), Proc., Int. conf. ecol. impact acid precip. Norway 1980, SNSF-project: 292-293.
- Beamish, R. J. & H. H. Harvey, 1972. Acidification of the La Cloche Mountain Lakes, Ontario, and resulting fish mortalities. J. Fish Res. Bd Can. 31: 49-54.
- Dejours, P. & J. Armand, 1982. Variations de l'équilibre acidebase de l'hemlymphe d'Ecrivisse en fonction des changements de certaines proprietes physico-chimiques des l'eau ambiante. C.R. Acad. Sci. Paris 295: 509-512.
- Dickson, W., 1978. Some effects of the acidification of Swedish lakes. Verh. Internat. Verein. Limnol. 20: 851-856.

- Driscoll, C. T., J. P. Baker, J. J. Bisogni, Jr. & C. L. Schofield, 1980. Effects of aluminium speciation of fish in dilute acidified waters. Nature 284: 161–164.
- Fugelli, K. & T. Vislie, 1980. Physiological responses to acid water in fish. 1. Effects from changes in plasma osmoality and ion concentration on heart ventricle water content and intracellular concentrations of K<sup>+</sup> and taurine in the brown trout Salmo trutta. In D. Drablös & A. Tollan (eds.), Proc., Int. conf. ecol. impact acid precip. Norway, 1980, SNSFproject: 346-347.
- Greenaway, P., 1972. Calcium regulation in the freshwater crayfish Austropotamobius pallipes (Lereboullet). J. exp. Biol. 57: 471-487.
- Greenaway, P., 1974. Calcium balance at the postmoult stage of the freshwater crayfish Austropotamobius pallipes (Lereboullet). J. exp. Biol. 61: 31-45.
- Havas, M. & T. C. Hutchinson, 1982. Aquatic invertebrates from the Smoking Hills, N.W.T.: effect of pH and metals on mortality. Can. J. Fish. Aquat. Sci. 39: 890-903.
- Höglund, L. B. & J. Härdig, 1969. Reactions of young salmonids to sudden changes of pH, carbon dioxide tension and oxygen content. Rep. Inst. Freshwat. Res., Drottningholm 49: 76-119.
- Leivestad, H., G. Hendrey, I. P. Muniz & E. Snekvik, 1976. Effects of acid precipitation on freshwater organism. In Impact of Acid Precip. on Forest and Freshwat. Ecosystems in Norway, SNSF-project: 86-111.
- Malley, D. F., 1980. Decreased survival and calcium uptake by the crayfish Orconectes virilis in low pH. Can. J. Fish. Aquat. Sci. 37: 364-372.
- Mantel, L. H. & L. L. Farmer, 1983. Osmotic and ionic regulation. In L. H. Mantel (ed.), The Biology of Crustacea Vol. 5. Academic Press: 53-161.
- McDonald, D. G. & C. M. Wood, 1981. Branchial and renal acid and ion fluxes in the rainbow trout, *Salmo gairdneri*, at low pH. J. exp. Biol. 93: 101–118.
- McMahon, B. R. & D. O. Morgan, 1983. Acid toxicity and physiological responses to sub-lethal acid exposures in crayfish, pp. 71-85. In C. R. Goldman (ed.), Freshwater Crayfish V. Papers from the Fifth International Symposium on Freshwater crayfish. Davis, California, USA, 1981.
- McWilliams, P. G., 1980. Effects of pH on sodium uptake in Norwegian brown trout Salmo trutta from an acid river. J. exp. Biol. 88: 259-267.
- Morgan, D. O. & D. R. McMahon, 1982. Acid tolerance and effects of sublethal acid exposure on iono-regulation and acid-base status in two crayfish *Procamabarus clarkii* and *Orconectes rusticus*. J. exp. Biol. 97: 241-252.
- Muniz, I. P. & H. Leivestad, 1980a. Acidification effects on freshwater fish. In D. Drablös & A. Tollan (eds.), Proc., Int. conf. ecol. impact. acid precip. Norway 1980, SNSF-project: 84-98.
- Muniz, I. P. & H. Leivestad, 1980b. Toxid effects of aluminium on the brown trout, Salmo trutta L. In D. Drablös & A. Tollan (eds.), Proc., Int. conf. ecol. impact acid precip. Norway 1980, SNSF-project: 320-321.
- Neville, C. M., 1979. Ventilatory response of rainbow trout Salmo gairdneri to increased H<sup>+</sup> ion concentration in blood and water. Comp. Biochem. Physiol. 63A: 373-376.
- Packer, R. K. & W. A. Dunson, 1970. Effects of low environ-

mental pH on blood and sodium balance on brook trout. J. exp. Biol. 174: 65-72.

- Runn, P., 1982. Effects of low pH on the development of early stages of fish. Ph.D. Thesis, Acta Univ. Upsal.: 647.
- Schofield, L. C., 1976. Acid precipitation: effects on fish. Ambio 5: 228-230.
- SAS Institute Inc., 1982. SAS User's Guide: Statistics, 1982 Edition. Cary. N.C. SAS Institute Inc., 1982: 584 pp.
- Shaw, J., 1959. The absorption of sodium ions by the crayfish Astacus pallipes Lereboullet. 1. The effect of external and internal sodium concentrations. J. exp. Biol. 36: 126-145.
- Shaw, J., 1960. The absorption of chloride ions by the crayfish Astacus pallipes Lereboullet. J. exp. Biol. 37: 557-573.

Shaw, J., 1964. The control of salt balance in the crustacea.

Symp. Soc. exp. Biol. 18: 237-256.

- Smith, V. J. & K. Söderhäll, 1983. Beta-1.3-glucan activation of crustacean hemocytes in vitro and in vivo. Biol. Bulletine, vol. 164: 299-314.
- Svärdson, G., 1974. Översikt över laboratoriets verksamhet med plan för år 1974. Inf. från Sötvattenslaboratoriet, Drottningholm 1: 1–26.
- Söderhäll, 1982. Prophenoloxidase activating system and melanization – a recognition mechanism of arthropods? A review. Developmental Comparative Immunology 6: 601–611.

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