



How Climate Shapes the Functioning of Tropical Montane Cloud Forests

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Abstract

Purpose of Review Tropical Montane Cloud Forest (TMCF) is a highly vulnerable ecosystem, which occurs at higher elevations in tropical mountains. Many aspects of TMCF vegetation functioning are poorly understood, making it difficult to quantify and project TMCF vulnerability to global change. We compile functional traits data to provide an overview of TMCF functional ecology. We use numerical models to understand the consequences of TMCF functional composition with respect to its responses to climate and link the traits of TMCF to its environmental conditions.

Recent Findings TMCF leaves are small and have low SLA but high Rubisco content per leaf area. This implies that TMCF maximum net leaf carbon assimilation (A_n) is high but often limited by low temperature and leaf wetting. Cloud immersion provides important water and potentially nutrient inputs to TMCF plants. TMCF species possess low sapwood specific conductivity, which is compensated with a lower tree height and higher sapwood to leaf area ratio. These traits associated with a more conservative stomatal regulation results in a higher hydraulic safety margin than nearby forests not affected by clouds. The architecture of TMCF trees including its proportionally thicker trunks and large root systems increases tree mechanical stability.

Summary The TMCF functional traits can be conceptually linked to its colder and cloudy environment limiting A_n , growth, water transport and nutrient availability. A hotter climate would drastically affect the abiotic filters shaping TMCF communities and potentially facilitate the invasion of TMCF by more productive lowland species.

Keywords Climate change · Cloud forests · Functional traits · Plant hydraulics · Photosynthesis

Introduction

Tropical Montane Cloud Forest (TMCF) is a rare ecosystem type, which covers only 0.26% of Earth's land surface [1].

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Despite its restricted distribution, TMCF hosts a large biodiversity [1, 2, 3•] and provides important ecosystem services in mountainous regions [1, 3•, 4, 5]. Cloud immersion events are the main climatic attribute defining TMCF [6, 7]. The increase in the cloud immersion frequency at higher altitudes produces a progressively shorter vegetation with smaller leaves, and trunks and branches covered by epiphytes [4, 7, 8]. This shift in vegetation structure is accompanied by changes in the floristic community composition [9•, 10, 11•]. There is a marked drop in the abundance of Fabaceae [10, 11•], while Myrtaceae and several Magnoliid families become more abundant [8, 12–14].

TMCF exhibits an extremely high plant species richness per unit area [3•]. For example, Mexican TMCF cover less than 1% of its territory but contain 650 genera of vascular plants with at least one species endemic or preferentially associated to TMCF [15]. While overall tree diversity tends to decline with altitude [2], pteridophytes and epiphytic bryophytes become highly abundant and diverse at higher-altitude TMCF [3•]. The peculiar hydroclimatic environment in TMCF associated with its fragmented nature also favours

high levels of endemism. Gentry [2] estimates that 10–24% of the plant species in South American TMCF are endemic to this ecosystem. The species diversity and endemism in TMCF makes these ecosystems valuable gene pools for the improvement of commercial crop species [1]; in addition, a high diversity of mammal, amphibian and bird species is found primarily in TMCF [3•, 16, 17].

The low transpiration rates and high cloud water input in TMCF contribute to the maintenance of the streamflow in mountainous regions during the dry season [3•, 4]. The cloud water input in TMCF usually ranges from 15 to 20% of the rainfall, but in some sites, it can contribute as much as 50–60% [1]. The hydrological function of TMCF is important for the water supply of major cities in mountainous regions [1]. Besides its direct influence on streamflow, TMCF also acts to naturally filter water, which contributes to a higher water quality in the streamflow [5].

Changes in climate and land-use are major threats to TMCF [16, 18–20]. Model simulations predict that increases in land surface temperature could increase the height of cloud formation in tropical mountains [16, 19, 20]. Changes to TMCF cloud immersion regime holds major implications for its vegetation physiology and ecosystem processes [21, 22•, 23–26], and its decline threatens the integrity of these ecosystems. The fauna of TMCF are also highly vulnerable to changes in the TMCF cloud regime [18]. Pounds et al. [18] attributes the loss of 40% of the frog species in a Costa Rican TMCF to the increase in the number of days without rainfall or fog. Mountain environments are also subject to increased rates of climate warming [27]. Increases in temperature of up to 4 °C are predicted to occur in TMCF [22•], which should aggravate the water deficit associated with the cloud uplift.

Predicting TMCF responses to climate change requires a mechanistic understanding of how TMCF hydroclimatic conditions determine its community composition and functioning. Functional traits provide a theoretical bridge to link plant physiological responses to environmental gradients/conditions and community assembly [28]. Information on plant functional traits can be incorporated into process-based Dynamic Global Vegetation Models (DGVM) to predict large scale vegetation shifts in response to climate change [29, 30]. Whereas certain vegetation traits are widely associated with TMCF such as small and thick leaves that form canopies with low Leaf Area Index (LAI) and low stature trees [6, 8, 16, 31], little is known about more detailed aspects of TMCF photosynthetic and hydraulic functioning. This information is essential to predict plant responses to climate [32, 33]. In this review, we address this important knowledge gap by compiling functional trait data from TMCF communities. We use this functional trait information to characterize TMCF and understand what makes these communities functionally different from nearby forests not affected by clouds. In addition, we use the TMCF functional trait information to parametrize

process-based models that are used to understand how the climate drives water transport, stomatal regulation and photosynthesis in the TMCF vegetation. The main questions we intend to address in this review are: (i) What are the functional traits of the species that dominate TMCF and how do they differ from humid tropical forests not affected by clouds? (ii) How do these functional traits modulate TMCF responses to climate? and (iii) What mechanisms can explain the predominance of certain functional traits in TMCF? We also aim to identify critical knowledge gaps about TMCF, which currently limit our capacity to respond these questions.

A Case Study of South/Southeast Brazilian Cloud Forest Functional Composition

We start this review with a case study focused on TMCF forest from South/Southeast Brazil (SSBCF) (Table 1; Fig. S1). We use floristic data from 10 TMCF sites and 8 non-cloud affected Atlantic forests sites in South/Southeast Brazil to illustrate the floristic and functional differences between TMCF and nearby non-cloud affected tropical forests. For this compilation, we chose studies that fulfilled the following conditions: (1) the study was conducted within the area of interest (South/Southeast Brazil), (2) the study provided information on the species relative abundance and (3) the sites were classified

Table 1 Cloud forest and non-cloud affected Atlantic forest sites in South/Southeast Brazil. Sites are classified as Cloud Forests (CF) or Atlantic forests not affected by clouds (nCF) based on its source study

Code	Coordinates	Altitude (m)	Source
CF1	28° 08' S 49° 28' W	1590	Suhs et al. [34]
CF2	25° 54' S 48° 56' W	1610	Koehler et al. [35]
CF3	25° 41' S 49° 02' W	1390	Koehler et al. [35]
CF4	25°32' S 48° 56' W	1545	Koehler et al. [35]
CF5	25°32' S 48° 56' W	1460	Koehler et al. [35]
CF6	25°21' S 48° 54' W	1590	Koehler et al. [35]
CF7	22°41' S 45° 25' W	2000	Oliveira et al. <i>unpublished</i>
CF8	22°26' S 44° 51' W	2250	Meireles & Shepherd [13]
CF9	21°58' S 43° 52' W	1300	Valente et al. [36]
CF10	21°46' S 46° 24' W	1387	Costa et al. [37]
nCF1	21°59' S 43° 53' W	1000	Valente et al. [36]
nCF2	28°36' S 49° 33' W	178	Colonetti et al. [38]
nCF3	25°30' S 48° 38' W	485	Silva [39]
nCF4	24°14' S 48° 04' W	108	Guilherme et al. [40]
nCF5	24°00' S 47° 55' W	650	Dias & Couto [41]
nCF6	23°21' S 45° 05' W	371	Rochelle et al. [42]
nCF7	22°40' S 42° 30' W	150	Carvalho et al. [43]
nCF8	23°20' S 44° 50' W	55	Prata et al. [44]

either as Cloud Forests or Atlantic forests not affected by clouds. We use the term TMCF throughout the text to refer to tropical and subtropical montane forests exposed regularly to clouds, including Lower Montane Cloud Forests (LMCF), Upper Montane Cloud Forests (UMCF) and Elfin/Dwarf Cloud Forests [16]. All TMCF sites used in this section were located at altitudes higher than 1000 m and in locations frequently exposed to clouds (Fig. S1). The TMCF study sites have a mean annual temperature (MAT) on average 5 °C lower than non-TMCF sites, with a MAT lapse rate of 0.4 °C per 100 m (Fig. S2). The SSBCF sites were dominated by characteristic TMCF genera (Fig. S2), such as *Drymis*, *Ilex*, *Weinmannia* and several Myrtaceae genera [9•, 12, 13]. The exceptions were the more northern sites, CF9 and CF10, which were dominated by Euphorbiaceae and Solanaceae (Fig. S3). We used genus (or family) level means of plant functional traits compiled from the Choat et al. [45] and Kattge et al. [46] datasets together with the species abundance at each site (Table 1, Fig. S3) to compute community weighted average (CWA) trait values for each studied site. See Appendix S1 for details on our methodology. This approach assumes the existence of a strong phylogenetic signal [47], which was found in most of the studied functional traits (Table 2). We adopted this indirect approach to circumvent the lack of functional trait data for Brazilian TMCF, which highlights the urgent need for more trait surveys in these forests. While these indirect CWA estimates must be interpreted carefully, we show in the next section they largely agree with published values collected in situ in TMCF worldwide.

We conducted a cluster analysis on the CWA traits using the first two principal axes from a Principal Component Analysis (PCA) to identify a possible functional convergence among these TMCF sites. The sites can be grouped into two clusters, which maximize the data average silhouette width, that is, minimize the dissimilarity between points within a cluster [47] (Fig. 1). The blue cluster contains 8 out of the 10 TMCF sites used in this analysis (Table 1). The main

CWA traits that define the blue cluster sites are low Specific Leaf Area (SLA), low leaf nitrogen content on a mass basis (N_m), low sapwood specific hydraulic conductivity (K_s) and high Huber Value (HV). The sapwood density (ρ) is also lower in the blue cluster, but it is mostly associated with the within cluster variability along the second PCA axis. The red cluster contains all the Atlantic forest sites not affected by clouds in addition to the two most northern TMCF sites (CF9 and CF10). The different CWA in the northern TMCF sites reflects its floristic composition distinct from the other TMCF sites (Fig. S3).

The values of the functional traits predominant in SSBCF are associated with more conservative ecological strategies, that is, plants with slower rates of resource use and acquisition [48]. In the next section, we assess the generality of this finding by contrasting the results from our indirect phylogenetic approach with data collected in situ from TMCF around the globe. We review sequentially the functional traits of TMCF leaves, wood and roots. Whereas, most of the discussion in the next sections is focused on the traits present in Fig. 1, other traits relevant to understanding TMCF functioning are also discussed.

Cloud Forest Leaves and Canopy

Leaf Structure, Stoichiometry and Photosynthesis

The apparent xeromorphism of TMCF leaves has intrigued plant ecologists for several decades [6–8, 49], given the humid TMCF environment, albeit the generality of this assumption is questionable as TMCF can occur across a wide range of rainfall regimes [49, 50•] and high atmospheric aridity [51–53]. As expected, the dominant genera in SSBCF communities had leaves with an SLA 1.86 (CI95%: 0.01 to 3.72) $m^2 kg^{-1}$ lower than non-TMCF communities (Fig. 1a). These findings are corroborated by numerous studies reporting a decline in

Table 2 Description of traits evaluated at the cloud forest sites in South/Southeast Brazil and measurements of the trait phylogenetic signal with the Pagel’s λ . When Pagel’s $\lambda = 0$, there is no phylogenetic signal, that is, the trait evolved independently of phylogeny; when $\lambda = 1$, the trait evolution followed a pure Brownian model of evolution

Trait	Units	Description	λ	p
SLA	$m^2 kg^{-1}$	Specific leaf area	0.33	< 0.01
N_m	$mg g^{-1}$	Nitrogen content on a leaf mass basis	0.71	< 0.01
A_n	$\mu mol m^{-2} s^{-1}$	Leaf net carbon assimilation rate	0.12	0.03
Ψ_{50}	MPa	Xylem/Tracheid water potential when plant loses 50% of its maximum hydraulic conductivity	0.51	< 0.01
Ψ_{min}	MPa	Minimum leaf water potential at the field	0.21	0.07
π_{tip}	MPa	Leaf turgor loss point	< 0.01	1
K_s	$kg m^{-1} s^{-1} MPa^{-1}$	Xylem/Tracheid specific conductivity	0.1	0.06
HV	$cm^2 m^{-2}$	Huber Value, the ratio between sapwood area and leaf area	< 0.01	1
H_{max}	m	Maximum canopy height	0.75	< 0.01
ρ	$g cm^{-3}$	Basic wood density	0.42	< 0.01

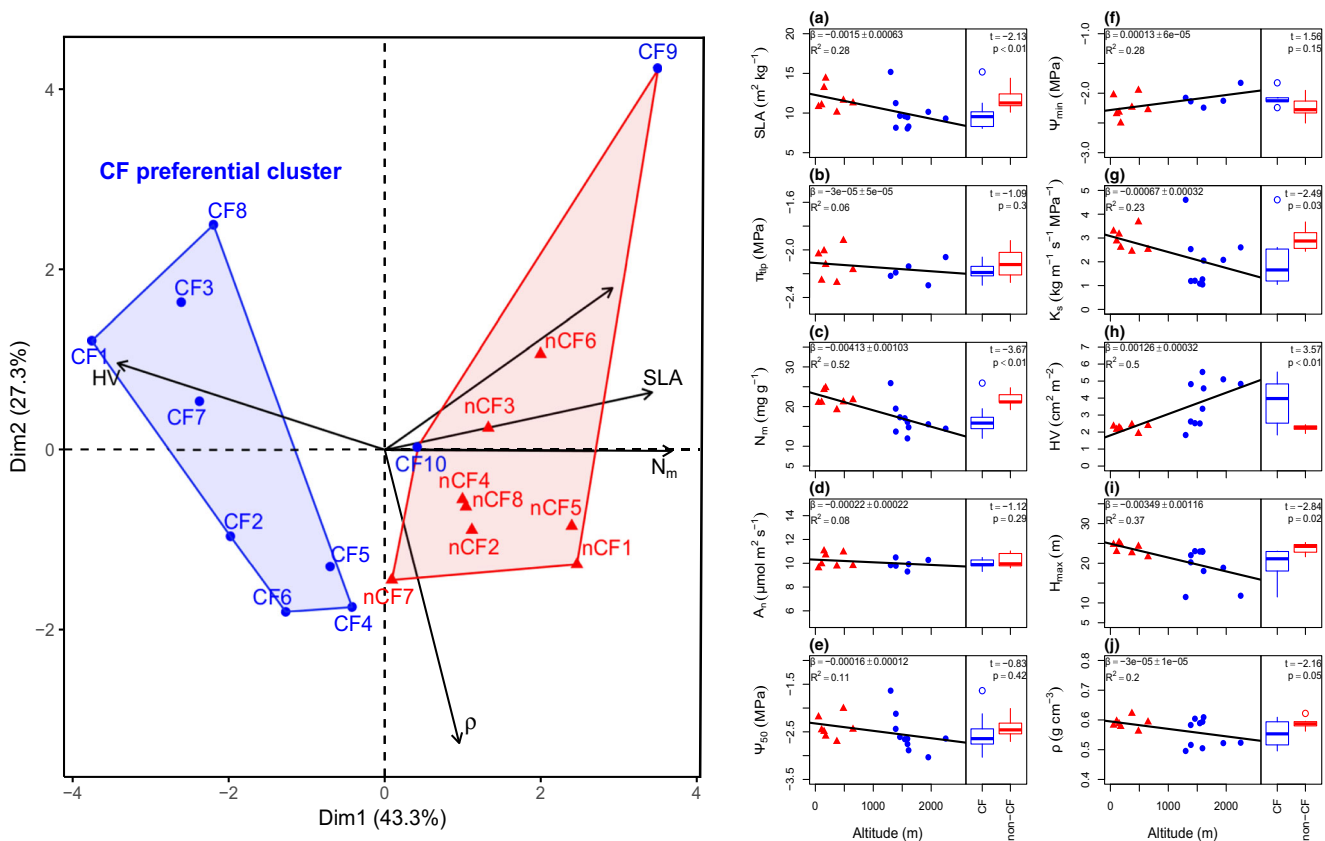


Fig. 1 Biplot of a Principal Component Analysis (PCA) and altitudinal trends of the community averaged functional traits from the South/Southeast Brazil sites. In the PCA biplot on the left, the arrows show the five functional traits most strongly associated with the first two PCA components (HV Huber value, K_s sapwood specific conductivity, SLA specific leaf area, N_m leaf nitrogen on a mass basis, ρ sapwood density). The red and blue clusters were defined using a k -means clustering

algorithm based on the data silhouette width. In the panels on the right, the sites in blue are Tropical Montane Cloud Forests and the sites in red are other types of Atlantic forests (see Table 1). The meaning of the functional traits' acronyms is given in Table 2. We only used the sites from Table 1 where we could find genus-level trait data enough to cover at least 50% of the community total stem basal area

SLA with increasing altitude in tropical mountains [10, 54–56]. Kitayama & Aiba [55] have found a mean SLA of 4.95 and 3.98 $\text{m}^2 \text{kg}^{-1}$ in two Bornean Upper Montane Forest sites, which were on average 2.61 $\text{m}^2 \text{kg}^{-1}$ lower than nearby Lowland Rainforests (LRF). Van de Weg et al. [10] reports a SLA of 7.47 ± 1.1 (mean \pm SE) $\text{m}^2 \text{kg}^{-1}$ in four Peruvian TMCF sites, on average 4 $\text{m}^2 \text{kg}^{-1}$ lower than a nearby LRF. According to Grubb [8] UMCF have SLA values ranging from 4.5 to 7 $\text{m}^2 \text{kg}^{-1}$ and LMCF can reach 8 $\text{m}^2 \text{kg}^{-1}$, while LRF ranges from 9 to 13 $\text{m}^2 \text{kg}^{-1}$. The TMCF CWA SLA of 9.9 (CI 95%: 8.6 to 11.2 $\text{m}^2 \text{kg}^{-1}$; Fig. 1a) for SSBCF are on the higher end of Grubb [8] and Van de Weg et al. [10] observations. As noted by Bruijnzeel & Veneklaas [31], the lower SLA in TMCF makes its total leaf biomass closer to LRF, despite the large LAI difference between communities. The TMCF LAI can be as low as 2 $\text{m}^2 \text{m}^{-2}$ in dwarf TMCF [57] but typically ranges from 5 to 6 $\text{m}^2 \text{m}^{-2}$ in UMCF [8, 31], while LRF LAI can reach 9 $\text{m}^2 \text{m}^{-2}$ [8]. The leaf biomass (kg) per m^2 of soil in TMCF (computed as $1/\text{SLA} \times \text{LAI}$) ranges from 0.71 to 1.25 $\text{kg} \text{m}^{-2}$ assuming an SLA between 4 and

7 $\text{m}^2 \text{kg}^{-1}$ and a LAI of 5 $\text{m}^2 \text{m}^{-2}$. This is potentially higher than LRF leaf biomass, which should range from 0.69 to 1 $\text{kg} \text{m}^{-2}$ assuming its SLA ranges from 9 to 13 $\text{m}^2 \text{kg}^{-1}$ [8] and the LAI is 9 $\text{m}^2 \text{m}^{-2}$.

SLA is the product of leaf thickness and density and is one of the traits at the core of the fast-slow continuum in plant ecological strategies [48, 58]. A lower SLA implies a higher investment in structural and defense tissues, which increases the leaf resistance to herbivory and disturbances, resulting in a longer lifespan [58]. The low SLA in TMCF species, as well as many other TMCF traits are often attributed to low nutrient availability [11•, 50•]. However, the effect of nutrient availability on SLA is relatively small if compared with the effect of irradiance [59]. High irradiance and atmospheric aridity can also explain TMCF leaf structure [51–53]. Smith & Geller [52] model simulations shows that bigger leaves, which usually have higher SLA, would quickly overheat at higher altitude because of the higher radiation loads. The thick cell walls, radial sclereids and fibres of low SLA plants allow the maintenance of cell turgor during dehydration [60], which

results in a more negative π_{tip} [61]. However, the leaves from the dominant genera in SSBCF lose turgor at similar Ψ than non-TMCF, around -2.15 MPa (Fig. 1b). Despite the weak phylogenetic signal found in π_{tip} (Table 2), our CWA π_{tip} are within the range reported for TMCF and LRF. The value we estimated is in the range of observations from TMCF in New Zealand, Colombia and Hawaii, which report a π_{tip} ranging from -1.34 to -2.6 MPa [62–64]. Marechaux et al. [62] reports a wider interval for 71 LRF species, with the π_{tip} ranging from -1.4 to -3.2 MPa.

SLA is strongly correlated with leaf N_m and longevity, forming the classic leaf economic spectrum of Wright et al. [58]. The dominant genera in our TMCF sites follow this classic trade-off possessing on average 5.25 (CI95%: 2.19 to 8.33) mg g^{-1} less N_m than non-TMCF communities (Fig. 1c). Tanner et al. [54] shows that leaf N_m declines at a rate of 0.12 mg g^{-1} per 100 m increase in altitude across several tropical forest sites ranging from 0 to 3700 m of altitude. Van de Weg et al. [10] study over a 220 – 3360 m altitudinal gradient in Peru a shows a decline in leaf N_m of 0.26 mg g^{-1} per 100 m of altitude. Our observed leaf N_m decline rate of 0.4 mg g^{-1} per 100 m (Fig. 1b) is considerably higher than both studies. Grubb [8] shows the N_m across several UMCF sites can range from 8.1 to 16.1 mg g^{-1} , while LMCF can reach 17.6 mg g^{-1} . Soethe et al. [56] found a leaf N_m of 15.5 ± 3.16 (mean \pm SE) mg g^{-1} for three Ecuadorian TMCF sites. These values are close to our TMCF CWA N_m of 16.6 (CI95%: 14.2 to 19) mg g^{-1} (Fig. 1c).

Plants grown in lower temperatures typically have higher amounts of nitrogen-rich photosynthetic enzymes to compensate for the lower activity of the enzymes at low temperatures [63]. Several studies report high leaf-level A_n in TMCF species, which are close to non-pioneer LRF species A_n [31]. Letts & Mulligan [64] measured seven pairs of congeneric species in a Colombian LMCF and UMCF. They have found light saturated maximum A_n of 10.6 and 10.2 $\mu\text{mol m}^{-2} \text{s}^{-1}$ for the LMCF and UMCF, respectively. van de Weg et al. [65] reports light saturated mean A_n rates of 7.04 ± 0.33 (mean \pm SE) $\mu\text{mol m}^{-2} \text{s}^{-1}$ for five TMCF species in Peru. Our findings for SSBCF are in agreement with these studies as the dominant genera in our TMCF sites have similar A_n to non-TMCF communities (Fig. 1d). The TMCF CWA A_n was 9.93 (CI95%: 9.61 to 10.26) $\mu\text{mol m}^{-2} \text{s}^{-1}$, while non-TMCF communities CWA A_n was 10.26 (CI95%: 9.91 to 10.71) $\mu\text{mol m}^{-2} \text{s}^{-1}$. However, Wittich et al. [66] reports the light saturated A_n of 170 species from 18 sites in altitudes ranging from sea level to 4000 m decreases by 0.13 $\mu\text{mol m}^{-2} \text{s}^{-1}$ per 100 m increase in altitude.

We can use the ratio between N_m and leaf phosphorus content (N:P) to assess nutrient limitations to plant growth [67, 68]. Aerts & Chapin [68] classifies a N:P ratio lower than 14 as indicative of N limitation, whereas N:P higher than 16 indicates P-limitation. Gusewell [67] defines that leaf N:P

lower than 10 or higher than 20 are indicative of N and P-limitation, respectively. We did not have leaf P in the dataset used for the analysis of the SSBCF sites. However, we compiled published data of leaf N:P from 31 LMCF and UMCF, six LRF and two subalpine sites, to evaluate the evidence supporting the hypothesis of nutrient limitation in TMCF. We found no significant differences in the leaf N:P among LMCF, UMCF and LRF (Fig. S4). There was also no relationship between leaf N:P and altitude for the 19 sites where the altitude data was available (Fig. S4; $R^2 = 0.07$, $p = 0.23$). Only two out of 15 UMCF sites showed sign of N limitation (i.e. $\text{N:P} < 14$), whereas 44% of the LMCF showed sign of N limitation. Only three LMCF sites had a N:P lower than 10 , and they were all from Hawaii [69, 70]. Despite our findings, some studies have experimentally demonstrated that TMCF productivity is limited by nutrients [11•, 54, 71•]. As noted by Gusewell [72], as high altitude plants tend to have higher leaf N than low elevation plants, they might reach higher N:P ratios even in N-limited environments.

The Role of Leaves in Water Acquisition

An important characteristic of TMCF leaves that received considerable recent attention is the capacity of TMCF leaves to acquire directly the water condensed on its surface through foliar water uptake (FWU) [21, 24, 25, 73]. This process is driven by a water potential (Ψ) gradient between the water outside leaves and the water inside, with the water flowing through the stomata [74, 75], cuticle [24, 76] and/or specialized structures [24, 77]. Eller et al. [24, 25] showed through greenhouse experiments that FWU allows saplings of three Brazilian TMCF species to sustain gas exchange, leaf turgor and growth during soil drought. The total amount of water absorbed by FWU is small but not insignificant, ranging from 5 to 26% of maximum transpiration fluxes [78]. Importantly, Goldsmith et al. [21] reported a Ψ increment of 0.67 ± 0.02 (mean \pm SE) MPa in the leaves of 12 TMCF species after 1 h of experimental leaf wetting, which was higher than the Ψ increment in a nearby submontane forest of 0.55 ± 0.12 (mean \pm SE) MPa. As noted by Oliveira et al. [22•], this magnitude of water input and Ψ increment can be very important to maintain the hydraulic integrity and survival of plants in certain TMCF during seasonal and interannual droughts. However, Berry et al. [78] show that FWU is a ubiquitous process found in plants worldwide, including LRF [79]. Binks et al. [79] observed a mean leaf Ψ increment in Amazon tree species of 0.63 MPa after 1 h of artificial wetting, which is close to observations of Goldsmith et al. [21] for TMCF. More studies are necessary to understand the differences between the occurrence and significance of FWU for LRF and TMCF. Whereas TMCF can likely benefit from high FWU rates due to the persistence of cloud immersion events wetting its canopy, Dawson & Goldsmith [80] show that plants in most

biomes in the world are also exposed to long periods with wet leaves, including LRF.

Oliveira et al. [22•] note that the ecophysiological importance of FWU to TMCF plants varies from site to site and seasonally. Whereas most TMCF occurs in sites with high rainfall (2000 to 2600 mm from Jarvis & Mulligan [50•]), a significant number of TMCF sites might occur in lower rainfall locations. Jarvis & Mulligan [50•] show that 6% out of 477 TMCF sites from the UNEP-WCMC global database (UNEP-WCMC, 2004) occur in locations that receive less than 1000 mm of rainfall annually. Some sites receive as little as 405 mm annually and can be exposed to significant rainfall seasonality. Additionally, the shallow soils usually found in TMCF [81, 82] coupled with a potentially high atmospheric aridity due to higher incident shortwave irradiance and diffusivity of water vapour and heat in air [51–53] can cause water shortage for TMCF plants during seasonal or interannual periods with reduced rainfall. Mortality events in TMCF vegetation have been reported following severe droughts [83, 84]. These conditions make the vegetation in drier TMCF sites highly dependent on cloud water input, both through direct cloud interception and FWU, to sustain its physiological activity during the dry season [26, 85–87]. Acquiring the water condensed on the plant canopy through FWU before it drips to the forest floor is a method to bypass the belowground competition for water with other plants and the possibility of interception by epiphytes and understory vegetation.

The TMCF arboreal component should only experience water stress regularly in more arid TMCF sites. However, the epiphytic community of every TMCF regularly experiences water stress due to the limited soil volume available and high radiation in the canopy environment [88]. Gostch et al. [23] shows that FWU provides large amounts of water to some epiphytes in a Costa Rican TMCF. During 1 month of observations, the seven epiphyte species studied by Gostch et al. [23] absorbed through FWU on average 70% of its transpired water, with some species absorbing up to 96% of its transpired water. Not all epiphytes are capable of FWU [77]; therefore, more studies are necessary to assess the FWU capabilities of different TMCF epiphytes. The epiphytic community is an essential component of TMCF water and nutrient cycles [16, 88], and its reliance on FWU makes this process extremely important for TMCF functioning.

Besides acquiring water through FWU, TMCF leaves might also contribute to the plant nutrient acquisition through N foliar uptake [89]. Cloud water from forests in Southern Chile can contain significant amount of organic and inorganic N [90]. Additionally, the canopy of TMCF hosts microbes and epiphytes capable of fixing atmospheric N₂ [91, 92]. We postulate that plants capable of accessing these resources through direct FWU would have an important competitive advantage in N-limited TMCF. We could not find any studies on the relevance

of FWU for nutrient uptake in TMCF; therefore, we consider this topic a priority for future TMCF research.

Cloud Forest Sapwood Structure and Hydraulics

Resistance to Embolism and Hydraulic Safety

Contrary to TMCF leaves, which possess well-defined characteristics differentiating them from LRF, much less is understood about the structure and function of TMCF wood. Wood functional traits, particularly the hydraulics of xylem (vessel and tracheid based), are a fundamental aspect of plant physiology, which determine plant responses to climate [32, 33, 93]. The scarcity of studies investigating the hydraulic traits of TMCF communities is a major gap in our understanding of TMCF ecophysiology and limits our capacity to predict their response to climate change. In this section, we compiled the available studies on this topic together with the functional analysis of SSBCF to provide an initial picture of TMCF wood hydraulic/functional traits.

The Ψ when a vessel or tracheid-containing stem loses 50% of its maximum hydraulic conductivity (Ψ_{50}) can be used as reference for the plant's capacity to withstand drought-induced embolism [45, 94]. The Ψ_{50} of evergreen plants is often positively related with plant water availability [95, 96]. We did not detect significant differences between the Ψ_{50} of the dominant genera in SSBCF and non-TMCF communities (Fig. 3e). Most studies assessing in situ branch xylem Ψ_{50} values of TMCF species have found values similar to our TMCF CWA Ψ_{50} of -2.54 (CI95%: -2.29 to -2.79) MPa (Table S1). In Oliveira et al. [22•], we reported a Ψ_{50} for the vesselless angiosperm *Drimys brasiliensis* located in CF7 (Fig. S1) of -1.56 MPa. More recently, Eller et al. [100] measured the Ψ_{50} of seven additional TMCF species at the same site and found a Ψ_{50} of -2.79 ± 0.37 (mean \pm SE) MPa across all species. Hacke et al. [101] and Sperry et al. [102] conducted studies in vesselless and basal angiosperm hydraulics, and measured the Ψ_{50} of 12 TMCF species from Costa Rica, New Caledonia and the North of Australia. Pooling together the TMCF species from these two studies results in a Ψ_{50} of -2.42 ± 0.27 (mean \pm SE) MPa.

Plant K_s loss is traditionally described using a sigmoidal function with two parameters, such as:

$$\frac{K_s}{K_{s\max}} = \frac{1}{[1 + (\Psi/\Psi_{50})^a]} \quad (1)$$

where the $K_{s\max}$ is the xylem or tracheid maximum K_s and a determines the shape of the curve. A low a implies K_s starts declining at lower Ψ but with a small K_s loss rate, so $K_s > 0$ even when Ψ is much lower than the Ψ_{50} . A high a will

produce a clear Ψ threshold at Ψ_{50} where the plant suddenly shifts from $K_s \approx K_{s_{\max}}$ to $K_s \approx 0$. There is considerable variation in the shape of vulnerability curves in plants globally [103], which implies that the Ψ_{50} by itself does not provide a complete picture of xylem or tracheid resistance to drought induced embolism [104]. Despite the importance of the vulnerability curve shape for modelling plant hydraulic and stomatal functioning [32, 33] few studies report this parameter. We could not find a single published a value for a TMCF species in the literature. We used published vulnerability curve data from the eight TMCF species from Eller et al. [100] to compute the linear gradient between Ψ_{50} and the Ψ where K_s loses 88% of its $K_{s_{\max}}$ (Ψ_{88}), and compared this gradient with the gradient from LRF using 13 tree and shrub species from the Choat et al. [45] dataset. We have found similar gradients between communities, with the TMCF species $K_s/K_{s_{\max}}$ dropping 0.32 (CI95%: 0.13 to 0.52) per MPa, while the LRF species $K_s/K_{s_{\max}}$ would drop 0.33 (CI95%: 0.18 to 0.49) per MPa. Clearly more data on the vulnerability curve shape is needed before we can make firm conclusions about how the plants of these communities lose K_s in response to water stress. Based on the currently available data, TMCF and LRF species vulnerability to embolism is remarkably similar, both in curve shape and Ψ_{50} (Fig. 1d).

There was a gradual increase in the leaf minimum Ψ (Ψ_{\min}) with altitude at a rate of 0.01 MPa per 100 m increase in altitude in the SSBCF sites (Fig. 1f). This increment rate implies that at our highest TMCF site at 2250 m, trees would have a xylem hydraulic safety margin (HSM, calculated as $\Psi_{\min} - \Psi_{50}$) of 0.54 MPa, which is 0.45 MPa higher than at sea-level, assuming the communities have their respective mean Ψ_{50} from Fig. 1d. There are very few studies reporting HSM values for TMCF, but Eller et al. [100] have found a mean HSM of 1.31 ± 0.24 MPa, which is substantially higher than the global tropical forest HSM median of 0.33 MPa (Choat et al. [45]). These observations can be interpreted as evidence that some TMCF rely on a more conservative stomatal regulation to maintain a high Ψ , relative to its Ψ_{50} , resulting in a relatively large HSM.

Hydraulic Efficiency and Architecture

While some TMCF communities might possess a safer hydraulic system (i.e. higher HSM) when compared with LRF communities, the TMCF hydraulic system is generally less efficient when expressed on a sapwood area basis. We found that the dominant genera in SSBCF had a K_s on average 0.97 (CI95%: 0.12 to 1.81) $\text{kg m}^{-1} \text{s}^{-1} \text{MPa}^{-1}$ lower than non-TMCF communities (Fig. 1g). The low K_s in TMCF can be partly attributed to the abundance of vesselless basal angiosperms in the TMCF community, such as *Drimys*, and other species with primitive vessel morphology, such as *Weinmannia*, which have low sapwood specific conductivity

[101, 102]. Zotz et al. [105] found that trees in a Panamanian TMCF have sapwood hydraulic conductivity, expressed on a leaf area basis, 0.08 to 1.4 $\text{kg m}^{-1} \text{s}^{-1} \text{MPa}^{-1}$ lower than LRF trees. In terms of absolute values, Feild & Holbrook [106] have found K_s varying from 0.12 to 0.65 $\text{kg m}^{-1} \text{s}^{-1} \text{MPa}^{-1}$ in eight TMCF species. These values are considerably lower than our TMCF CWA K_s of 1.97 (CI95%: 1.28 to 2.65) $\text{kg m}^{-1} \text{s}^{-1} \text{MPa}^{-1}$ (Fig. 1g). Low K_s is associated with conduits of smaller diameters [107], as explained by Poiseuille's law. Small diameter conduits are more resistant to ice nucleation [108], which might indicate a temperature mediated selection on TMCF hydraulic traits. Freezing temperatures can occur in some TMCF [50, 109] and potentially induce freeze-thaw embolism in species with wider conduits.

The TMCF hydraulic architecture is also distinct from non-TMCF communities (Fig. 1h–i). The dominant genera in South/Southeast TMCF possess, on average, 1.68 (CI95%: 0.57 to 2.5) cm^2 more sapwood area per leaf area (i.e. HV) than non-TMCF communities (Fig. 1h). The differences in HV between TMCF and LRF species have been reported previously in the literature [105, 110]. Zotz et al. [105] reports TMCF trees have on average 3 cm^2 more sapwood area per m^2 of leaf area than LRF. Santiago et al. [110] have found a TMCF population of *M. polymorpha* had 0.4 cm^2 more xylem area per m^2 of leaf area than a LRF population. The HV computed for the eight TMCF species by Feild & Holbrook [106] range from 5.3 to 20.8 $\text{cm}^2 \text{m}^{-2}$, which are above our TMCF CWA HV of 3.77 (CI95%: 2.94 to 4.6) $\text{cm}^2 \text{m}^{-2}$ (Fig. 1h) and might reflect the weak phylogenetic signal on HV (Table 2).

The length of the hydraulic path linking root and leaves, which is associated with the tree height (H), is another important aspect of tree hydraulic architecture [111, 112]. Low tree stature is a defining characteristic of TMCF vegetation [8, 16]. Accordingly, we have found that the SSBCF communities are composed by genera with a significantly lower H_{\max} than non-TMCF communities (Fig. 1i). The mean difference in H_{\max} between TMCF and non-TMCF is 4.28 (CI95%: 0.97 to 7.6) m, with a TMCF CWA H_{\max} of 19.45 (CI95%: 16.6 to 22.2) m. Our observations fall within the TMCF tree height interval defined in the literature as ranging from 1.5 to 20 m [4, 8, 31].

The low stature of TMCF trees associated with its higher HV results in a higher tree diameter/height ratio than LRF [71]. This type of tree architecture is also observed in trees exposed to intense mechanical perturbations [113, 114]. Strong winds are common in high-altitude environments, being particularly common in TMCF located in exposed ridges [81, 82, 114, 115], and Arriaga [81] shows that wind is a major cause of vegetation mortality in TMCF. Therefore, a thigmomorphogenetic response in TMCF vegetation, that is, the plant growth patterns shift to increase its capacity to

withstand mechanical perturbations [113, 114], is another possible explanation for the TMCF stunted architecture.

The effect of strong winds on trees also depend on its wood properties [115–117]. Higher wood density (ρ) provides a better combination of elasticity and mechanical strength for wood to withstand high winds [115–117]. However, low ρ allows trees to produce thicker trunks for a given height with a smaller carbon investment, which are more resistant to wind damage [116]. The dominant genera in SSBCF have wood slightly less dense than non-TMCF communities, with ρ decreasing 0.0025 g cm^{-3} per 100 m increase in altitude (Fig. 1j). Our data shows only a weak (but statistically significant) relationship. However, Chave et al. [118] show a similar negative relationship between ρ and altitude in a large-scale study with 2456 Neotropical tree species. A lower ρ implies TMCF species can rely on an increased diameter/height ratio to resist wind damage. Several studies report tree diameter/height ratio increases with altitude (see Fahey et al. [119] for a review), which corroborates this hypothesis.

Cloud Forest Roots

There are considerably less studies investigating the belowground traits of TMCF than its aboveground traits, which reflects the technical and logistic challenges of measuring belowground traits and processes. However, the observations available indicate TMCF allocate a large fraction of its assimilated carbon to root production [8, 55, 120–122], indicating that belowground organs have a central role in TMCF. We had no root-related information in the dataset used for the SSBCF analysis; therefore, we focus this section on compiling data available in the literature regarding the structure and function of TMCF roots.

According to Grubb [8], TMCF can reach a belowground biomass from 40 to 72 Mg C ha^{-1} , whereas LRF ranges from 11 to 67 Mg C ha^{-1} . According to this Grubb [8] data, the ratio of belowground to aboveground biomass in TMCF is 0.27, which is more than twice the LRF ratio (0.12). Similarly, Girardin et al. [121] found the fine root to stem biomass ratio increases from 0.02 at 194 m to 0.11 at 3020 m across an altitudinal gradient in Peru. Leuschner et al. [120] observed even greater carbon allocation changes in Ecuadorian montane forests, with the root to aboveground biomass ratio increasing from 0.04 at 1050 m to 0.43 at 3060 m.

The increased belowground carbon allocation in TMCF is often attributed to a nutrient limitation on plant growth [11•]. Plants tend to allocate carbon in order to maximize the acquisition of limiting resources [123]; therefore, N or P limitations tend to increase plant root:shoot ratios [124]. There is some evidence that the fine roots of TMCF possess morphological traits to facilitate nutrient acquisition [125]. Girardin et al. [125] shows that Peruvian Andes TMCF has a higher specific

fine root area (SFRA) and specific fine root length (SFRL) than LRF forests from Metcalfe et al. [126]. Higher SFRA and SFRL allow plants to explore a bigger volume of soil per mass of carbon invested in root production. However, other mechanisms might also favour high SFRA and SFRL, such as a denser soil structure [127] or a decrease in root herbivory [128]. Lower temperatures can decrease root nutrient uptake capacity [129, 130]; therefore, the higher investment in fine roots for nutrient acquisition can compensate for a lower nutrient absorption rate per unit of root area.

A higher biomass allocation to roots can also be attributed to increased need for mechanical stability in the TMCF environment. Higher-altitude TMCF sites usually have a higher proportion of uprooted and snapped trees than lower elevation sites [81, 82]. Soethe et al. [82] have found that the coarse roots from an Ecuadorian elfin forest are more asymmetrical and expand more horizontally in the soil than trees at lower altitudes. These coarse root morphological traits improve tree capacity to withstand the irregular mechanical loads associated with the TMCF shallower and unstable soils, steep slopes and strong winds [82]. In addition to the wind and the tree's own weight, the TMCF root system needs to support a considerable biomass of epiphytes and associated canopy humus [131, 132]. Hofstede et al. [132] reports that a single *Weinmannia mariquitae* tree in a Colombian UMCF would hold 115 kg of epiphytic biomass, which was equivalent to 12% of the tree biomass. The total epiphytic mat weight in an area basis can reach up to 44 Mg ha^{-1} in UMCF [132].

Environmental Drivers of Carbon and Water Fluxes in Cloud Forests

Climatic Controls on Leaf-Level Photosynthesis

In this section, we use the functional traits compiled previously to parametrize a photosynthesis model [99] for TMCF vegetation and evaluate the main abiotic factors controlling TMCF A_n . A very clear pattern observed in the TMCF functional traits is the decline of SLA and leaf N_m at increasing altitudes (Fig. 1). As SLA decreases with increasing altitude (Fig. 1a), less N_m is necessary to achieve a given N content per leaf area (N_a). This makes leaf N_a less sensitive to altitude changes than N_m , and in some cases, it might be even higher in TMCF than in non-TMCF communities [10, 55]. This distinction is important to understand the mechanisms controlling TMCF leaf-level carbon assimilation as the maximum rate of Rubisco carboxylation at 25 °C ($V_{\text{cmax}25}$) is a linear function of N_a [133]. Using the intercept and slope values for tropical trees from Harper et al. [97] and the CWA N_a from Fig. 1 (i.e. $N_a = N_m \times 1/\text{SLA}$), we estimate a mean $V_{\text{cmax}25}$ for TMCF of 40.48 $\mu\text{mol m}^{-2} \text{ s}^{-1}$, which is just 2.9 $\mu\text{mol m}^{-2} \text{ s}^{-1}$ lower than the non-TMCF mean $V_{\text{cmax}25}$. Our TMCF $V_{\text{cmax}25}$

estimates are lower than the V_{cmax25} of 55.6 ± 23.03 (mean \pm SD) $\mu\text{mol m}^{-2} \text{s}^{-1}$ measured by van de Weg et al. [65] in a Peruvian TMCF, which is higher than the V_{cmax25} typically found in LRF [134–136]. This pattern can be interpreted as one type of acclimation of the plant photosynthetic apparatus to lower temperatures. These findings suggest that it is unlikely that TMCFA_n is limited by its V_{cmax25} . To understand how abiotic factors control TMCFA_n, we used the Collatz et al. [99] photosynthesis model to simulate TMCFA_n responses to altitudinal gradients (Fig. 2). The A_n of C₃ plants can be described as the minimum of three limiting processes, the Rubisco carboxylation capacity (J_c), the light-dependent Rubisco regeneration (J_l) and the capacity to export or utilize

the photoassimilates (J_e). These processes can be represented following Collatz et al. [99]:

$$A_n = \min \left\{ \begin{array}{l} J_c = V_{cmax} \left[\frac{c_i - \Gamma}{c_i + K_c(O_a/K_o)} \right] \\ J_l = \alpha(1 - \omega) I_{par} \left(\frac{c_i - \Gamma}{c_i + 2\Gamma} \right) \\ J_e = 0.5 V_{cmax} \end{array} \right. - R_d \quad (2)$$

where V_{cmax} is the temperature adjusted Rubisco maximum carboxylation rates, c_i is the leaf internal CO₂ concentration, Γ is the photocompensation point, K_c and K_o are Michaelis-

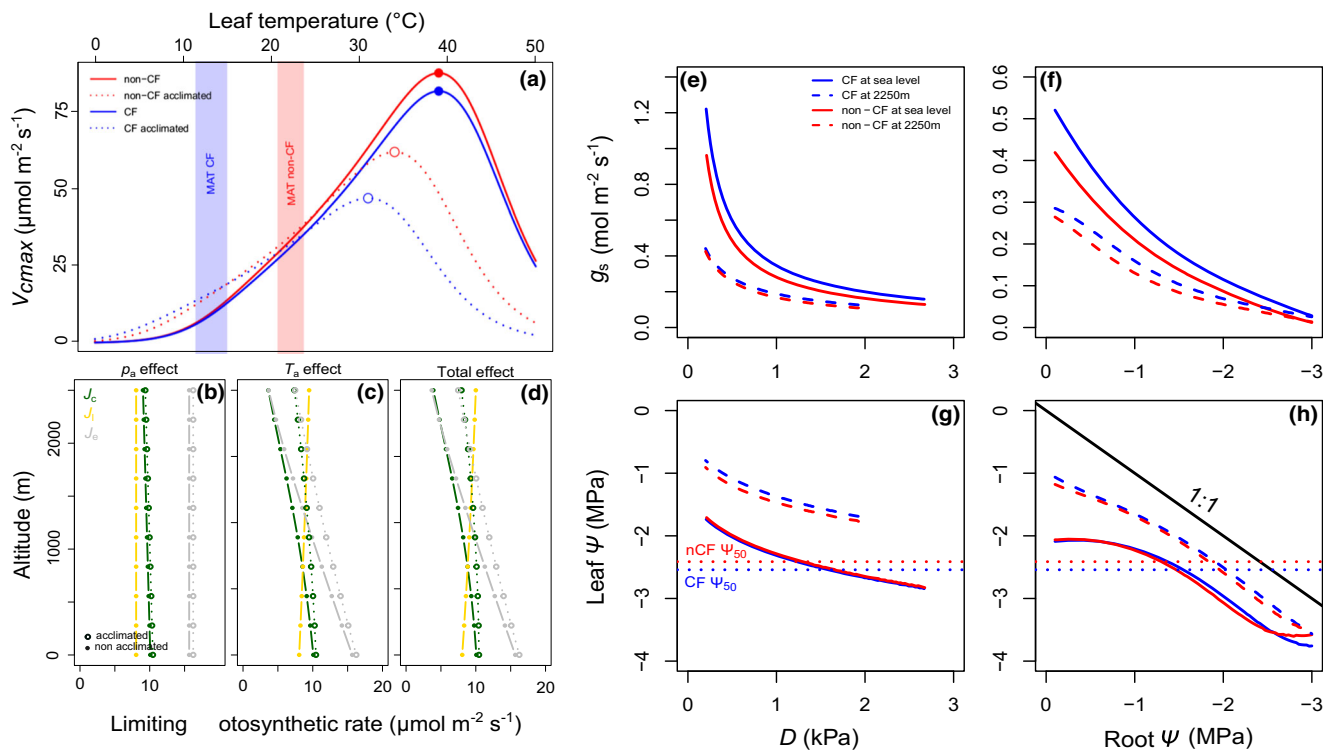


Fig. 2 Rubisco maximum carboxylation rate (V_{cmax}) responses to leaf temperature (T_l) in Tropical Montane Cloud Forest (CF) and non-CF species (a). The continuous lines is the response of non-acclimated species, that is, the T_{low} and T_{upp} parameters from Eq. 2 are assumed to be equal to the parameters used in Harper et al. [97]. In the dotted lines the equations from Kattge & Knorr [98] were used to simulate plant acclimation to the observed mean annual temperature at the sites (95% confidence interval for CF and non-CF sites represented as the blue and red shaded regions, respectively). In b to d, we show the predicted altitudinal trend of each of the limiting steps in the process of photosynthesis (A_n) according to Collatz et al. [99]. The yellow line is the light-limited rate (J_l), the green line is the Rubisco carboxylation limited rate (J_c) and the grey line is the transport limited rate (J_e). The J_l rate was computed to represent low radiation conditions (Incident photosynthetic active radiation of $180 \mu\text{mol m}^{-2} \text{s}^{-1}$). The continuous lines assume no thermal acclimation of V_{cmax} , and the dotted lines are the acclimated responses. In panel (b), only the atmospheric pressure (p_a), and consequently the partial pressure of CO₂ (c_a) and O₂ declines with altitude. In (b), only air temperature (T_a) declines with altitude. In c, all

variables change with altitude. The leaf internal CO₂ partial pressure in the model was assumed to be 0.7 of the c_a and leaf temperature was assumed equal to air temperature. The model V_{cmax} was computed based on the observed leaf N and SLA for TMCF (see details in text), and the other photosynthetic parameters were set equal to Harper et al. [97]. On the right, we show the modelled stomatal conductance (g_s ; e, f) and leaf water potential (Ψ ; g-h) responses to leaf to air vapour pressure deficit (D) and root Ψ . The blue lines represent CF trees, and the red lines non-CF trees. The dashed lines represent the environmental conditions at 2250 m and the continuous lines are the environmental conditions at sea-level. The hydraulic and photosynthetic parameters used in the model are derived from Fig. 1. The differences in temperature between seal level and 2250 were based on the lapse rate from Fig. S2, and altitudinal changes in the incident shortwave radiation were modelled following Leuschner [53]. The mean wind speed was constant at 2 m s^{-1} . In the panels e–g, the root Ψ was constant at -0.1 MPa and relative humidity changed from 1 to 95%. In the panels f–h the relative humidity was constant at 80% and root Ψ changed from -0.1 to -3 MPa . The full model description is given in Appendix S2

Menten constants for CO_2 and O_2 , respectively, O_a is oxygen partial pressure in the atmosphere, α is the intrinsic quantum efficiency of CO_2 uptake, ω is the leaf light scattering coefficient, I_{par} is the photosynthetically active radiation incident on the leaf and R_d is the leaf dark respiration. Our simulations indicate that at higher altitude, A_n tends to be limited by temperature effects on J_c or J_e (Fig. 2). Cloudiness in TMCF can reduce I_{par} (Fig. S1), which limit A_n in lower altitudes, but even a reduction of 90% in the I_{par} is not enough to reduce J_1 to lower levels than J_c and J_e at higher-altitude TMCF (Fig. 2b–d). Bittencourt et al. [137] shows that fog and rain events attenuate, on average, from 74 to 80% of the incoming radiation in a SSBCF; therefore, our simulations represent a particularly strong cloud effect on I_{par} . It is important to note that the leaf wetting associated with rain and fog can have a stronger effect on leaf A_n than what we can predict with Eq. 1. Leaf wetting can directly restrict the CO_2 diffusion to the leaf interior due to the formation of a water film over the stomata [64, 138]. This effect would reduce plant c_i affecting J_c and, to a lesser extent, J_1 .

The total effect of changes in CO_2 partial pressure with altitude on A_n are very small (Fig. 2b). The decline in the atmospheric CO_2 partial pressure is mostly compensated with a lower O_a in J_c , which decreases Γ (computed as $O_a/2\tau$ after Collatz et al. [99]), where τ is the Rubisco affinity for CO_2 relative to O_2). The modelled decrease in J_c or J_e , which dominate A_n responses to altitude is caused by the temperature effects on $V_{\text{cmax}25}$ (Fig. 2a). The V_{cmax} value used in Eq. 1 is calculated as a function of $V_{\text{cmax}25}$ and leaf temperature (T_l) following Clark et al. [139]:

$$V_{\text{cmax}} = \frac{V_{\text{cmax}25} [2^{0.1(T_l-25)}]}{[1 + e^{0.3(T_l-T_{\text{upp}})}] [1 + e^{0.3(T_{\text{low}}-T_l)}]} \quad (3)$$

where T_{low} and T_{upp} are the parameters that define the lower and upper limits of V_{cmax} . The T_{low} and T_{upp} for evergreen broadleaved tropical trees is 13 and 43 °C according to Harper et al. [97], which implies that Rubisco operates at its maximum efficiency when $T_l = 39$ °C (Fig. 2a). Plants can adjust their photosynthetic apparatus to lower or higher temperature through changes to enzymatic content and structure [63, 98]. Given the large difference in MAT between TMCF and LRF (Fig. 2; Fig. S2), it is likely that TMCF species would have different optimum temperatures for Rubisco activity than LRF species. However, we could not find studies measuring the response of TMCF photosynthetic parameters to temperature, which imposes a major constraint on our capability to simulate $\text{TMCF}A_n$. We used the linear relationship between V_{cmax} and MAT from Kattge & Knorr [98] to estimate the optimum V_{cmax} temperature for cold acclimated TMCF plants at 2250 m. This approach predicts that plants

acclimated for the TMCF lower temperatures would have an optimum V_{cmax} at 31 °C. The acclimation of V_{cmax} temperature responses makes TMCF A_n decline only $0.02 \mu\text{mol m}^{-2} \text{s}^{-1}$ per 100 m increase in altitude, whereas if we assume no acclimation, the decline rate reaches $0.17 \mu\text{mol m}^{-2} \text{s}^{-1}$ per 100 m (Fig. 2b). This large difference highlights the need for data on the TMCF V_{cmax} temperature responses so that the $\text{TMCF}A_n$ can be correctly represented in vegetation models [140]. According to our simulations, cold-acclimated TMCF leaves should be capable of reaching high A_n under favourable climatic conditions. However, the frequent leaf wetting events restrict the amount of time TMCF leaves are close to their optimum A_n [64].

Our simulations indicate that J_c and J_e co-limit $\text{TMCF}A_n$ at higher altitude (Fig. 2). However, the limiting effect of J_e could be underestimated by our simulations. The J_e equation from the Collatz et al. [99] model does not explicitly represent the effects of phloem functioning on the export of photosynthetic products from the leaves [141]. Phloem transport rates are dependent on the sap viscosity, which increases at low temperatures [119, 142]. Besides, the sink activity of plant meristematic tissues is strongly inhibited by lower temperatures [143]. As plant growth is interrupted, phloem unloading and transport rates should decline, which can make temperature limitations on A_n at higher altitude more important than what can be predicted in our simulations with Eq. 2.

Stomatal Responses to Climate

The previous section shows the environmental controls on the biochemical and photochemical processes of A_n . In this section, we focus on the role of stomatal responses to climate controlling plant carbon assimilation and hydraulic safety in TMCF. We used the leaf and hydraulic traits from Fig. 1 to parameterize a stomatal optimization model based on Eller et al. [33] (full description in Appendix S2). The main assumption of the Eller et al. [33] model is that plant stomata evolved to balance instantaneous A_n with the loss of hydraulic conductance, derived from Eq. 1. The model predicts that both TMCF and non-TMCF plants will adopt a more conservative stomatal regulation at higher-altitude climate (Fig. 2). The modelled g_s response to soil and atmospheric drought is very similar between TMCF and non-TMCF plants, with TMCF reaching a slightly higher g_s than non-TMCF, especially at a low-altitude environment (D ; Fig. 2a, b). Even so, the TMCF leaf Ψ is similar to non-TMCF (Fig. 2c, d) due to its higher HV and low tree height, which compensates for its low K_s (Fig. S4). TMCF plants at high altitude have a g_s , on average, 40% lower than at sea-level, while non-TMCF g_s is, on average, 32% lower at high altitude. This low g_s results in a higher leaf Ψ and HSM at high altitude, which is partially supported by our observations (Fig. 1f). Whereas at sea-level the TMCF leaf Ψ reaches its Ψ_{50} when $D = 1.6$ kPa, at high-altitude, it

would still maintain an HSM of 0.85 MPa even at the highest simulated D at 2250 m (1.9 kPa). Similarly, in response to soil drying the model predicts that, at low altitude, plants would always maintain a smaller root to leaf Ψ gradient, except during extreme drought, that is at root $\Psi \approx -3$ MPa (Fig. 2d). At higher altitude, TMCF leaf Ψ would reach its Ψ_{50} when root $\Psi = -2$ MPa, whereas at sea-level, TMCF would reach HSM = 0 with a root Ψ 0.5 MPa higher

The conservative stomatal behaviour predicted by the model can be attributed to the temperature response of V_{cmax25} (Fig. 2a), which decreases the potential A_n for a given g_s . The lower rates of carbon assimilated at high elevation do not compensate for the hydraulic conductance lost, resulting in a more conservative stomatal regulation. Our simulations assume no acclimation in the V_{cmax} response to temperature, an acclimation of the magnitude as the one shown in Fig. 2a would result in a similar stomatal behaviour at low and high altitudes. A more conservative water use and stomatal regulation have been observed in many TMCF sites [24, 144–147], but there is also evidence of TMCF species that respond very little to drought [148]. As discussed in the previous section and in Oliveira et al. [22•], it is important to consider the role that FWU can have in maintaining the leaf Ψ in certain TMCF plants during drought. The water acquired by FWU might compensate for a less conservative stomatal regulation in some TMCF species and contribute to the maintenance of leaf turgor [24] and a higher HSM [22•].

Climatic Limitations to Water Transport

In many species, the HV increases with tree height to compensate for the increased hydraulic resistance caused by the increased distance between roots and leaves [111]. Our results indicate that TMCF are an exception to this trend as they have both low H_{max} and high HV (Fig. 1). McDowell et al. [111] uses a simple hydraulic model based in Darcy’s Law to explain why HV declines in taller trees:

$$HV = \frac{H\eta g_w D}{p_s \Delta\Psi} \tag{4}$$

where p_s is the sapwood permeability, $\Delta\Psi$ is the soil to leaf Ψ gradient accounting for the gravitationally induced Ψ drop, H is tree height, η is water viscosity, g_w is the combined g_s and boundary layer conductance to water. This model predicts that, assuming all other parameters are constant, sustaining a given g_s at increasingly high H requires a higher HV (Fig. 3). We can use Eq. 4 to understand the differences in HV between TMCF and non-TMCF communities based on the biotic and environmental differences between these communities. The colder temperatures observed in TMCF [50•]

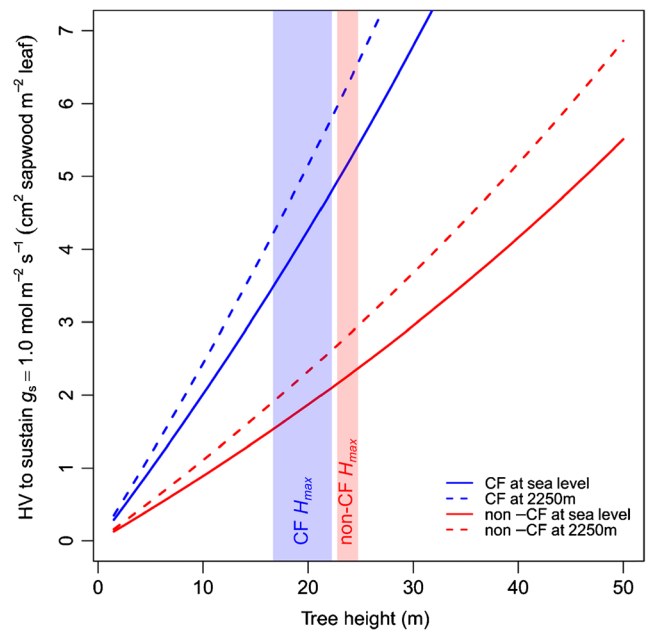


Fig. 3 Huber Value (HV) necessary to sustain a stomatal conductance of $1 \text{ mol m}^{-2} \text{ s}^{-1}$ plotted in function of tree height. The blue lines represent Cloud Forests (CF) trees, and the blue lines represent non-CF trees. The dashed lines represent the environmental conditions at 2250 m. The differences in temperature between seal level and 2250 were based on the lapse rate from Fig. 2, and altitudinal changes in the incident shortwave radiation were modelled following Leuschner [53]. The red and blue shaded regions in the plot are the 95% confidence intervals of the mean tree height observed in CF and non-CF, respectively

cause η to increase from $9.54 \times 10^{-4} \text{ Pa s}$ at sea-level (MAT = 22.2 °C) to $1.19 \times 10^{-3} \text{ Pa s}$ at our highest site at 2250 m (MAT = 13.2 °C). The increased viscosity restricts xylem water transport and requires more sapwood per leaf area (higher HV) to sustain a given g_s . While the lower temperature at higher elevations reduce air D (assuming constant air humidity), the increases in solar irradiance and water vapour diffusion coefficient can counteract the temperature effect [51–53]. We included a boundary layer and leaf energy budget model in the D calculation in Eq. 4 to account for these effects (equations S2–S4 in Appendix S2). Equation 4 shows that the high-altitude TMCF environment requires a higher investment in HV to sustain a given g_s than the LRF environment (Fig. 3). An increase in 1 m in tree height will require 0.027 cm^2 more sapwood per m^2 of leaf to sustain a $g_s = 1 \text{ mol m}^{-2} \text{ s}^{-1}$ in the high-altitude TMCF environment than at sea level (Fig. 3). When we include the functional differences between TMCF and LRF (Fig. 1) in Eq. 4, the investment in sapwood per unit of leaf area becomes even more important for TMCF plants. The lower K_s in TMCF (Fig. 1g) and more conservative stomatal regulation (Fig. 2) results in a less negative Ψ_{min} (Fig. 1f). This makes an increase in 1 m in TMCF tree height require 0.15 cm^2 more sapwood per m^2 of leaf area in comparison with LRF to sustain a $g_s = 1 \text{ mol m}^{-2} \text{ s}^{-1}$ (Fig. 3).

Abiotic Filters Shaping Cloud Forest Communities

In the previous sections, we described TMCF functional composition (Fig. 1) and the consequences of these traits for TMCF ecophysiological processes (Figs. 2 and 3). In this section, we propose hypotheses to explain why these particular traits and processes are prevalent in TMCF vegetation. We employ the concept of environmental filters [149] to explain how the peculiar TMCF hydroclimatic environment selects a set of plant traits, which, ultimately, determine the functioning of TMCF ecosystems and its response to global change. We represent the postulated environmental filters in TMCF as a conceptual model in Fig. 4, which can provide a roadmap for the representation of TMCF in DVGM.

The concept of environmental or abiotic filter assumes the environment functions as a metaphorical “sieve” that only allow species with certain traits to establish and persist [150]. We postulate that lower temperatures are a fundamental environmental filter in TMCF (Fig. 4). Temperature will directly affect several TMCF plant processes, besides indirectly driving other important abiotic filters in TMCF, such as clouds and nutrient availability (Fig. 4). Lower temperatures favour cloud formation due to its effect on the lifting condensation level (Fig. S2). Lower temperatures also decrease the soil nutrient availability by affecting nitrogen mineralization rates [151], besides directly affecting root nutrient uptake capacity [129, 130]. Air temperature also affect directly many aspects

of TMCF plant physiological processes, such as reducing the leaf A_n of non-acclimated species (Fig. 2) and making xylem and phloem transport more difficult due to its effect on water viscosity [119, 142]. These effects can trigger a series of compensatory traits, such as a more conservative stomatal functioning (Fig. 2), a higher allocation to wood production over leaf area (Fig. 1h) and shorter path lengths between leaves and roots (Fig. 1i) to facilitate the canopy water supply. In colder TMCF sites, temperature can have an even stronger selective effect on the community hydraulic traits. The occurrence of colder winters that can freeze xylem/tracheid water, even if rare, could have a lasting impact on a TMCF community. Freezing induced embolism could favour species with small diameter and low conductivity xylem/tracheids (Fig. 1g), which are resistant to freezing embolism [108].

The persistence of cloud immersion events and low nutrient availability form another important layer of TMCF abiotic filters (Fig. 4). The occurrence of cloud immersion might favour competitively species capable of accessing and utilizing the resources made available by the frequent leaf wetting events (water and potentially nutrients), as well as select against species vulnerable to leaf pathogens. Leaf wetness favours the establishment of bacteria, fungi and other organisms, which might damage leaves [80]; therefore, species with low structural investment on leaf tissues (high SLA) would be more vulnerable to leaf infection [152]. The leaf wetness associated with cloud immersion events also limits leaf A_n and can constrain the occurrence of fast acquisitive species [48,

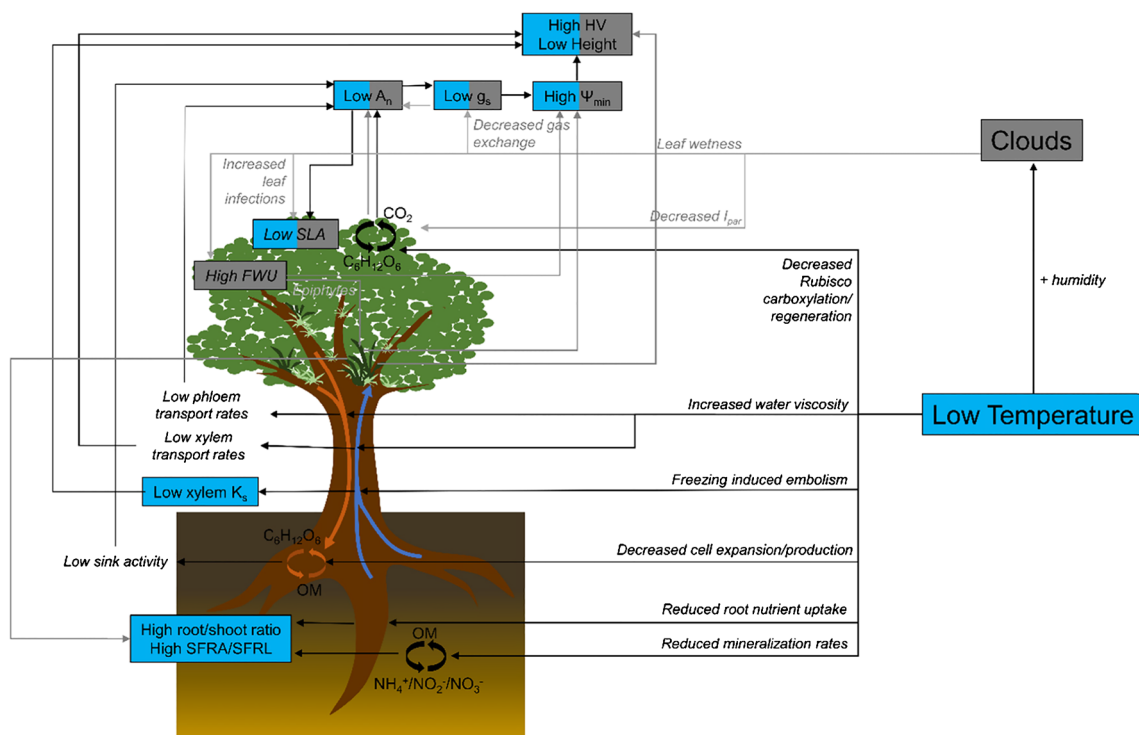


Fig. 4 Schematic representation of the relationships between low temperatures and clouds on the Cloud Forest (CF) defining functional traits. The functional traits in blue boxes can be linked to low temperatures, while traits in grey boxes can be linked to cloud immersion

58], that is, species with both high A_n and SLA. Species with a fast leaf tissue turnover would depend on high A_n to quickly compensate its low carbon investment in its short-lived leaves. However, even though TMCF vegetation can potentially reach high A_n when climatic conditions are favourable due to its high leaf N content and V_{cmax25} (Figs. 1 and 2), it cannot consistently maintain a high A_n due to temperature related limitations to A_n (Fig. 2) and the frequent wet leaves interrupting leaf gas exchange [64, 138]. The low nutrient availability in TMCF will also favour low SLA species [153] and species capable of sustaining a high investment in root production [123, 124].

Besides the postulated main filters represented in Fig. 4, we propose other TMCF typical environmental conditions reinforce the selective effect of lower temperatures, clouds and low soil fertility. During clear periods, high-altitude environments are exposed to high levels of shortwave irradiance [52, 53], which select against larger leaves with high SLA and more prone to overheating [52]. Higher transpiratory rates could provide an alternative mechanism to cool down leaves [154]. However, the limitations to water transport and stomatal regulations discussed previously prevent this strategy. Mechanical stress can also be a strong driver of plant form and function in certain TMCF [113, 114]. Wind-induced stress can trigger thigmomorphogenetic responses that include low SLA, low tree stature (Fig. 1i) and high stem diameter/height ratio [155–157]. The existence of multiple independent environmental conditions selecting a similar set of traits can increase the resistance of the TMCF community to certain changes to environmental conditions. For example, traits such as low SLA would still be dominant in a TMCF community even in a site protected from wind, as the high irradiance and leaf wetting events would still favour low SLA species.

Multiple overlapping environmental filters can enhance TMCF resistance against certain types of environmental change, but the interactions between key TMCF environmental factors, such as temperature, cloudiness and nutrient availability (Fig. 4), make TMCF highly vulnerable to hotter climates. General circulation models predict an increase in the TMCF MAT from 2 to 4 °C in the next decades [22]. A higher surface temperature can increase the cloud base formation height up to 1634 m in some TMCF sites [23]. An increase in 4 °C also would increase the litter decomposition and nutrient mineralization rates by 53% [151]. It is unlikely that TMCF communities could resist the simultaneous removal of the three main filters shaping their structure and driving their processes (Fig. 4). The lack of leaf wetting and higher temperatures events would allow plants to sustain a consistently high A_n , and the increased nutrient availability would decrease the need for high fine root investments. These changes can make TMCF environments more favourable to larger LRF species with highly acquisitive traits and fast tissue turnover. In this scenario, the current TMCF community would be

restricted to higher-altitude elevations [16], where the strength of other TMCF filters, such as high irradiance loads and water vapour diffusion in air [51–53], compensates for the higher temperature and less clouds. Particularly windy TMCF sites with shallow soils and steep slopes could also be refugia for TMCF communities as they would still impose mechanical restrictions to the establishment of larger and fast growing LRF species. However, even these sites can be invaded by grassland species and tropical shrubs, which can thrive in drier sites [158]. These hypotheses can subsequently be tested in a process based DGVM framework using the data presented here to represent TMCF vegetation.

Abiotic filters are not the only elements determining the structure and function of ecosystems. Biotic interactions and dispersal limitations often have a significant role on community assembly [159]. These effects are important for understanding TMCF dynamics, especially due to the role of epiphytes in TMCF biogeochemistry [85]. The epiphytic community of TMCF is exposed to a more arid environment in the canopy than the rest of the TMCF community [88]; therefore, it should respond faster to a decrease in cloud immersion and increased temperatures. Declines in the epiphyte's abundance can accelerate TMCF responses to climate change by decreasing TMCF cloud water interception and nutrient acquisition combined with changes in the canopy fauna. A decrease in the heavy epiphytic load characteristic of TMCF will also decrease the need for investment resources to increase the tree mechanical stability, which should facilitate the establishment of species with a typical LRF architecture.

Conclusions

Our findings show that the unique hydroclimatic conditions in TMCF selects a community functionally distinct from other tropical forests (Fig. 1), which results in different ecophysiological responses to climate (Figs. 2 and 3). The TMCF functional composition can be interpreted as the result of many interacting and overlapping environmental filters (Fig. 4). These filters impose restrictions on the establishment of larger and fast-growing lowland species through hydraulic and mechanic restrictions on plant height (Fig. 3), coupled to temperature and leaf-wetting related restrictions on leaf structure, stomatal functioning and carbon assimilation (Figs. 1 and 2). The TMCF community also have proportionally large root systems with a high density of fine roots, which reflects a heavy investment in nutrient acquisition.

The functional uniqueness of TMCF we show in this review must be considered in DGVM and Earth System models to quantify the potentially large contributions of this ecosystem global and regional biogeochemical cycles and climate [150]. Our conceptual framework based on TMCF functional traits (Fig. 4) suggests that TMCF structure and function are

highly vulnerable to increases in temperature, which are likely to occur in the next decades [22, 27]. Our findings provide a roadmap for the inclusion of TMCF in DGVMs, which should enable the assessment of TMCF vulnerability to climate change scenarios at the global scale. Predicting TMCF vulnerability worldwide is the first step to establishing TMCF conservation priorities and prepare human communities for the potential loss of TMCF and the services provided by these ecosystems.

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Compliance with Ethical Standards

Conflict of Interest Cleiton B Eller, Leonardo D Meireles, Stephen SO Burgess, Stephen Sitch, Rafael S Oliveira declare that they have no conflict of interest.

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