¹³⁷Cs uptake in spring wheat (*Triticum aestivum* L. cv. Tonic) at varying K supply

II. A potted soil experiment

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Abstract

Spring wheat (Triticum aestivum L. cv. Tonic) was grown for 16 days in a sandy loam soil which was contaminated with ¹³⁷Cs. The soil was fertilised with K at three rates (0,1 and 2 mmol K per 950 g dry soil) and with NO_3^- -N at two rates (0 and 2 mmol per 950 g dry soil) in a factorial design. The ¹³⁷Cs Activity Concentration (AC) in the shoot tissue significantly reduced 8.2-fold (nil N treatment, p<0.001) and 9.3-fold (highest N dose, p<0.001) with increasing K supply. In contrast, the K application increased the ¹³⁷Cs AC in soil solution 1.7 fold (nil N treatment) or had no significant effect (highest N dose). At similar K application, the application of N increased the ¹³⁷Cs AC in the shoot compared to the control. This effect is most probably due to the increased NH₄⁺ concentration in soil solution which increased the ¹³⁷Cs AC in soil solution composition (¹³⁷Cs and K concentration) in the rhizosphere was estimated from the average soil solution composition at day 16 and solute transport calculations. The ¹³⁷Cs AC in the shoot tissue was predicted from the estimated soil solution composition in the rhizosphere and the relationship between K concentration and ¹³⁷Cs uptake derived from a nutrient solution experiment. The predictions of ¹³⁷Cs AC's in the shoot are qualitatively correct for the fertiliser effects but underestimate the observations between 1.4 and 9.9 fold.

Introduction

Potassium fertilisation has been shown to be efficient in reducing ¹³⁷Cs availability to crops growing in contaminated soils. In comparative studies on countermeasures, K fertilisation is generally more efficient than liming (e.g. Lönsjö et al., 1990; Mitchell et al., 1990; Rosén, 1992) and has a more immediate effect than application of zeolites which are used to immobilise radiocaesium (Jones et al.,1994). A large number of field experiments on the effect of K fertilisation have been carried out in the area affected by the Chernobyl accident of 1986 such as Ukraine (Prister et al.,1993), Sweden (Lönsjö et al., 1990; Rosén,1992) and the upland areas of the UK (Jackson and Nisbet, 1990; Jones et al., 1994; Mitchell et al., 1990). The reduction factors (ratio of ¹³⁷Cs activity concentration in plants of untreated versus treated plots) vary widely with the type of soil and crop used. In peat soils or peat bogs, reduction factors for grasslands vary from 1.1 (Mitchell et al., 1990) to over 12 (treatments up to 200 kg K ha⁻¹, Lönsjö et al., 1990). Potassium fertilisation can also reduce ¹³⁷Cs transfer from peat soils to barley (Rosén, 1992). In mineral soils, reduction factors vary from almost unity in leek and carrots grown on a silt loam soil to over 5 in barley grown in a sandy loam (Rosén, 1992). Larger reduction factors of over 30 have been found in long-term studies (1961-1981) with oats, peas and white mustard grown in mineral soils (Lönsjö et al., 1990). These factors may perhaps not reflect common agricultural conditions as the control plots were probably strongly depleted in nutrients.

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The reduction in ¹³⁷Cs availability by increased K supply has also been found in a number of pot trials (Belli et al., 1995; Fawaris and Johanson, 1994; Fowler and Christenson, 1959; Grauby et al., 1990; Jackson et al.,1965; Nishita et al.,1960, 1962). It is generally found in these studies that enhanced K supply reduces ¹³⁷Cs uptake substantially at low K availability in soil and reduction factors up to 100 have been demonstrated (Grauby et al.,1990). At high K availability in soil however, enhanced K supply has only little effect or even stimulates ¹³⁷Cs uptake (Fowler and Christenson, 1959; Grauby et al.,1990; Nishita et al.,1960).

Net ¹³⁷Cs uptake from soil results from a combination of the effects of the added cations on maintaining ¹³⁷Cs in (soil) solution and on inhibition of uptake and transport processes (Jackson et al., 1965). The two processes act in an opposite way. No attempt has yet been made to separate the effect of K supply on ¹³⁷Cs desorption from soil and its effect on the root uptake process. In mineral soils, increased K in solution increases the ¹³⁷Cs activity in solution. This has been quantified in the radiocaesium interception potential theory which state that the ¹³⁷Cs solid/liquid distribution constant K_D equals

$$K_D = \frac{RIP}{m_{\rm K} + 5m_{\rm NH_4}} \tag{1}$$

(De Preter, 1990) in which $m_{\rm K}$ and $m_{\rm NH_4}$ represent the K and $\rm NH_4^+$ concentrations in solution and RIP is the radiocaesium interception potential, a soil characteristic which can be measured experimentally (Sweeck et al.,1990). If $\rm NH_4^+$ concentration in soil is low, K additions to a soil will result in a proportional increase in both K and 137 Cs concentration in soil solution. In organic soils (e.g. peat soils) a considerable fraction of 137 Cs is adsorbed to organic matter. In these soils, the effect of K addition on the 137 Cs activity in solution varies from almost absent to a significant increase. This increase is however less than proportional to the K concentration increase (Nisbet et al.,1994).

Solution culture studies showed that increasing K concentrations in solution decrease ¹³⁷Cs uptake sharply at micromolar concentrations of K (Shaw et al.,1992). Previously, we demonstrated that between 50 and 250 μ M K, ¹³⁷Cs uptake was more than proportionally affected by $1/m_{\rm K}$ (Smolders et al., 1996). From this observation it follows that, even if ¹³⁷Cs levels in soil solution rise along with increasing K concentration, a reduction in net uptake by increased K supply could be possible because of the larger effect on the uptake process.

In this paper, we present results of a potted soil experiment in which the effect of increased K supply on (a) ¹³⁷Cs uptake in wheat and on (b) soil solution ¹³⁷Cs activity was measured. It is assessed whether the effect can be predicted based on the solution culture observations (Smolders et al., 1996) and soil solution composition (¹³⁷Cs and K concentrations). The potted soil experiment was carried out at similar environmental conditions as the solution culture experiment. The soil solution composition at the soil/root interface is approximated from the average soil solution composition with a solute transport calculation. The experiment was set-up in a factorial design with three rates of K and two rates of N (added as NO_3^-). The NO_3^- application was included to assess the effect of an increased plant growth rate independent of K nutrition.

Materials and methods

Soil sampling and treatments

A silt-loam soil (USDA) was sampled from the E horizon of a luvisol (FAO,1990). This mineral soil was collected in a forest and was selected for its low K availability at which the effect of added K on ¹³⁷Cs availability can be large. The soil was sieved (3 mm) and amended with lime (10 mmol Ca(OH)₂ per kg dry soil) and phosphorous (1.05 mmol CaHPO₄ kg⁻¹ dry soil). Treatments were three rates of K (0,1 and 2 mmol K 950 g^{-1} dry soil, the quantity of soil to fill one pot) and two rates of NO_3^- -N (0 and 2 mmol NO_3^- /pot) in a factorial design with 5 fold replication. K was added as a Cl salt at the nil N treatment and as a NO_3^- salt in the 2 mmol N treatment. NO_3^- was added as a mixture of Ca and K salts depending on the K application rate. Selected characteristics of the amended soil at the nil K and N treatment are given in Table 1. The soil amendments (application of lime, P, N and K),¹³⁷Cs contamination and moistening to 21% (g g^{-1} oven dry soil, 86% of water holding capacity) were carried out by mixing 5.5 kg soil with appropriate volumes of stock solutions (N, K, and ¹³⁷Cs), amount of dry salts (lime and P) and volumes of distilled water. Mixing was made using a conventional kitchen mixer. Application of carrier-free ¹³⁷Cs was about 1 MBq per pot, added through a 10 mL solution per 5.5 kg soil. Per treatment, another 10 uncontaminated replicates were prepared. These replicates (further denoted as 'unlabelled') were used to quantify shoot and root growth rate at the different treatments. These biomeTable 1. Selected characteristics of the silt loam soil after liming and P application at zero N and K application

pH(KCl)	4.7
% Clay	9
% Silt	61
% Sand	30
C(%)	1.1
Macronutrients(cmol kg ⁻¹)	
Р	0.064
K ^a	0.17
Mg^a	0.04
Ca ^a	1.2
Na ^a	0.034
NH_4^{+b}	0.029
NO_3^{-b}	0.016
-	
CEC^c (cmol ⁺ kg ⁻¹)	6.5

^a In 0.1 *M* ammoniumlactate extract (1:20) and measured colorimetrically by ammonium vanadate (P) or by atomic absorption spectophotometry (K, Mg, Ca, Na). ^b In 1 *N* KCl extract and measured colori-

metrically using sulfanylamide and N-(1naftyl)ethylenediaminedihydrochloride (NO₃) or indophenylblue method (NH₄). ^c Ammoniumacetate method.

tric data are essential for the solute transport calculation (see below).

All potted soils were installed in a growth cabinet at 25 °C, 40% relative humidity and high light intensity (650 μ mol photons m⁻² s⁻¹). Soils were allowed to dry in these condition for 4 days, remoistened to 24% and redried for 3 days. Soils were then finally remoistened to 21 % after which they were covered and incubated for 4 days in darkness at 18.5 °C prior to seed transplanting.

Plant growth

Seeds of spring wheat (*Triticum aestivum* L., cv. Tonic) were soaked in water and germinated on moist tissue paper in a plastic bag at 25 °C. One day later, germinated seeds were transplanted to the pots (two per pot). The soil surface was covered with polythene beads to reduce surface evaporation. Plants were grown in a growth cabinet (Weiss, 18'SP/+5Ju-Pa) with a 12h/12h day/night cycle. Air temperature was 18.5 °C during the day and 14.5 °C during the night. Relative humidity was 75% throughout. Light intensity at

canopy height was 650 (\pm 30) μ mol photons m⁻² s⁻¹. Daily water loss from the pots was recorded by weighing and adjusted with deionised water. At day 7 (day 0 is day of germination), plants were thinned to one per pot.

Harvest

Unlabelled plants were harvested at day 13 and day 17, removing 5 replicates at each occasion. Labelled plants (shoots only) were harvested at day 16. The shoot fresh weight of the labelled plants was recorded, transferred into 5 mL counting vials and their gamma activity measured (Minaxi, 5530 auto- γ ,1 h per sample). ¹³⁷Cs activities ranged between 18 cpm and 632 cpm per plant. Plants were subsequently dried at 65 °C and reweighed. The shoots of the unlabelled plants were harvested and their roots were isolated (two replicates per treatment only). Roots were isolated by gently washing the soil from roots over a 2 mm sieve. Roots were further cleaned from soil and debris in a Na₂P₂O₇ solution and subsequently weighed. Root fresh weight was recorded. A subsample of about 0.2 g fresh weight (15 to 50% of the root sample) was coloured with safranine, cut to 1-2 cm pieces and root length was measured with the line intersection method on a 0.5 cm grid (Tennant, 1975). Root diameter was measured on 20 roots per sample. The subsample was then combined with the remaining root sample. All plant tissues were dried (65 °C) and reweighed.

Soil solution isolation

Soil solution was sampled from all labelled replicates at harvest. Per pot, the majority of the roots were removed by hand-picking and the soil was homogenised. A syringe from which the plunger was removed was filled with glass wool and 50 g soil. A holder with a 0.22 μ m membrane filter (Millipore) was connected to the syringe outlet. The syringe was transferred to a centrifuge tube and centrifuged for 1 h at about 1100 RCF. For each pot, two soil samples were centrifuged and the collected solutions (about 2.5 mL each) were combined. The ¹³⁷Cs activities in the combined solutions were counted for 1 h and varied from 12 to 78 cpm per sample.

Chemical analyses

The soil solution was analysed for K, Ca and Mg by Atomic Absorption Spectrophotometry (AAS), for

 NH_4^+ by colorimetry (indophenol-blue method) and for Cl^- , NO_3^- and SO_4^{2-} by Ion Chromatography (Dionex, AS4A column). The plant tissues (all replicates for the labelled plants, and two plants per harvest and treatment for unlabelled plants) were digested in HNO₃ and K in the digest was measured by AAS. No clear digest was obtained of the root tissues of the unlabelled replicates. K concentrations in these digests were low and irreproducible, hence no K content of the roots could be obtained.

Data analysis and soil solute transport calculation

Data were subjected to a two way ANOVA for the effect of K and N fertilisation and their interaction.

The soil solution composition (K and ¹³⁷Cs concentration) at the root surface was calculated from the average soil solution composition and soil and plant parameters. The approach of estimating rhizosphere composition by calculation is similar to that used by Rengel and Robinson (1990). The approximation of Baldwin et al. (1973) was used to calculate the steadystate concentration profile around the root.

For a randomly dispersed root at constant root density in a finite volume of soil, assuming only radial transport in the solution phase by diffusion and mass flow, the concentration of solute X at the root surface, $m_{X,a}$, is given by Baldwin et al. (1973) as:

$$m_{X,a} = m_X \left[\frac{\alpha}{\nu} + \left(1 - \frac{\alpha}{\nu} \right) \left(\frac{2}{2 - \frac{a\nu}{Db}} \right) \frac{\left(\left(\frac{x}{a} \right)^{2 - a\nu/Db} - 1 \right)}{\left(\left(\frac{x}{a} \right)^2 - 1 \right)} \right]$$
(2)

 m_X is the average concentration in soil solution, α is the root absorption power (solute uptake rate on a root area basis divided by solute concentration, m day⁻¹), ν is the water flux at the root surface (m day⁻¹), a is the root radius (m), D is the effective diffusion coefficient of the solute in the soil (m² day⁻¹), b is the buffer power (see below) and x is the radius of the soil cylinder that can be exploited by the root. The radius x is calculated as

$$x = \frac{1}{\sqrt{\Pi L}} \tag{3}$$

L is the root density (root length/soil volume, $m m^{-3}$).

Whereas in the solute transport model of Baldwin and co-workers the root absorption power is an input parameter derived from solution culture studies (e.g. Bhat et al.,1979a and b) it was derived here from the observed uptake rate in this soil experiment. This method corresponds to that used by Rengel and Robinson (1990). For this purpose, Equation (2) was modified by replacing α by U_a/m_{xa} in which U_a is the solute uptake rate per unit root surface. This modified equation was solved for $m_{X,a}$ In this way, the steadystate concentration profiles of K and ¹³⁷Cs in soil solution around the root were calculated from their average concentrations in soil solution and the observed uptake rates. The input parameters were found as follows:

 m_x : the average of the soil solution concentrations of K and ¹³⁷Cs respectively at day 16.

 U_a : for both K and ¹³⁷Cs, U_a at day 16 was calculated from the ratio of the uptake rate per unit root dry weight (average for each treatment) and the average root area per root dry weight (average of all treatments and harvests). For K, the uptake rate was calculated from the difference in K content (mmol plant $^{-1}$) between the successive harvests of unlabelled replicates. As mentioned in previous section, K analysis on roots failed, and, based on the solution culture experiment (Smolders et al., 1996), an average content of 1.4 mmol g^{-1} dry weight root tissue was assumed. The uptake rate per root weight was derived by dividing the uptake rate by the average root weight of successive harvests. This implies that root growth rate is assumed to be linear between the two harvests. The average root surface area per unit root dry weight was calculated from the average root radius of each subsample and its total root length per root dry weight. From all samples (all treatments and harvests), an average of 0.150 (± 0.034) m² g⁻¹ dry weight was derived. The ¹³⁷Cs uptake rate was calculated from the shoot and root growth rate of the unlabelled replicates (averages per treatment) and, for the shoot, a ¹³⁷Cs AC equal to that observed at day 16 (averages per treatment). Solution culture studies showed that in wheat, concentration changes with growth stage were small (Smolders and Shaw,1995). Root concentrations of ¹³⁷Cs were not measured as it was anticipated that contamination of roots by soil particles would make root activities difficult to interpret. Therefore, it was assumed that ¹³⁷Cs concentration in the root tissue is 5 fold higher than the shoot concentration, corresponding to solution culture observations at 250 μM K (Smolders et al., 1996). This value was found to increase at lower K concentrations, but it affects our calculation results slightly in only one treatment (see Results). ¹³⁷Cs uptake rate per root area was calculated from ¹³⁷Cs uptake rate as for K.

v: The net daily water uptake in the plant was measured from the difference in daily water loss between planted pots and the initial daily water loss when plants were small. The cumulative net water uptake of each plant was plotted to the shoot biomass and a linear regression was made. The slope of the line was 190 mL g⁻¹ dry weight(±50). This value was converted to average water intake per unit root area, v, using the shoot relative growth rate (average per treatment), the shoot:root dry weight ratio (average per treatment of second harvest), and the root area per root dry weight (average of all samples).

a : an average root radius of 0.17 mm (± 0.03) was derived from all treatments and harvests.

Db: the product Db equals $D_1\Theta f_1$ (Baldwin et al., 1973) in which D_1 is the diffusion coefficient in water, θ is the volumetric moisture content and f_1 is the tortuosity factor. D_1 of ¹³⁷Cs was assumed equal to that of K as 1.74 10⁻⁵ cm² s⁻¹ (Robinson and Stokes,1965). At the volumetric soil moisture content of 0.22, a tortuosity factor of 0.18 was selected for the silt-loam soil (Barraclough and Tinker, 1981).

x: The root length per soil volume (900 mL pot⁻¹) was averaged for each treatment at each harvest of the unlabelled plants. The root density at day 16 was then estimated from the regression line of the root densities at day 13 and 17. The value of x was then derived according to Equation 3 and varied between 5 and 6 mm according to the treatment.

In our calculations, we assume that the concentration profile around the roots is in a steady-state. It can be demonstrated that the concentration profile readily reaches a steady-state if the value of the ratio $\alpha a/Db$ is small. Solute uptake calculated from the steady state concentration profile in a stepwise manner (i.e. by adjusting the concentration profile in each step to account for the amount of solute absorbed by the root) has been compared to numerical non-steady state calculations. Good correspondence between both methods was found when the ratio $\alpha a/Db$ is below 5.0 (Baldwin, 1973). Using our parameter values, the ratio for K and ¹³⁷Cs was maximum 0.004. Hence, the use of the steady-state approximation is justified to calculate the K and ¹³⁷Cs concentration gradients for our conditions.

Results

Shoot growth was significantly enhanced by the N application whereas the K application only slightly

Table 2). Roc

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enhanced growth at the nil N treatment (Table 2). Root biomass in contrast was unaffected by the fertiliser treatments (Table 3, unlabelled replicates). Root length per plant was not significantly affected by fertilisation although a decreasing trend with increasing K application was found at the 2N treatment. The root:shoot weight ratio was always highest at the nil N treatment compared to the N treatment. The data of the specific root length (length per root dry weight) and the root radius were subjected to a statistical analysis. No treatment effects (p>0.05) were found on both parameters except for a significant N-K interaction on specific root length at the first harvest: at increasing K supply, specific root length decreased at the 2N treatment only (details not shown).

Increasing K supply increased the K concentration in soil solution about fourfold (Table 2). A similar increase was found for the NH_4^+ concentration at the nil N treatment. In the NO_3^- supplied soil, NH_4^+ concentrations (over 1 m*M*) were high and increased slightly with K application. Ca and Mg concentrations in soil solution also increased by K application (figures not shown). The high NH_4^+ concentration at the 2N treatment were rather unexpected as N was given as $NO_3^$ only. Inhibition of nitrification of mineralised N by high NO_3^- concentrations and low soil pH values may explain this result.

The ¹³⁷Cs AC in the shoot decreased sharply by increasing K application rate (Fig. 1). Application of NO_3^- -N enhanced the ¹³⁷Cs AC. The reduction factors at a K application of 2 mmol pot^{-1} were 8.2 at the nil N treatment and 9.3 at the 2 N treatment compared to the nil K treatment. In contrast, ¹³⁷Cs AC's in soil solution either increased 1.7 fold (nil N treatment) or was unaffected (2 N treatment) by the K application (Fig. 2). According to Equation 1 and soil solution composition, the ¹³⁷Cs AC in soil solution is expected to increase 3.8-fold (nil N treatment) and 1.3-fold (2 N treatment) following K application. ¹³⁷Cs AC's were higher in the N supplied soils than at no N application. Equation 1 predicts a 3.3 fold increase of the ¹³⁷Cs AC in solution by the N application at no K gift whereas a 2 fold increase is observed. Hence, in a quantitative way, the effects of K and NH_4^+ on the ¹³⁷Cs activity in soil solution did not fully obey Equation 1. It has however been found that the product $K_D(m_K + 5m_{NH_4})$ ls slightly increased in the micromolar range of K and NH_4^+ (Wauters, 1994).

Solute transport calculations reveal a considerable K depletion at the root surface compared to the bulk soil solution at the 0K treatments (Table 4). At the

N-K application (mmol pot ⁻¹	Shoot dry weight(g)	K shoot(mmol g^{-1} dry weight)	K soil solution (mmol L ⁻¹)	NH ⁺ soil solution(mmol L ⁻¹)
0-0	0.032(0.006)	1.05(0.14)	0.17(0.02)	0.32(0.11)
0-1	0.036(0.005)	1.04(0.13)	0.42(0.07)	0.80(0.28)
0-2	0.041(0.009)	1.13(0.19)	0.81(0.11)	1.18(0.21)
2-0	0.048(0.007)	0.97(0.07)	0.21(0.03)	1.13(0.21)
2-1	0.051(0.004)	1.00(0.02)	0.46(0.05)	1.20(0.34)
2-2	0.045(0.013)	0.90(0.19)	0.94(0.05)	1.32(0.23)
Treatment				
effect ^a				
N	***	ns	**	***
К	ns	ns	***	***
N-K	ns	***	ns	*

Table 2. Shoot dry weight, shoot K content (mmol g^{-1} dry weight) and the K and NH⁺₄ concentrations in isolated soil solution at the various K and N application rates (in mmol pot⁻¹). Averages and standard deviations(between brackets) of the 5 replicates observed at day 16. Treatment effects (including the N-K interaction) are assessed by two way ANOVA

^{a*},**,*** Significant at the 0.05, 0.01 or 0.001 level respectively; ns, not significant at the 0.05 level.

Tuble 3. Root dry weight, the root to shoot dry weight ratio and the root length of 13 and 17 days old wheat plants grown at the various K and N application rates (in mmol pot⁻¹). Averages and standard deviations (between brackets) of 2 replicates. Treatment effects (including the N-K interaction) are assessed by two way ANOVA

N-K application	Root dry weight(g)		Root:shoot dry weight ratio		Root length per plant(m)	
(mmol pot ⁻¹)	Day 13	Day 17	Day 13	Day 17	Day 13	Day 17
0-0	0.053(0.005)	0.074(0.015)	1.95(0.44)	2.04(0.62)	6.5(0.7)	13.5(4.3)
0-1	0.060(0.008)	0.066(0.011)	1.49(0.01)	1.62(0.31)	8.7(0.7)	12.0(3.7)
0-2	0.070(0.005)	0.082(0.014)	1.82(0.22)	1.91(0.14)	7.2(0.7)	13.2(1.8)
2-0	0.048(0.010)	0.098(0.023)	1.31(0.23)	1.28(0.10)	7.5(2.1)	13.4(2.8)
2-1	0.058(0.018)	0.074(0.058)	1.25(0.11)	1.04(0.41)	7.6(2.3)	11.7(8.4)
2-2	0.054(0.017)	0.065(0.013)	1.48(0.18)	1.25(0.23)	4.4(0.8)	8.4(0.8)
Treatment						
effect ^a						
N	ns	ns	*	*	ns	ns
К	ns	ns	ns	ns	ns	ns
N-K	ns	ns	ns	ns	ns	ns

^a Significant at the 0.05 level; ns not significant at the 0.05 level.

other treatments, K was not so strongly depleted from the bulk solution because a K concentration gradient $(m_{\rm K} - m_{\rm K,a})$, below 100 μM K was sufficient to match the observed K uptake rate. This concentration gradient is relatively small compared to the bulk solution concentration. The calculated ¹³⁷Cs AC's at the root surface were not lower than 50% of the bulk soil solution AC's (Table 4). A parameter sensitivity analysis showed that the K concentrations at the root surface in the nil K treatment are most sensitive to changing input parameters (details not shown). The fact that potassium is more readily depleted around roots than 137 Cs is related to the higher value of α for K than for 137 Cs. This is confirmed by the observation that in solution culture, K is more efficiently absorbed by plant roots than 137 Cs (see e.g. Smolders et al., 1996).

The solution composition at the root surface was used to compare the results of the potted soil experiments with those of a solution culture experiment (Smolders et al., 1996). In the solution culture

N-K application	Average soil solution		Soil solution at root surface		¹³⁷ Cs activity in shoot(cpm g ⁻¹ dry weight)	
(mmol pot ⁻¹)	Κ(μΜ)	¹³⁷ Cs (cpm mL ⁻¹)	Κ (μ <i>M</i>)	¹³⁷ Cs (cpm mL ⁻¹)	Predicted	Observed
0-0	170	4.7	44	2.8	4444	6191
0-1	420	5.7	354	5.2	278	2765
0-2	810	7.8	776	7.7	249	755
2-0	210	9.2	49	4.5	5665	11831
2-1	460	9.8	403	9.4	451	3219
2-2	940	9.5	918	9.5	285	1267

Table 4. Predicted and observed 137 Cs activity in wheat shoots grown at different N-K application rates. The predictions are based on the K and 137 Cs concentrations in soil solution at the root surface and 137 Cs uptake characteristics measured in solution culture. The soil solution composition at the root surface is calculated with a solute transport model



Figure 1. ¹³⁷Cs activity (kcpm g^{-1}) in the shoot of 16 day old wheat plants grown at varied K and NO₃⁻-N application rates. Data points represent averages and standard deviations (error bars) of 5 replicates. Error bars are not visible when standard deviations are smaller than the symbol size.



K rate (mmol/kg soil)

Figure 2. ¹³⁷Cs activity concentration in soil solution (cpm mL⁻¹) at the different K and NO₃⁻¹ - N application rates. Averages and standard deviations (error bars) of 5 replicates.

experiment,¹³⁷Cs Transfer Factor (TF, the ratio of the ¹³⁷Cs AC in the tissue to that in solution) was measured in wheat plants grown at K concentrations varying between 25 μ M K and 1000 μ M K. The shoot ¹³⁷Cs transfer factor was regressed to $m_{\rm K}$ (a non-linear regression of the form log TF = $(A - B m_{\rm K})/(C + m_{\rm K})$ in which A, B and C are regression parameters). This curve, combined with the K and ¹³⁷Cs concentration at the root surface, was used to predict ¹³⁷Cs AC's in shoots (Table 4).

The calculations predict a decrease in the ¹³⁷Cs AC with increasing K application and an increase in the ¹³⁷Cs AC in the plants with N supply at corresponding K application. The calculated reduction factors by the K application are however twice as large as observed. The predicted AC's in the plant are always higher than the observed AC's. The continued decrease in ¹³⁷Cs uptake with increasing K supply above 1 mmol K pot⁻¹ could hardly be predicted.

Discussion

As in most other potted soil experiments referred to in the introduction of this paper, a substantial reduction in the ¹³⁷Cs AC in the plant by K application is found. As demonstrated here, the K application reduces the ¹³⁷Cs AC in the plant in contrast to the soil solution AC's which are either enhanced or unaffected by the K supply. The reduction in ¹³⁷Cs availability is most probably related to the sharp reduction in the root ¹³⁷Cs absorption power at increasing K supply which has been found in solution culture experiments (Shaw et al.,1992; Smolders et al., 1996). Other plant parameters such as root length per plant or root area per plant influence ¹³⁷Cs uptake from soil but are not sufficiently affected to explain our results.

The reduction factors found here can be compared to those found in field trials. An application of 2 mmol K pot⁻¹ should be compared to an application rate in the field of 204 kg K ha⁻¹ if it is assumed that the K dose on 1 ha is spread into 3000 tons of soil (0-25 cm plough layer, ρ =1.2 g cm⁻³). Such a K dose does not represent an exceptionally high fertiliser rate, yet a strong effect is found on ¹³⁷Cs transfer. In field trials, reduction factors for pasture grass at 200 kg K ha⁻¹ vary between 2 and 12 (Lönsjö et al., 1990; Rosén, 1992). The K content of the unfertilised soils used by Lönsjö et al.(1990) varied between 1.1 and 4.0 μ mol g⁻¹ soil (peat, sand and loamy sand soils) a range which covers the K content of the soil used here (Table 1).

The increase of ¹³⁷Cs uptake with N application was rather unexpected as all N was applied as NO_3^- . Reports on the effect of nitrate application in potted soil or fields have not been found. NH_4^+ or urea application is known to increase ¹³⁷Cs transfer (e.g. Belli et al., 1995; Jackson et al.,1965). The soil solution data of our experiment however showed that NH_4^+ concentrations increased with nitrate application (Table 2). Since NH_4^+ has been shown to have much less competitive effect on ¹³⁷Cs uptake than does K (Shaw et al.,1992), the increased ¹³⁷Cs uptake by N supply can be explained by the mobilisation of ¹³⁷Cs in soil by NH_4^+ . This hypothesis is supported by the fact that the ratio of ¹³⁷Cs AC's in plant to that in soil solution is almost equal at equal K supply in both N treatments.

The calculated soil solution composition at the root surface is most probably a better indicator to compare uptake of elements from soil with uptake from nutrient solution (Nye et al., 1975). Our predictions of ¹³⁷Cs uptake made in this way are biased to an overestimation of the reduction factors by K application and a general underestimation of the ¹³⁷Cs transfer (Table 4). A number of assumptions in this calculation can obviously be questioned. The calculated rhizosphere solution composition is only based on soil solution data at the day of plant harvest. The composition could have been different during earlier plant growth stages. The total uptake of K and ¹³⁷Cs from the pot by the biomass is however very small (respectively maximum 12% of the ion exchangeable K and below 0.1% of the total amount of ¹³⁷Cs in soil, calculation details not shown). Therefore, uptake of elements by the plant may not have caused a varying composition during plant growth. The ¹³⁷Cs

AC could however have been reduced during plant growth because of a slow fixation in the soil of the freshly added ¹³⁷Cs. The solute transport calculation (steady-state approximation) can also be questioned. A numerical non-steady state calculation was made (details can be obtained from senior author) and results were compared with those obtained with the approximative method. Between 2 and 5 days after the start of the simulated uptake, concentrations at the root surface reached almost stable values which showed good correspondence with the values given in Table 4. Refining the numerical calculation by including the effect of K concentration on the 137 Cs K_d and including a kinetic term for ¹³⁷Cs desorption would both yield an even lower ¹³⁷Cs uptake from soil, yet current predictions underestimate the observations. Rhizosphere effects and possible effects of mycorhizza are here not accounted for. Because of a lack of consistent views on to the influence of these on 137 Cs availability these are not further discussed here. It may be critical in our calculations to assume that uptake of ¹³⁷Cs and K is occurring over the whole root system. Confining the uptake of both cations to a fraction of the root length because of root morphology or physiology, root age or partial soil solution/root contact would yield lower calculated element concentration at these active roots. For reasons given above, it is thought that the ¹³⁷Cs activity depletion is generally lower than the K concentration depletion. As a result, predictions of ¹³⁷Cs AC's would be higher and the effect of K fertilisation would persist between 1 and 2 mmol K, in line with the observations.

Our current knowledge on interionic effects on ¹³⁷Cs uptake and behaviour in soil can be used to understand K fertiliser effects on ¹³⁷Cs availability in field and pot trials. It is generally found that K fertilisation reduces ¹³⁷Cs availability at low K concentration in soil whereas at high K availability, K fertilisation has little or even a stimulation effect on ¹³⁷Cs uptake. At low K availability, the ¹³⁷Cs absorption capacity of the roots can be substantially reduced by increased K supply, even though ion exchange reactions increase ¹³⁷Cs mobility in soil. At high K availability, increasing K concentrations (above 1 mM) have been found to have no effect on ¹³⁷Cs uptake characteristics of the root (Cline and Hungate, 1960) whereas ¹³⁷Cs mobility can be increased, resulting in an increased ¹³⁷Cs AC in the crop.

Based on the observations in solution culture (Shaw et al., 1992; Smolders et al., 1996), we think that K fertilisation will reduce 137 Cs availability if micromo-

lar K concentrations occur at the root surface. Such concentrations do not only depend on the K intensity in soil solution but also on the K buffering capacity of the soil. Also, the K depletion around root depends on K demand rate of the crop which is correlated with growth rate.

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References

- Baldwin J P, Nye P H and Tinker P B 1973 Uptake of solutes by multiple root systems from soil. III. A model for calculating the solute uptake by a randomly dispersed root system developing in a finite volume of soil. Plant and Soil 38, 621–635.
- Barraclough P B and Tinker 1981 The determination of ionic diffusion coefficients in soils. I. Diffusion coefficients in sieved soil in relation to water content and bulk density. J. Soil Sci. 32, 225-236.
- Belli M, Sansone U, Ardiani R, Feoli E, Scimone M, Menegon S and Parente G 1995 The effect of fertiliser application on ¹³⁷Cs uptake by different plant species and vegetation types. J. Environ. Radioactivity 27, 75–89.
- Bhat K K S, Brereton A J and Nye P H 1979a The possibility of predicting solute uptake and plant growth response from independently measured soil and plant characteristics. VI. The growth and uptake of rape in solutions of constant nitrate concentrations. Plant and Soil 53,137–167.
- Bhat K K S, Brereton A J and Nye P H 1979b The possibility of predicting solute uptake and plant growth response from independently measured soil and plant characteristics. VII. The growth and nitrate uptake of rape in soil at two nitrate concentrations and a comparison of the results with model predictions. Plant and Soil 53,169–191.
- Cline J F and Hungate F P 1960 Accumulation of potassium, caesium-137 and rubidium-86 in bean plants grown in nutrient solutions. Plant Physiol. 35, 826–829.
- De Preter P 1990 Radiocesium retention in aquatic, terrestrial and urban environments: a quantitative and unifying analysis. Dissertationes de Agricultura, no. 190, Katholieke Universiteit Leuven, Belgium.
- FAO 1990 FAO-Unesco Soil Map of the World. Revised Legend. Soils Bulletin 60. FAO, Rome, Italy. 119 p.
- Fawaris B H and Johanson K J 1994 Uptake of ¹³⁷Cs from coniferous soils by sheep fescue in pot experiments. In Soil-Plant Relationship. Proceedings of the XXIVth annual ESNA/IUR meeting. Ed. M H Gerzabek. pp 197–203. Forschungszentrum Seibersdorf, Germany.

- Fowler E B and Christenson C W 1959 Factors affecting uptake of radioactive cesium by lettuce, grass and alfalfa. J. Agric. Food Chem. 7,847–849.
- Grauby A, Jouve A and Legrand B 1990 Study of the possibility of attenuating soil-plant transfer after an accident by application of manure to the soil and by foliar spraying. *In* Transfer of Radionuclides in natural and semi-natural Environments. Eds.
 G Desmet, P Nassimbeni and M Belli. pp 403-410. Elsevier Applied Science, London, UK.
- Jackson D and Nisbet A F 1990 The effect of fertiliser treatment, soil pH and grazing on the transfer of radiocaesium to upland fell vegetation. *In* Transfer of Radionuclides in natural and seminatural Environments. Eds. G Desmet, P Nassimbeni and M Belli. pp 395-402. Elsevier Applied Science, London, UK.
- Jackson W A, Craig D and Lugo H M 1965 Effects of various cations on cesium uptake from soils and clay suspensions. Soil Sci. 99, 345–353.
- Jones D R, Paul L and Mitchell N G 1994 Field investigation of the effect of ameliorative treatments applied to unimproved upland vegetation in the UK contaminated by radiocaesium. *In* Soil-Plant Relationship. Proceedings of the XXIVth annual ESNA/IUR meeting. Ed. M H Gerzabek. pp 197–203. Forschungszentrum Seibersdorf, Germany.
- Lönsjö H, Haak E and Rosén K 1990 Effects of remedial measures on long term transfer of radiocaesium from soil to agricultural products as calculated from Swedish field experimental data. *In* Proceedings of a Symposium on Environmental Contamination following a Major Nuclear Accident. pp 151–162. STI/PUB/825, IAEA, Vienna, Austria.
- Mitchell N G, Coughtrey P J and Kirkton J A 1990 Effects of calcium, potassium, rubidium and various fertilisers on radiocaesium transfer in field and experimental conditions. *In* Transfer of Radionuclides in natural and semi-natural Environments. Eds. G Desmet, P Nassimbeni and M Belli. pp 387-394. Elsevier Applied Science, London, UK.
- Nisbet A F, Mocanu N and Shaw S 1994 Laboratory investigation into the potential effectiveness of soil-based countermeasures for soils contaminated with radiocaesium and radiostrontium. Sci. Tot. Environ.149,145–154.
- Nishita H, Romney E M, Alexander G V and Larson K H 1960 Influence of K and Cs on release of Cs-137 from three soils. Soil Sci. 84,167–176.
- Nishita H, Dixon D and Larson K H 1962 Accumulation of Cs and K and growth of bean plants in nutrient solution and soils. Plant and Soil 17, 221–242.
- Nye P H, Brewster J L and Bhat K K S 1975 The possibility of predicting solute uptake and plant growth response from independently measured soil and plant characteristics. I. The theoretical basis of the experiments. Plant and Soil 42,161–170.
- Prister B S, Perepelyatnikov G P and Perepelyatnikova L V 1993 Countermeasures used in Ukraine to produce forage and animal food products with radionuclide levels below intervention limits after the Chernobyl accident. Sci. Tot. Environ. 137, 183–198.
- Rengel Z and Robinson D L 1990 Modeling magnesium uptake from an acid soil. I. Nutrient relationships at the soil-root interface. Soil Sci. Soc. Am. J. 54,785–791.
- Robinson R A and Stokes R H 1965 Electrolyte Solutions, 2nd edn., revised. Butterworths, London, UK.
- Rosén K 1992 Effects of potassium fertilisation on caesium transfer to grass barley and vegetables after Chernobyl. In The Chernobyl fallout in Sweden. Research from a research program on environmental radiology. Ed. L Möberg. Swedish Radiation Protection Institute, Stockholm, Sweden.

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- Shaw G, Hewamanna R, Lillywhite J and Bell J N B 1992 Radiocaesium uptake and translocation in wheat with reference to the transfer factor concept and ion competition effects. J. Environ. Radioactivity 16,167–180.
- Smolders E and Shaw G 1995 Changes in radiocaesium uptake and distribution in wheat during plant development: a solution culture study. Plant and Soil 176,1-6.
- Smolders E, Kiebooms L, Buysse J and Merckx R 1996¹³⁷ Cs uptake in spring wheat (*Triticum aestivum* L. cv. Tonic) at varying K supply. 1. The effect in solution culture. Plant and Soil 181.
- Sweeck L, Wauters J, Valcke E and Cremers A 1994 The specific interception potential for radiocaesium. *In* Transfer of Radionuclides in natural and semi-natural Environments. Eds. G Desmet, P Nassimbeni and M Belli. pp 249–258. Elsevier Applied Science, London, UK.
- Tennant D 1975 A test of a modified line intersection method of estimating root length. J. Ecol. 63, 995–1001.
- Wauters J 1994 Radiocesium in aquatic sediments: sorption, remobilization and fixation. Ph.D. Thesis, Dissertationes de Agricultura, no. 246, Katholieke Universiteit Leuven, Belgium.

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